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1 **Interaction between climate and management on beta diversity components of** 2 **vegetation in relation to soil properties in arid and semi-arid oak forests**

3
4 **Mehdi Heydari, Fatemeh Aazami, Marzban Faramarzi, Reza Omidipour, Masoud Bazgir,**
5 **David Pothier Bernard Prévosto**

6 7 **Abstract**

8
9
10 This study aimed to investigate the interaction between regions with different climatic conditions
11 (arid vs. semi-arid) and management (protected vs. unprotected) on the turnover and nestedness of
12 vegetation in relation to physical, chemical and biological properties of soils in the Ilam province of
13 Iran. In each of the two regions, we sampled eight sites (4 managed and 4 unmanaged sites) within
14 each of which we established four circular plots (1000 m²) that were used to inventory woody species,
15 while two micro-plots (1×1 m²) were established in each 1000-m² plot to inventory herbaceous
16 species. In each sample unit, we also extracted three soil samples (0–20 cm depth) for measuring soil
17 properties. The results indicated that the interaction between region and **conservation** management
18 significantly affected the percent canopy cover of Persian oak (*Quercus brantii* Lindl), soil
19 respiration, substrate-induced respiration, as well as beta and gamma diversities and turnover. The
20 percent canopy cover of oak was positively correlated with **soil silt**, electrical conductivity, **available**
21 **potassium**, and alpha diversity, whereas it was negatively correlated with plant turnover. In addition,
22 plant turnover was positively related to **available phosphorus** while nestedness was positively related
23 to organic carbon and total N. According to these results, we conclude that physical, chemical and
24 biological characteristics of limited ecological niche generally influence plant diversity. **Also, this**
25 **study has indicated the major contribution of the β -diversity on γ -diversity, especially in the semi-arid**
26 **region, because of a higher heterogeneity in this area.**

27

28

29 **Keywords:** climatic conditions; conservation; Beta diversity; oak forests

30

31 **1. Introduction**

32 Forest ecosystems have a critical role to mitigate soil loss and erosion, regulate weather and
33 maintain habitats and biodiversity (Fathizadeh et al., 2017). Forest degradation and/or destruction had
34 many negative effects on plant and soil microorganism diversity and composition (Yacht et al., 2017;
35 Moradi Behbahani et al., 2017) and soil properties (Feng et al., 2017). Conservation management can
36 be a key factor in preventing worldwide destruction of forest ecosystems and promoting the
37 sustainability of these valuable resources.

38 Biodiversity indices are generally used to assess and evaluate ecosystem degradation (Wilson and
39 Tilman, 2002). However, with the increasing importance of biodiversity conservation, the use of new
40 and effective methods to investigate plant biodiversity has become increasingly popular. Diversity
41 partitioning (additive vs. multiplicative) is one of these methods that was frequently used during the
42 two recent decades by ecologists devoted to biodiversity conservation (Erfanzadeh et al., 2015).
43 According to this method, total regional diversity can be partitioned additively ($\gamma = \alpha +$
44 β ; Lande, 1996) or multiplicatively ($\gamma = \alpha \times \beta$) into within (α diversity) and
45 among (β diversity) components across different spatial/temporal scales (Crist et al., 2003). The
46 additive partitioning method has been increasingly used in rangeland and forest studies in relation to
47 plant species diversity (i.e. Chávez and Macdonald, 2012; Zhang, et al., 2014; Erfanzadeh et al.,
48 2015; Schulze, et al., 2016).

49 Beta diversity is an important biodiversity index that is defined as the variation in community
50 composition from one place to another. Baselga (2010) demonstrated mathematically that beta
51 diversity can be decomposed in two main processes: turnover and nestedness. Spatial turnover is
52 defined as the continuous or sudden replacement of species along an environmental gradient due to
53 environmental pressure, climate change or competition (Baselga, 2010; Lafage et al., 2015). For its
54 part, nestedness refers to communities in which species composing a low-diversity community are a
55 subset of a high-diversity community that was subjected to species migration or extinction (Calderón-
56 Patrón et al., 2013). By analyzing beta diversity components and their respective contribution to the
57 total beta diversity, we can detect the main mechanism involved in changes in community
58 composition.

59 Semi-arid forests of the Zagros Mountains are dominated by Persian oak (*Quercus brantii* Lindl.)
60 and cover an area of about 5 million ha (40% of Iranian forests; Sagheb-Talebi et al., 2014). They
61 form one of the most important ecosystems of Iran due to their high level of biodiversity and the
62 presence of endemic species (Marvi Mohadjer, 2005). The Zagros region has a great impact on
63 people's livelihood due to the dependency of people on local resources (Salehi et al., 2013).

64 Unfortunately, during the three last decades, the vegetation of the Zagros forests was highly degraded
65 due to incorrect management practices and human activities such as human-caused fire, over-grazing,
66 excessive cutting, firewood harvesting and land use changes (Heydari et al., 2016).

67 While forest degradation could create important changes in vegetation composition and forest
68 structure, forest soils can also be impacted. Soils are an important constituent of ecosystems and play
69 a critical role in the development of forest vegetation, which, in turn, have noticeable effects on the
70 development of physical, chemical and biological soil properties (Kooch et al., 2007; Onyekwelu et
71 al., 2006; Wang, 2007).

72 Biodiversity conservation in Zagros forests should take advantage of new management practices to
73 help reduce the progression of its degradation caused by drought-induced oak mortality. Conservation
74 plan including exclosures is one management solution that was effective in Zagros forests (Heydari et
75 al., 2013a). Tárrega et al. (2009) reported that different management regimes in dehesa ecosystems
76 (grazed, ungrazed, ungrazed with shrub cutting) had a critical effect on soil properties and vegetation
77 in oak (*Q. pyrenaica*) stands, and that conservational practices led to improvement in soil fertility.
78 This is in accordance with results indicating that conservational practices (grazed vs. no grazed) had
79 positive effects on soil properties and vegetation (Strandberg et al., 2005). In addition, Sheklabadi et
80 al. (2007) observed that in areas degraded by grazing, the amount of organic carbon, total nitrogen
81 and microbial respiration were lower than in non-degraded areas.

82 Climate is another factor that has profound effects on vegetation diversity and composition as well
83 as on soil properties (Kardol et al., 2010). Climatic factors, particularly temperature and
84 precipitation, affect the formation and evolution of soils (Binkley and Fisher, 2012), the
85 accumulation of organic carbon (Wang et al., 2013), the cycling rate of nutrients (Auyeung et al.,
86 2013) and the soil biological activity (Sardans and Peñuelas, 2005; Steinweg et al., 2013).

87 Due to the interaction between [soil characteristics, climatic conditions and plant species](#) (von
88 [Lutzow & Kogel-Knabner, 2009](#); [Mirzaei and Moradi, 2017](#); Toure et al., 2015), changes in plant
89 species diversity and composition between regions with different climatic conditions are expected.
90 On the other hand, plant communities and their characteristics such as life form spectra (Raunkiaer,
91 1934) gradually reach an equilibrium with climatic conditions to form so-called climatic climax
92 communities.

93 The effects of forest degradation and management strategies on plant diversity and soil attributes
94 have been recently investigated in the Zagros forest of Iran (Parma & Shataee Jouybari, 2010;

95 Heydari et al., 2017). However, no study has addressed the effects of climate with or without
96 interaction with management practices on diversity components (alpha and beta diversity) as well as
97 on beta diversity components (turnover and nestedness). This study aims to fill this gap by evaluating
98 the effects of climate, management strategies and their interaction on soil physical, chemical and
99 biological properties as well as on biodiversity components.

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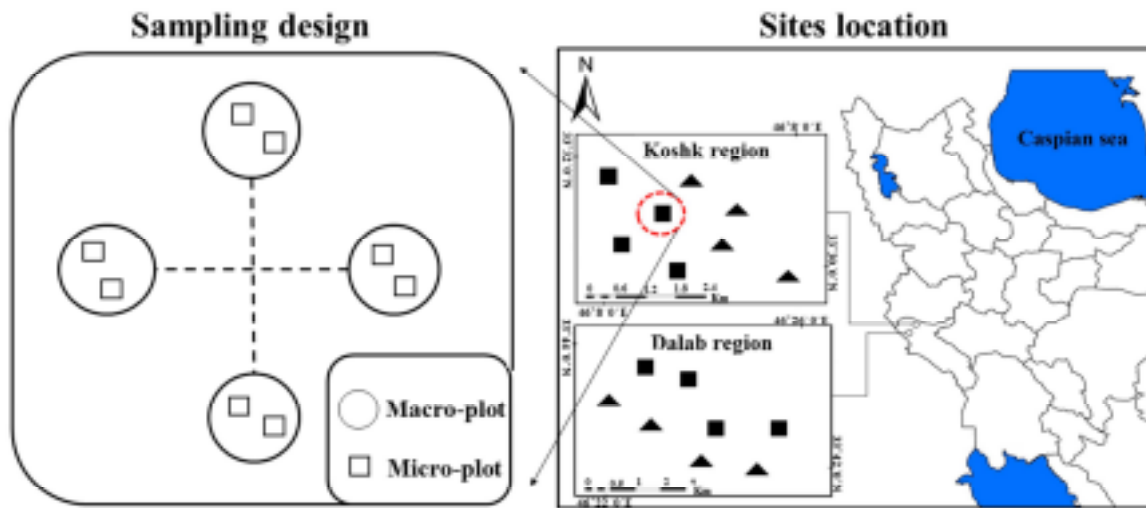
102 **2. Material and methods**

103 2.1. Study area

104 This study took place in managed and unmanaged (see details below) forests dominated by Persian
105 oak (*Quercus brantii* Linddl) in western Iran. These forests were located in two regions (Fig. 1) with
106 different climatic conditions. According to the classification system of de Martonne (1925), the
107 climate of these regions are both Mediterranean, but the semi-arid Dalab region is characterized by
108 annual precipitation of 590 mm and mean annual temperature of 16.7 °C, while the mean annual
109 precipitation and temperature of the arid Koshk region are 408 mm and 18.5 °C, respectively. These
110 regions are both part of the Zagros mountains, and the geology and the pedogenesis of the two
111 regions are similar. Calcareous soils with a clay loamy texture, dominate the soils of these regions.

112 In each region, we selected four managed and four unmanaged sites that were described by
113 Heydari et al. (2013 b). In managed sites, a 20-year period of strict conservation was applied
114 following long-term human disturbances. For their part, unmanaged sites were never protected
115 against long-term anthropogenic land degradation such as tree and shrub cutting for firewood and
116 development of arable lands as well as animal husbandry (Heydari et al., 2013 b). The vegetation of
117 the managed sites of Dalab (MD) was dominated by *Cerasus microcarpa*, *Daphne mucronata*,
118 *Acantholimon bromoifolium*, *Teucrium polium*, *Alyssum marginatum*, *Centaurea depressa*, *Centaurea*
119 *amadanensis*, *Centaurea koeieana*. In the unmanaged sites of Dalab (UMD), the dominant
120 vegetation was composed of *Capparis parviflora*, *Astragalus fasciculifolius*, *Astragalus*
121 (*Leucocercis*) *curviflorus*, *Bromus tomentellus*, *Gundelia Tournefortii*, *Erodium cicutarium*,
122 *Stellaria media*, *Valerianella dactylophylla*. In the managed sites of Koshk (MK), the dominant
123 species were *Pistacia khinjuk*, *Amygdalus lycioides*, *Populus euphratica*, *Medicago rigidula*, *Sinapis*
124 *arvensis*, *Trifolium purpureum*, *Ziziphora capitata*, *Anthemis pseudocotula*. Finally, in the
125 unmanaged sites of Koshk (UMK), the vegetation was dominated by *Nerium oleander*, *Vitex*

126 *pseudo-Negundo Medicago rigidula*, *Sinapis arvensis*, *Phlomis olivieriv*, *Crepis katschyana*,
 127 *Scabiosa leucactic* (Aazami, 2016).
 128



129
 130 Fig. 1 Study area location and sampling method; ■ : managed and ▲ : unmanaged sites;
 131 ○ : Macro-plot and □ : Micro-plot

132
 133 2.2. Experimental design and data collection

134 The combination of our two regions (D=Dalab, K=Koshk) with the two types of management
 135 (M=managed, UM=unmanaged) results in four treatments (MD, UMD, MK, UMK) and with
 136 different climatic conditions (semi-arid Dalab region vs. arid Koshk region). In each treatment, we
 137 selected four independent sites that were considered as replicates. However, to decrease variability,
 138 geographically adjacent sites were selected in each region. In the center of each site, we used a
 139 systematic sampling method to establish a cluster of four circular, 1000-m² plots located 100 m
 140 apart (Fig. 1). These plots were used to estimate the percent cover of all woody species while two
 141 squared 1-m² microplots, located within each main plot, were randomly established to inventory
 142 herbaceous species (each species was given a value of abundance) in May 2015. In total, we have 16
 143 sampling sites (8 in Koshk and 8 in Dalab), 64 macroplots (1000-m²) and 128 microplots.

144

145 2.3. Soil sampling and laboratory methods

146 We extracted soil samples at 0-20 cm depth in May 2015. Physical, chemical and biological soil
147 properties were determined from composite samples, each consisting of three sub-samples that were
148 gathered from three randomly selected points in each main plot (for a total of 64 composite soil
149 samples). Soils were sieved (2-mm mesh) to remove roots and debris prior to laboratory analyses
150 and divided into two subsamples.

151 The first subsample was air-dried for measuring physico-chemical properties. Soil texture was
152 measured using a hydrometer (Bouyoucos, 1962). Soil organic carbon (OC) was determined by
153 dichromate oxidation using the Walkley-Black method (Nelson and Sommers, 1982). Soil pH and
154 electrical conductivity (EC) were determined electrometrically (in H₂O, 2:1 v/m) with a
155 conductivity probe in filtered extracts (Kalra and Maynard, 1991), respectively. Soil cation
156 exchange capacity (CEC) was determined following extraction in buffered sodium acetate (NaOAc,
157 at pH 8.2; [Sumner and Miller, 1996](#)). Total nitrogen was measured by Kjeldahl digestion (Bremner,
158 1996). Available phosphorus (P_{ava}), as ortho-PO₄⁻², was determined using the method of Bray and
159 Kurtz (1945). Available potassium (K_{ava}) was determined following ammonium acetate (pH 7)
160 extraction and quantified by flame photometry (Black, 1986). Lime percentage, expressed using the
161 Total Neutralizing Value (T.N.V), was determined using the NaOH titration method.

162 The second subsample was maintained at field moisture and stored at 4 °C for subsequent
163 measurements of soil microbial activity. Soil basal respiration (BR) was measured by trapping and
164 quantifying CO₂ that was emitted from soil samples over a five-day period (Alef and Nannipieri,
165 1995). Substrate-induced respiration (SIR) was determined using glucose (1 %) as the substrate and
166 the evolved CO₂ was determined after 8 h incubation. Evolved CO₂ was adsorbed in 1 M NaOH and
167 measured by 0.1 M HCl titration (Anderson and Domsch, 1978).

168 Additional undisturbed soil cores distributed in the four treatments were collected for the
169 determination of bulk density (BD) in the 0-15 cm mineral layer (Blake and Hartge, 1986). Soils
170 were immediately sieved (2 mm) and kept in a plastic box to avoid evaporation. Soil moisture
171 content or WC (water mass/soil dry mass) was measured gravimetrically (oven-drying at 105 °C for
172 24 h; [Famiglietti et al., 1998](#))

173

174 2.4. Diversity partitioning

175 To estimate the diversity components, we used the additive partitioning method as shown in
176 equation (1) (Crist et al., 2003):

$$\gamma = \alpha_1 + \sum_{i=1}^m \beta_i \quad [1]$$

178

179

180 where γ is the total species diversity, m is the number of scales, α_1 is the average diversity within
181 sample units in each site, and β_i is the beta diversity at each scale i . This method enabled us to
182 estimate the total species richness in each region (γ), which was partitioned into α_1 (the mean
183 number of species found per 1 m² plot) and β_1 (the difference between α_1 and γ_{site} , in which γ_{site}
184 was the total species richness of each site).

185 Then, we additively partitioned the β -diversity into two components, i.e. the spatial turnover and
186 nestedness. To do this, we used the method suggested by Baselga (2010), which is based on the
187 multiple dissimilarity derived from the Sørensen coefficient of dissimilarity (Baselga 2010). This
188 analysis was performed using the betapart package (Baselga and Orme 2012) within the R software
189 (R Development Core Team 2013).

190

191 2.5. Statistical analysis

192 First, the normality of each variable was verified using the Kolmogorov–Smirnov test as well as
193 variance homogeneity (Levene Test). Mathematical transformations of data were used when
194 necessary to correct deviations from normality and heterogeneity of variance. Then we used two-
195 way ANOVAs to detect significant differences between variable means associated with treatments,
196 regions, and the interaction between treatments and regions. If the two-way ANOVAs detected
197 significant differences, we applied a Duncan test for pairwise comparisons. Pearson correlation
198 coefficients were calculated to investigate possible relationships between oak cover, soil properties
199 and diversity components. Finally, relationships between beta diversity components and soil factors
200 were analyzed using linear regressions. All analyses were performed with the R statistical software
201 (R Core Team 2013).

202

203

204

205 **3. Results**

206 3.1. Effects of management measures and climate

207 The results showed that management (protected vs. unprotected), climatic conditions (arid vs. semi-
 208 arid) and their interaction had a significant (P -value < 0.05) effects on oak canopy cover (Table 1).
 209 These variables also produced significant effects on water content, silt, BR and SIR. However, for
 210 T.N.V, clay, Pava, Kava, total N and OC, only management and climatic conditions were significant
 211 and their interaction was not significant (Table 1). The percent of sand and BD were solely affected
 212 by management, while EC and pH were affected only by climatic conditions.

213 Diversity components (alpha, beta and gamma) and beta diversity components (turnover and
 214 nestedness) showed different trends in relation to management, climatic conditions and their
 215 interaction (Table 1). We detected significant effects of management and climatic conditions for all
 216 diversity components but a significant influence of the interaction was revealed only for beta and
 217 gamma diversity as well as turnover. Therefore, the effects of management on alpha diversity and
 218 nestedness were similar across the two regions (Table 1).

219 Table 1 Results of two way-ANOVA (F-value and significance level) testing for regions with different
 220 climatic conditions and management on oak canopy cover percentage, soil properties, diversity components
 221 (alpha, beta and gamma) and beta diversity components (spatial turnover and nestedness); (EC: electrical
 222 conductivity; OC: organic carbon; BD: Bulk density; WC: Water content; T.N.V: total neutralizing Value;
 223 N_{tot} :total nitrogen; K_{ava} : available potassium; P_{ava} : available phosphorus; BR: basal respiration; SIR:
 224 substrate induced respiration)

	Variable	Management	Climate condition	Management × Climate condition
Oak canopy cover and soil physical properties	Oak canopy cover (%)	11.662**	8.396**	6.04*
	BD (g/cm ³)	1.037*	2.846	2.732
	Water content (%)	15.859**	11.017**	5.475*
	Sand (%)	5.988*	0.828	0.198
	Clay (%)	3.562*	0.608*	0.007
	Silt (%)	0.463*	3.795**	0.372*
Soil chemical properties	T.N.V (%)	26.447***	14.393**	2.084
	OC (%)	62.331***	43.721***	0.388
	Pava (ppm)	13.529	394.731***	0.438
	EC (dS/m)	1.752	6.209*	5.1*
	pH	4.134	62.152***	1.145
	N_{tot} (%)	62.331***	43.721***	0.388
	K_{ava} (ppm)	0.219	43.522***	0.005
Soil biological	BR (mg CO ₂ - C/kg soil/day)	177.92***	368.656***	112.515***

variables	SIR (mg CO ₂ -C/kg soil/day)	105.622***	162.689***	33.996***
Diversity	Alpha diversity	160.163***	4.379*	0.054
	Beta diversity	136.289***	39.472***	37.768***
	Gamma diversity	370.559***	20.931***	35.477***
	Turnover	11.626**	183.214***	42.374***
	Nestedness	13.636**	4.909*	2.182

225 *** Significant at $P = 0.001$; ** Significant at $P = 0.01$ and * significant at $P = 0.05$

226

227

228

229 3. 2. Oak canopy cover and soil physico-chemical properties

230 Oak canopy cover was significantly higher in the Dalab region than in the Koshk region. This
 231 variable was also significantly higher in managed than in unmanaged plots for both Dalab and Koshk
 232 regions (Table 2). Available K was higher in the Dalab than in the Koshk region while the reverse
 233 was true for P_{ava} .

234 EC was higher in the unmanaged than in the managed plots of the Koshk region while it was lower
 235 in the Dalab region within which no significant differences were found between managed and
 236 unmanaged plots. pH was unaffected by management in both regions but was higher in the Koshk
 237 than in the Dalab region. Clay content was significantly higher in MD than in the other treatments,
 238 whereas silt content was higher in MK. Sand content was higher in unmanaged plots but the
 239 difference was significant in the Koshk region only. The treatment order for total nitrogen and organic
 240 matter contents was MD > UMD = MK > UMK whereas that of T.N.V was UMK > UMD > MK > MD
 241 (Table 2). Water content and bulk density were highest in the managed plots of the Dalab region and
 242 in the unmanaged plots of the Koshk region, respectively.

243

244 3. 3. Soil biological variables

245 The mean values of BR and SIR in the Dalab region were higher than those of the Koshk region but
 246 the influence of management was significant in both regions with higher values of BR and SIR in
 247 managed than in unmanaged plots (Table 2).

248 Table 2 Oak canopy cover, soil physico-chemical and microbial properties (mean \pm SE) measured in the
 249 different regions; (EC: electrical conductivity; OC: organic carbon; BD: Bulk density; WC: Water content;
 250 T.N.V: total neutralizing Value; N_{tot}: total nitrogen; K_{ava}: available potassium; P_{ava}: available phosphorus;
 251 BR: basal respiration; SIR: substrate induced respiration)

252

Variable	Koshk		Dalab	
	Managed	Unmanaged	Managed	Unmanaged
Oak canopy cover (%)	10.18 \pm 2.33 b	7.52 \pm 1.32b	20.13 \pm 3.92a	8.94 \pm 1.22b
BD (g/cm ³)	1.37 \pm 0.02b	1.48 \pm 0.03a	1.36 \pm 0.01b	1.38 \pm 0.02b
Sand (%)	47.87 \pm 1.61b	55.65 \pm 2.45a	48.58 \pm 1.24b	52.71 \pm 2.34ab
Clay (%)	27.71 \pm 1.51b	24.43 \pm 0.51b	31.04 \pm 2.16a	27.04 \pm 3.2b
Silt (%)	24.42 \pm 1.64a	19.92 \pm 2.04b	20.38 \pm 1.24b	20.25 \pm 0.91b
WC (%)	40.87 \pm 0.59b	38.06 \pm 0.85b	50.54 \pm 3.09a	39.74 \pm 0.99b
pH	7.38 \pm 0.01a	7.37 \pm 0.01a	7.14 \pm 0.04b	7.20 \pm 0.01b
EC (ds/m)	0.38 \pm 0.03 b	0.52 \pm 0.03a	0.29 \pm 0.01c	0.32 \pm 0.01c
Total nitrogen (%)	0.23 \pm 0.010b	0.18 \pm 0.010c	0.33 \pm 0.010a	0.21 \pm 0.006b
OC (%)	3.37 \pm 0.31b	2.81 \pm 0.55c	5.7151 \pm 0.32a	3.55 \pm 0.11b
T.N.V (%)	43.93 \pm 0.91b	52.40 \pm 0.47a	35.83 \pm 3.39c	48.85 \pm 1.87ab
K _{ava} (ppm)	263.09 \pm 15.00bc	169.80 \pm 8.47c	400.79 \pm 17.56a	387.67 \pm 39.55a
P _{ava} (ppm)	9.95 \pm 0.07a	9.09 \pm 0.13a	7.60 \pm 0.21b	7.07 \pm 0.05b
BR (mg CO ₂ - C/kg soil/day)	12.21 \pm 0.35c	10.49 \pm 0.21c	34.01 \pm 0.41a	16.40 \pm 1.34b
SIR (mg CO ₂ -C/kg soil/day)	23.61 \pm 0.62b	21.84 \pm 1.48c	49.40 \pm 1.95a	24.58 \pm 1.46b

253 **Within rows, means with the same letters do not significantly differ (p > 0.05)**

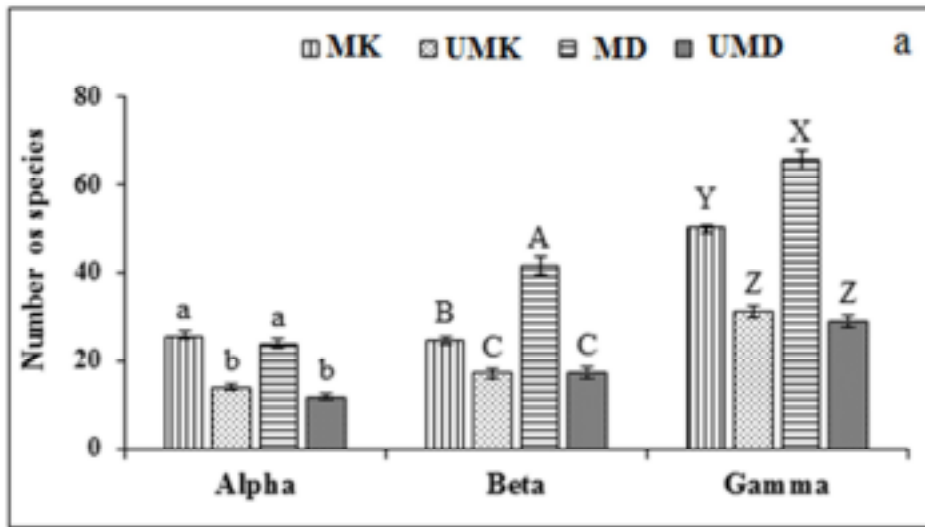
254

255 3.4. Alpha, Beta and Gama diversity and plant beta diversity components

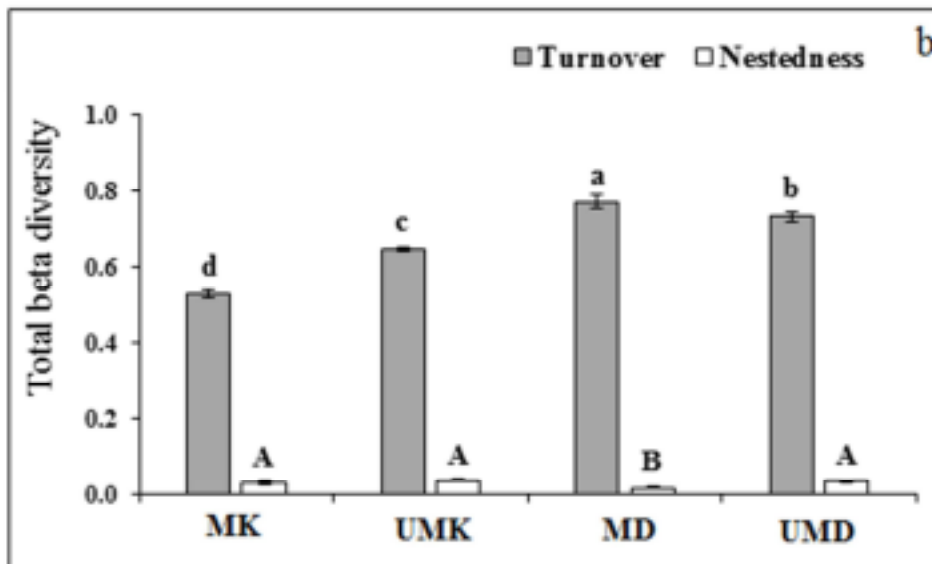
256 Alpha, beta and gamma diversity values were higher in managed than in unmanaged plots in both
 257 regions. Diversity values were significantly higher (beta and gamma) or similar (alpha) between
 258 managed Dalab and managed Koshk whereas no significant differences were detected between the
 259 two regions in the unmanaged plots (Fig. 2 a).

260 Between the two components of the beta diversity, spatial turnover was largely higher than
 261 nestedness in each treatment (Fig. 2 b). Turnover significantly decreased according to the following
 262 order MD > UMD > UMK > MK. By contrast, nestedness was lowest in MD while there were no
 263 significant differences among the other treatments.

264



265



266

267

268 Fig. 2 Variations of the diversity components (mean \pm SE) according to the treatments for a) Alpha,
 269 Beta and Gamma components and b) Turnover and Nestedness. Means with the same letters for a
 270 given diversity component are not significantly different based on Duncan's multiple range test ($p >$
 271 0.05)

272

273 3.5. Relationships between oak canopy cover, soil properties and diversity components

274 The oak canopy cover was positively correlated with silt, pH, K_{ava} and alpha diversity, while it was
 275 negatively correlated with spatial turnover (Table 3).

276

277

278 Table 3 Pearson correlation coefficients between the oak canopy cover and soil properties and components
 279 of diversity; (EC: electrical conductivity; OC: organic carbon; BD: Bulk density; WC: Water content;
 280 T.N.V: total neutralizing Value; N_{tot}: total nitrogen; K_{ava}: available potassium; P_{ava}: available phosphorus;
 281 BR: basal respiration; SIR: substrate induced respiration)
 282

Variables	oak canopy cover	Variables	oak canopy cover	Variables	oak canopy cover
BD (g/cm ³)	- 0.14 ^{ns}	P _{ava} (ppm)	0.14 ^{ns}	Total nitrogen (%)	- 0.02 ^{ns}
WC (%)	- 0.08 ^{ns}	K _{ava} (ppm)	0.36 [*]	Alpha	0.36 [*]
Sand (%)	- 0.23 ^{ns}	TNV (%)	0.09 ^{ns}	Beta	- 0.28 ^{ns}
Clay (%)	- 0.03 ^{ns}	BR (mg CO ₂ - C/kg/day)	- 0.25 ^{ns}	Gamma	- 0.09 ^{ns}
Silt (%)	0.14 [*]	SIR (mg CO ₂ -C/kg soil/day)	- 0.28 ^{ns}	Spatial turnover	- 0.62 [*]
pH	0.32 [*]	EC (dS/m)	0.13 ^{ns}	Nestedness	0.12 ^{ns}

283

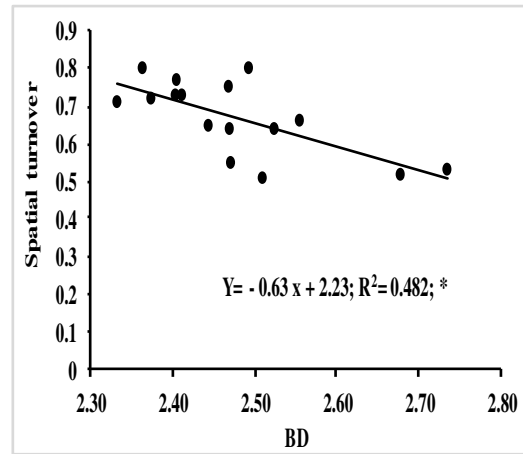
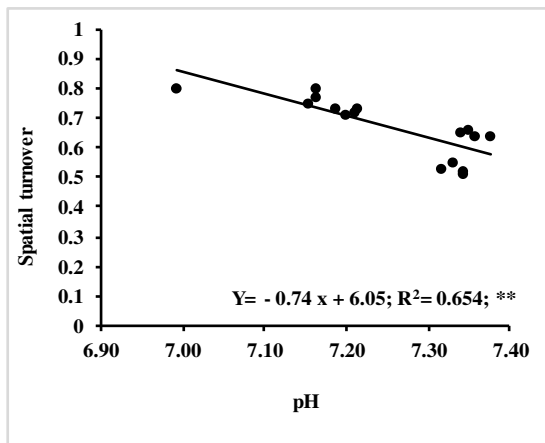
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285 3.6. Linear regressions between beta diversity components and soil properties

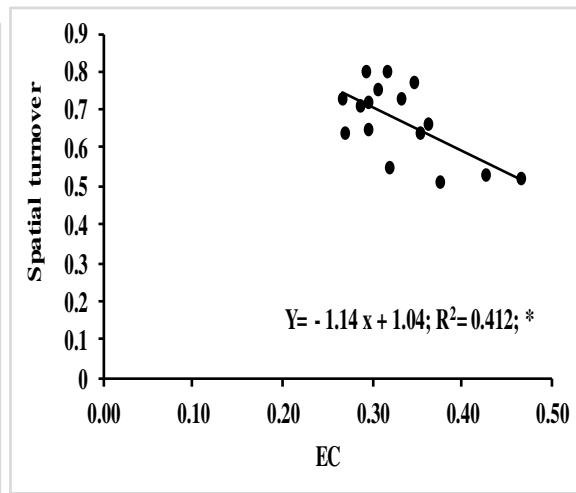
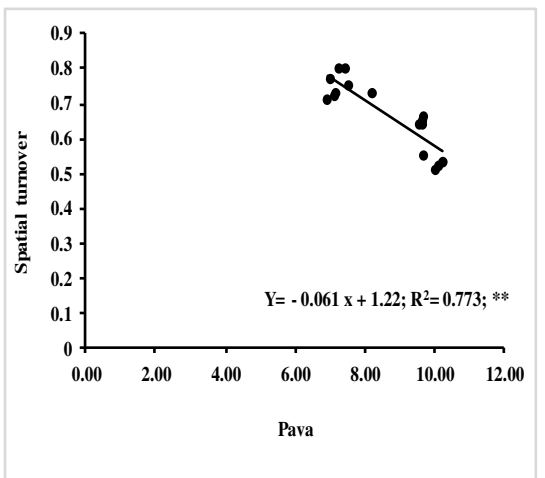
286 There were strong negative relationships between spatial turnover and BD, EC, pH and P_{ava}. By
 287 contrast, spatial turnover was strongly and positively related to Kava, BR and SIR (Fig. 3).

288

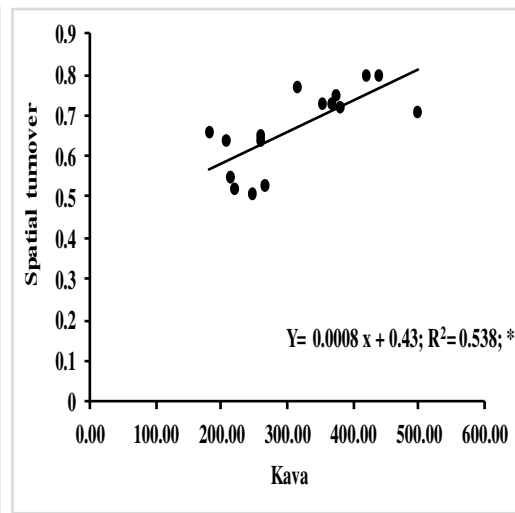
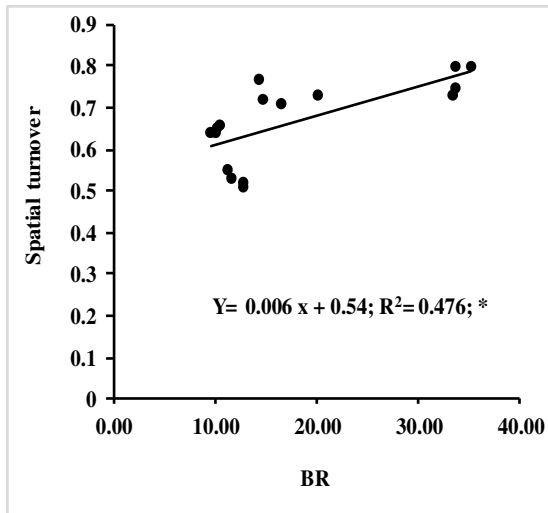
289

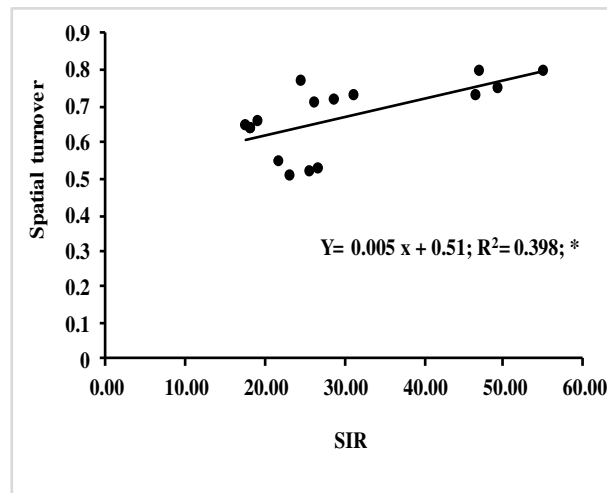


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293 Fig. 3 Simple linear regressions between spatial turnover of plant species and soil properties;

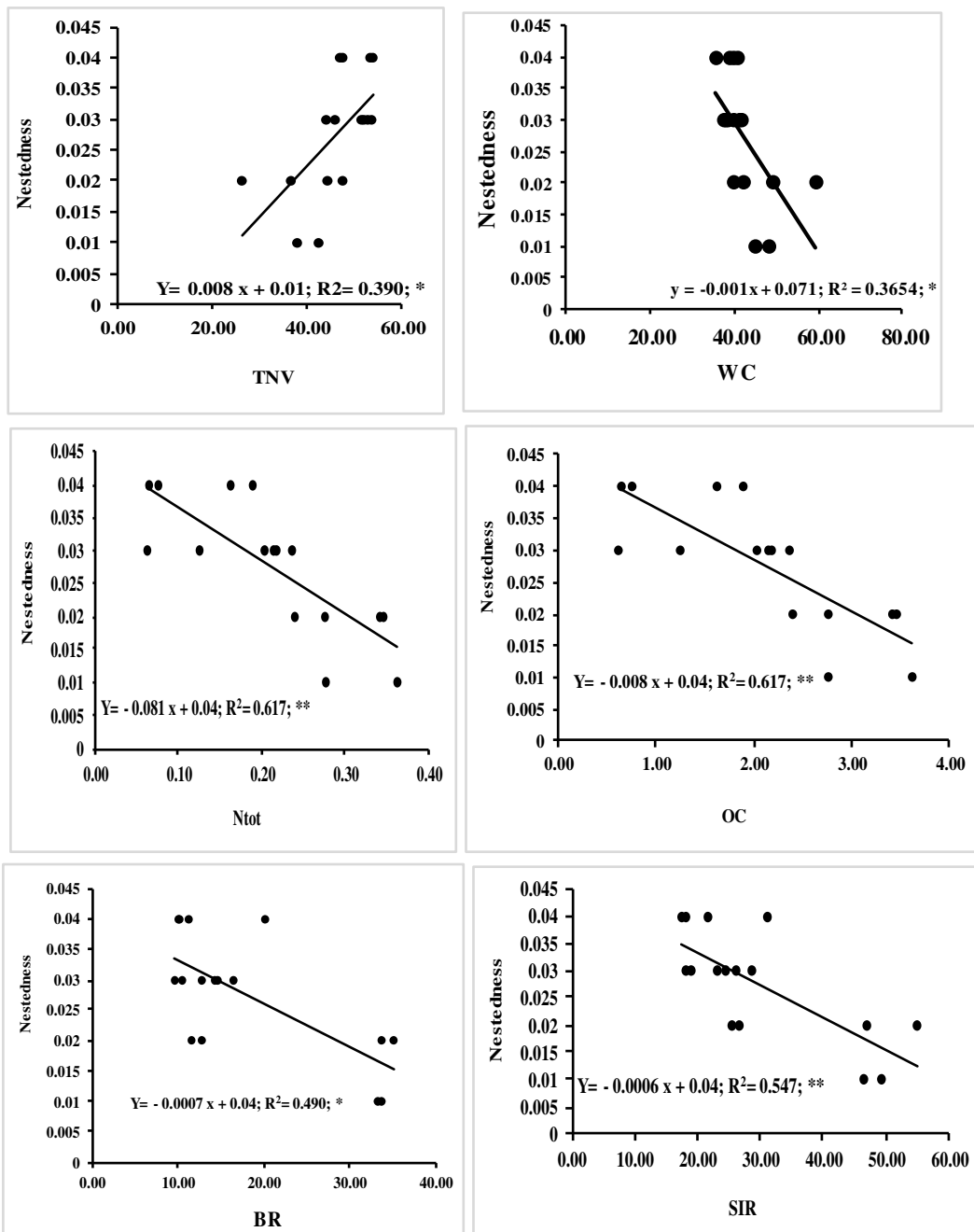
294 * p-value < 0.05 and ** p-value <0.01

295

296 The nestedness of plant species was significantly and positively related to T.N.V., and negatively
 297 related to WC, OC, N_{tot}, BR, and SIR (Fig.4).

298

299



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303 Fig. 4 Simple linear regressions between nestedness of plant species and soil properties; * p-
 304 value < 0.05 and ** p-value < 0.01

305

306

307

308 5. Discussion

309 5.1. Effect of management and climatic conditions on soil properties

310 We found that some soil attributes were influenced by climatic conditions while others responded to
311 conservation measures. For instance, pH was probably mainly influenced by climatic conditions and
312 showed significantly lower values in the more humid Dalab region than in Koshk. However, we cannot
313 exclude that the higher organic matter content in soils of Dalab has stimulated the microbial activity, in
314 particular respiration, which has led to enhanced carbon dioxide production and the formation of
315 carbonic acid forms reducing soil pH (Brady and Weil, 2008).

316 Similarly, soil K_{ava} was higher in the semi-arid Dalab region than in the arid Koshk region whereas the
317 reverse was observed for P_{ava} . In accordance, Sardans and Peñuelas (2007) emphasized the influence of
318 climatic conditions, especially drought, in modifying soil phosphorus and potassium accumulation
319 patterns in Mediterranean woodlands.

320 Total nitrogen and organic carbon were higher in the more humid Dalab region than in the arid Koshk
321 region. Accordingly, climatic conditions influence soil organism activities, which play an important role
322 in improving the soil nutrient cycle (Frossard et al., 2000). For example, temperature affects the
323 decomposition of organic matter and thus influences the physical, chemical and biological soil
324 characteristics (von Lutzow & Kogel-Knabner, 2009). Therefore, the higher organic carbon content in
325 the Dalab region is likely related to soil nutrient availability and thus plays an important role in soil
326 fertility (Bationo et al., 2006). The decrease in nitrogen and organic carbon in unmanaged sites was
327 consistent with the result reported by Dahlgren et al. (2003) and Heydari et al. (2014) who showed that
328 human disturbances and site degradation decreased the amount of organic matter and soil nutrients.
329 Higher cover of woody and herbaceous species in managed sites protects soils from erosion (Mekuria
330 and Aynekulu, 2013) while degradation agents such as livestock grazing in unmanaged Koshk decrease
331 plant biomass accumulation, which in turn affects soil fertility (Savadogo et al., 2007).

332 EC was higher in the Koshk than in the Dalab region and unmanaged plots were associated with higher
333 EC values in Koshk. Some disturbances, such as overgrazing or tree cutting, reduce the protection of
334 soil by vegetation and tend to dry out the soil surface and increase [electrical conductivity](#) (Binkley and
335 Fisher, 2012), a process that is especially pronounced in arid and semi-arid regions (Pan et al., 2014).
336 This explanation also holds for the soil water content, which is higher in managed than in unmanaged
337 areas (Kidron and Gutschick, 2013).

338 Bulk density was higher in unmanaged than in managed areas although the difference was only
339 significant in the Koshk region. It is well known that soil compaction that is induced by overgrazing and
340 low soil organic carbon increases bulk density (Li et al., 2011, Chaudhari et al., 2013) as observed in the
341 Koshk region.

342 Finer soil particles (i.e. clay and silt contents) were significantly higher in managed than in unmanaged
343 sites, while sand was higher in unmanaged ones. Due to a higher percentage of bare soil and a lower
344 vegetation cover, soils of unmanaged sites are more exposed to run off and erosion favouring the loss of
345 silt and clay particles (Tessema et al., 2011).

346 Lime percentage (T.N.V) was higher in unmanaged sites than in managed ones in both arid and semi-
347 arid climates. Accordingly, Celik (2005) showed that the amount of lime increased with forest
348 degradation and land-use change, particularly from forestland to agriculture. Changes in the content of
349 soil lime can also be influenced by other factors such as soil organic carbon, biological activities (soil
350 respiration) and CO₂ pressure (pCO₂). Indeed, higher amounts of soil organic matter enhance CO₂
351 pressure due to higher heterotrophic and autotrophic respiration and then increase carbonate dissolution
352 (Brady and Weil, 2008).

353

354 5.2. Soil biological variables

355 Basal respiration (BR) and substrate-induced respiration (SIR) in the the Dalab region were higher
356 than those obtained from the Koshk region. **More abundant precipitations and lower temperatures in the**
357 **semi-arid Dalab region compared to the arid Koshk region** are favourable to vegetation development and
358 thus increase soil organic matter (Binkley and Fisher, 2012). This process promotes microorganism
359 activities, which can explain the higher BR and SIR values. This result is consistent with the findings of
360 Sardans and Peñuelas (2005) who showed that drought reduces soil enzyme activity in a Mediterranean
361 holm oak forest while more generally, mesic climatic conditions are favourable to the activity of soil
362 organisms (e.g. Qiu et al., 2005; Mansourzadeh & Raiesi, 2012). The same process applies to explain
363 the higher BR and SIR values measured in managed plots compared to unmanaged plots. Indeed, the
364 lower vegetation cover in unmanaged areas caused by human disturbances led to lower organic matter
365 content and reduced respiration values (Cheng et al., 2013).

366

367 5.3. Alpha, Beta and Gama diversity and beta diversity components

368 Alpha and gamma diversity values as well as both components of beta diversity (turnover and
369 nestedness) were significantly affected by management type and climatic conditions and, to a lesser
370 degree, by the interaction between these two variables. This is in accordance with the observations that
371 changes in climatic and management conditions influence species occurrence, diversity and distribution
372 in various ecosystems (Adler & Levine, 2007; van der Putten, 2013), particularly in arid and semi-arid
373 rangelands (Erfanzadeh et al. 2015) .

374 However, our results clearly indicated that management had a major influence on alpha, beta and
375 gamma diversity (Fig 2a) whereas climatic conditions had no significant effect in our unmanaged
376 Mediterranean ecosystems contrary to the results of some previous studies (e.g. Rodriguez et al., 2017).
377 Conservation management positively affected the different components of species diversity in our study
378 areas likely because it provided sufficient opportunities for vegetation rehabilitation and plant
379 community reconstruction (Miura et al., 2003; Coetzee et al., 2014) as well as preservation of native
380 plant species diversity (Cox and Underwood, 2011).

381 The fact that beta and gamma diversities were not influenced by climatic conditions when there are no
382 conservation measures indicate that the role of climate on vegetation composition and diversity is
383 considerably reduced compared to the influence of anthropogenic disturbances. In fact, gamma
384 diversity, which corresponds to the number of species in each region, peaked at 65 and 50 in the
385 managed Dalab and Koshk regions, respectively, whereas alpha diversity was approximately the same
386 (24 and 25 respectively in MD and MK). Beta diversity, which represents the difference between the
387 plots diversity (alpha), is therefore of 41 species for Dalab and 25 for Koshk. This indicates that the beta
388 diversity in managed areas of both regions has a similar (Korshk) or a higher (Dalab) contribution to
389 total diversity than alpha diversity. The reduction of the negative impacts of degradation agents by the
390 conservation management strategy resulted not only in a higher diversity at the plot level (alpha
391 diversity; Erfanzadeh et al., 2015) but also increased species composition at larger scales (Hermy and
392 Verheyen, 2007). Our results are consistent with the findings of Tang et al. (2011) who observed a
393 significant positive effect of conservation management on plant species diversity and those of
394 Erfanzadeh et al. (2015) who showed that conservation management can increase plant diversity at the
395 regional scale in arid regions.

396 Our results emphasize the major contribution of spatial turnover in beta diversity in comparison to
397 nestedness whose contribution was limited in all treatments. These results imply that the variation in
398 species assemblages in the studied regions is less explained by the presence of subsets of species from

399 richer sites than by the high degree of species replacement from one location to the other (Baselga,
400 2010). First, the more favourable environmental conditions of the Dalab region can explain its higher
401 turnover rate compared to the Koshk region because better soil qualities can positively affect plant
402 species distribution and migration (Harrison et al., 2006; LaManna et al., 2017) and result in increased
403 turnover. Second, the clear influence of conservation measures in both areas suggest that they should be
404 applied more extensively rather than limited to the richest sites (Balsega, 2010).

405

406 5.4. Relationships between percent oak canopy cover, soil properties and diversity components

407 Overstory canopy cover is an important ecological characteristic of Zagros oak forests, which was
408 significantly reduced in disturbed areas (Heydari et al., 2014). Previous studies indicated that canopy
409 cover controls the biological, physical and chemical soil properties as well as the herbaceous coverage,
410 which in turn, influence the seedling recruitment and development (Moradi et al., 2017; Caldeira et al.,
411 2014).

412 Our results show that oak canopy cover was positively correlated with silt, pH, Kava and alpha
413 diversity while it was negatively correlated with turnover (Table 3). The higher organic matter content
414 in soils under a more developed canopy cover can improve the cation exchange capacity and thus
415 increase the soil pH (Dahlgren et al., 2003). Oak trees, thanks to their deep rooting system, can bring up
416 nutrients from lower soil layers where they can be recycled via leaf litter, a process that can explain
417 higher soil K content with increasing canopy cover (Moreno et al., 2007).

418 The canopy cover and litter, by creating a suitable microclimate and by limiting leaching, plays an
419 important role in seed storage of plant species and thus increase the diversity of plant species (Arriaga &
420 Mercado, 2004; Heydari et al. 2013 a). This can explain the positive relationship between oak canopy
421 cover and the alpha diversity of plant species. As already observed by Heydari et al. (2015) in the same
422 region, human degradation and lack of protective measures significantly and negatively affect overstory
423 canopy, which, in turn, decrease species diversity.

424 In contrast, the negative correlation observed between oak canopy cover and turnover of plant species
425 could reflect the modification of the microclimatic conditions below canopy, particularly light
426 availability, which is beneficial to some shade-tolerant species but detrimental to many light-demanding
427 plants (Benítez et al., 2015). Sabatini et al (2014) proposed that the ecological factors driving species

428 composition and turnover, like canopy cover, forest structure or topography, are system-specific and
429 thus vary according to forest types.

430 Species turnover is negatively correlated with BD, EC, pH and Pava, but positively with Kava, BR and
431 SIR. Turnover means the continuous or sudden replacement of species along an environmental gradient,
432 and that removal or replacement is controlled by environmental variables and competition (Lafage *et al.*,
433 2015; Baselga, 2010). Past studies have shown that the reduction of habitat quality could lead to the
434 elimination of the most sensitive and non-resistant species (extinction) or species with narrow ecological
435 niche range (Tilman and Lehman, 2001). In contrast, the improvement of soil conditions linked to
436 conservation management efforts, can cause increased migration, development and distribution of
437 species (Harrison *et al.*, 2006; LaManna *et al.*, 2017) and may increase the turnover.

438 Nestedness is observed when species of a specific habitat are a subset of species of a richer habitat (in
439 term of species), and is usually explained by species extinction or selective migration (Calderón-Patrón,
440 *et al.*, 2013; Baselga, 2010). In our study, nestedness is significantly and positively related to TNV, and
441 negatively related to WC, OC, N_{tot} , BR, and SIR. Changes in these factors can lead to limitation in
442 ecological niches -for instance, highly calcareous soils are not adapted to lime-intolerant species- and
443 can increase the nestedness due to extinction of species (Barnes *et al.*, 1998). In contrast, more fertile
444 soils (e.g. with higher NPK or OC contents) may increase the competition and thus species replacement,
445 which can lead to a higher turnover.

446 Given that turnover was by far a greater contributor to total beta diversity than nestedness, our results
447 support the hypothesis that site productivity (Harrison *et al.*, 2006; LaManna *et al.*, 2017 but see
448 Zemunik *et al.*, 2016) and environmental heterogeneity (Chase and Leibold, 2002; Veech and Crist,
449 2007; Stegen *et al.*, 2013) promote beta diversity.

450 **6. Conclusion**

451 Our results indicate that conservation management results in positive changes in the physical,
452 chemical and biological soil properties as well as in canopy cover of oak trees, which improve plant
453 species diversity in both studied regions.

454 This confirms the effectiveness of conservation management to increase plant species diversity
455 independently of the climatic conditions although we have noted a more pronounced effect in the less
456 arid area. More specifically, this study has shown the major contribution of the β -diversity on γ -
457 diversity, especially in the semi-arid region, because of a higher heterogeneity in this area.

458 Knowledge of the contribution of the components of beta diversity can be a valuable guide for
459 managers in adopting appropriate management decisions. The high spatial turnover and the low
460 importance of nestedness observed in this study imply that conservation measures should be applied
461 to a large number of sites of various qualities rather than to a limited number of the richest sites.
462 Besides, the negative correlation between oak cover and turnover suggests that oak forests could
463 benefit from a silviculture aiming at enhancing micro-environmental heterogeneity, such as the
464 creation of gaps, in order to promote species conservation and diversity.

465

466 **Acknowledgements**

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