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Birds and bats contribute to natural regulation of the millet head miner in tree-crop agroforestry systems

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\textbf{Abstract}

A better understanding of environmental factors driving natural pest regulation is a major challenge for designing sustainable cropping systems. The objective of the present study was to assess the association between vegetation openness in traditional tree-crop agroforestry systems in Senegal, richness and
abundance of vertebrates including insectivorous birds and bats, and their contribution to the natural regulation of crop pests. The millet head miner (MHM), *Heliocheilus albipunctella* (Lepidoptera, Noctuidae), a major constraint to increasing millet crop productivity in sub-Saharan Africa, was selected as a model system. Ten sites separated by at least one km from each other were selected in a 100 km² study area in the Peanut basin in Senegal. In each site, a pair of millet fields distant from at least 100 m each was selected according to a gradient of vegetation openness within a 100-m radius buffer with sampling plot (5 x 5 m) at the center. Nine insectivorous bird and bat species were recorded in millet fields over the 2017 cropping season and their predator status was confirmed by direct observation or DNA detection in fecal samples. Grain losses were reduced when panicles were accessible to birds and bats, confirming their net contribution to pest regulation. At a local scale, tree diversity and vegetation openness were important predictors of the abundance of insectivorous village weaver birds and grey-headed sparrows, respectively. Some tree species (soapberry trees and neems) indirectly contributed to natural regulation of the MHM likely by providing refuges to insectivorous vertebrates whereas other tree species (baobabs) provided disservices as possible refuges for the MHM moths. Further research is needed to better understand relationships between tree cover, food web interactions and natural pest suppression, so that specific conservation measures such as habitat management can be designed to improve pest control.

1. Introduction

During the last decades, worldwide losses of biodiversity have occurred at an unprecedented rate and agricultural intensification has been a major driver (Cardinale et al., 2012; Philpott et al., 2008; Tilman et al., 2002; Wilby and Thomas, 2002). The dramatic land use changes include the simplification of agricultural landscapes by deforestation and monoculture and the extensive use of agricultural inputs such as broad-spectrum insecticides to control crop pests, with detrimental impact on ecosystem health.
Along with the loss of functional biodiversity, particularly biological control agents, such cultivated ecosystems have become more and more susceptible to pest outbreaks. This is particularly pregnant in southern contexts where arthropod pests pose a permanent threat to food security (Garrity et al., 2010). New models for crop protection are needed to reduce the dependence of agricultural systems on pesticides that stimulate ecological processes of natural regulation of crop pests (Brévault and Clouvel, 2019). One approach is to use the benefits of agroforestry systems to enhance or restore functional biodiversity (here, natural enemies of crop pests) to boost ecosystem services such as biological control of crop pests. A meta-analysis conducted by Pumarino et al. (2015) showed that agroforestry practices generally result in higher abundance of natural enemies of pests, lower pest abundance and less plant damage in perennial crops (e.g., coffee, cocoa and plantain) but not in annual crops (e.g., maize, rice and beans). However, studies carried out in Kenya showed that abundance of stem borers was significantly lower in *Leucaena leucocephala* (Lam.) (Fabales, Mimosaceae) alley cropping than in a maize monocrop (Ogil et al., 1999).

Among natural enemies of crop pests, vertebrates including insectivorous bats and birds can provide valuable ecosystem services such as suppression of insect pests (Maas et al., 2013; Maine and Boyles, 2015; Pejchar et al., 2018). By protecting crops through feeding on herbivorous pests, they can provide substantial increases in crop yields (Bael et al., 2008; Karp et al., 2013) and a true alternative to chemical insecticides (Bianchi et al., 2006; Tscharntke et al., 2012). The abundance and diversity of semi-natural habitats within agricultural landscapes, particularly trees, has the potential to increase the abundance and diversity of birds and bats (Harvey et al., 2006; Redlich et al., 2018) and thus natural regulation of crop pests (e.g., nesting and breeding site, alternative preys). However, birds and bats can also negatively impact natural regulation of crop pests by consuming natural enemies of crop pests, i.e., intra-guild predation (Grass et al. 2017; Tschumi et al., 2018). On the other hand, trees can also benefit pests by providing food and refuges (Schroth et al., 2000). More research is thus needed on ecological issues...
to better understand how tree-crop agroforestry systems can contribute to enhance ecosystem services such as biological control of crop pests. Although a substantial body of research has been performed on the role of bats and birds in perennial cropping systems such as coffee plantations (Bael et al., 2008), few studies have focused on annual cropping systems as an unstable mosaic of habitats (Kennedy and Storer, 2000). More research is thus needed to deepen our knowledge of the life system of insectivorous birds and bats including seasonal population dynamics (e.g., breeding cycles, migration), multitrophic interactions within food webs (e.g. intraguild predation; see Grass et al., 2017), and temporal suitability of habitats in agricultural landscapes (Maas et al., 2016).

Agricultural landscapes in the “Peanut basin” in Senegal (West Africa) are generally structured by traditional agroforestry parkland systems (AFS), in which trees (mainly apple-ring acacias, *Faidherbia albida*) (Fabales, Fabaceae) are distributed in cultivated fields (Seyler, 1993). Even though the impact of those trees on soil fertility has been widely demonstrated, their effect on crop pests and biological control has been poorly studied (Hadgu et al., 2009; Umar et al., 2013). At the landscape scale, they offer a greater diversity of ecological niches in time and space than a simple mosaic of annual crops. The pearl millet, *Pennisetum glaucum* (L.) R. Br. (Cyperales, Poaceae), is the main annual crop in those agricultural systems. However, a key insect pest, the millet head miner (MHM) (*Heliocheilus albipunctella* de Joannis) (Lepidoptera, Noctuidae) is a permanent threat for increasing crop yields (Gahukar, 1984; Sow et al., 2018; Vercambre, 1978). Moths emerge from the soil one to two months after the beginning of the rainy season. After mating, females lay eggs in the millet panicle on the silk of the flower involucre or under stamens, but also on the floral peduncles or on the spine (Nwanze and Harris, 1992). The young larvae perforate the glumes and consume flowers, while the older ones cut off the floral peduncles (Gahukar, 1984; Vercambre, 1978), thus preventing grain formation. In the absence of any insecticide application by farmers, millet production relies on pest regulation by natural enemies (Bhatnagar, 1987; Nwanze and Sivakumar, 1990; Sow et al., 2019, 2017). A two-year monitoring of 90 millet fields and their surrounding
landscape (2000 m-buffer around millet fields) showed that abundance of trees at the landscape scale significantly increased natural regulation of the MHM (Soti et al., 2019). Using high throughput molecular tools, the food web structures and ecological processes underlying natural regulation of crop pests by arthropods were documented and a large diversity of parasitoids and predators among arthropods were identified (Sow et al., 2018).

In the present study, we investigated the contribution of vertebrates, particularly insectivorous birds and bats, to the natural regulation of the MHM in tree-crop agroforestry systems in Senegal. We hypothesized that (i) abundance and diversity of trees at a local and landscape scale enhance the abundance and diversity of insectivorous birds, and that (ii) abundance and diversity of insectivorous birds at a local scale reduce pest incidence. We used satellite and drone imagery, remote sensing and geographical information system (GIS), to map and quantify the abundance and diversity of trees species at local and landscape scale around a set of 20 millet fields. Pest incidence and grain loss were evaluated by direct field observations on millet panicles while natural regulation was assessed using exclusion cages. In addition, the abundance and diversity of birds and bats was monitored in and around millet fields and fecal samples were collected to check for the presence of the MHM based on DNA analyses. Results are discussed in the light of relationships between tree diversity and abundance, role of vertebrates in food webs (ecosystem services vs. disservices) and biological control of crop pests for the deployment of ecologically-intensive cropping systems.

2. Materials and Methods

2.1. Study area

The field survey was conducted in 2017 in the northern area of the “Peanut basin”, near Bambey (14°43'0.79"N ; 16°30'5.56"O) in Senegal. This area is under the influence of a Sudano-Sahelian climate...
with a short rainy season (400-600 mm) from July to October (monsoon). The landscape is composed of
tree-crop agroforestry systems mainly including apple-ring acacias (*Faidherbia albida*), soapberry trees
(*Balanites aegyptica*; Sapindales, Zygophyllaceae), baobabs (*Adansonia digitata*; Malvales,
*Bombacaceae*), tamarinds (*Tamarindus indica*; Fabales, Caesalpiniaceae) and red acacias (*Acacia seyal*;
Fabales, Fabaceae). Crops mainly include staple crops such as pearl millet, peanuts, and cowpea.

2.2. Study design

A total of 10 sites distant from at least one km each other (Fig. 1) were selected across the landscape
(100 km²) according to the land-use proportion of tree cover (2.8-8.7%) at a landscape scale (one-km
radius buffer around millet fields), following previous treatment of very high resolution satellite imagery
(Soti et al., 2018). In each site, two millet fields distant from at least 100 m each were selected according
to vegetation openness (open vs. closed) within a 100-m radius buffer with sampling plot (5 x 5 m) at the
center. The minimum-maximum distance between paired sampling points was 100 m (Fig. 1). Sampling
points captured a large variation in the abundance (number of trees and land-use proportion of tree
cover) and diversity (SHDI) of trees (Table 1). Tree cover around sampling plots (100-m radius buffer) was
mainly composed of *A. indica*, *B. aegyptiaca*, *F. albida*, and *A. digitata* (Table 1). Situations of closed
vegetation were mostly found around “compound” (close-to-home) millet fields with high abundance of
neems, whereas “bush” millet fields presented more open vegetation (Fig. S1).

2.3 Landcover mapping

Drone photogrammetry and object-based image analysis procedures were used to characterize each site
following the method described in Sarron *et al.* (2018). Briefly, each site was overflowed using a UAV (DJI
Mavic Pro, DJI Inc, China) in order to compute a high resolution orthomosaic (1.30 cm.pixel⁻¹) and a
Canopy Height Model (CHM) using a photogrammetry procedure. The two layers were then analysed in
eCognition Developer 9 software (Trimble Geospatial, Munich, Germany) through an object-based classification workflow to obtain land cover maps. These landcover maps contained 5 object classes: millet crops, bare soil, trees, grasslands, and building (Fig. 1). Tree species were manually identified within the 100-m radius buffer (photointerpretation) and corrections were made when needed. Land cover maps were expressed as geospatial vectors with each object, including the delineated trees, corresponding to a geolocalized polygon and described in an adjunct database (.shapefile format) to be imported in GIS and processed in Fragstats software (McGarigal and Marks, 1995) for spatial analysis.

2.4. Data collection

The incidence of the MHM was monitored in the 20 sampling plots by counting eggs on 25 randomly selected panicles at panicle emergence. Counting is done with a round trip on the surface of the panicle with a fine forceps. Females of the MHM usually lay their eggs on emerging panicles up to the male flowering stage, between involucral bristles and glumes.

At the male flowering stage of millet, i.e. after completion of egg development, sampling plots were divided into two subplots of each 2.5 x 5 m. One subplot designated bird exclosure, while the other served as control. Bird exclosures (2.5 m height) were built with wooden frames covered with a polyethylene nylon mesh (25 x 25 mm cell size) that allowed predatory arthropods to enter but that excluded birds. In each cage, eight panicles were protected with a mesh bag to exclude natural enemies, while 16 other panicles were left open (without mesh bag) to exclude only birds. The same design was reproduced outside the cage in the second subplot with eight panicles protected with a mesh bag to exclude natural enemies, and 16 other panicles left open as a control. As a result, a total of 960 panicles were monitored in the study design. Damages caused by the MHM (number and length of mines) were recorded from each panicle just before harvest, i.e., 21 days later. Conversion of damages to grain losses was done using allometric relations (Thiaw et al., 2017).
Field monitoring of birds and bats was performed on the five sites located in the southern part of the study area, over a period of 15 days (3 days in each site) from millet grain filling to maturity. Observations of birds were carried out around sampling plots in the morning (7.00 to 9.00 am UTC) and in the late afternoon (4.30 to 6.30 pm UTC). The species and food diet (millet grain vs. MHM larvae) of birds observed on millet panicles were recorded. Bird abundance in millet fields was calculated by using the frequency of occurrence (C) proposed by Linsdale (1928). Bird species were identified using the “Guide of West African birds” (Borrow and Demey, 2015). Bats were counted at regular time intervals (5 min) at dusk (from 6.30 to 7.40 pm UTC).

2.5. Analysis of fecal samples

To check the predation of the MHM by vertebrates, fecal samples of the village weaver (*Ploceus cucullatus*; Passeriformes, Ploceidae) and the Mauritian tomb bat (*Taphozous mauritianus*; Chiroptera, Emballonuridae) were collected during 24 hours by using a plastic tarpaulin placed under nesting trees, neem (*Azadirachta indica*; Sapindales, Meliaceae) and palm trees (*Borassus aethiopum*; Arecales, Arecaceae), respectively. Samples were collected at three millet phenological stages corresponding to successive development of the immature stages of the MHM: (1) panicle emergence to female flowering for eggs, (2) panicles with grains at the milk stage for the 1\textsuperscript{st} and 2\textsuperscript{nd} instar larvae, and (3) panicles with grains at the dough stage for 3\textsuperscript{rd} and 4\textsuperscript{th} instar larvae, from late August to late September. A total of 80 and 92 fecal samples for birds and bats, respectively, were individually collected using clean cotton buds and placed in 2-ml microtubes with silica-gel granules to absorb moisture and prevent the development of molds and DNA degradation. Detection of the MHM in fecal samples was performed by amplification of the 133 bp fragment of cytochrome c oxidase I (COI mini-barcode, Gillet et al., 2015) and sequencing on a MiSeq Illumina platform (Appendix A).
2.6. Statistical analyses

The R program (version 3.6.1) was used for statistical analyses (R Core team, 2014). We used generalized linear models (\textit{glm} function from the \textit{MASS} package) to test the effect of bird abundance and diversity on pest incidence (including the number of eggs per panicle and grain loss). Models were fitted using the appropriate distribution type and link function: binomial for proportion of infested panicles or grain loss (\textit{link = logit}), and Poisson for abundance (count data, \textit{link = log}). We then tested with the same type of models the effect of tree cover at a local and landscape scale on the abundance and diversity of birds, and on pest incidence. Multi-model inference approach was performed to examine the relative importance of each explanatory variable (Burnham and Anderson, 2002). The \textit{MuMIn} package was used to perform model selection and model averaging. Akaike weights were calculated according to the Akaike information criterion (AICc) which is a bias-corrected version of the AIC recommended in analyses where the sample size is small relative to the number of fitted parameters (Burnham and Anderson, 2002; Rusch et al., 2011; Soti et al., 2019). The indirect effect of natural enemy exclosure (three treatments: mesh bag, cage, and open as control) on grain loss was analysed using a linear mixed model (\textit{lmer} function in the \textit{lme4} package) with site as a random effect to account for the variation due to this factor. Multiple post-hoc comparisons were carried out with Tukey contrasts using the \textit{glht} function from the \textit{multcomp} package.

3. Results

3.1. Abundance and diversity of insectivorous birds and bats

Sampling points captured a large variation in the abundance (number) and diversity (SHDI) of observed birds (Table 1). A total of eight insectivorous birds belonging to five families (Corvidae, Passeridae, Ploceidae, Pycnonotidae and Sturnidae) were observed in millet fields (Table 2). The grey-headed sparrow (\textit{Passer griseus}, Passeriformes, Passeridae), the village weaver (\textit{P. cuculatus}), and the little weaver (\textit{Ploceus luteolus}; Passeriformes, Ploceidae) were the most frequently observed species (Table 2).
The frequency of occurrence of the five other bird species observed was very low (<3%). Direct field observations showed that the encountered bird species were omnivorous (both granivorous and insectivorous), except Lamprotornis pulcher (Muller) (Passeriformes, Sturnidae) which was observed feeding exclusively on insects. We identified two bat species foraging around millet sampling plots: a frugivorous species, the African straw-colored fruit bat (Eidolon helvum; Chiroptera, Pteropodidae) and an insectivorous species, the Mauritian tomb bat (T. mauritianus) (Table 2).

Using a metabarcoding approach, we successfully detected DNA sequences of the MHM in fecal samples of the village weaver bird and the Mauritian tomb bat. The frequency of detection of the MHM DNA in fecal samples from village weavers indicated that these birds primarily feed on late instar larvae (Fig. 2). Detection of the MHM DNA in bats indicated that they feed on MHM moths (egg oviposition period) and probably also on late instar larvae.

### 3.2. Effect of tree cover on insectivorous birds

The diversity of trees was the most important predictor explaining the abundance of birds observed in and around millet sampling plots (Fig. 3A). The abundance of birds increased with the diversity of trees in a 100-m radius around sampling plots. The abundance of the village weavers followed the same trend, whereas the abundance of the grey-headed sparrow was negatively correlated to the abundance of trees (Fig. 3B, 3C). No effect of tree cover on the abundance or diversity of birds was detected at a landscape scale (1000 and 1750-m radius buffer around sampling points). Direct field observations showed that palm trees served as preferential nesting site for the Mauritian tomb bat.

### 3.3. Effect of insectivorous vertebrates on pest incidence

The MHM incidence in millet fields was generally high (53.4% egg-infested panicles, 96% damaged panicles), but variable among fields (24-80% egg-infested panicles, 87-100% damaged panicles). Grain losses due to the MHM varied from 9.7 to 29.3% among sampling plots (Table 1). The abundance and
diversity of birds did not contribute significantly in explaining pest incidence levels (Fig. 4). Grain loss was significantly reduced when panicles were accessible to arthropods only (-17.4%), and even more (-35.7%) when they were accessible to both arthropods and birds (Fig. 5).

3.4. Effect of tree cover on MHM incidence

The abundance of baobabs at a local scale, i.e., 100-m radius buffer around sampling plots, increased MHM egg infestation on panicles and grain losses (Fig. 6). Conversely, the abundance of soapberry trees in the same area decreased egg infestation and grain losses, whereas the abundance of neems decreased egg infestation only (Fig. 6). Egg infestation and grain loss did not depend on the abundance or diversity of trees at a landscape scale.

4. Discussion

Among natural enemies of crop pests, vertebrates including insectivorous bats and birds can provide valuable ecosystem services such as suppression of insect pests, particularly in agroforestry systems (Karp et al., 2013, 2016; Maas et al., 2013; Maine and Boyles, 2015). However, they can also disrupt natural pest regulation by intraguild competition with predatory arthropods in agroecosystems (Grass et al., 2017; Martin et al., 2013). In the present study, we investigated the association between vegetation openness, richness and abundance of insectivorous birds and bats, and their contribution to the natural regulation of the MHM in tree-crop agroforestry systems in Senegal.

The exclusion of natural enemies including arthropods and vertebrates from millet panicles showed that grain losses significantly decreased when panicles were accessible to arthropods only, and even more when they were accessible to both arthropods and birds. We did not observed significant damage on millet panicles due to seed predation by granivorous birds. The vertebrate predatory community was composed of eight bird species of Passeriformes mostly represented by weavers (Ploceus spp.) and sparrows (Passer griseus). Direct field observations showed that encountered bird species were both
granivorous and insectivorous, except *L. pulcher*, which was observed feeding exclusively on insects. Two bat species were seen foraging in the surrounding environment of millet sampling plots: a frugivorous species, the African straw-colored fruit bat (*Eidolon helvum*) and an insectivorous species, the Mauritian tomb bat (*Taphozous mauritianus*). The analysis of fecal samples of birds and bats indicated that village weavers primarily fed on late instar larvae, whereas bats fed on MHM moths (egg oviposition period) and probably also on late instar larvae. According to Goodman et al. (2007), the Mauritian tomb bat is a nocturnal hunter, with moths as preferred food, though during daylight hours they can occasionally prey upon butterflies and termites. The Mauritian tomb bat possibly contributes more than birds to the natural regulation of MHM populations, because unlike birds, insectivorous bats can feed on a large quantity of moths (Kingdon, 1974), thus limiting egg-laying on millet panicles. Birds can exert high predation on MHM larvae during the reproductive season to feed their chicks. Contrary to adults who feed mainly on seeds to which they add insects, chicks feed exclusively on insects including lepidopteran larvae (Bruggers et al., 1985; Petersen et al., 2007).

The abundance and diversity of natural habitats in agroforestry systems is a key factor for the conservation of natural enemies, particularly vertebrates in agricultural landscapes (Bianchi et al., 2006; Holland et al., 2016; Rusch et al., 2016). The positive impact of trees species richness on bats and insectivorous birds communities in agricultural landscape has been demonstrated (Harvey et al., 2006). In our study, the abundance of birds (all species together) was positively related to the diversity of trees around millet sampling plots. Populations of village weavers showed the same trend, whereas grey-headed sparrows were preferentially observed in open habitats. Village weavers are gregarious (Collias and Collias, 1970) and their colonies settle a few weeks after the start of the rainy season, just before the emergence of millet panicles (Bruggers et al., 1985). They build their nests on different tree species (mainly neems, palm trees, and balanites) located near residential areas and forage in neighbor millet fields to collect insects, particularly late instar larvae on panicles at the time of chick breeding. Unlike the
village weaver, the grey-headed sparrow is solitary and it can be encountered in a wide range of open
habitats, including open woodlands and human habitation. No effect of tree cover was detected at a
landscape scale (1000 and 1750-m radius buffer around sampling points), indicating short-distance
interaction of insectivorous birds with the surrounding landscape. Direct field observations showed that
palm trees served as preferential nesting site for the Mauritian tomb bat. The species is characterized by
very effective hunting aptitudes including excellent vision, nocturnal hunting, echolocation of prey, and
fast flight (Fenton, 1990; Dengis, 1996), which make it an excellent predator of noctuid moths. Their
hunting season coincides with the moment when MHM females leave their shelter sites at dusk to lay
eggs on millet heads. Our field observations revealed that this generalist predator was able to hunt
insects under the street lights in villages, in addition to its ability to hunt effectively in total darkness (A.
Sow, unpublished data).

When addressing the effect of tree cover on the incidence of MHM, we found that relative contribution
of trees to natural pest regulation was probably species-dependent. In the present study, we observed a
negative effect of the abundance of soapberry trees and neems on egg infestation. As these trees are
often used as nesting sites by village weaver colonies, MHM moths could have been deterred from
resting on such trees, as the analysis of fecal samples indicated that those birds primarily feed on late
instar larvae. An alternative hypothesis is that soapberry trees and neems offer refuges for other natural
enemies groups capable of affecting pest incidence through the direct consumption of eggs (e.g.,
predatory arthropods) or moths (e.g., bats). On the other hand, the presence of certain tree species can
negatively affect crop yield by increasing the population density of the pest (Schroth et al., 2000). This is
possibly the case for baobabs, as our results showed that MHM incidence (egg infestation on panicles
and grain losses) increased with baobab density at a local scale. Baobabs could provide resources for
insect pests, such as refuges and food.
5. Conclusion

The present study provides key information on hypotheses addressed about the link between abundance and diversity of trees at a local and landscape scale and the abundance and diversity of insectivorous birds and bats, and their net contribution to natural pest control and reduction of millet grain loss. At a local scale, the abundance of insectivorous village weaver birds increased with tree diversity whereas that of grey-headed sparrows decreased with the abundance of trees. Results also showed that trees indirectly contribute to natural regulation of the MHM likely by providing refuges to insectivorous vertebrates (e.g., palms/bats or neems/village weavers). However, other tree species can provide disservices as refuges for pests (e.g., baobabs/MHM moths). To capture the variability of such natural pest regulation and the effects of environmental perturbations (climate, land use, deforestation, pesticides, etc.), it would be relevant to carry out this study on a pluriannual scheme. More research is also needed to deeply examine the links between tree species (habitats), natural enemies (taxonomic and functional traits of communities), pest regulation (e.g., intraguild predation, temporal dynamics) and crop yield. This should assist the design of specific conservation measures such as habitat management to promote conservation biological control. Next steps include farmer-participatory approaches to design farmscapes fostering natural pest regulation.

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Competing interests

The authors declare no competing interests.

References


Agronomie Tropicale 33, 62–79.

Appendix A. Supplementary Data

Figure S1

Figure captions

Figure 1. Map of the 20 paired millet sampling plots within the study area (Bambey, Senegal). Example (on the right) of a digitized landscape around one sampling plot.

Figure 2. Detection rate (%) of the millet head miner DNA in bird (*Ploceus cucullatus*) and bat (*Taphozous mauritianus*) fecal samples collected at different phenological stages of the millet crops corresponding to the presence of different development stages of the millet head miner (moths and eggs, moths and early instar larvae, late instar larvae). Bars represent the 95% confidence intervals (CI95).

Figure 3. Relative importance of the abundance and diversity of trees (at a 100-m radius buffer around sampling plots) in explaining variation of (A) the abundance of birds, (B) the abundance of the grey-headed sparrow, and (C) the abundance of the village weaver. Variables were ranked according to their sum of Akaike weights, which are the probabilities that the given predictor would appear in the best fitting model. Black bars indicate the most important variables. Estimated parameter values and their standard error (SE) are presented on the right of each bar.

Figure 4. Relative importance of the abundance and diversity of birds (at a 100-m radius buffer around sampling plots) in explaining variation of (A) egg infestation and (B) grain losses due to the millet head miner (MHM). Variables were ranked according to their sum of Akaike weights (w + (j)), which are the probabilities that the given predictor would appear in the best fitting model. Estimated parameter values and their standard error (SE) are presented on the right of each bar.
Figure 5. Grain loss (%) on egg-infested panicles (i) covered with a mesh bag to exclude natural enemies, (ii) covered with a polyethylene nylon mesh (25 x 25 mm cell size) to exclude vertebrates but not predatory arthropods or parasitoids, or (iii) exposed to both parasitoids, predatory arthropods and vertebrates (open panicles). Means followed by different letters are significantly different (post hoc Tukey test). Mean grain (±SE) loss: mesh bag (26.3 ± 1.4%), cage (21.7 ± 1.3%) and open control (16.9 ± 1.1%).

Figure 6. Relative importance of the abundance and diversity of trees (at a 100-m radius buffer around sampling plots) in explaining variation of (A) egg infestation and (B) grain losses due to the millet head miner (MHM). Variables were ranked according to their sum of Akaike weights \((w + j)\), which are the probabilities that the given predictor would appear in the best fitting model. Black bars indicate the most important variables. Estimated parameter values and their standard error (SE) are presented on the right of each bar.
Table 1. Tree cover and diversity, bird abundance and diversity, and pest incidence of the 20 sampling plots in the study area. Tree cover is characterized at the local (100-m radius buffer around sampling plots) and landscape scale (1-km radius buffer around sampling plots).

<table>
<thead>
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<th>Variables</th>
<th>Range (min-max)</th>
<th>Mean ± CI95</th>
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<td></td>
<td></td>
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<tr>
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<td>Tree cover (%)</td>
<td>3.5-18.3</td>
<td>9.3 ± 1.8</td>
</tr>
<tr>
<td>Number of trees</td>
<td>21-94</td>
<td>52.3 ± 9.3</td>
</tr>
<tr>
<td>Number of <em>Azadirachta indica</em></td>
<td>0-53</td>
<td>15.8 ± 6.5</td>
</tr>
<tr>
<td>Number of <em>Balanites aegyptiaca</em></td>
<td>0-29</td>
<td>13.4 ± 3.2</td>
</tr>
<tr>
<td>Number of <em>Faidherbia albida</em></td>
<td>2-18</td>
<td>8.2 ± 1.9</td>
</tr>
<tr>
<td>Number of <em>Adansonia digitata</em></td>
<td>0-31</td>
<td>8.3 ± 3.7</td>
</tr>
<tr>
<td>Number of <em>Borassus aethiopium</em></td>
<td>0-7</td>
<td>1.3 ± 0.8</td>
</tr>
<tr>
<td>Tree SHDI</td>
<td>0.9-1.7</td>
<td>1.3 ± 0.1</td>
</tr>
<tr>
<td>At a landscape scale</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Tree cover (%)</td>
<td>2.8-8.7</td>
<td>5.7 ± 1.0</td>
</tr>
<tr>
<td><strong>Bird abundance and diversity</strong></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Number of birds per sampling plot</td>
<td>1-83</td>
<td>49 ± 18</td>
</tr>
<tr>
<td>Bird SHDI</td>
<td>0.0-1.3</td>
<td>0.9 ± 0.3</td>
</tr>
<tr>
<td><strong>Pest incidence</strong></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Egg infested panicles (%)</td>
<td>24.0-80.0</td>
<td>53.4 ± 6.8</td>
</tr>
<tr>
<td>Number of eggs per panicle</td>
<td>0.4-3.8</td>
<td>1.9 ± 0.4</td>
</tr>
<tr>
<td>Damaged panicles (%)</td>
<td>86.7-100</td>
<td>96.2 ± 1.9</td>
</tr>
<tr>
<td>Grain loss (%)</td>
<td>9.7-29.3</td>
<td>16.9 ± 2.1</td>
</tr>
</tbody>
</table>
Table 2. Inventory of major birds and bats observed in millet-based agroecosystems in the North of the Peanut basin in Senegal.

<table>
<thead>
<tr>
<th>Order Family</th>
<th>Common name</th>
<th>Latin name</th>
<th>Food diet</th>
<th>Frequency of occurrence (%)</th>
<th>Relative abundance (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Passeriformes</td>
<td>Corvidae</td>
<td>African piapiac</td>
<td>Ptilostomus afer (Swainson, 1837)*</td>
<td>I/G</td>
<td>0.4</td>
</tr>
<tr>
<td></td>
<td>Passeridae</td>
<td>Grey-headed sparrow</td>
<td>Passer griseus (Vieillot, 1817)*</td>
<td>I/G</td>
<td>32.5</td>
</tr>
<tr>
<td></td>
<td>Ploceidae</td>
<td>Village weaver</td>
<td>Ploceus cucullatus (Statius Müller, PL, 1776)**</td>
<td>I/G</td>
<td>34.2</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Little weaver</td>
<td>Ploceus luteolus (Lichtenstein, MHK, 1823)*</td>
<td>I/G</td>
<td>17.1</td>
</tr>
<tr>
<td></td>
<td>Pycnonotidae</td>
<td>Common bulbul</td>
<td>Pycnonotus barbatus (Desfontaines, 1789)*</td>
<td>I/G</td>
<td>2.5</td>
</tr>
<tr>
<td></td>
<td>Sturnidae</td>
<td>Long-tailed glossy starling</td>
<td>Lamprotornis caudatus (Statius Müller, 1776)*</td>
<td>I/G</td>
<td>0.4</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Chestnut-bellied starling</td>
<td>Lamprotornis pulcher (Müller, 1776)*</td>
<td>I</td>
<td>0.8</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Lesser blue-eared starling</td>
<td>Lamprotornis chloropterus (Swainson, 1838)*</td>
<td>I/G</td>
<td>0.8</td>
</tr>
<tr>
<td>Chiroptera</td>
<td></td>
<td>African straw-colored fruit bat</td>
<td>Eidolon helvum (Kerr, 1792)**</td>
<td>F</td>
<td>-</td>
</tr>
<tr>
<td>Emballonurida</td>
<td>Mauritain tomb bat</td>
<td>Taphozous mauritianus (E. Geoffroy, 1818)**</td>
<td>I</td>
<td>-</td>
<td>-</td>
</tr>
</tbody>
</table>

Millet phenological stage

- Panicle emergence to late flowering
- Grain filling
- Grain maturation

Presence of the MHM in fecal samples (%)

- Bat (Taphozous mauritianus)
- Bird (Ploceus cucullatus)

MHM development stage

- Moths and eggs
- Moths and early instar larvae
- Late instar larvae
A

- All insectivorous birds: $-0.13 \pm 0.11$
- Village weavers: $-0.26 \pm 0.22$
- Grey-headed sparrows: $0.01 \pm 0.00$
- SHDI birds: $-0.01 \pm 0.00$

B

- All insectivorous birds: $0.01 \pm 6.11$
- Grey-headed sparrows: $0.04 \pm 0.33$
- Village weavers: $-0.02 \pm 0.37$
- SHDI birds: $-0.12 \pm 0.34$
Grain loss (%)

- Mesh bag
- Cage
- Open

Box plots showing the grain loss for different storage conditions.
A

- Baobabs: 0.03 ± 0.02
- Soapberry trees: -0.05 ± 0.02
- Neems: -0.04 ± 0.02
- All trees: 0.01 ± 0.02
- Palmyra palms: 0.11 ± 0.07
- Apple-ring acacias: 0.001 ± 0.03
- SHDI trees: -0.72 ± 0.52

B

- Baobabs: 0.39 ± 0.37
- Soapberry trees: -0.27 ± 0.29
- Palmyra palms: 1.12 ± 1.08
- All trees: -0.08 ± 0.28
- Neems: 0.05 ± 0.37
- SHDI trees: -2.95 ± 7.94
- Apple-ring acacias: 0.26 ± 0.52

Relative variable importance
**Highlights**

- Vertebrates, including birds and bats, are effective predators of the millet head miner.
- They contribute to natural pest control and reduction of millet grain loss.
- The abundance of insectivorous birds is enhanced by tree diversity at a local scale.
- The contribution of trees for natural pest control is species-dependent. Trees can serve as refuges for pest (eg. baobab/moths) and predators (eg. palm tree/bats or neem/village weavers).
Declaration of interests

☒ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

☐ The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: