

Mining ecophysiological responses of European beech ecosystems to drought

Fabio Gennaretti, Jerome Ogee, Julien Sainte-Marie, Matthias Cuntz

▶ To cite this version:

Fabio Gennaretti, Jerome Ogee, Julien Sainte-Marie, Matthias Cuntz. Mining ecophysiological responses of European beech ecosystems to drought. Agricultural and Forest Meteorology, 2020, 280, pp.1-9. 10.1016/jagrformet.2019.107780 . hal-02557077

HAL Id: hal-02557077 https://hal.inrae.fr/hal-02557077

Submitted on 21 Dec 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



1	Mining ecophysiological responses of European beech ecosystems to				
2	drought				
3					
4	Fabio Gennaretti ^{a,b,*} ; Jérôme Ogée ^c ; Julien Sainte-Marie ^a ; Matthias Cuntz ^a				
5					
6	^a Université de Lorraine, AgroParisTech, INRA, UMR1434 Silva, 54000 Nancy, France.				
7	^b Institut de recherche sur les forêts, UQAT, Rouyn-Noranda, Québec, J9X 5E4, Canada.				
8	^c INRA Centre de Bordeaux Aquitaine, UMR1391 ISPA, 33140 Villenave d'Ornon, France.				
9					
10	*Corresponding author e-mail address and telephone number:				
11	fabio.gennaretti@libero.it				
12	+33 776692938				
13	*Corresponding author full postal address:				
14	INRA Centre Grand Est – Nancy, UMR1434 Silva, Rue d'Amance, 54280 Champenoux, France				

Mining ecophysiological responses of European beech ecosystems to

drought

15

16

17

18

19

20

21

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

Abstract

The most accurate understanding of forest functioning during drought is crucial to improve the forecast of future forest productivity. Here we investigate the ecophysiological responses (i.e. primary production, evapotranspiration and water use efficiency) of European beech to drought events with the ecosystem model MuSICA, using as benchmark the observed fluxes at the experimental forest Hesse (France). We show that MuSICA is able to realistically simulate observed drought-induced limitations. Subsequently we use simulation experiments to provide: (1) a quantification of the reduction of ecosystem fluxes during the 2003 drought, (2) a partitioning of heat stress and water limitations during droughts, (3) an analysis of the impact of specific drought trajectories, and (4) an evaluation of the potential impact of projected climate change on the studied forest and (5) over the beech distributional range. Our results show that the 2003 drought resulted in a 17% reduction of annual gross primary production and in a 21% reduction of evapotranspiration at Hesse. The studied forest ecosystem is mostly sensitive to negative precipitation anomalies (82% of the reduced forest productivity in 2003) and almost insensitive to heat stress due to high temperatures (16%). Moreover, we show that the ecosystem fluxes are limited more by fast drought onsets in the early growing season (June-July) than by onsets later in the season. Deciphering the impact of future climate change on beech productivity is complicated by large uncertainties in projected future precipitation and in the severity of extreme dry years. Drastic reduction of ecosystem fluxes is only predicted with climate projections that show marked reductions in precipitation. However, increased CO₂ fertilization in the future will counterbalance negative drought impacts. This modelling-based study improves our understanding of the functioning of

- 37 an emblematic European tree species during extreme events and informs on potential future forest
- 38 responses to projected climate change.

39 Keywords

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

40 Drought; carbon flux; Fagus sylvatica L.; ecosystem model; MuSICA; projected climate change

1. Introduction

The frequency of extreme climate events, and especially of droughts, is supposed to increase with future climate change (Prudhomme et al. 2014; Zhao and Dai 2017). Central and Southern Europe in particular are regional hotspots where drought severity and frequency will increase the most (Prudhomme et al. 2014). This implies that European forest ecosystems will face new and extreme conditions driving changes in species productivity and distribution (Lindner et al. 2010). Indeed, severe droughts are linked to significant reductions of forest productivity (Granier et al. 2007; Reichstein et al. 2007). Speciesspecific responses to drought should hence be analyzed thoroughly, in addition to possible interactions between the impact of drought and other projected changes (e.g. increased plant water use efficiency with elevated CO₂; Battipaglia et al. 2013). Among the emblematic European tree species, European beech (Fagus sylvatica L.) may be severely impacted by drought during the growing season, resulting in xylem embolism and reduced productivity (Gessler et al. 2007). Beech has already experienced locally increased growth limitations over the 20th century due to drought, which has reduced tree competitiveness in relation to other species (Scharnweber et al. 2011). This loss of competitiveness is especially evident at the southern edge of the species distributional range (Peñuelas et al. 2007). A process-based assessment of the ecophysiological responses of European beech to drought would allow an analysis of the species vulnerability to climate

change and would potentially improve future projections of beech growth performance over its distributional range.

Such process-based assessment could be performed with ecophysiological models. These models can play an important role in investigations and forecasts of drought impacts on forests because they ensure a coupled representation of the processes behind carbon and water fluxes (Bonan and Doney 2018; Ogée et al. 2003). The selected models should guarantee plausible simulations also at the margin or outside of their calibration environment. Indeed, plausible extrapolations beyond the observed boundary conditions are crucial to accurately predict future ecosystem functioning and carbon budget during extreme events, such as future droughts (Clark et al. 2011).

Some studies have already quantified the impact of drought on ecosystem productivity globally (Zhao and Running 2010), on specific regions (Ciais et al. 2005), or on specific forest ecosystems (Reichstein et al. 2007). However, droughts are the result of two components: an increase in temperature-induced plant stress (e.g. effects of temperature on vapor pressure deficit, or the effects on carboxylation rate) and a decrease in water availability driven by low precipitation (Zhao and Dai 2017). The partitioning of heat stress and water limitations during drought events for specific ecosystems is still to be further analyzed. Moreover, each drought has its own climate trigger and its timing of occurrence during the year (Mo 2011), potentially inducing a very specific impact on forests. For example, fast drought onsets in spring followed by average summer conditions can have different impacts on annual ecosystem performance than lesser but constant reduction in water availability during the whole growing season. Impacts due to different drought trajectories should also be better analyzed. Here we focus on the ecophysiological responses (i.e. forest water and carbon fluxes, and ecosystem water use efficiency) of European beech forests using simulation experiments. First, we ensure that the used ecosystem model (MuSICA; Ogée et al. 2003) is able to realistically simulate drought-induced flux reductions at an experimental site, even for much drier periods than used during model calibration. Second, we re-

compute the local impact of the 2003 drought on ecosystem fluxes using a novel simulation-based approach. Third, we quantify the partitioning of heat stress and water limitations during drought. Fourth, we investigate the impact of different drought trajectories. Fifth, we evaluate the potential impact of projected climate change on the productivity and water use of beech forests locally and, sixth, evaluate the potential impact over the entire species distributional range in Europe.

2. Materials and Methods

2.1. The Hesse forest experimental station

We used the data of the Hesse forest experimental station (Granier et al. 2008) to optimize the MuSICA ecosystem model for beech forests. The Hesse site (+48.7° latitude, +7.1° longitude, 300 m elevation) is a European beech forest located in north-eastern France (Moselle) in the center of the species distributional range. Hesse belongs to the ICOS infrastructure (www.icos-ri.eu) and there are 20 years of ecosystem fluxes and ecophysiological measurements available at the site. Hesse is a drought-sensitive ecosystem and extreme dry years, such as 2003, produce significant anomalies on the observed carbon dynamics (Granier et al. 2007). We used measured ecophysiological traits for the initialization of the model parameters (Dufrêne et al. 2005) and soil water content at three soil depths (SWC; 0.1, 0.3 and 0.55 m), carbon fluxes (net ecosystem exchange - NEE; and gross primary production - GPP; NEE was partitioned into GPP and ecosystem respiration using the night-time approach of Reichstein et al. 2005) and evapotranspiration (ET) for further parameter optimization and validation.

2.2. The MuSICA ecosystem model

MuSICA is a mechanistic ecosystem model primary developed to simulate water, carbon and energy exchanges over the soil-vegetation-atmosphere continuum (Domec et al. 2012; Giuggiola et al. 2016; Klein et al. 2014; Ogée et al. 2003; Ogée et al. 2004; Wingate et al. 2010). MuSICA discretizes the ecosystem in several air and soil layers. In this project, we used 15 layers in the soil and 10 layers in the

vegetation. Carbon fluxes depend on the coupling between photosynthesis (Farquhar et al. 1980) and stomatal conductance (Leuning 1995), as well as on the simulation of respiration, leaf light interception (Sinoquet et al. 1990), convection and turbulent transport (Raupach 1989) in each air layer. Soil water fluxes are a mixture of preferential and matrix flows, the latter ones being described by Richards' equations (Braud et al. 1995).

2.3. Model calibration and validation

MuSICA was optimized using as target half-hourly SWC, NEE and ET measurements of the year 2002, which had normal climate conditions. ET observations were corrected to ensure energy balance closure as proposed by Mauder et al. (2013). All MuSICA parameters were initialized according to our best knowledge of the ecophysiological traits of the Hesse experimental forest. Subsequently, we used Morris parameter screening (Morris 1991) to quantify the sensitivities of the target variables to the individual model parameters of MuSICA (Cuntz et al. 2015). The Morris parameter screening analyses the variations of the target variables with respect to variations in the model parameters (Saltelli et al. 2004). The identified sensitive parameters were further optimized using a differential evolution algorithm (Abbass et al. 2001; Mier 2017). Finally, the 2003 dry year was used to verify that MuSICA realistically simulates drought-induced ecosystem limitations as well as daily variations of observed water-related variables during extreme dry periods. This calibration/validation scheme was specifically designed because drought severity is supposed to increase with future climate change and, therefore, plausible model extrapolations are crucial.

2.4. Local simulation experiments

Once MuSICA was optimized for Hesse, we performed several local simulations modifying the 2003 temperature and precipitation input data. These simulations were grouped in four experiments (Table 1). Experiment 1 quantifies the reduction of fluxes during the 2003 drought. Experiment 2 provides partitioning of heat stress and water limitations during drought. Experiment 3 studies the impact of

alternative drought trajectories having the same severity than the 2003 drought but different timing of occurrence over the growing season (i.e. faster or slower drought onset). Experiment 4 analyzes the impact of future climate change on the studied forest. All simulations use the 2003 half-hourly variability to avoid potential impacts of different structures of rainy/dry, warm/cold, sunny/cloudy days. Only the monthly temperature and precipitation mean values were modified and were specific to each simulation:

$$T_{new_{i,y}} = T_{ref_i} + \Delta_{i,y}$$
 (1)

$$P_{new_{i,y}} = P_{ref_i} \cdot \lambda_{i,y}$$
 (2)

where $T_new_{i,y}$ ($P_new_{i,y}$) is the new temperature (precipitation) mean value of the month i and the simulation y, T_ref_i (P_ref_i) is the reference temperature (precipitation) mean value of the month i of the median observed climate at Hesse (1997-2013), $\Delta_{i,y}$ is the climate anomaly of temperature for the specific simulation y, and $\lambda_{i,y}$ is the climate anomaly factor of precipitation for simulation y. Specific simulations and their climate anomalies are detailed in Table 1. Each simulation was preceded by a one year spin-up period, which was subsequently discarded from following analyses, where MuSICA was forced by the meteorology of an average year.

The impact of future climate change in experiment 4 was analyzed using an ensemble of six EURO-CORDEX simulations forced by an RCP8.5 (Table 2; Jacob et al. 2014; data downloaded from the Pierre Simon Laplace Institute node of the Earth System Grid Federation). This extreme RCP was selected to emphasize the potential impact of climate change on forest ecophysiology and analyze forest responses under extreme climate conditions. Furthermore, the role of increased CO_2 concentration in the future was isolated by the production of MuSICA simulations at Hesse with both current and future CO_2 values. Indeed, a 2.17-time increase in CO_2 concentration is forecast at the 2080-2100 horizon in RCP8.5 scenarios. The annual ecosystem fluxes (F) of the obtained MuSICA simulations y were compared

between each other by percent change (PC) computations relative to the fluxes of the reference simulation (median climate):

154
$$PC = \frac{F_y - F_{Ref}}{F_{Ref}} 100 \, (\%)$$
 (3)

2.5. Regional simulation experiments

Experiment 4 described in the previous section was extended to analyze the potential impact of projected climate change on European beech ecophysiology over the entire species distributional range as defined by Caudullo et al. (2017). The MuSICA temperature and precipitation input data were modified using the same EURO-CORDEX simulations (Table 2; all EURO-CORDEX data were extracted over a regular 0.5° grid) and the same equations described earlier (Eq. 1 and 2). This climate forcing was used to simulate European beech GPP and ET during extreme dry years at two time horizons (2006-2026 and 2080-2100; an extreme year was defined with median monthly temperature and precipitation of six extreme dry years, one per simulation). We also took into account the impact of spatially different soil properties. We used USDA soil classification and soil texture data available at "soilgrids.org" to estimate MuSICA soil parameters, such as van Genuchten water retention parameters and saturated hydraulic conductivity, using the Rosetta computer program (Schaap et al. 2001).

3. Results

3.1. Validation of MuSICA simulations during dry periods

MuSICA accurately simulates daily variation of GPP, ET and SWC at the Hesse European beech forest (Figs. 1 and S1-S2). The typical annual cycle of drying and rewetting of the soil profile is well caught by the model (Fig. 1, red and blue points). Furthermore, MuSICA simulations show similar performance over both, the 2002 calibration year and the 2003 validation year (Fig. 1), the latter was characterized by reduced precipitation and a severe 4-month drought starting in June and ending in late September (Fig.

S3). Finally, observed drought-induced ecosystem limitations are well simulated, such as the limitation of evapotranspiration by soil water availability (Fig. 2). At Hesse, ecosystem fluxes begin to be limited by water availability once soil water in the top 70 cm drops below about 250 kg/m³, corresponding to values of relative extractable soil water (REW) lower than 0.3.

3.2. Mining forest responses to drought with simulation experiments

174

175

176

177

178

179

180

181

182

183

184

185

186

187

188

189

190

191

192

193

194

195

196

for ET.

The first simulation experiment allowed a precise quantification of the reduction of ecosystem fluxes at Hesse during the 2003 drought. Comparison of the simulation with a median climate but diurnal variations of 2003 (2003_median_climate, red) and the full 2003 dry climate (2003, black) in Figure 3 shows that there is a reduction of cumulated annual GPP by 17% and a reduction in total ET by 21% at Hesse due to higher temperatures and less precipitation. This suggests that such an important drought had only a slight impact on water use efficiency (evapotranspiration decreased slightly more than productivity). Interestingly, the divergence between the two simulations started only 1.5 months after the beginning of the 2003 heatwave in early June (Fig. S3). This is the time needed to dry out the soil profile to the threshold of about 250 kg/m³ in the top 70 cm, or REW = 0.3, when plants start experiencing water stress (Fig. 2). The second experiment allowed the partitioning of heat stress and water limitations at Hesse during drought. Comparing the simulation with a median climate but diurnal variations of 2003 (2003_median_climate, red), the simulation with median climate and the 2003 temperature anomalies only (2003_T_limited, blue), the simulation with median climate and the 2003 precipitation anomalies only (2003_P_limited, green), and the simulation with full 2003 dry climate (2003, black) in Figure 3, we can observe that water limitations caused by low precipitations are responsible for most of the observed

reductions in GPP and ET. Heat stress due to high temperature is instead very low for GPP and negligible

Subsequently, we show that droughts with same severity but diverse trajectories (i.e. timing of drought onset and recovery over the growing season) have different impacts on the ecosystem (Experiment 3; Table 1 and Fig. 4). The beech forest in Hesse shows very similar annual sums of GPP and ET if there is still some precipitation in June-July, which was the case in 2003 and is the case in the slow-onset scenario (black and purple lines). However, the forest is sensitive to fast drought onset. If there is a different precipitation timing with almost no precipitation in June-July, SWC decreases earlier than in 2003, leading to a longer drought-limited period, which cannot be compensated during drought release later in the year (orange line). Note that all simulations in Fig. 4 have the same precipitation amount over the growing season (Table 1).

Our fourth experiments allowed an evaluation of the impact of projected climate change at the studied forest. Fig. 5 shows the expected changes between present and future forest performance over median

forest. Fig. 5 shows the expected changes between present and future forest performance over median climatological years (*Median_futur vs. 2003_median_climate*; dark blue) and extreme dry years (*Xtreme_futur vs. Xtreme_today*; light blue). The same figure also allows the comparison of the expected impact of future extreme dry years (*Xtreme_futur vs. 2003_median_climate*; green) and of present extreme dry years (*Xtreme_today vs. 2003_median_climate*; red) relative to present median conditions. All CORDEX simulations used in Fig. 5 agree in projecting an important increase in temperature over the growing season (median increase of +3.0 °C; Fig. S4). Conversely, there is a large spread in the magnitude and sign of the precipitation change (Fig. 5; dark and light blue), which is the most influential parameter modifying ecosystem fluxes in MuSICA at Hesse. MuSICA, forced by these climate inputs and by stable, current CO₂ concentrations, projects only slight changes in GPP and ET during a median year (Fig. 5, dark blue) and an extreme year (Fig. 5, light blue) of 2080-2100. Although EURO-CORDEX climate simulations show hotter and drier extreme years in the future than today (Fig. 5, green and red), the reduction of the ecosystem fluxes relative to present median conditions remains very similar. Indeed, the simulated climate of extreme dry years in EURO-CORDEX is less severe than the observed extreme

year 2003, so that the soil water threshold of 250 kg/ m^3 in the top 70 cm, or REW = 0.3, is rarely reached in the MuSICA simulations. The picture changes drastically if we include the future increase in CO2 concentration projected by RCP8.5 simulations at the 2080-2100 horizon (2.17 current CO₂). Over a median year (Fig. 5, dark blue), GPP increases by 36% in this case, ET decreases by 17%, leading to very high water use efficiency (+64%). We computed that a 1.22-time increase in CO₂ concentration (+89 ppm) would be enough to counterbalance future negative drought impacts on carbon fluxes (i.e. the green horizontal bar in the fourth column of Fig. 5, showing the percent change in GPP with increased atmospheric CO₂, which would then be located at the zero line). We also compared the performance of potential European beech forests under current and future extreme dry years over the entire species distributional range (Fig. 6). The results show that the species is potentially vulnerable to climate change at its southern and western range, where projected precipitation changes are large (Fig. S4). European beech forests in the Alps and in the north-eastern part of the distribution range are insensitive to temperature and precipitation changes (Fig. 6, top row). Indeed, water availability over the growing season is not a limiting factor for European beech in the Alps, while CORDEX simulations project a precipitation increase in the north-eastern part of the species range (Fig. S4). As it was the case for the Hesse forest, beech productivity is highly enhanced (the first and the third quartiles of the GPP change are +40 and +55 %, respectively) if we include in our simulations the

4. Discussion

221

222

223

224

225

226

227

228

229

230

231

232

233

234

235

236

237

238

239

240

241

242

243

Drought frequency over the 21st century will increase by more than 20% in Central and Southern Europe and the dynamic response of vegetation to these droughts is highly uncertain (Prudhomme et al. 2014). We thus need ecophysiological models able to capture key forest processes and to provide plausible

increase of future CO2 concentrations, with a clear southwest-northeast gradient of increased forest

productivity due to projected precipitation changes (Fig. 6, bottom row).

extrapolations over severe dry periods in order to assess the impact of increased drought frequency on forest functioning and carbon budgets (Clark et al. 2011; Williams et al. 2018). In our case, we used the MuSICA ecosystem model (Ogée et al. 2003) to analyze the ecophysiological responses of European beech to drought events. The model showed high skills in reproducing key variables and droughtinduced ecosystem limitations over an independent verification year with severe drought (year 2003; Figs. 1-2). MuSICA thus seems an appropriate tool to extrapolate the ecosystem responses to future droughts. Our work refines the results of previous studies and points out important forest processes and dynamical responses in a climate change context. Previous studies quantified a reduction of ecosystem fluxes (GPP and ET) of more than 30% at the Hesse beech forest in 2003 (Granier et al. 2007; Reichstein et al. 2007). Our results give instead a much more conservative estimate (-17% for GPP and -21% for ET). Previous studies used the preceding 2002 year as reference for their quantifications. We rather used a novel anomaly-based approach, attenuating the impact of specific years and of different climate structures such as of wet/dry, warm/cold, or sunny/cloudy days between the dry year 2003 and the reference. We confirm here the key role of water limitations for reduced forest productivity during the 2003 drought (Granier et al. 2007; Reichstein et al. 2007). Furthermore, we are able to disentangle the partitioning between heat stress and water limitations during drought. The results represented in Figure 3 estimate that 82% of the reduced forest productivity in 2003 can be attributed directly to low precipitation. Only 16% of the reduced forest productivity comes from heat stress in the model, while only 2% comes from combined effect of heat stress and water limitations. This low impact of temperature is surprising because high temperatures influence plant biochemistry and increase vapor pressure deficit, dampening stomatal conductance and enhancing soil water evaporation. Granier et al.

(2000) measured a difference of about 23% between ecosystem evapotranspiration and stand

244

245

246

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

262

263

264

265

266

transpiration at Hesse. They concluded that this difference could at least partly be due to soil evaporation. They argued, though, that measurement uncertainties in both, ecosystem evapotranspiration and stand transpiration, are so large that they cannot be proven to be significantly different. This led Betsch et al. (2011) to neglect soil evaporation in their analysis of drought effects on the Hesse beech forest. However, high temperatures should also affect plant biochemistry. Optimum temperatures for biochemical reactions, such a carboxylation, were found to be above 35 °C for Fagus sylvatica (Kattge and Knorr 2007), which are rarely reached in Hesse. However, maximum daily temperatures of the sunlit leaves increased from 39 (August 1st) to 48 °C (August 8th) during the hottest spell of summer 2003, inducing a progressive and almost complete cessation of photosynthesis and transpiration, especially in the upper canopy layer (Fig. S5). Such high temperatures have the potential to damage sun-exposed leaves (e.g. burned epidermis). Although these carry-over effects of heat stress are not accounted for in the model, the rather good agreement of MuSICA with ecosystem-scale water and CO₂ flux observations during and after the 2003 heat wave, indicates that these effects have only a minor influence at the ecosystem scale and that water limitations remain crucial. An important point of our study is the demonstration that rain seasonality matters, i.e. specific drought trajectories over the growing season can potentially drive different impacts to ecosystem fluxes. European beech forests are more sensitive to fast drought onset in the early growing season (Fig. 4) because the soil profile dries out earlier, leading to water limitations during times of highest incoming radiation and hence highest ecosystem fluxes. This result is similar to what D'Orangeville et al. (2018) found for eastern North American forests using tree-ring data but partially contrasts with the findings of Granier et al. (2007) and Arend et al. (2016) pointing out that rainfall events during drought can rapidly restore ecosystem fluxes. Although fluxes increase after rainfall in the fast onset scenario, they cannot close the overall gap with the fluxes of alternative drought trajectories having the same severity but slower onset (i.e. the actual 2003 drought trajectory and the experiment 2003_slow_onset). For this

268

269

270

271

272

273

274

275

276

277

278

279

280

281

282

283

284

285

286

287

288

289

290

reason, a more in-depth analysis of the projected changes in drought seasonality should be sought (see Mo 2011). For example, do we have any indication that crucial months for beech forest productivity, such as June, are becoming drier relative to other months? How much precipitation deficit is needed during these crucial months to significantly reduce ecosystem fluxes and how long does it take to the soil profile to dry (e.g. 1.5 months in 2003 at Hesse; Fig. 3)? Such drought analysis should not be performed with simplistic drought indices, which respond mainly to temperature changes and result in biased interpretations (e.g. Thornwaite equation for potential ET; see Sheffield et al. 2012).

292

293

294

295

296

297

298

299

300

301

302

303

304

305

306

307

308

309

310

311

312

313

314

315

Our comparison of the impact of current and future extreme dry years on European beech forests reveals important key findings. First, the large uncertainties in the future trends of precipitation limit the predictability of forest responses at specific sites and at European scale (see Fig. 5). This known issue cannot fully be resolved due to the impact of high internal climate variability on precipitation (Knutti and Sedláček 2013). Furthermore, we confirmed that caution should be paid when using climate model simulations to analyze extreme years (see Asadieh and Krakauer 2015); climate models are constructed to simulate climate and hence dampen interannual variations of meteorology. Observed severe droughts at specific sites are more extreme than those simulated at the corresponding grid cells (see red and black dots in Fig. 5). Consequently, drastic reductions of ecosystem fluxes at the Hesse forest (a humid site only sensitive to extreme dry years) are only projected for those simulations with lower precipitation amounts at the 2080-2100 horizon compared to today (Fig. 5; see points surrounded by circles). This is consistent with the important role of critical thresholds in relative extractable water determining forest responses to drought (Granier et al. 2007). These thresholds are clearly reached more often over the southern and the western part of the beech distributional range (Fig. 6), where the species has already shown its vulnerability to climate change (Hacket-Pain et al. 2016; Peñuelas et al. 2007). Although our analysis of the potential impact of future droughts on beech forests relies on MuSICA's performance for Hesse, drought response traits of beech trees only show small differences according to tree provenance (Aranda et al. 2014). In all cases, increased CO_2 fertilization in the future (+2.17-time increase of CO_2 concentrations in the RCP8.5 scenarios at the 2080-2100 horizon) will compensate negative drought impacts (Fig. 6). This response is consistent with retrospective studies based on tree-ring $\delta^{13}C$ data, which show important increases in water use efficiency of beech trees over the past century (up to 44%; Duquesnay et al. 1998). This response is also consistent with the increased productivity of European forests projected by other modelling studies over the next century (Davi et al. 2006; Gea-Izquierdo et al. 2017). However, controlled experiments under a CO_2 -enriched atmosphere seem to show a more moderate CO_2 fertilization effect with almost no changes in productivity (Bader et al. 2013). The effects of nutrient limitations should hence be included in future studies (see Luo et al. 2004; Norby et al. 2010).

5. Conclusion

In this study, we used the MuSICA ecosystem model (Ogée et al. 2003) to analyze the potential impact of present and future droughts on fluxes and processes of European beech forests. Using MuSICA, we were able to simulate realistic carbon and water fluxes at the leaf- and soil-level, and to quantify precisely the impact of drought. Severe droughts may reduce European beech GPP and ET annual fluxes by 20%. We demonstrated that low precipitation limits beech forest productivity much more than heat stress. Beech forest ecosystems are also sensitive to the timing of drought onset over the growing season. Finally, our results show a good adaptation of European beech ecosystems to projected climate change because increased CO₂ concentration in the future will enhance plant water use efficiency counterbalancing negative drought impacts. However, large uncertainties exist as simulated trends in ecosystem fluxes depend mainly on projected trends and anomalies of precipitation, which are both largely uncertain in climate model projections. This study represents a step forward to a more complete ecophysiological understanding of the potential impact of drought on the productivity and seasonal

dynamics of European beech forests, which should be enhanced in the future by the inclusion of additional processes such as nitrogen limitations, legacy/carry-over effects after extreme events, or growing season and phenological changes.

Acknowledgements

We acknowledge André Granier, Bernard Longdoz, Patrick Gross, Jean-Baptiste Lily, and Pascal Courtois for their support with the data from the Hesse forest experimental station. This work was supported by a grant overseen by the French National Research Agency (ANR) as part of the "Investissements d'Avenir" program (ANR-11-LABX-0002-01, Lab of Excellence ARBRE), and by funding from the "Conseil Régional Grand Est" within the program "soutien aux jeunes chercheurs". Funding was also provided by the ForRISK project (ERA-NET Sumforest). Sumforest was funded by the European Union under Grant Agreement No. 606803. The ERA-NET Sumforest project ForRISK was funded in France through the French National Research Agency ANR (Grant No. ANR-16-SUMF-0001-01).

Competing interests

Declarations of interest: none.

References

356	Abbass HA, Sarker R, Newton C (2001) PDE: a Pareto-frontier differential evolution approach for multi
357	objective optimization problems. In: Proceedings of the Congress on Evolutionary Computation
358	2001. IEEE, pp 971-978. doi:10.1109/CEC.2001.934295
359	Aranda I et al. (2014) Variation in photosynthetic performance and hydraulic architecture across
360	European beech (Fagus sylvatica L.) populations supports the case for local adaptation to wate
361	stress. Tree Physiology 35:34-46.
362	Arend M, Sever K, Pflug E, Gessler A, Schaub M (2016) Seasonal photosynthetic response of European
363	beech to severe summer drought: Limitation, recovery and post-drought stimulation
364	Agricultural and Forest Meteorology 220:83-89.
365	Asadieh B, Krakauer N (2015) Global trends in extreme precipitation: climate models versus
366	observations. Hydrology and Earth System Sciences 19:877-891.
367	Bader MKF, Leuzinger S, Keel SG, Siegwolf RT, Hagedorn F, Schleppi P, Körner C (2013) Central European
368	hardwood trees in a high-CO2 future: synthesis of an 8-year forest canopy CO2 enrichmen
369	project. Journal of Ecology 101:1509-1519.
370	Battipaglia G, Saurer M, Cherubini P, Calfapietra C, McCarthy HR, Norby RJ, Cotrufo MF (2013) Elevated
371	CO2 increases tree-level intrinsic water use efficiency: insights from carbon and oxygen isotope
372	analyses in tree rings across three forest FACE sites. New Phytologist 197:544-554
373	doi:10.1111/nph.12044
374	Betsch P, Bonal D, Breda N, Montpied P, Peiffer M, Tuzet A, Granier A (2011) Drought effects on wate
375	relations in beech: the contribution of exchangeable water reservoirs. Agricultural and Fores
376	Meteorology 151:531-543.
377	Bonan GB, Doney SC (2018) Climate, ecosystems, and planetary futures: The challenge to predict life in
378	Earth system models. Science 359:eaam8328.

379	Braud I, Dantas-Antonino AC, Vauclin M, Thony JL, Ruelle P (1995) A simple soil-plant-atmosphere
380	transfer model (SiSPAT) development and field verification. Journal of hydrology 166:213-250.
381	doi:https://doi.org/10.1016/0022-1694(94)05085-C
382	Caudullo G, Welk E, San-Miguel-Ayanz J (2017) Chorological maps for the main European woody species.
383	Data in Brief 12:662-666. doi:https://doi.org/10.1016/j.dib.2017.05.007
384	Ciais P et al. (2005) Europe-wide reduction in primary productivity caused by the heat and drought in
385	2003. Nature 437:529.
386	Clark JS et al. (2011) Inferential ecosystem models, from network data to prediction. Ecological
387	Applications 21:1523-1536. doi:10.1890/09-1212.1
388	Cuntz M et al. (2015) Computationally inexpensive identification of noninformative model parameters
389	by sequential screening. Water Resources Research 51:6417-6441.
390	D'orangeville L et al. (2018) Drought timing and local climate determine the sensitivity of eastern
391	temperate forests to drought. Global Change Biology 24:2339-2351.
392	Davi H et al. (2006) Sensitivity of water and carbon fluxes to climate changes from 1960 to 2100 in
393	European forest ecosystems. Agricultural and Forest Meteorology 141:35-56.
394	Domec J-C et al. (2012) Interactive effects of nocturnal transpiration and climate change on the root
395	hydraulic redistribution and carbon and water budgets of southern United States pine
396	plantations. Tree Physiology 32:707-723.
397	Dufrêne E, Davi H, François C, Maire Gl, Dantec VL, Granier A (2005) Modelling carbon and water cycles
398	in a beech forest: Part I: Model description and uncertainty analysis on modelled NEE. Ecological
399	Modelling 185:407-436. doi:http://dx.doi.org/10.1016/j.ecolmodel.2005.01.004
400	Duquesnay A, Breda N, Stievenard M, Dupouey J (1998) Changes of tree-ring δ13C and water-use
401	efficiency of beech (Fagus sylvatica L.) in north-eastern France during the past century. Plant,
402	Cell & Environment 21:565-572.

403	Farquhar GD, von Caemmerer S, Berry JA (1980) A biochemical model of photosynthetic CO2
404	assimilation in leaves of C3 species. Planta 149:78-90. doi:10.1007/bf00386231
405	Gea-Izquierdo G, Nicault A, Battipaglia G, Dorado-Liñán I, Gutiérrez E, Ribas M, Guiot J (2017) Risky
406	future for Mediterranean forests unless they undergo extreme carbon fertilization. Global
407	Change Biology 23:2915-2927.
408	Gessler A, Keitel C, Kreuzwieser J, Matyssek R, Seiler W, Rennenberg H (2007) Potential risks for
409	European beech (Fagus sylvatica L.) in a changing climate. Trees 21:1-11.
410	Giuggiola A, Ogée J, Rigling A, Gessler A, Bugmann H, Treydte K (2016) Improvement of water and light
411	availability after thinning at a xeric site: which matters more? A dual isotope approach. New
412	Phytologist 210:108-121. doi:10.1111/nph.13748
413	Granier A, Biron P, Lemoine D (2000) Water balance, transpiration and canopy conductance in two
414	beech stands. Agricultural and Forest Meteorology 100:291-308.
415	Granier A, Bréda N, Longdoz B, Gross P, Ngao J (2008) Ten years of fluxes and stand growth in a young
416	beech forest at Hesse, North-eastern France. Annals of Forest Science 65:704.
417	Granier A et al. (2007) Evidence for soil water control on carbon and water dynamics in European forests
418	during the extremely dry year: 2003. Agricultural and Forest Meteorology 143:123-145.
419	Hacket-Pain AJ, Cavin L, Friend AD, Jump A (2016) Consistent limitation of growth by high temperature
420	and low precipitation from range core to southern edge of European beech indicates
421	widespread vulnerability to changing climate. European Journal of Forest Research 135:897-909.
422	Jacob D et al. (2014) EURO-CORDEX: new high-resolution climate change projections for European
423	impact research. Regional environmental change 14:563-578.
424	Kattge J, Knorr W (2007) Temperature acclimation in a biochemical model of photosynthesis: a
425	reanalysis of data from 36 species. Plant, Cell & Environment 30:1176-1190.

426	Klein T et al. (2014) Quantifying transpirable soil water and its relations to tree water use dynamics in a			
427	water-limited pine forest. Ecohydrology 7:409-419.			
428	Knutti R, Sedláček J (2013) Robustness and uncertainties in the new CMIP5 climate model projections.			
429	Nature Climate Change 3:369.			
430	Leuning R (1995) A critical appraisal of a combined stomatal-photosynthesis model for C3 plants. Plant,			
431	Cell & Environment 18:339-355. doi:10.1111/j.1365-3040.1995.tb00370.x			
432	Lindner M et al. (2010) Climate change impacts, adaptive capacity, and vulnerability of European forest			
433	ecosystems. Forest Ecology and Management 259:698-709.			
434	Luo Y et al. (2004) Progressive nitrogen limitation of ecosystem responses to rising atmospheric carbon			
435	dioxide. BioScience 54:731-739.			
436	Mauder M et al. (2013) A strategy for quality and uncertainty assessment of long-term eddy-covariance			
437	measurements. Agricultural and Forest Meteorology 169:122-135.			
438	Mier PR (2017, August 25) pablormier/yabox: v1.0.3 (Version v1.0.3). Zenodo.			
439	doi:http://doi.org/10.5281/zenodo.848679			
440	Mo KC (2011) Drought onset and recovery over the United States. Journal of Geophysical Research:			
441	Atmospheres 116:D20106.			
442	Morris MD (1991) Factorial sampling plans for preliminary computational experiments. Technometrics			
443	33:161-174.			
444	Norby RJ, Warren JM, Iversen CM, Medlyn BE, McMurtrie RE (2010) CO2 enhancement of forest			
445	productivity constrained by limited nitrogen availability. Proceedings of the National Academy of			
446	Sciences 107:19368-19373.			
447	Ogée J, Brunet Y, Loustau D, Berbigier P, Delzon S (2003) MuSICA, a CO2, water and energy multilayer,			
448	multileaf pine forest model: evaluation from hourly to yearly time scales and sensitivity analysis.			
449	Global Change Biology 9:697-717. doi:10.1046/j.1365-2486.2003.00628.x			

450	Ogée J et al. (2004) Partitioning net ecosystem carbon exchange into net assimilation and respiration
451	with canopy-scale isotopic measurements: An error propagation analysis with 13CO2 and
452	CO18O data. Global Biogeochemical Cycles 18:GB2019.
453	Peñuelas J, Ogaya R, Boada M, S. Jump A (2007) Migration, invasion and decline: changes in recruitment
454	and forest structure in a warming-linked shift of European beech forest in Catalonia (NE Spain).
455	Ecography 30:829-837.
456	Prudhomme C et al. (2014) Hydrological droughts in the 21st century, hotspots and uncertainties from a
457	global multimodel ensemble experiment. Proceedings of the National Academy of Sciences
458	111:3262-3267.
459	Raupach M (1989) A practical Lagrangian method for relating scalar concentrations to source
460	distributions in vegetation canopies. Quarterly Journal of the Royal Meteorological Society
461	115:609-632.
462	Reichstein M et al. (2007) Reduction of ecosystem productivity and respiration during the European
463	summer 2003 climate anomaly: a joint flux tower, remote sensing and modelling analysis. Global
464	Change Biology 13:634-651. doi:10.1111/j.1365-2486.2006.01224.x
465	Reichstein M et al. (2005) On the separation of net ecosystem exchange into assimilation and ecosystem
466	respiration: review and improved algorithm. Global Change Biology 11:1424-1439.
467	Saltelli A, Tarantola S, Campolongo F, Ratto M (2004) Sensitivity analysis in practice: a guide to assessing
468	scientific models. John Wiley & Sons. doi:10.1002/0470870958
469	Schaap MG, Leij FJ, Van Genuchten MT (2001) Rosetta: A computer program for estimating soil hydraulic
470	parameters with hierarchical pedotransfer functions. Journal of hydrology 251:163-176.
471	Scharnweber T, Manthey M, Criegee C, Bauwe A, Schröder C, Wilmking M (2011) Drought matters—
472	declining precipitation influences growth of Fagus sylvatica L. and Quercus robur L. in north-
473	eastern Germany. Forest Ecology and Management 262:947-961.

474	Sheffield J, Wood EF, Roderick ML (2012) Little change in global drought over the past 60 years. Nature
475	491:435.
476	Sinoquet H, Moulia B, Gastal F, Bonhomme R, Varlet-Grancher C (1990) Modeling the radiative balance
477	of the components of a well-mixed canopy: application to a white clover-tall fescue mixture.
478	Acta Oecologica 11:469-486.
479	Williams KE et al. (2018) Revisiting the First ISLSCP Field Experiment to evaluate water stress in
480	JULESv5.0. Geoscientific Model Development 2018:1-47. doi:10.5194/gmd-2018-210
481	Wingate L, Ogée J, Burlett R, Bosc A (2010) Strong seasonal disequilibrium measured between the
482	oxygen isotope signals of leaf and soil CO2 exchange. Global Change Biology 16:3048-3064.
483	Zhao M, Running SW (2010) Drought-induced reduction in global terrestrial net primary production from
484	2000 through 2009. Science 329:940-943.
485	Zhao T, Dai A (2017) Uncertainties in historical changes and future projections of drought. Part II: model-
486	simulated historical and future drought changes. Climatic Change 144:535-548.

487 Tables

Table 1. Local MuSICA simulations at Hesse used in the four experiments.

Experiments*	Simulation name	Monthly temperature	Monthly precipitation	Daily/hourly variability	Temperature/precipitation annual anomalies
1, 2, 3	2003	2003	2003	2003	2003
1, 2	2003_median_climate	Median climate (1997-2013)	Median climate (1997-2013)	2003	Median climate
2	2003_only_P_limited	Median climate (1997-2013)	2003	2003	Median climate/2003
2	2003_only_T_limited	2003	Median climate (1997-2013)	2003	2003/Median climate
3	2003_fast_onset	Anomalies added to June-September: +2, +1, -1, -2 (°C)	Anomalies of June-September: 10%, 20%, 30%, 40% of total precipitation over the period	2003	2003
3	2003_slow_onset	Anomalies added to June-September: -2, -1, +1, +2 (°C)	Anomalies of June-September: 40%, 30%, 20%, 10% of total precipitation over the period	2003	2003
4	Xtreme_today	Modified by deltas relative to extreme dry years in CORDEX simulations at the 2006-2026 horizon	Modified by change factors relative to extreme dry years in CORDEX simulations at the 2006-2026 horizon	2003	Extreme simulated years at the 2006-2026 horizon
4	Median_futur	Modified by deltas relative to median years in CORDEX simulations at the 2080-2100 horizon	Modified by change factors relative to median years in CORDEX simulations at the 2080-2100 horizon	2003	Median simulated years at the 2080-2100 horizon
4	Xtreme_futur	Modified by deltas relative to extreme dry years in CORDEX simulations at the 2080-2100 horizon	Modified by change factors relative to extreme dry years in CORDEX simulations at the 2080-2100 horizon	2003	Extreme simulated years at the 2080-2100 horizon

^{*} Experiments are defined in the section "Local simulation experiments". The same simulation can be used in more than one experiment.

489

490 Table 2. EURO-CORDEX RCP8.5 simulations used to define the projected climate change.

Regional climate model	Driving global climate model	Member
ALADIN53	CNRM-CERFACS-CNRM-CM5	r1i1p1
RCA4	ICHEC-EC-EARTH	r12i1p1
WRF331F	IPSL-IPSL-CM5A-MR	r1i1p1
CCLM4-8-17	MOHC-HadGEM2-ES	r1i1p1
REMO2009	MPI-M-MPI-ESM-LR	r1i1p1
REMO2015	NCC-NorESM1-M	r1i1p1

Figures

Figure 1. Simulated versus observed daily values of water-related variables and forest productivity at Hesse during the 2002 calibration year and the 2003 dry validation year. From top to bottom: gross primary productivity, evapotranspiration, soil water content in the top 70 cm. For soil water content, red (blue) points correspond to daily values before (after) August 15, which is in the middle of the yearly dry period at Hesse. Grey lines are the one-to-one lines.

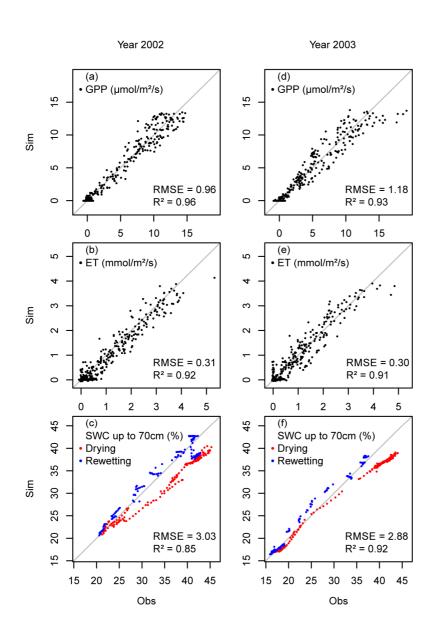


Figure 2. Impact of soil water availability on evapotranspiration (2001-2008): daily mean observed (black) and simulated (red) evapotranspiration versus soil water content in the top 70 cm for May to September. The upper envelope lines with larger circles show the 95th percentiles of the evapotranspiration data divided in 15 equally spaced bins of soil water content. Only sunny days with mean daily net radiation greater than 100 W/m² were considered.

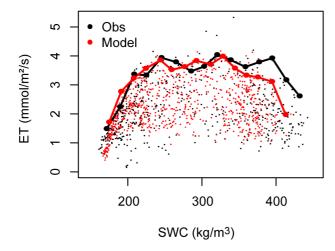


Figure 3. Quantification of the reduction of gross primary production and evapotranspiration due to heat stress, water limitations and combined effects during the 2003 drought. Lines are cumulative sums of daily gross primary productivity and evapotranspiration in 2003, simulated with the full 2003 dry climate (2003, black), simulated with a median climate but diurnal variations of 2003 (2003_median_climate, red), simulated with median climate and the 2003 precipitation anomalies only (2003_P_limited, green), simulated with median climate and the 2003 temperature anomalies only (2003_T_limited, blue). Horizontal grey arrows show the time lapse between the start of the drought and the divergence among simulations. Vertical black arrows indicate the 2003 reductions of fluxes between simulations with a median climate and the full 2003 dry climate.

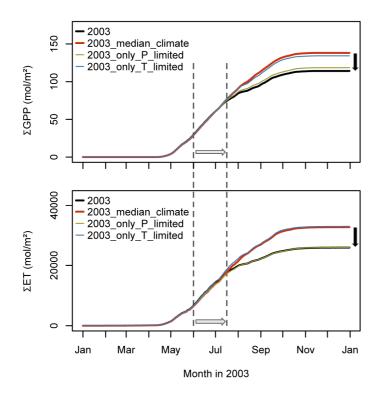


Figure 4. Quantification of the impact of different hypothetic drought trajectories. Lines are cumulative sums of daily gross primary productivity and evapotranspiration in 2003, simulated with the full 2003 dry climate (2003, black) and with modified versions of the 2003 climate. In the simulations "2003_fast_onset" and "2003_slow_onset", the 2003 temperature and precipitation values were modified to have a faster or slower drought onset but same overall anomalies over the growing season (see Table 1).

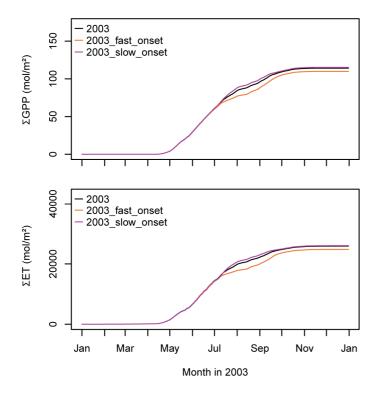


Figure 5. Comparison of annual gross primary productivity, evapotranspiration and water use efficiency (net assimilation divided by transpiration) at the Hesse beech forest under current and future (2080-2100) climate from six EURO-CORDEX simulations; future MuSICA simulations use either current atmospheric CO₂ levels or projected, increased atmospheric CO₂ concentrations (2.17-time increase). Each point represents the percent change (Eq. 3) of MuSICA input climate (May-September mean temperature and precipitation) from Regional Climate Models (Table 2) and of MuSICA simulated fluxes. The points for the simulation with driest projected climate are surrounded by circles. Horizontal bars represent the median of the six simulations. The specific MuSICA simulations are given in Table 1. **Xtreme_future vs. 2003_median_climate*, for example, means the percent change in the annual sum if MuSICA is driven by the most extreme year of specific CORDEX simulations over 2080-2100 compared to MuSICA driven by the median climate of 1997-2013. Black dots, showing the estimated impact of the observed 2003 drought at Hesse, are given for comparison.

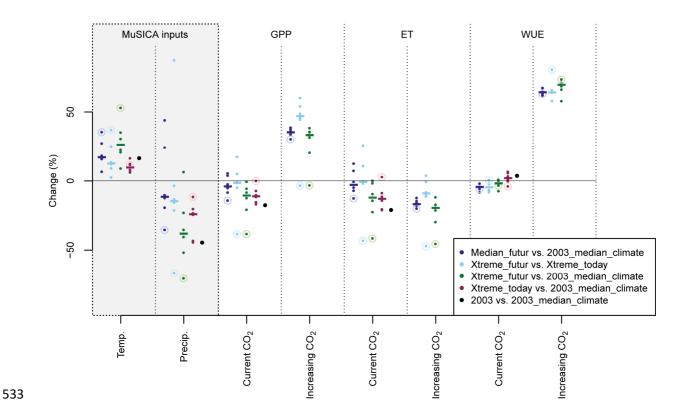


Figure 6. Comparison of gross primary productivity and evapotranspiration of potential European beech forests under current and future (2080-2100) extreme dry years over the species distributional range. Forcing temperature and precipitation used to drive MuSICA simulations are derived from the median of six extreme years, one pear each EURO-CORDEX simulation. Percent changes are computed as described in Eq. 3. The top row uses current atmospheric CO₂ concentration and the bottom row uses future CO₂ levels (2.17 current CO₂). The grey cross in Eastern France is the location of the Hesse forest experimental station.

