



**HAL**  
open science

## Birds and bats contribute to natural regulation of the millet head miner in tree-crop agroforestry systems

Ahmadou Sow, Djiby Seye, Emile Faye, Laure Benoit, Maxime Galan, Julien Haran, Thierry Brévault

► **To cite this version:**

Ahmadou Sow, Djiby Seye, Emile Faye, Laure Benoit, Maxime Galan, et al.. Birds and bats contribute to natural regulation of the millet head miner in tree-crop agroforestry systems. *Crop Protection*, 2020, 132, pp.105127. 10.1016/j.cropro.2020.105127 . hal-02558065

**HAL Id: hal-02558065**

**<https://hal.inrae.fr/hal-02558065>**

Submitted on 29 Apr 2020

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial - NoDerivatives 4.0 International License

1 Original article: Special issue Agroforestry

2 **Birds and bats contribute to natural regulation of the millet head miner in tree-**  
3 **crop agroforestry systems**

4 Ahmadou Sow<sup>a\*</sup>, Djiby Seye<sup>a</sup>, Emile Faye<sup>b,c,d</sup>, Laure Benoit<sup>e,f</sup>, Maxime Galan<sup>f,g</sup>, Julien Haran<sup>e,f</sup>, Thierry  
5 Brévault<sup>h,i</sup>

6 <sup>a</sup> UCAD, FST, Département de Biologie Animale, Dakar, Senegal

7 <sup>b</sup> CIRAD, UPR HortSys, Dakar, Senegal

8 <sup>c</sup> HortSys, Univ Montpellier, CIRAD, Montpellier, France

9 <sup>d</sup> ISRA, Centre pour le Développement de l'Horticulture, Dakar, Senegal

10 <sup>e</sup> CIRAD, UMR CBGP, 34398 Montpellier, France

11 <sup>f</sup> CBGP, Univ Montpellier, CIRAD, INRA, IRD, Montpellier SupAgro, Montpellier, France

12 <sup>g</sup> INRA, UMR CBGP, Montpellier, France

13 <sup>h</sup> CIRAD, UPR AIDA, Centre de recherche ISRA-IRD, Dakar, Senegal

14 <sup>i</sup> AIDA, Univ Montpellier, CIRAD, Montpellier, France

15

16 \*Corresponding author. E-mail address: s\_ahmadou@yahoo.fr (Ahmadou Sow)

17

18 **Abstract**

19 A better understanding of environmental factors driving natural pest regulation is a major challenge for  
20 designing sustainable cropping systems. The objective of the present study was to assess the association  
21 between vegetation openness in traditional tree-crop agroforestry systems in Senegal, richness and

22 abundance of vertebrates including insectivorous birds and bats, and their contribution to the natural  
23 regulation of crop pests. The millet head miner (MHM), *Heliocheilus albipunctella* (Lepidoptera,  
24 Noctuidae), a major constraint to increasing millet crop productivity in sub-Saharan Africa, was selected  
25 as a model system. Ten sites separated by at least one km from each other were selected in a 100 km<sup>2</sup>  
26 study area in the Peanut basin in Senegal. In each site, a pair of millet fields distant from at least 100 m  
27 each was selected according to a gradient of vegetation openness within a 100-m radius buffer with  
28 sampling plot (5 x 5 m) at the center. Nine insectivorous bird and bat species were recorded in millet  
29 fields over the 2017 cropping season and their predator status was confirmed by direct observation or  
30 DNA detection in fecal samples. Grain losses were reduced when panicles were accessible to birds and  
31 bats, confirming their net contribution to pest regulation. At a local scale, tree diversity and vegetation  
32 openness were important predictors of the abundance of insectivorous village weaver birds and grey-  
33 headed sparrows, respectively. Some tree species (soapberry trees and neems) indirectly contributed to  
34 natural regulation of the MHM likely by providing refuges to insectivorous vertebrates whereas other  
35 tree species (baobabs) provided disservices as possible refuges for the MHM moths. Further research is  
36 needed to better understand relationships between tree cover, food web interactions and natural pest  
37 suppression, so that specific conservation measures such as habitat management can be designed to  
38 improve pest control.

39

#### 40 **1. Introduction**

41 During the last decades, worldwide losses of biodiversity have occurred at an unprecedented rate and  
42 agricultural intensification has been a major driver (Cardinale et al., 2012; Philpott et al., 2008; Tilman et  
43 al., 2002; Wilby and Thomas, 2002). The dramatic land use changes include the simplification of  
44 agricultural landscapes by deforestation and monoculture and the extensive use of agricultural inputs  
45 such as broad-spectrum insecticides to control crop pests, with detrimental impact on ecosystem health

46 (Tscharrntke et al., 2005; Barzman et al., 2015). Along with the loss of functional biodiversity, particularly  
47 biological control agents, such cultivated ecosystems have become more and more susceptible to pest  
48 outbreaks. This is particularly pregnant in southern contexts where arthropod pests pose a permanent  
49 threat to food security (Garrity et al., 2010). New models for crop protection are needed to reduce the  
50 dependence of agricultural systems on pesticides that stimulate ecological processes of natural  
51 regulation of crop pests (Brévault and Clouvel, 2019). One approach is to use the benefits of agroforestry  
52 systems to enhance or restore functional biodiversity (here, natural enemies of crop pests) to boost  
53 ecosystem services such as biological control of crop pests. A meta-analysis conducted by Pumarino et al.  
54 (2015) showed that agroforestry practices generally result in higher abundance of natural enemies of  
55 pests, lower pest abundance and less plant damage in perennial crops (e.g., coffee, cocoa and plantain)  
56 but not in annual crops (e.g., maize, rice and beans). However, studies carried out in Kenya showed that  
57 abundance of stem borers was significantly lower in *Leucaena leucocephala* (Lam.) (Fabales,  
58 Mimosaceae) alley cropping than in a maize monocrop (Ogol et al., 1999).

59 Among natural enemies of crop pests, vertebrates including insectivorous bats and birds can provide  
60 valuable ecosystem services such as suppression of insect pests (Maas et al., 2013; Maine and Boyles,  
61 2015; Pejchar et al., 2018). By protecting crops through feeding on herbivorous pests, they can provide  
62 substantial increases in crop yields (Bael et al., 2008; Karp et al., 2013) and a true alternative to chemical  
63 insecticides (Bianchi et al., 2006; Tscharrntke et al., 2012). The abundance and diversity of semi-natural  
64 habitats within agricultural landscapes, particularly trees, has the potential to increase the abundance  
65 and diversity of birds and bats (Harvey et al., 2006; Redlich et al., 2018) and thus natural regulation of  
66 crop pests (e.g., nesting and breeding site, alternative preys). However, birds and bats can also  
67 negatively impact natural regulation of crop pests by consuming natural enemies of crop pests, i.e., intra-  
68 guild predation (Grass et al. 2017; Tschumi et al., 2018). On the other hand, trees can also benefit pests  
69 by providing food and refuges (Schroth et al., 2000). More research is thus needed on ecological issues

70 to better understand how tree-crop agroforestry systems can contribute to enhance ecosystem services  
71 such as biological control of crop pests. Although a substantial body of research has been performed on  
72 the role of bats and birds in perennial cropping systems such as coffee plantations (Bael et al., 2008), few  
73 studies have focused on annual cropping systems as an unstable mosaic of habitats (Kennedy and Storer,  
74 2000). More research is thus needed to deepen our knowledge of the life system of insectivorous birds  
75 and bats including seasonal population dynamics (e.g., breeding cycles, migration), multitrophic  
76 interactions within food webs (e.g. intraguild predation; see Grass et al., 2017), and temporal suitability  
77 of habitats in agricultural landscapes (Maas et al., 2016).

78 Agricultural landscapes in the “Peanut basin” in Senegal (West Africa) are generally structured by  
79 traditional agroforestry parkland systems (AFS), in which trees (mainly apple-ring acacias, *Faidherbia*  
80 *albida*) (Fabales, Fabaceae) are distributed in cultivated fields (Seyler, 1993). Even though the impact of  
81 those trees on soil fertility has been widely demonstrated, their effect on crop pests and biological  
82 control has been poorly studied (Hadgu et al., 2009; Umar et al., 2013). At the landscape scale, they offer  
83 a greater diversity of ecological niches in time and space than a simple mosaic of annual crops. The pearl  
84 millet, *Pennisetum glaucum* (L.) R. Br. (Cyperales, Poaceae), is the main annual crop in those agricultural  
85 systems. However, a key insect pest, the millet head miner (MHM) (*Heliocheilus albipunctella* de Joannis)  
86 (Lepidoptera, Noctuidae) is a permanent threat for increasing crop yields (Gahukar, 1984; Sow et al.,  
87 2018; Vercambre, 1978). Moths emerge from the soil one to two months after the beginning of the rainy  
88 season. After mating, females lay eggs in the millet panicle on the silk of the flower involucre or under  
89 stamens, but also on the floral peduncles or on the spine (Nwanze and Harris, 1992). The young larvae  
90 perforate the glumes and consume flowers, while the older ones cut off the floral peduncles (Gahukar,  
91 1984; Vercambre, 1978), thus preventing grain formation. In the absence of any insecticide application  
92 by farmers, millet production relies on pest regulation by natural enemies (Bhatnagar, 1987; Nwanze and  
93 Sivakumar, 1990; Sow et al., 2019, 2017). A two-year monitoring of 90 millet fields and their surrounding

94 landscape (2000 m-buffer around millet fields) showed that abundance of trees at the landscape scale  
95 significantly increased natural regulation of the MHM (Soti et al., 2019). Using high throughput molecular  
96 tools, the food web structures and ecological processes underlying natural regulation of crop pests by  
97 arthropods were documented and a large diversity of parasitoids and predators among arthropods were  
98 identified (Sow et al., 2018).

99 In the present study, we investigated the contribution of vertebrates, particularly insectivorous birds and  
100 bats, to the natural regulation of the MHM in tree-crop agroforestry systems in Senegal. We  
101 hypothesized that (i) abundance and diversity of trees at a local and landscape scale enhance the  
102 abundance and diversity of insectivorous birds, and that (ii) abundance and diversity of insectivorous  
103 birds at a local scale reduce pest incidence. We used satellite and drone imagery, remote sensing and  
104 geographical information system (GIS), to map and quantify the abundance and diversity of trees species  
105 at local and landscape scale around a set of 20 millet fields. Pest incidence and grain loss were evaluated  
106 by direct field observations on millet panicles while natural regulation was assessed using exclusion  
107 cages. In addition, the abundance and diversity of birds and bats was monitored in and around millet  
108 fields and fecal samples were collected to check for the presence of the MHM based on DNA analyses.  
109 Results are discussed in the light of relationships between tree diversity and abundance, role of  
110 vertebrates in food webs (ecosystem services vs. disservices) and biological control of crop pests for the  
111 deployment of ecologically-intensive cropping systems.

112

## 113 **2. Materials and Methods**

### 114 **2.1. Study area**

115 The field survey was conducted in 2017 in the northern area of the "Peanut basin", near Bambey  
116 (14°43'0.79"N ; 16°30'5.56"O) in Senegal. This area is under the influence of a Sudano-Sahelian climate

117 with a short rainy season (400-600 mm) from July to October (monsoon). The landscape is composed of  
118 tree-crop agroforestry systems mainly including apple-ring acacias (*Faidherbia albida*), soapberry trees  
119 (*Balanites aegyptica*; Sapindales, Zygophyllaceae), baobabs (*Adansonia digitata*; Malvales,  
120 Bombacaceae), tamarinds (*Tamarindus indica*; Fabales, Caesalpiniaceae) and red acacias (*Acacia seyal*;  
121 Fabales, Fabaceae). Crops mainly include staple crops such as pearl millet, peanuts, and cowpea.

## 122 **2.2. Study design**

123 A total of 10 sites distant from at least one km each other (Fig. 1) were selected across the landscape  
124 (100 km<sup>2</sup>) according to the land-use proportion of tree cover (2.8-8.7%) at a landscape scale (one-km  
125 radius buffer around millet fields), following previous treatment of very high resolution satellite imagery  
126 (Soti et al., 2018). In each site, two millet fields distant from at least 100 m each were selected according  
127 to vegetation openness (open vs. closed) within a 100-m radius buffer with sampling plot (5 x 5 m) at the  
128 center. The minimum-maximum distance between paired sampling points was 100 m (Fig. 1). Sampling  
129 points captured a large variation in the abundance (number of trees and land-use proportion of tree  
130 cover) and diversity (SHDI) of trees (Table 1). Tree cover around sampling plots (100-m radius buffer) was  
131 mainly composed of *A. indica*, *B. aegyptiaca*, *F. albida*, and *A. digitata* (Table 1). Situations of closed  
132 vegetation were mostly found around “compound” (close-to-home) millet fields with high abundance of  
133 neems, whereas “bush” millet fields presented more open vegetation (Fig. S1).

134

## 135 **2.3 Landcover mapping**

136 Drone photogrammetry and object-based image analysis procedures were used to characterize each site  
137 following the method described in Sarron *et al.* (2018). Briefly, each site was overflown using a UAV (DJI  
138 Mavic Pro, DJI Inc, China) in order to compute a high resolution orthomosaic (1.30 cm.pixel<sup>-1</sup>) and a  
139 Canopy Height Model (CHM) using a photogrammetry procedure. The two layers were then analysed in

140 eCognition Developer 9 software (Trimble Geospatial, Munich, Germany) through an object-based  
141 classification workflow to obtain land cover maps. These landcover maps contained 5 object classes:  
142 millet crops, bare soil, trees, grasslands, and building (Fig. 1). Tree species were manually identified  
143 within the 100-m radius buffer (photointerpretation) and corrections were made when needed. Land  
144 cover maps were expressed as geospatial vectors with each object, including the delineated trees,  
145 corresponding to a geolocalized polygon and described in an adjunct database (.shapefile format) to be  
146 imported in GIS and processed in Fragstats software (McGarigal and Marks, 1995) for spatial analysis.

#### 147 **2.4. Data collection**

148 The incidence of the MHM was monitored in the 20 sampling plots by counting eggs on 25 randomly  
149 selected panicles at panicle emergence. Counting is done with a round trip on the surface of the panicle  
150 with a fine forceps. Females of the MHM usually lay their eggs on emerging panicles up to the male  
151 flowering stage, between involucre bristles and glumes.

152 At the male flowering stage of millet, i.e. after completion of egg development, sampling plots were  
153 divided into two subplots of each 2.5 x 5 m. One subplot designated bird enclosure, while the other  
154 served as control. Bird enclosures (2.5 m height) were built with wooden frames covered with a  
155 polyethylene nylon mesh (25 x 25 mm cell size) that allowed predatory arthropods to enter but that  
156 excluded birds. In each cage, eight panicles were protected with a mesh bag to exclude natural enemies,  
157 while 16 other panicles were left open (without mesh bag) to exclude only birds. The same design was  
158 reproduced outside the cage in the second subplot with eight panicles protected with a mesh bag to  
159 exclude natural enemies, and 16 other panicles left open as a control. As a result, a total of 960 panicles  
160 were monitored in the study design. Damages caused by the MHM (number and length of mines) were  
161 recorded from each panicle just before harvest, i.e., 21 days later. Conversion of damages to grain losses  
162 was done using allometric relations (Thiaw et al., 2017).



163 Field monitoring of birds and bats was performed on the five sites located in the southern part of the  
164 study area, over a period of 15 days (3 days in each site) from millet grain filling to maturity.  
165 Observations of birds were carried out around sampling plots in the morning (7.00 to 9.00 am UTC) and  
166 in the late afternoon (4.30 to 6.30 pm UTC). The species and food diet (millet grain vs. MHM larvae) of  
167 birds observed on millet panicles were recorded. Bird abundance in millet fields was calculated by using  
168 the frequency of occurrence (C) proposed by Linsdale (1928). Bird species were identified using the  
169 "Guide of West African birds" (Borrow and Demey, 2015). Bats were counted at regular time intervals (5  
170 min) at dusk (from 6.30 to 7.40 pm UTC).

## 171 **2.5. Analysis of fecal samples**

172 To check the predation of the MHM by vertebrates, fecal samples of the village weaver (*Ploceus*  
173 *cucullatus*; Passeriformes, Ploceidae) and the Mauritian tomb bat (*Taphozous mauritanum*; Chiroptera,  
174 Emballonuridae) were collected during 24 hours by using a plastic tarpaulin placed under nesting trees,  
175 neem (*Azadirachta indica*; Sapindales, Meliaceae) and palm trees (*Borassus aethiopum*; Arecales,  
176 Arecaceae), respectively. Samples were collected at three millet phenological stages corresponding to  
177 successive development of the immature stages of the MHM: (1) panicle emergence to female flowering  
178 for eggs, (2) panicles with grains at the milk stage for the 1<sup>st</sup> and 2<sup>nd</sup> instar larvae, and (3) panicles with  
179 grains at the dough stage for 3<sup>rd</sup> and 4<sup>th</sup> instar larvae, from late August to late September. A total of 80  
180 and 92 fecal samples for birds and bats, respectively, were individually collected using clean cotton buds  
181 and placed in 2-ml microtubes with silica-gel granules to absorb moisture and prevent the development  
182 of molds and DNA degradation. Detection of the MHM in fecal samples was performed by amplification  
183 of the 133 bp fragment of cytochrome c oxidase I (COI mini-barcode, Gillet et al., 2015) and sequencing  
184 on a MiSeq Illumina platform (Appendix A).

## 185 **2.6. Statistical analyses**

186 The R program (version 3.6.1) was used for statistical analyses (R Core team, 2014). We used generalized  
187 linear models (*glm* function from the *MASS* package) to test the effect of bird abundance and diversity  
188 on pest incidence (including the number of eggs per panicle and grain loss). Models were fitted using the  
189 appropriate distribution type and link function: binomial for proportion of infested panicles or grain loss  
190 (link = logit), and Poisson for abundance (count data, link = log). We then tested with the same type of  
191 models the effect of tree cover at a local and landscape scale on the abundance and diversity of birds,  
192 and on pest incidence. Multi-model inference approach was performed to examine the relative  
193 importance of each explanatory variable (Burnham and Anderson, 2002). The *MuMIn* package was used  
194 to perform model selection and model averaging. Akaike weights were calculated according to the  
195 Akaike information criterion (AICc) which is a bias-corrected version of the AIC recommended in analyses  
196 where the sample size is small relative to the number of fitted parameters (Burnham and Anderson,  
197 2002; Rusch et al., 2011; Soti et al., 2019). The indirect effect of natural enemy exclusion (three  
198 treatments: mesh bag, cage, and open as control) on grain loss was analysed using a linear mixed model  
199 (*lmer* function in the *lme4* package) with site as a random effect to account for the variation due to this  
200 factor. Multiple post-hoc comparisons were carried out with Tukey contrasts using the *glht* function from  
201 the *multcomp* package.

## 202 **3. Results**

### 203 **3.1. Abundance and diversity of insectivorous birds and bats**

204 Sampling points captured a large variation in the abundance (number) and diversity (SHDI) of observed  
205 birds (Table 1). A total of eight insectivorous birds belonging to five families (Corvidae, Passeridae,  
206 Ploceidae, Pycnonotidae and Sturnidae) were observed in millet fields (Table 2). The grey-headed  
207 sparrow (*Passer griseus*, Passeriformes, Passeridae), the village weaver (*P. cuculatus*), and the little  
208 weaver (*Ploceus luteolus*; Passeriformes, Ploceidae) were the most frequently observed species (Table 2).

209 The frequency of occurrence of the five other bird species observed was very low (<3%). Direct field  
210 observations showed that the encountered bird species were omnivorous (both granivorous and  
211 insectivorous), except *Lamprotornis pulcher* (Muller) (Passeriformes, Sturnidae) which was observed  
212 feeding exclusively on insects. We identified two bat species foraging around millet sampling plots: a  
213 frugivorous species, the African straw-colored fruit bat (*Eidolon helvum*; Chiroptera, Pteropodidae) and  
214 an insectivorous species, the Mauritian tomb bat (*T. mauritanus*) (Table 2).

215 Using a metabarcoding approach, we successfully detected DNA sequences of the MHM in fecal samples  
216 of the village weaver bird and the Mauritian tomb bat. The frequency of detection of the MHM DNA in  
217 fecal samples from village weavers indicated that these birds primarily feed on late instar larvae (Fig. 2).  
218 Detection of the MHM DNA in bats indicated that they feed on MHM moths (egg oviposition period) and  
219 probably also on late instar larvae.

### 220 **3.2. Effect of tree cover on insectivorous birds**

221 The diversity of trees was the most important predictor explaining the abundance of birds observed in  
222 and around millet sampling plots (Fig. 3A). The abundance of birds increased with the diversity of trees in  
223 a 100-m radius around sampling plots. The abundance of the village weavers followed the same trend,  
224 whereas the abundance of the grey-headed sparrow was negatively correlated to the abundance of trees  
225 (Fig. 3B, 3C). No effect of tree cover on the abundance or diversity of birds was detected at a landscape  
226 scale (1000 and 1750-m radius buffer around sampling points). Direct field observations showed that  
227 palm trees served as preferential nesting site for the Mauritian tomb bat.

### 228 **3.3. Effect of insectivorous vertebrates on pest incidence**

229 The MHM incidence in millet fields was generally high (53.4% egg-infested panicles, 96% damaged  
230 panicles), but variable among fields (24-80% egg-infested panicles, 87-100% damaged panicles). Grain  
231 losses due to the MHM varied from 9.7 to 29.3% among sampling plots (Table 1). The abundance and

232 diversity of birds did not contribute significantly in explaining pest incidence levels (Fig. 4). Grain loss was  
233 significantly reduced when panicles were accessible to arthropods only (-17.4%), and even more (-35.7%)  
234 when they were accessible to both arthropods and birds (Fig. 5).

### 235 **3.4. Effect of tree cover on MHM incidence**

236 The abundance of baobabs at a local scale, i.e., 100-m radius buffer around sampling plots, increased  
237 MHM egg infestation on panicles and grain losses (Fig. 6). Conversely, the abundance of soapberry trees  
238 in the same area decreased egg infestation and grain losses, whereas the abundance of neems  
239 decreased egg infestation only (Fig. 6). Egg infestation and grain loss did not depend on the abundance  
240 or diversity of trees at a landscape scale.

## 241 **4. Discussion**

242 Among natural enemies of crop pests, vertebrates including insectivorous bats and birds can provide  
243 valuable ecosystem services such as suppression of insect pests, particularly in agroforestry systems  
244 (Karp et al., 2013, 2016; Maas et al., 2013; Maine and Boyles, 2015). However, they can also disrupt  
245 natural pest regulation by intraguild competition with predatory arthropods in agroecosystems (Grass et  
246 al., 2017; Martin et al., 2013). In the present study, we investigated the association between vegetation  
247 openness, richness and abundance of insectivorous birds and bats, and their contribution to the natural  
248 regulation of the MHM in tree-crop agroforestry systems in Senegal.

249 The exclusion of natural enemies including arthropods and vertebrates from millet panicles showed that  
250 grain losses significantly decreased when panicles were accessible to arthropods only, and even more  
251 when they were accessible to both arthropods and birds. We did not observed significant damage on  
252 millet panicles due to seed predation by granivorous birds. The vertebrate predatory community was  
253 composed of eight bird species of Passeriformes mostly represented by weavers (*Ploceus* spp.) and  
254 sparrows (*Passer griseus*). Direct field observations showed that encountered bird species were both

255 granivorous and insectivorous, except *L. pulcher*, which was observed feeding exclusively on insects. Two  
256 bat species were seen foraging in the surrounding environment of millet sampling plots: a frugivorous  
257 species, the African straw-colored fruit bat (*Eidolon helvum*) and an insectivorous species, the Mauritian  
258 tomb bat (*Taphozous mauritanus*). The analysis of fecal samples of birds and bats indicated that village  
259 weavers primarily fed on late instar larvae, whereas bats fed on MHM moths (egg oviposition period)  
260 and probably also on late instar larvae. According to Goodman et al. (2007), the Mauritian tomb bat is a  
261 nocturnal hunter, with moths as preferred food, though during daylight hours they can occasionally prey  
262 upon butterflies and termites. The Mauritian tomb bat possibly contributes more than birds to the  
263 natural regulation of MHM populations, because unlike birds, insectivorous bats can feed on a large  
264 quantity of moths (Kingdon, 1974), thus limiting egg-laying on millet panicles. Birds can exert high  
265 predation on MHM larvae during the reproductive season to feed their chicks. Contrary to adults who  
266 feed mainly on seeds to which they add insects, chicks feed exclusively on insects including lepidopteran  
267 larvae (Bruggers et al., 1985; Petersen et al., 2007).

268 The abundance and diversity of natural habitats in agroforestry systems is a key factor for the  
269 conservation of natural enemies, particularly vertebrates in agricultural landscapes (Bianchi et al., 2006;  
270 Holland et al., 2016; Rusch et al., 2016). The positive impact of trees species richness on bats and  
271 insectivorous birds communities in agricultural landscape has been demonstrated (Harvey et al., 2006).  
272 In our study, the abundance of birds (all species together) was positively related to the diversity of trees  
273 around millet sampling plots. Populations of village weavers showed the same trend, whereas grey-  
274 headed sparrows were preferentially observed in open habitats. Village weavers are gregarious (Collias  
275 and Collias, 1970) and their colonies settle a few weeks after the start of the rainy season, just before the  
276 emergence of millet panicles (Bruggers et al., 1985). They build their nests on different tree species  
277 (mainly neems, palm trees, and balanites) located near residential areas and forage in neighbor millet  
278 fields to collect insects, particularly late instar larvae on panicles at the time of chick breeding. Unlike the

279 village weaver, the grey-headed sparrow is solitary and it can be encountered in a wide range of open  
280 habitats, including open woodlands and human habitation. No effect of tree cover was detected at a  
281 landscape scale (1000 and 1750-m radius buffer around sampling points), indicating short-distance  
282 interaction of insectivorous birds with the surrounding landscape. Direct field observations showed that  
283 palm trees served as preferential nesting site for the Mauritian tomb bat. The species is characterized by  
284 very effective hunting aptitudes including excellent vision, nocturnal hunting, echolocation of prey, and  
285 fast flight (Fenton, 1990; Dengis, 1996), which make it an excellent predator of noctuid moths. Their  
286 hunting season coincides with the moment when MHM females leave their shelter sites at dusk to lay  
287 eggs on millet heads. Our field observations revealed that this generalist predator was able to hunt  
288 insects under the street lights in villages, in addition to its ability to hunt effectively in total darkness (A.  
289 Sow, unpublished data).

290 When addressing the effect of tree cover on the incidence of MHM, we found that relative contribution  
291 of trees to natural pest regulation was probably species-dependent. In the present study, we observed a  
292 negative effect of the abundance of soapberry trees and neems on egg infestation. As these trees are  
293 often used as nesting sites by village weaver colonies, MHM moths could have been deterred from  
294 resting on such trees, as the analysis of fecal samples indicated that those birds primarily feed on late  
295 instar larvae. An alternative hypothesis is that soapberry trees and neems offer refuges for other natural  
296 enemies groups capable of affecting pest incidence through the direct consumption of eggs (e.g.,  
297 predatory arthropods) or moths (e.g., bats). On the other hand, the presence of certain tree species can  
298 negatively affect crop yield by increasing the population density of the pest (Schroth et al., 2000). This is  
299 possibly the case for baobabs, as our results showed that MHM incidence (egg infestation on panicles  
300 and grain losses) increased with baobab density at a local scale. Baobabs could provide resources for  
301 insect pests, such as refuges and food.

## 302 **5. Conclusion**

303 The present study provides key information on hypotheses addressed about the link between abundance  
304 and diversity of trees at a local and landscape scale and the abundance and diversity of insectivorous  
305 birds and bats, and their net contribution to natural pest control and reduction of millet grain loss. At a  
306 local scale, the abundance of insectivorous village weaver birds increased with tree diversity whereas  
307 that of grey-headed sparrows decreased with the abundance of trees. Results also showed that trees  
308 indirectly contribute to natural regulation of the MHM likely by providing refuges to insectivorous  
309 vertebrates (e.g., palms/bats or neems/village weavers). However, other tree species can provide  
310 disservices as refuges for pests (e.g., baobabs/MHM moths). To capture the variability of such natural  
311 pest regulation and the effects of environmental perturbations (climate, land use, deforestation,  
312 pesticides, etc.), it would be relevant to carry out this study on a pluriannual scheme. More research is  
313 also needed to deeply examine the links between tree species (habitats), natural enemies (taxonomic  
314 and functional traits of communities), pest regulation (e.g., intraguild predation, temporal dynamics) and  
315 crop yield. This should assist the design of specific conservation measures such as habitat management  
316 to promote conservation biological control. Next steps include farmer-participatory approaches to  
317 design farmscapes fostering natural pest regulation.

318

## 319 **Acknowledgements**

320 This work was undertaken as part of, and funded by the CGIAR Research Program on Grain Legumes and  
321 Dryland Cereals (Program FP3-3.3, Managing functional biodiversity and crop pests) and supported by  
322 CGIAR Fund Donors. Funding support for this study was also provided by CIRAD (Action incitative CreSi  
323 2017). We are grateful to Amadou R.A. Etikokou for assistance to statistical analyses.

324 **Competing interests**

325 The authors declare no competing interests.

326 **References**

327 Bael, S.A.V., Philpott, S.M., Greenberg, R., Bichier, P., Barber, N.A., Mooney, K.A., Gruner, D.S., 2008.

328 Birds as predators in tropical agroforestry systems. *Ecology* 89, 928–934.

329 Barzman, M., Bàrberi, P., Birch, A.N.E., Boonekamp, P., Dachbrodt-Saaydeh, S., Graf, B., Hommel, B.,

330 Jensen, J.E., Kiss, J., Kudsk, P., Lamichhane, J.R., Messéan, A., Moonen, A.-C., Ratnadass, A., Ricci,

331 P., Sarah, J.-L., Sattin, M., 2015. Eight principles of integrated pest management. *Agronomy for*

332 *Sustainable Development* 35, 1199–1215.

333 Bhatnagar, V.S., 1987. Conservation and encouragement of natural enemies of insect pests in dryland

334 subsistence farming: Problems, progress and prospects in the Sahelian zone. *International Journal*

335 *of Tropical Insect Science* 8, 791–795.

336 Bianchi, F.J.J.A., Booij, C.J.H., Tscharrntke, T., 2006. Sustainable pest regulation in agricultural landscapes:

337 a review on landscape composition, biodiversity and natural pest control. *Proceedings of the Royal*

338 *Society B: Biological Sciences* 273, 1715–1727.

339 Borrow, N., Demey, R., 2015. *Guide des oiseaux de l’Afrique de l’Ouest*. Delachaux et Niestlé. Paris.

340 Brévault, T., Clouvel, P., 2019. Pest management: Reconciling farming practices and natural regulations.

341 *Crop Protection* 115, 1–6.

342 Bruggers, R.L., Jaeger, M.E., Jaeger, M.M., 1985. Tisserins gendarmes (*Ploceus cucullatus abyssinicus*) et

343 tisserins masqués (*Ploceus intermedius intermedius*) munis d’émetteurs radio et de rubans dans

344 une colonie de nidification du Sud de l’Ethiopie. *L’Oiseau et la Revue française d’ornithologie* 55,

345 81–92.



- 346 Burnham, K. P., Anderson, D. R., 2004. Multimodel inference: understanding AIC and BIC in model  
347 selection. *Sociological Methods and Research* 33, 261-304.
- 348 Cardinale, B.J., Duffy, J.E., Gonzalez, A., Hooper, D.U., Perrings, C., Venail, P., Narwani, A., Mace, G.M.,  
349 Tilman, D., Wardle, D.A., Kinzig, A.P., Daily, G.C., Loreau, M., Grace, J.B., Larigauderie, A.,  
350 Srivastava, D.S., Naeem, S., 2012. Biodiversity loss and its impact on humanity. *Nature* 486, 59–67.
- 351 Collias, N.E., Collias, E.C., 1967. A Quantitative Analysis of Breeding Behavior in the African Village  
352 Weaverbird. *The Auk* 84, 396–411.
- 353 Dengis, C. A., 1996. *Taphozous mauritanus*. *Mammalian Species* 522, 1-5.
- 354 Fenton, M. B., 1990. The foraging behaviour and ecology of animal-eating bats. *Canadian Journal of*  
355 *Zoology* 68, 411-422.
- 356 Gahukar, R.T., 1984. Insect pests of pearl millet in West Africa: a review. *International Journal of Pest*  
357 *Management* 30, 142–147.
- 358 Garrity, D.P., Akinnifesi, F.K., Ajayi, O.C., Weldesemayat, S.G., Mowo, J.G., Kalinganire, A., Larwanou, M.,  
359 Bayala, J., 2010. Evergreen Agriculture: a robust approach to sustainable food security in Africa.  
360 *Food Security* 2, 197–214.
- 361 Gillet, F., Tiouchichine, M.-L., Galan, M., Blanc, F., Némoy, M., Aulagnier, S., Michaux, J.R., 2015. A new  
362 method to identify the endangered Pyrenean desman (*Galemys pyrenaicus*) and to study its diet,  
363 using next generation sequencing from faeces. *Mammalian Biology - Zeitschrift für Säugetierkunde*  
364 80, 505–509.
- 365 Grass, I., Lehmann, K., Thies, C., Tschardtke, T., 2017. Insectivorous birds disrupt biological control of  
366 cereal aphids. *Ecology* 98, 1583–1590.
- 367 Hadgu, K.M., Kooistra, L., Rossing, W.A.H. et al. 2009. Assessing the effect of *Faidherbia albida* based  
368 land use systems on barley yield at field and regional scale in the highlands of Tigray, Northern  
369 Ethiopia. *Food Security* 1, 337–350.

- 370 Harvey, C.A., Medina, A., Sánchez, D.M., Vílchez, S., Hernández, B., Saenz, J.C., Maes, J.M., Casanoves, F.,  
371 Sinclair, F.L., 2006. Patterns of Animal Diversity in Different Forms of Tree Cover in Agricultural  
372 Landscapes. *Ecological Applications* 16, 1986–1999.
- 373 Holland, J.M., Bianchi, F.J., Entling, M.H., Moonen, A.-C., Smith, B.M., Jeanneret, P., 2016. Structure,  
374 function and management of semi-natural habitats for conservation biological control: a review of  
375 European studies. *Pest management science* 72, 1638–1651.
- 376 Karp, D.S., Mendenhall, C.D., Sandí, R.F., Chaumont, N., Ehrlich, P.R., Hadly, E.A., Daily, G.C., 2013. Forest  
377 bolsters bird abundance, pest control and coffee yield. *Ecology letters* 16, 1339–1347.
- 378 Kennedy, G. G., Storer, N. P., 2000. Life systems of polyphagous arthropod pests in temporally unstable  
379 cropping systems. *Annual Review of Entomology* 45, 467-493.
- 380 Kingdon, J., 1974. *East African Mammals: An Atlas of Evolution in Africa*, Academic Press. ed. Jonathan  
381 Kingdon, London.
- 382 Linsdale, J.M., 1928. A Method of Showing Relative Frequency of Occurrence of Birds. *The Condor* 30,  
383 180–184.
- 384 Maas, B., Clough, Y., Tscharntke, T., 2013. Bats and birds increase crop yield in tropical agroforestry  
385 landscapes. *Ecology letters* 16, 1480–1487.
- 386 Maas, B., Karp, D.S., Bumrungsri, S., Darras, K., Gonthier, D., Huang, J.C.-C., Lindell, C.A., Maine, J.J.,  
387 Mestre, L., Michel, N.L., Morrison, E.B., Perfecto, I., Philpott, S.M., Şekerciöğlü, Ç.H., Silva, R.M.,  
388 Taylor, P.J., Tscharntke, T., Bael, S.A.V., Whelan, C.J., Williams-Guillén, K., 2016. Bird and bat  
389 predation services in tropical forests and agroforestry landscapes. *Biological Reviews* 91, 1081–  
390 1101.
- 391 McGarigal, K., Marks, B.J., 1995. FRAGSTATS: spatial pattern analysis program for quantifying landscape  
392 structure. Gen. Tech. Rep. PNW-GTR-351. U.S. Department of Agriculture, Forest Service, Pacific  
393 Northwest Research Station, Portland, Oregon, USA, 122 pp.

- 394 Maine, J.J., Boyles, J.G., 2015. Bats initiate vital agroecological interactions in corn. Proceedings of the  
395 National Academy of Sciences USA 112, 12438–12443.
- 396 Martin, E.A., Reineking, B., Seo, B., Steffan-Dewenter, I., 2013. Natural enemy interactions constrain pest  
397 control in complex agricultural landscapes. Proceedings of the National Academy of Sciences USA  
398 110, 5534–5539.
- 399 Nwanze, K.F., Harris, K.M., 1992. Insect pests of pearl millet in West Africa. Review of Agricultural  
400 Entomology 80, 1132–1155.
- 401 Nwanze, K.F., Sivakumar, M.V.K., 1990. Insect pests of pearl millet in Sahelian West Africa—II. *Raghuva*  
402 *albipunctella* De Joannis (Noctuidae, Lepidoptera): Distribution, population dynamics and  
403 assessment of crop damage. Tropical Pest Management 36, 59–65.
- 404 Ogol, C.K.P.O., Spence, J.R., Keddie, A., 1999. Maize stem borer colonization, establishment and crop  
405 damage levels in a maize-leucaena agroforestry system in Kenya. Agriculture, Ecosystems and  
406 Environment 76, 1–15.
- 407 Pejchar, L., Gallo, T., Hooten, M.B., Daily, G.C., 2018. Predicting effects of large-scale reforestation on  
408 native and exotic birds. Diversity and Distributions 24, 811–819.
- 409 Petersen, B.S., Christensen, K.D., Jensen, F.P., 2007. Bird population densities along two precipitation  
410 gradients in Senegal and Niger. Malimbus 29, 101–121.
- 411 Philpott, S.M., Arendt, W.J., Armbrecht, I., Bichier, P., Diestch, T.V., Gordon, C., Greenberg, R., Perfecto,  
412 I., REYNOSO-SANTOS, R., SOTO-PINTO, L., 2008. Biodiversity loss in Latin American coffee  
413 landscapes: review of the evidence on ants, birds, and trees. Conservation Biology 22, 1093–1105.
- 414 Pumariño, L., Sileshi, G.W., Gripenberg, S., Kaartinen, R., Barrios, E., Muchane, M.N., Midega, C., Jonsson,  
415 M., 2015. Effects of agroforestry on pest, disease and weed control: A meta-analysis. Basic and  
416 Applied Ecology 16, 573–582.

- 417 R Core Team, 2014. R: A language and environment for statistical computing. R Foundation for Statistical  
418 Computing, Vienna, Austria. URL <http://www.R-project.org/>
- 419 Redlich, S., Martin, E.A., Wende, B., Steffan-Dewenter, I., 2018. Landscape heterogeneity rather than  
420 crop diversity mediates bird diversity in agricultural landscapes. *PloS one* 13, e0200438.
- 421 Rusch, A., Valantin-Morison, M., Sarthou, J. P., & Roger-Estrade, J., 2011. Multi-scale effects of landscape  
422 complexity and crop management on pollen beetle parasitism rate. *Landscape Ecology* 26, 473-  
423 486.
- 424 Rusch, A., Chaplin-Kramer, R., Gardiner, M.M., Hawro, V., Holland, J., Landis, D., Thies, C., Tschardt, T.,  
425 Weisser, W.W., Winqvist, C., 2016. Agricultural landscape simplification reduces natural pest  
426 control: A quantitative synthesis. *Agriculture, Ecosystems and Environment* 221, 198–204.
- 427 Sarron, J., Malézieux, ., Sané, C.A.B., Faye, É., 2018. Mango Yield Mapping at the Orchard Scale Based on  
428 Tree Structure and Land Cover Assessed by UAV. *Remote Sensing* 10, 1900.
- 429 Schroth, G., Krauss, U., Gasparotto, L., Aguilar, J.A.D., Vohland, K., 2000. Pests and diseases in  
430 agroforestry systems of the humid tropics. *Agroforestry Systems* 50, 199–241.
- 431 Seyler, J.R., 1993. A systems analysis of the status and potential of *Acacia Albida* (Del.) in the North  
432 Central Peanut Basin of Senegal. Michigan State University. Department of Forestry. 30 p.
- 433 Soti, V., Lelong, C., Goebel, F.-R., Brévault, T., 2018. Designing a field sampling plan for landscape-pest  
434 ecological studies using VHR optical imagery. *International Journal of Applied Earth Observation*  
435 and *Geoinformation* 72, 26–33.
- 436 Soti, V., Thiaw, I., Debaly, Z.M., Sow, A., Diaw, M., Fofana, S., Diakhate, M., Thiaw, C., Brévault, T., 2019.  
437 Effect of landscape diversity and crop management on the control of the millet head miner,  
438 *Heliocheilus albipunctella* (Lepidoptera: Noctuidae) by natural enemies. *Biological Control* 129,  
439 115–122.

- 440 Sow, A., Brévault, T., Benoit, L., Chapuis, M.-P., Galan, M., D'acier, A.C., Delvare, G., Sembène, M., Haran,  
441 J., 2019. Deciphering host-parasitoid interactions and parasitism rates of crop pests using DNA  
442 metabarcoding. *Scientific Reports* 9, 3646.
- 443 Sow, A., Brévault, T., Delvare, G., Haran, J., Benoit, L., Coeur d'Acier, A., Galan, M., Thiaw, C., Soti, V.,  
444 Sembène, M., 2018. DNA sequencing to help identify crop pests and their natural enemies in agro-  
445 ecosystems: the case of the millet head miner *Heliocheilus albipunctella* (Lepidoptera: Noctuidae)  
446 in sub-Saharan Africa. *Biological Control* 121, 199–207.
- 447 Sow, A., Brévault, T., Delvare, G., Haran, J., Mbacké, S., 2017. Régulation naturelle des populations de la  
448 mineuse de la chandelle de mil, *Heliocheilus albipunctella* (Lepidoptera, Noctuidae), dans le bassin  
449 arachidier au Sénégal. *AFPP, CIRAA* 1–9.
- 450 Thiaw, C., Brévault, T., Diallo, N. F., Sow, A., Ngom, D., Soti et al. 2017. Incidence et régulation naturelle  
451 de la chenille mineuse de l'épi de mil, *Heliocheilus albipunctella* de Joannis (Lepidoptera,  
452 Noctuidae) à Bambey dans le bassin arachidier au Sénégal. *Agronomie Africaine* 29, 83-95.
- 453 Tilman, D., Cassman, K.G., Matson, P.A., Naylor, R., Polasky, S., 2002. Agricultural sustainability and  
454 intensive production practices. *Nature* 418, 671–677.
- 455 Tschumi, M., Ekroos, J., Hjort, C., Smith, H. G., & Birkhofer, K., 2018. Rodents, not birds, dominate  
456 predation-related ecosystem services and disservices in vertebrate communities of agricultural  
457 landscapes. *Oecologia* 188, 863-873.
- 458 Tschardtke, T., Klein, A.M., Kruess, A., Steffan-Dewenter, I., Thies, C., 2005. Landscape perspectives on  
459 agricultural intensification and biodiversity–ecosystem service management. *Ecology letters* 8,  
460 857–874.
- 461 Umar, B. B., Aune, J. B., & Lungu, O. I., 2013. Effects of *Faidherbia albida* on the fertility of soil in  
462 smallholder conservation agriculture systems in eastern and southern Zambia. *African Journal of*  
463 *Agricultural Research* 8, 173-183.

464 Vercambre, B., 1978. *Raghuva* spp. et *Masalia* sp., chenilles des chandelles du mil en zone sahélienne.

465 Agronomie Tropicale 33, 62–79.

466 Wilby, A., Thomas, M.B., 2002. Natural enemy diversity and pest control: patterns of pest emergence

467 with agricultural intensification. Ecology Letters 5, 353–360.

468

469

470

Journal Pre-proof

471 **Appendix A. Supplementary Data**

472 **Figure S1**

473 **Figure captions**

474 **Figure 1.** Map of the 20 paired millet sampling plots within the study area (Bambey, Senegal). Example  
475 (on the right) of a digitized landscape around one sampling plot.

476 **Figure 2.** Detection rate (%) of the millet head miner DNA in bird (*Ploceus cucullatus*) and bat (*Taphozous*  
477 *mauritanus*) fecal samples collected at different phenological stages of the millet crops corresponding to  
478 the presence of different development stages of the millet head miner (moths and eggs, moths and early  
479 instar larvae, late instar larvae). Bars represent the 95% confidence intervals (CI95).

480 **Figure 3.** Relative importance of the abundance and diversity of trees (at a 100-m radius buffer around  
481 sampling plots) in explaining variation of (A) the abundance of birds, (B) the abundance of the grey-  
482 headed sparrow, and (C) the abundance of the village weaver. Variables were ranked according to their  
483 sum of Akaike weights, which are the probabilities that the given predictor would appear in the best  
484 fitting model. Black bars indicate the most important variables. Estimated parameter values and their  
485 their standard error (SE) are presented on the right of each bar.

486 **Figure 4.** Relative importance of the abundance and diversity of birds (at a 100-m radius buffer around  
487 sampling plots) in explaining variation of (A) egg infestation and (B) grain losses due to the millet head  
488 miner (MHM). Variables were ranked according to their sum of Akaike weights ( $w + (j)$ ), which are the  
489 probabilities that the given predictor would appear in the best fitting model. Estimated parameter values  
490 and their standard error (SE) are presented on the right of each bar.

491 **Figure 5.** Grain loss (%) on egg-infested panicles (i) covered with a mesh bag to exclude natural enemies,  
492 (ii) covered with a polyethylene nylon mesh (25 x 25 mm cell size) to exclude vertebrates but not  
493 predatory arthropods or parasitoids, or (iii) exposed to both parasitoids, predatory arthropods and  
494 vertebrates (open panicles). Means followed by different letters are significantly different (post hoc  
495 Tukey test). Mean grain ( $\pm$ SE) loss: mesh bag ( $26.3 \pm 1.4\%$ ), cage ( $21.7 \pm 1.3\%$ ) and open control ( $16.9 \pm$   
496  $1.1\%$ ).

497 **Figure 6.** Relative importance of the abundance and diversity of trees (at a 100-m radius buffer around  
498 sampling plots) in explaining variation of (A) egg infestation and (B) grain losses due to the millet head  
499 miner (MHM). Variables were ranked according to their sum of Akaike weights ( $w + (j)$ ), which are the  
500 probabilities that the given predictor would appear in the best fitting model. Black bars indicate the most  
501 important variables. Estimated parameter values and their standard error (SE) are presented on the right  
502 of each bar.



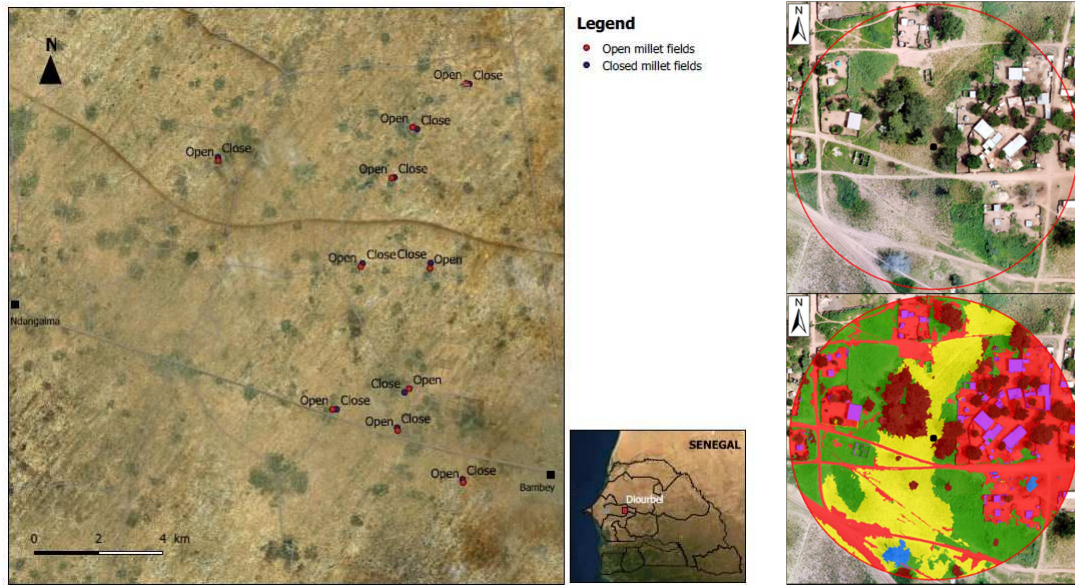
**Table 1.** Tree cover and diversity, bird abundance and diversity, and pest incidence of the 20 sampling plots in the study area. Tree cover is characterized at the local (100-m radius buffer around sampling plots) and landscape scale (1-km radius buffer around sampling plots).

Variables		Range (min-max)	Mean $\pm$ CI95
<b>Tree cover</b>			
At a local scale	Tree cover (%)	3.5-18.3	9.3 $\pm$ 1.8
	Number of trees	21-94	52.3 $\pm$ 9.3
	Number of <i>Azadirachta indica</i>	0-53	15.8 $\pm$ 6.5
	Number of <i>Balanites aegyptiaca</i>	0-29	13.4 $\pm$ 3.2
	Number of <i>Faidherbia albida</i>	2-18	8.2 $\pm$ 1.9
	Number of <i>Adansonia digitata</i>	0-31	8.3 $\pm$ 3.7
	Number of <i>Borassus aethiopium</i>	0-7	1.3 $\pm$ 0.8
	Tree SHDI	0.9-1.7	1.3 $\pm$ 0.1
At a landscape scale	Tree cover (%)	2.8-8.7	5.7 $\pm$ 1.0
<b>Bird abundance and diversity</b>			
	Number of birds per sampling plot	1-83	49 $\pm$ 18
	Bird SHDI	0.0-1.3	0.9 $\pm$ 0.3
<b>Pest incidence</b>			
	Egg infested panicles (%)	24.0-80.0	53.4 $\pm$ 6.8
	Number of eggs per panicle	0.4-3.8	1.9 $\pm$ 0.4
	Damaged panicles (%)	86.7-100	96.2 $\pm$ 1.9
	Grain loss (%)	9.7-29.3	16.9 $\pm$ 2.1

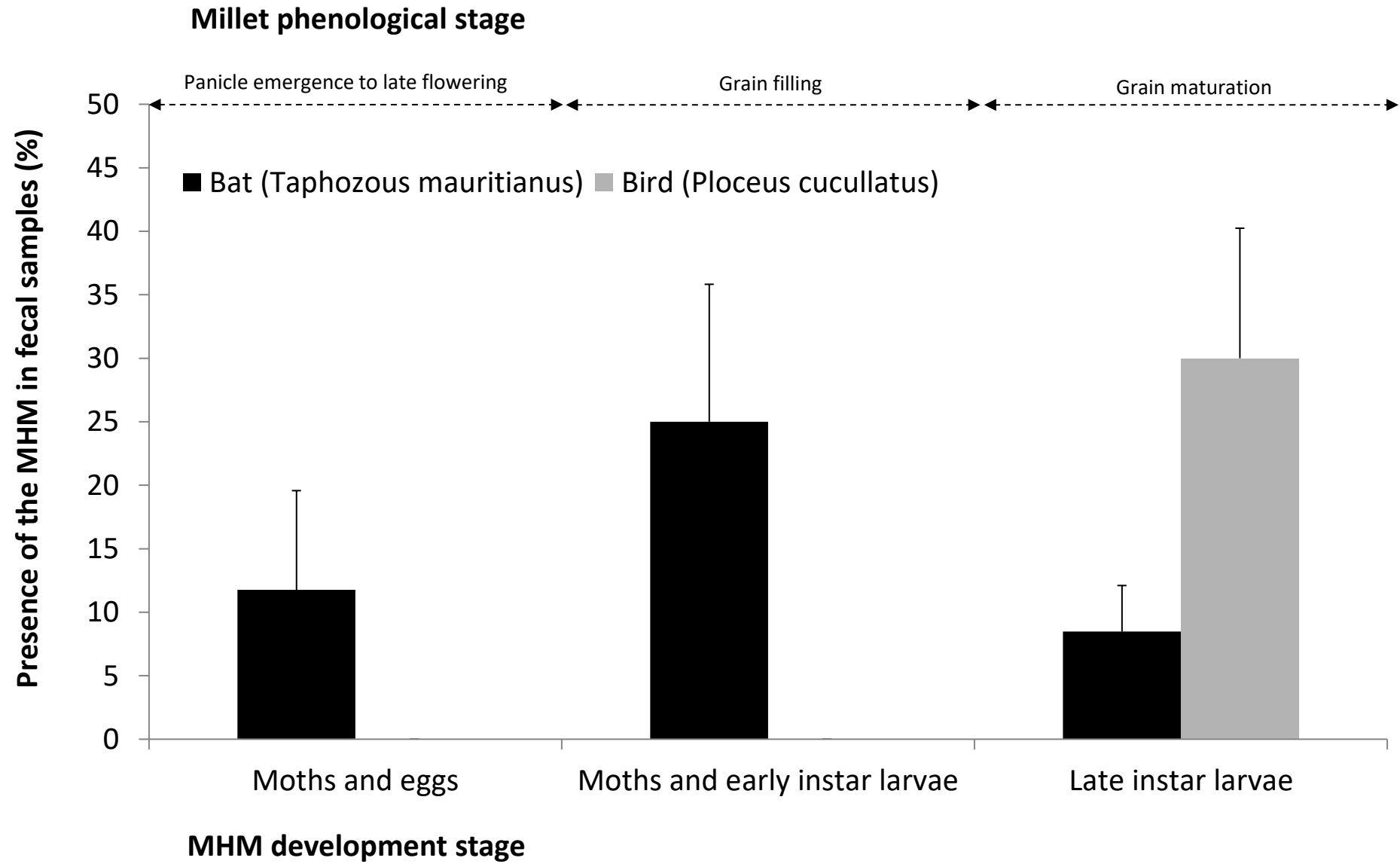
**Table 2.** Inventory of major birds and bats observed in millet-based agroecosystems in the North of the Peanut basin in Senegal.

Order Family	Common name	Latin name	Food diet	Frequency of occurrence (%)	Relative abundance (%)	*Predat ion evidenc e on MHM by direct observa tions. **Pred ation evidenc e on MHM confirm ed by molecul ar
<b>Passeriformes</b>						
Corvidae	African piapiac	<i>Ptilostomus afer</i> (Swainson, 1837)*	I/G	0.4	1.5	
Passeridae	Grey-headed sparrow	<i>Passer griseus</i> (Vieillot, 1817)*	I/G	32.5	42.6	
Ploceidae	Village weaver	<i>Ploceus cucullatus</i> (Statius Müller, PL, 1776)**	I/G	34.2	37.4	
	Little weaver	<i>Ploceus luteolus</i> (Lichtenstein, MHK, 1823)*	I/G	17.1	12.3	
Pycnonotidae	Common bulbul	<i>Pycnonotus barbatus</i> (Desfontaines, 1789)*	I/G	2.5	1.9	
Sturnidae	Long-tailed glossy starling	<i>Lamprotornis caudatus</i> (Statius Müller, 1776)*	I/G	0.4	3.3	
	Chestnut-bellied starling	<i>Lamprotornis pulcher</i> (Müller, 1776)*	I	0.8	0.4	
	Lesser blue-eared starling	<i>Lamprotornis chloropterus</i> (Swainson, 1838)*	I/G	0.8	0.6	
<b>Chiroptera</b>						
Pteropodidae	African straw-colored fruit bat	<i>Eidolon helvum</i> (Kerr, 1792)**	F	-	-	
Emballonuridae	Mauritian tomb bat	<i>Taphozous mauritanus</i> (E. Geoffroy, 1818)**	I	-	-	

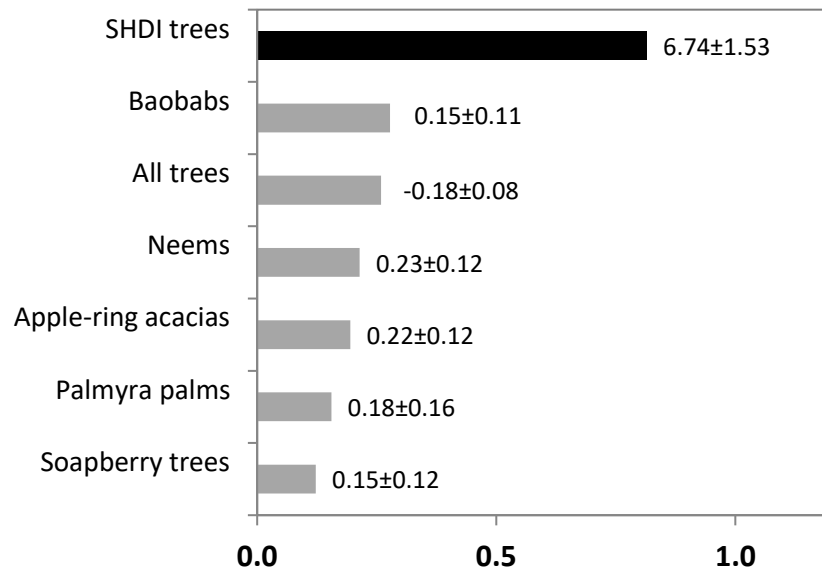
analysis. I: insectivorous. F: frugivorous. G: granivorous.



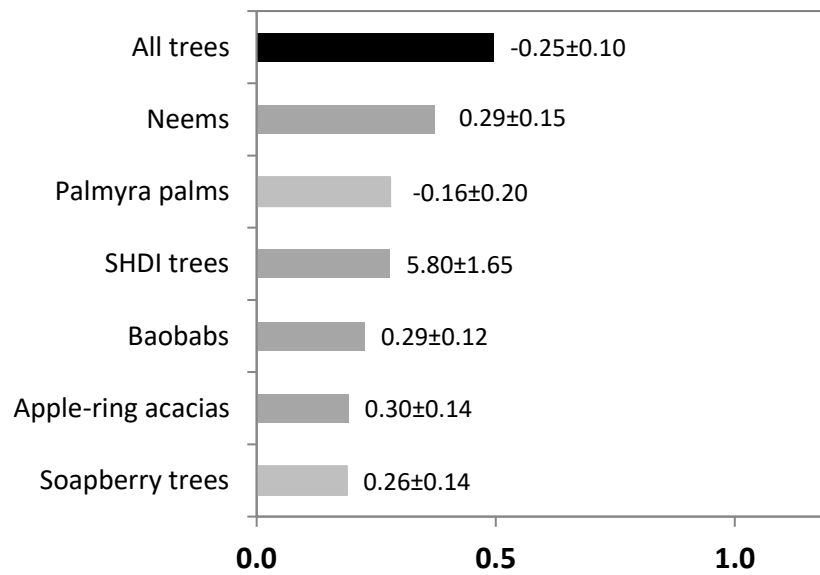
Journal Pre-proof



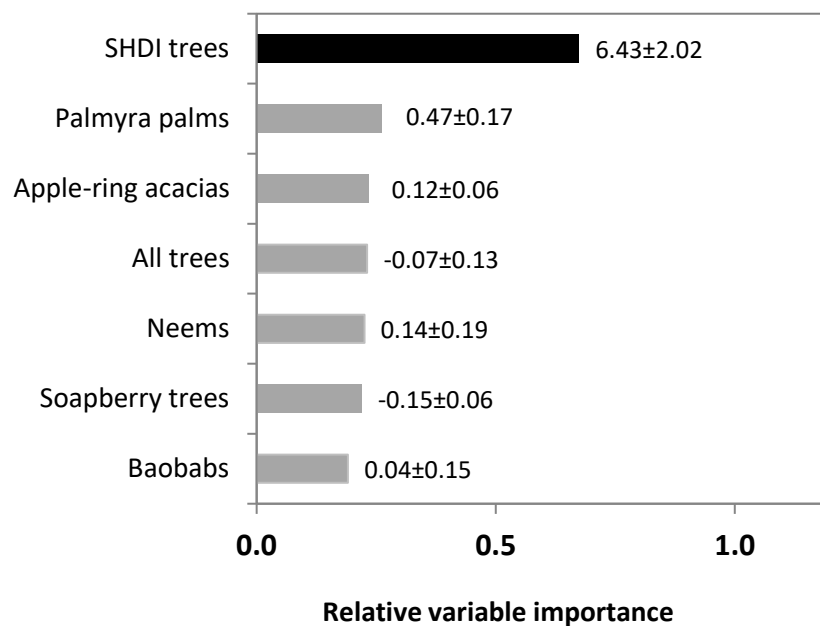
A



B

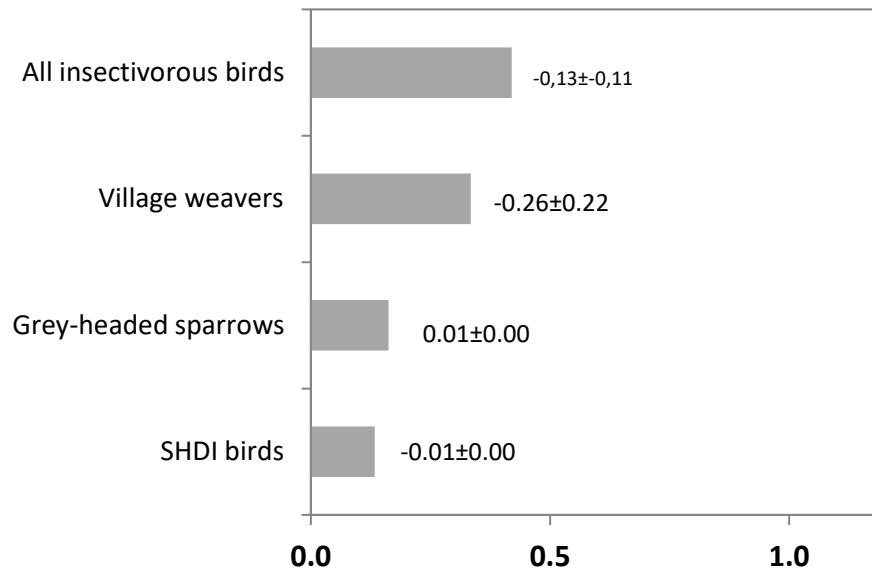


C

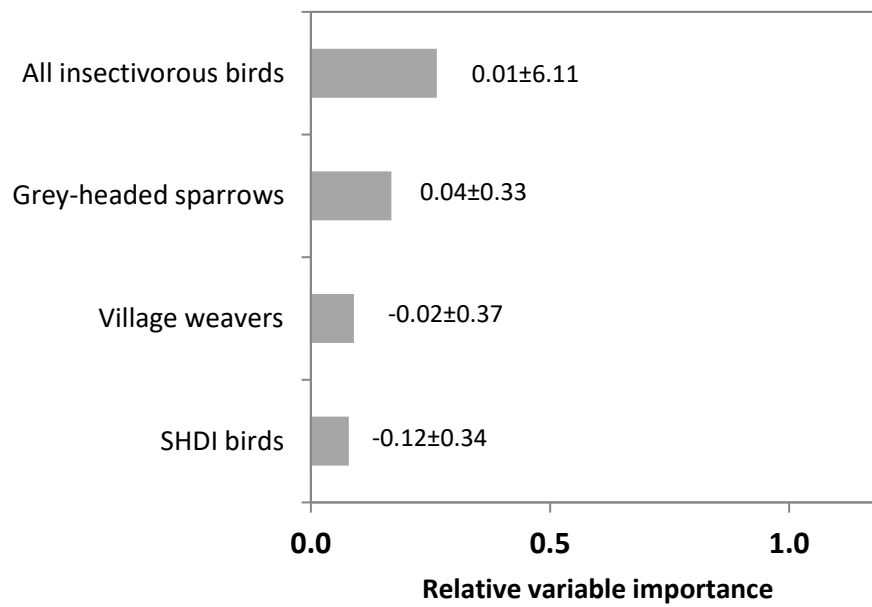


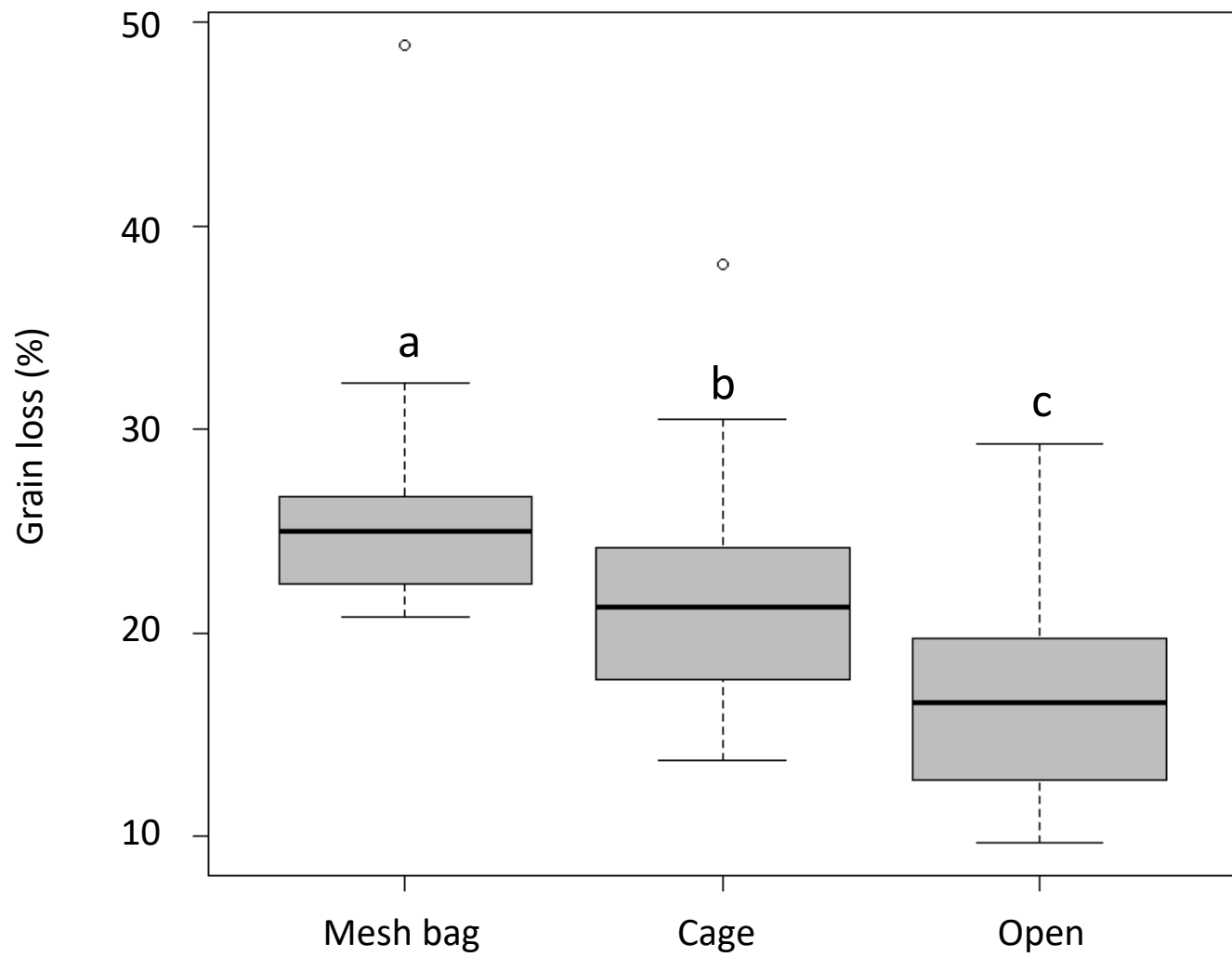
Journal Pre-proof

A



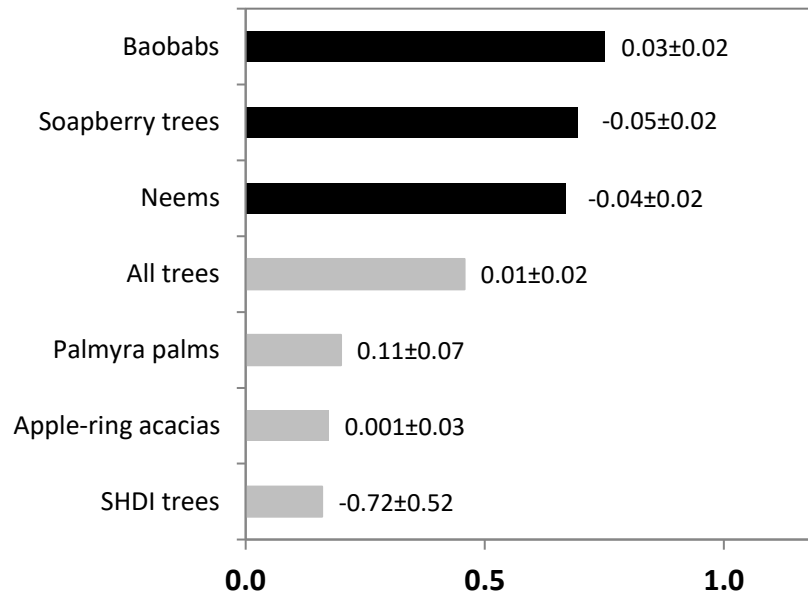
B



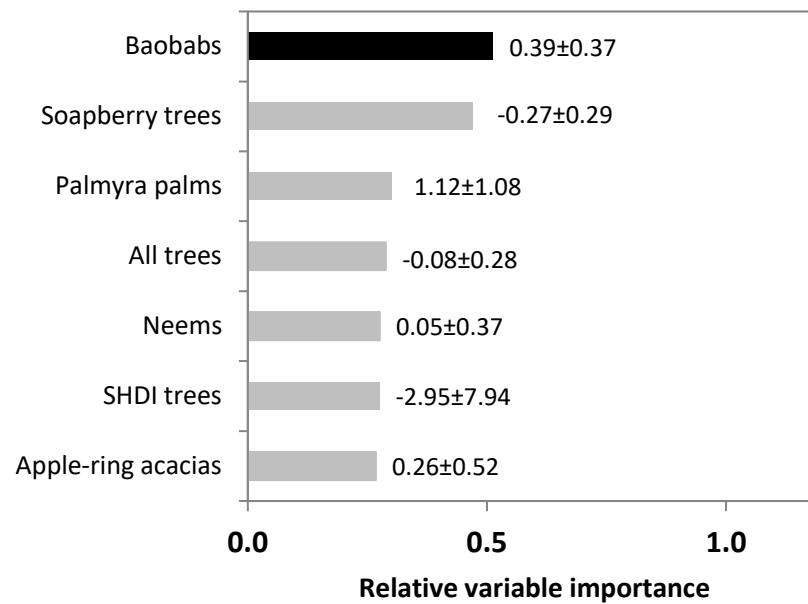




A

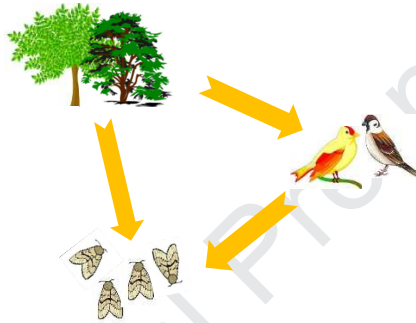


B



## Highlights

- Vertebrates, including birds and bats, are effective predators of the millet head miner.
- They contribute to natural pest control and reduction of millet grain loss.
- The abundance of insectivorous birds is enhanced by tree diversity at a local scale.
- The contribution of trees for natural pest control is species-dependent. Trees can serve as refuges for pest (eg. baobab/moths) and predators (eg. palm tree/bats or neem/village weavers).



**Declaration of interests**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Journal Pre-proof