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Combining Landsat observations with hydrological modelling for improved surface water monitoring of small lakes

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Abstract

1 Small reservoirs represent a critical water supply to millions of farmers
2 across semi-arid regions, but their hydrological modelling suffers from data
3 scarcity and highly variable and localised rainfall intensities. Increased avail-
4 ability of satellite imagery provide substantial opportunities but the moni-
5 toring of surface water resources is constrained by the small size and rapid
6 flood declines in small reservoirs. To overcome remote sensing and hydro-
7 logical modelling difficulties, the benefits of combining field data, numerical
8 modelling and satellite observations to monitor small reservoirs were inves-
9 tigated. Building on substantial field data, coupled daily rainfall-runoff and
10 water balance models were developed for 7 small reservoirs (1-10 ha) in semi
11 arid Tunisia over 1999-2014. Surface water observations from MNDWI clas-

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sifications on 546 Landsat TM, ETM+ and OLI sensors were used to update model outputs through an Ensemble (n=100) Kalman Filter over the 15 year period. The Ensemble Kalman Filter, providing near-real time corrections, reduced runoff errors by modulating incorrectly modelled rainfall events, while compensating for Landsat's limited temporal resolution and correcting classification outliers. Validated against long term hydrometric field data, daily volume root mean square errors (RMSE) decreased by 54% to 31 200 m³ across 7 lakes compared to the initial model forecast. The method reproduced the amplitude and timing of major floods and their decline phases, providing a valuable approach to improve hydrological monitoring (NSE increase from 0.64 up to 0.94) of flood dynamics in small water bodies. In the smallest and data-scarce lakes, higher temporal and spatial resolution time series are essential to improve monitoring accuracy.

Keywords: Remote sensing, Water balance, Rainfall-runoff model, Data assimilation, Ensemble Kalman Filter, Water harvesting

1. Introduction

1.1. Hydrology of small water bodies

Small reservoirs have developed across semi-arid areas to reduce transport of eroded soil and mobilise water resources for local users. Their reduced costs favoured significant bottom-up development, resulting in several million small reservoirs worldwide (Lehner et al., 2011). Due to their modest size and large numbers, field monitoring of small water bodies remains rare except for scientific purposes (Albergel and Rejeb, 1997), limiting their hydrological understanding.

Local studies in Sub-Saharan Africa (Desconnets et al., 1997; Martin-Rosales and Leduc, 2003), Brazil (Molle, 1991), Mexico (Avalos, 2004), India (Massuel et al., 2014b) and Tunisia (Grunberger et al., 2004; Zammouri and Feki, 2005) performed water balance modelling to quantify available resources and hydrological processes illustrated in figure 1. These exploit field measurements of rainfall, reservoir stage and pan evaporation but difficulties occur due to the uncertainties in estimating inflow, infiltration and groundwater inflow, withdrawals and lake evaporation (Li and Gowing, 2005), which must be modelled, extrapolated and/or neglected based on reasonable assumptions. Inflow due to diffuse runoff is often assessed indirectly by closing the water balance or through rainfall-runoff modelling. The latter notably suffer from the spatial variability of semi-arid rainfall regimes, leading to model performance of NSE=0.5 or less, even with site specific field data (Lacombe et al., 2008; Neppel et al., 1998; Ogilvie, 2015). Difficulties increase when seeking to upscale site specific data and model water resources in ungauged small reservoirs (Cudennec et al., 2007; Hrachowitz et al., 2013).

As a result, limited information exists on their water resources, preventing the optimisation of farming practices and local stakeholder investments (Wisser et al., 2010). Capturing runoff and favouring evaporation and infiltration, these reservoirs also modify the spatio-temporal distribution of resources. Hydrological studies have shown these can reduce downstream flows by up to 80 % in small catchments and highlighted their cumulative influence in larger catchments (Ma et al., 2010; Nyssen et al., 2010). Studies in China (Gao et al., 2011; He et al., 2003) and Tunisia (Kingumbi et al., 2007; Lacombe et al., 2008; Ogilvie et al., 2016b) on catchments over 1000 km^2

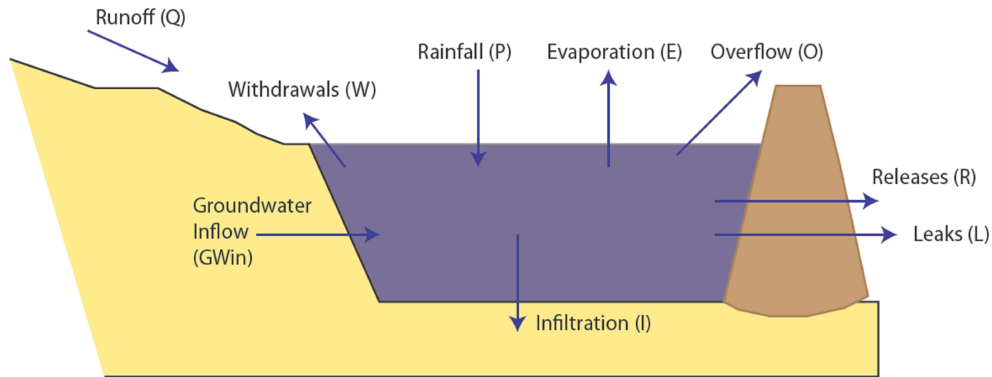


Fig. 1. Water balance fluxes in small reservoirs

59 identified reductions ranging between 1 and 50% over the same periods and
60 catchments, highlighting the uncertainties resulting in part from hydrological
61 data scarcity on small reservoirs.

62 1.2. Remote sensing and data assimilation of small water bodies

63 Satellite imagery is increasingly exploited to provide input data or to
64 calibrate hydrological models, with remotely sensed values of evaporation,
65 rainfall and soil moisture (Soti et al., 2010; Zribi et al., 2011) or assessments
66 of surface water areas (Leauthaud et al., 2013; Ogilvie et al., 2015; Swenson
67 and Wahr, 2009), lake and river stages (da Silva et al., 2014), and lake water
68 volumes (Baup et al., 2014; Crétaux et al., 2015; Frappart et al., 2018). Used
69 extensively across large wetlands, lakes or rivers, and at continental or global
70 scales, remote sensing has also been applied to provide insights across smaller
71 water bodies.

72 Studies using Landsat 30 m or pansharpened 14.5 m (Feng et al., 2016)
73 notably enabled mapping numerous water bodies and their storage capacities

(Liebe et al., 2005; Sawunyama et al., 2006). Long term Landsat time series have also recently been used to monitor surface water variations over time. Pekel et al. (2016) developed a publicly available global data set of surface water at a monthly scale over 1984-2015. Ogilvie et al. (2018) showed the benefits of a specific approach to monitor small reservoirs (< 10 ha) and account for the greater presence of flooded vegetation (Mueller et al., 2016; Yamazaki and Trigg, 2016) and difficulties resulting from limited spatial (30 m) and temporal resolution (up to 8 day from the combination of Landsat 8 and Landsat 7 satellites). These succeeded in reducing mean surface water RMSE to $9\ 300\ \text{m}^2$ (NRMSE = 24%) but the presence of clouds reduced image availability reducing the method's ability to detect rapid floods and reproduce coherent flood declines.

Data assimilation seeks to combine external sources of data or observations to beneficially correct or calibrate in real time (i.e. as observations become available) model outputs. Widely relied on in meteorology, it has become increasingly used in other scientific fields, including hydrology (Beven and Freer, 2001; Boulet et al., 2002; Clark et al., 2008; Emery et al., 2017; Moradkhani et al., 2005; Xie and Zhang, 2010) notably to combine the benefits of increasingly available and valuable (precise, accurate, higher temporal and spatial resolution) remote sensing data.

To overcome the difficulties in monitoring surface water variations in small reservoirs through hydrological modelling and satellite imagery, the benefits of combining field data, numerical modelling and remote sensing were investigated here. A daily hydrological model to simulate volumetric changes in small reservoirs combined with an Ensemble Kalman filter to reevaluate in

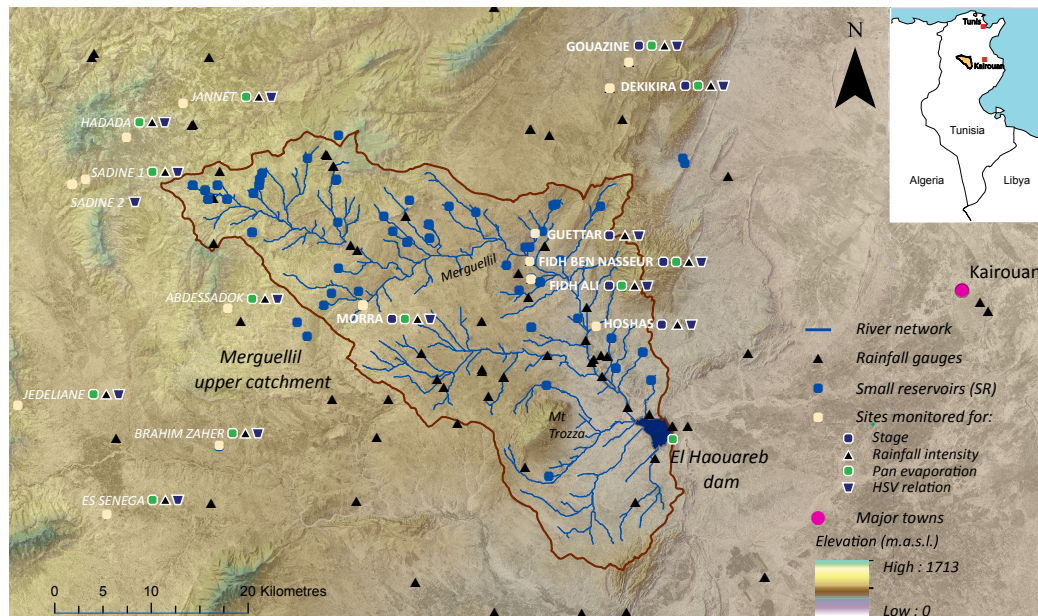


Fig. 2. Location of Merguellil upper catchment and of neighbouring hydrometeorological data used in the paper. In bold, the 7 modelled reservoirs.

99 real time model outputs based on Landsat observations was developed here.
100 The benefits of this combined model on daily values and mean annual avail-
101 ability were assessed against field data on 7 gauged reservoirs and compared
102 with results obtained using only hydrological modelling or Landsat observa-
103 tions. Finally, the sensitivity of the approach to downgrading the confidence
104 in input values and moving towards conditions found on ungauged reservoirs
105 was investigated.

106 2. Methods

107 2.1. Study sites

108 This research focussed on seven small reservoirs in semi-arid central Tunisia
109 (figure 2) benefiting from long term hydroclimatic data acquired through re-

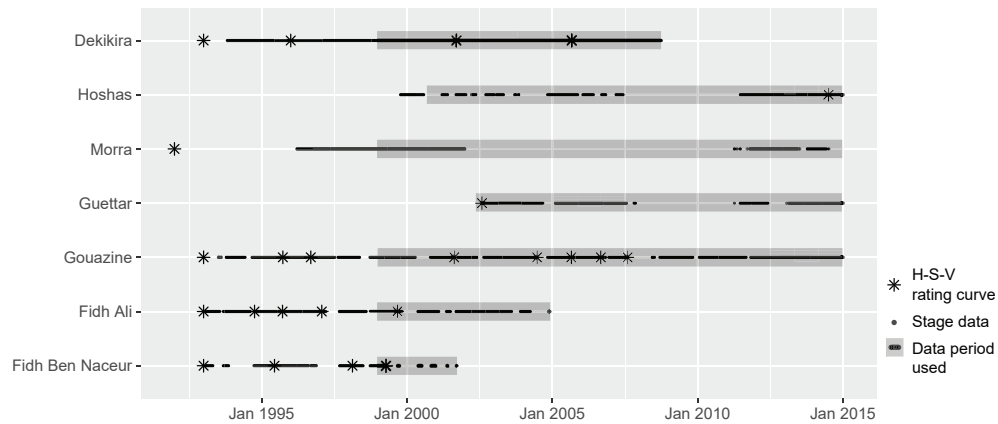


Fig. 3. Availability of stage field data and rating curves over modelling periods of the 7 lakes

110 search collaboration with government agencies (Albergel and Rejeb, 1997;
111 Leduc et al., 2007; Ogilvie, 2015). Field instrumentation on each lake con-
112 sisted of automatic stage pressure transducers and tipping bucket rainfall
113 gauges, supplemented by daily limnimetric (ladder) and rainfall readings by
114 local observers. Thirteen lakes in the vicinity had also been equipped with
115 evaporation pans. Complementary pressure transducers and automatic rain-
116 fall gauges were installed as part of this research in 2011 on three lakes
117 (Hoshas, Morra, Guettar) to extend time series (figure 3) and tend to the
118 declining monitoring network exacerbated by the Tunisian revolution.

119 Stage and surface area were converted using site specific Height-Surface-
120 Volume relations (figure 3) acquired and updated since the 1990s to account
121 for silting (Albergel and Rejeb, 1997). Complementary surveying was also
122 carried out on Hoshas in 2014. Figure 4 illustrates the shift in the rating
123 curves from silting, which can be used to assess the level of uncertainty
124 associated with volumes in recent years. On Gouazine, after 6 years (2001

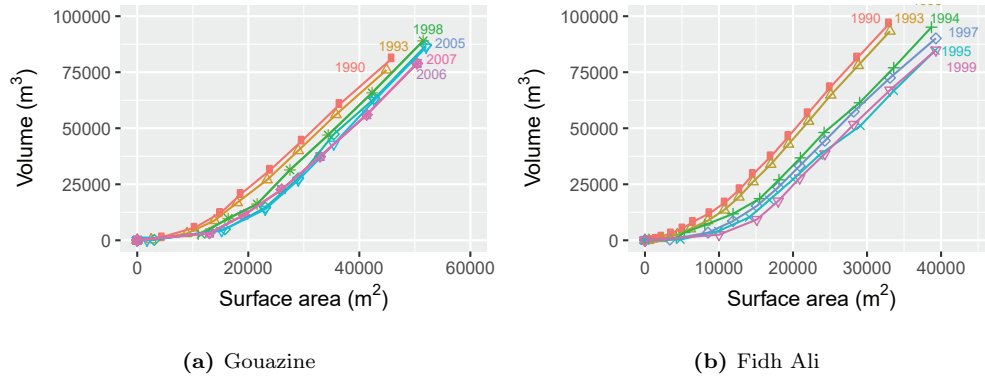


Fig. 4. Change over time of surface area - volume rating curves for two small lakes

125 vs. 2007) the obsolescence of the rating curve results in a mean RMSE
126 of 4900 m^3 , while on Fidh Ali it reaches $25\,000 \text{ m}^3$ on volumes under $80\,000 \text{ m}^3$. On lakes where rating curves could not be updated (Guettar and
127
128 Morra) for logistical reasons (cost, access to lakes and presence of water
129 and/or vegetation on lake bed), GPS contours nevertheless highlighted that
130 errors in the H-S rating curves only reached 11-12% after 12 and 22 years
131 respectively (Ogilvie et al., 2018).

132 These are inferior to errors generated from extrapolating capacity loss
133 based on studies on 15 nearby surveyed reservoirs (figure 2), due to the
134 strong disparities in silting rates and the difficulties in erosion modelling,
135 especially over extended periods (Albergel and Rejeb, 1997; Baccari et al.,
136 2008; Hentati et al., 2010; Lacombe, 2007; Ogilvie, 2015). The Gouazine
137 reservoir benefited from the longest and most reliable time series (figure
138 3) due to regular maintenance, field observations and six updates to the
139 stage-surface-volume rating curves but results on other reservoirs enabled to
140 confront the method on lakes of different capacities ranging between 50 m^3

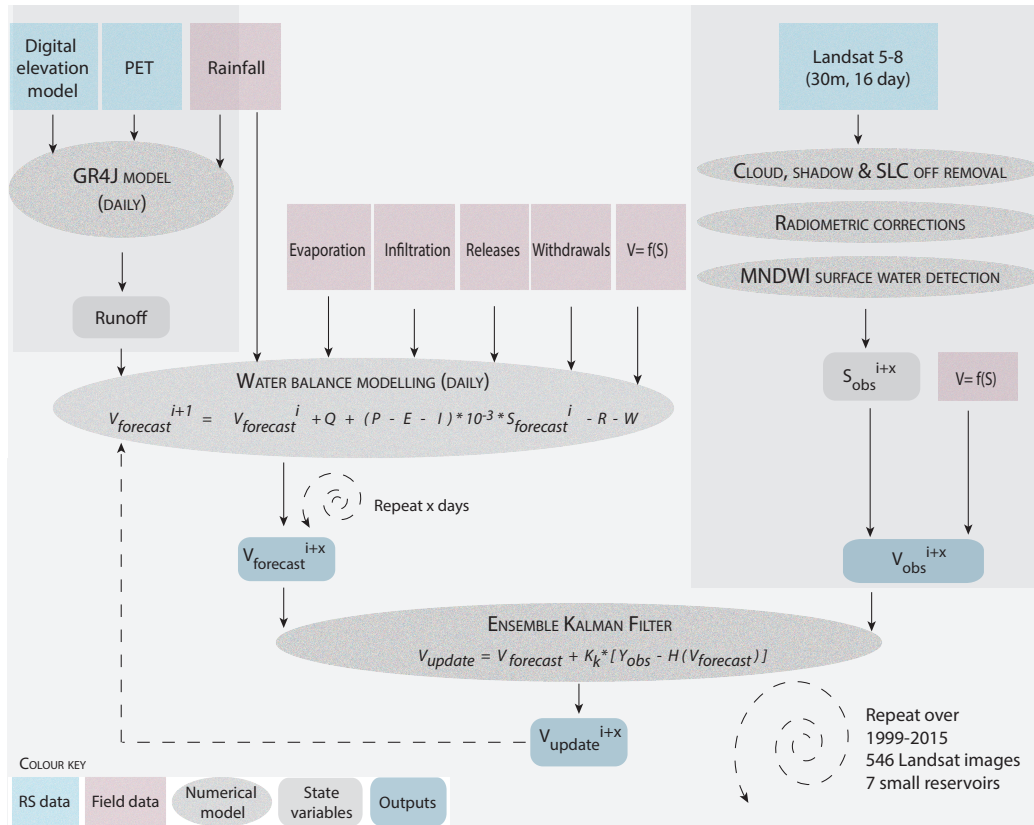


Fig. 5. Schematic representation of methodology for assimilation of Landsat observations into hydrological model with the Ensemble Kalman Filter

141 and 700 000 m³ (table 1).

142 The field data collected were used to estimate the multiple fluxes in the
143 water balance (WB) of small lakes and develop rainfall-runoff models for their
144 catchments. Site specific hydrological models were developed over 1999-2014
145 for seven lakes, before evaluating the benefits of integrating earth observation
146 data, as illustrated in figure 5.

147 2.2. Water balance modelling of 7 small water bodies

148 2.2.1. Rainfall inputs

149 Daily rainfall (P , mm/day) over the 7 lakes was interpolated over 1999-
150 2014 from the 50 manual and automatic rainfall gauges situated at the lakes
151 and within and around their catchments (figure 2). Inverse Distance Weight-
152 ing (IDW) interpolation was used after tests showed the marginal benefit
153 (error reduction by 1 mm) (Ogilvie, 2015) of geostatistical methods such
154 as Kriging with external drift (Hengl et al., 2007) compared to the lengthy
155 treatment times. Mean rainfall varied between 299 mm/year \pm 108 mm/year
156 to 396 mm/year \pm 124 mm/year (table 1). The homogeneous distribution of
157 the rainfall gauges in this catchment inherently accounts for the altitudinal
158 gradient within subcatchments (Feki et al., 2012; Ogilvie et al., 2016b; Van
159 Der Heijden and Haberlandt, 2010; Wackernagel, 2004).

160 2.2.2. Lake evaporation

161 Lake evaporation rates (E , mm/day) were IDW interpolated based on
162 field observations from Colorado type sunken pans on 13 lake shores over
163 1999-2008 (figure 2). Evaporation time series were completed to 2015 based
164 on linear regressions between each lake and a reference station with contin-
165 uous observations (El Haouareb), assuming homogeneous evaporation varia-
166 tions across the basin ($R^2 = 0.92$). Potential lake evaporation varied across
167 lakes between 1776 mm/year \pm 143 mm/year to 2019 mm/year \pm 198 mm/year
168 (table 1). A pan coefficient (C_t) of 0.8 based on water bodies of similar sizes
169 in semi-arid areas was used (Alazard et al., 2015; Cadier, 1996; Linacre, 1994;
170 McMahon et al., 2013; Molle, 1991; Riou, 1972).

171 2.2.3. Infiltration rules

172 Infiltration (I , mm/day) was modelled based on equation 1 where Z_{water}
173 is the absolute head of water (mm), a the slope, and i_0 (mm/day) the inter-
174 cept values provided in table A.1. These were extracted from Lacombe (2007)
175 and estimated for Guettar, Dekikira and Hoshas (figure 6) during depletion
176 phases (respectively 1262, 651 and 1546 days) when other fluxes are absent
177 (rainfall, runoff, withdrawals, releases) based on stage monitoring and esti-
178 mated evaporation (Lacombe, 2007; Ogilvie, 2015). Mean daily infiltration
179 varied between 2 mm and 28 mm for a lake on gravelly soil (table 1). Re-
180 cent data do not indicate a noticeable change in infiltration properties from
181 silting over time, confirming past observations (Lacombe, 2007). Similarly,
182 uncertainties from silting on the absolute head of water used in infiltration
183 rules are estimated on average at 12.5% per metre, and may be lower due
184 to partial silting of the lake floor and constant infiltration rates observed
185 on four of these lakes (Ogilvie, 2015). Groundwater and subsurface inflow
186 are often neglected in water budgets (Lacombe, 2007; Li and Gowing, 2005)
187 as these are minor fluxes and their quantification requires intense monitor-
188 ing and geochemical methods (Massuel et al., 2014b; Montoroi et al., 2002).
189 Accordingly, infiltration estimates provided here may in some cases corre-
190 spond to the combination of infiltration, leaks and groundwater inflow. On
191 Gouazine, groundwater inflow was shown to reach 50 m³/day (Grunberger
192 et al., 2004), meaning absolute infiltration may be up to 2.5 mm/day greater
193 when the lake is 2 ha and less when surface area rises (Ogilvie, 2015).

$$I = i_0 + a * Z_{water} \quad (1)$$

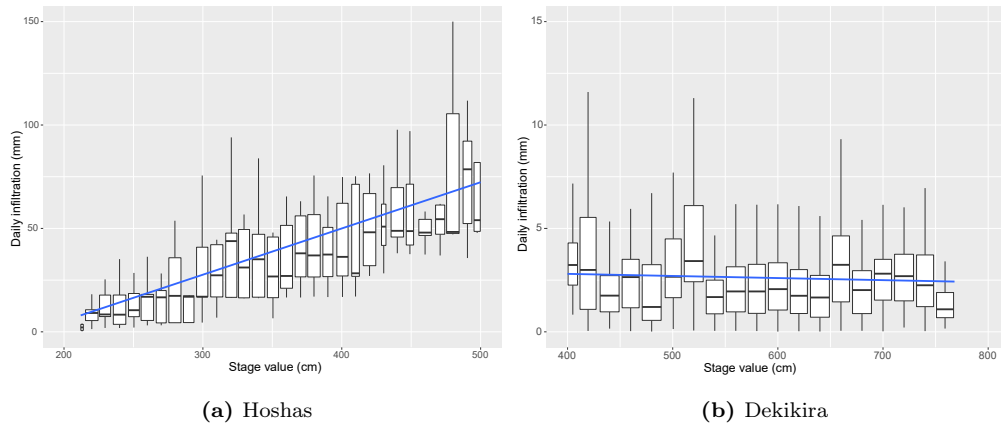


Fig. 6. Infiltration values as a function of stage in the lake estimated during depletion periods

2.2.4. Modelling releases and overflows

Semi-structured interviews with the dam operators revealed the absence of strict rules to protect the infrastructure as releases depended on further storm forecasts, government advice, presence of lakes downstream, technical problems with the valve and pressure from users to maintain maximal resources for the dry season (Ogilvie, 2015). Releases (R , m^3/day) were detected on two lakes after only the most significant events (1% of all events). Based on the extraordinary decline rates witnessed in instantaneous (15 min) hydrometric data, releases were modelled on the basis of a $10\,000\,\text{m}^3/\text{day}$ release to reach 80% of V_{max} if and when the latter is exceeded (Lacombe, 2007; Ogilvie, 2015). This also accounts for overflows through the spillway. Minor releases to flush out sediments and vegetation from the conduit were increasingly rare and remain of the order of $1000\text{--}5000\,\text{m}^3/\text{year}$.

Table 1: Characteristics of the 7 small reservoirs modelled and their catchments

Lake	Catchment size (km ²)	Altitude (m)	Initial capacity (10 ³ m ³)	Rainfall (mm/year, 1999-2014)	Evaporation (mm/year, 1999-2014)	Infiltration (mm/day, 1999-2014)
Dekikira	3.31	406	219	396	1842	2.7
Hoshas	7.90	306	130	302	2003	28
Guettar	4.98	393	150	339	1994	10
Gouazine	16.64	397	237	387	1776	9
Fidh Ali	2.74	350	134	324	2019	3.6
Fidh Ben Nasseur	1.82	368	47	327	2016	7.8
Morra	11.69	588	705	299	1917	2

2.2.5. Modelling withdrawals

Regular field visits and quantitative questionnaires with 48 farmers on 22 lakes (Ogilvie, 2015) revealed the extreme heterogeneity of pumping practices across lakes and years but highlighted the absence, or reduced importance of withdrawals (W , m³/day) on most lakes. These represented less than 40 m³/day in the summer months, compared to the 340 m³/day from infiltration (of 7 mm) and evaporation (10 mm) on a small (2 ha) surface area (Lacombe, 2007; Ogilvie, 2015). On Guettar and Morra lakes however, withdrawals to water fruit trees were estimated to reach over 130 m³/day over April to October. No withdrawal restrictions to preserve the resource as it wanes were observed and thus modelled (Ogilvie, 2015).

2.3. Runoff estimation through GR4J catchment modelling

Runoff (Q , m³/day) into small reservoirs was assessed using a daily GR4J rainfall-runoff model developed for each reservoir's catchment. This lumped conceptual model is well suited to the relative scarcity of data and used

222 across semi-arid catchments of comparable size (Perrin et al., 2003). A daily
223 time step was used to capture the intense rainfall events and corresponds
224 to the available rainfall and runoff data, as availability of sub-daily data
225 is extremely limited. It relies on a simple two reservoir structure and four
226 parameters:

- 227 • $X1$ production store capacity (mm)
- 228 • $X2$ groundwater exchange coefficient (mm/day)
- 229 • $X3$ routing store capacity (mm)
- 230 • $X4$ unit hydrograph time constant (day)

231 Input variables consist of catchment size delineated using 1 arc second
232 SRTM digital elevation model, rainfall (P , mm/day) IDW interpolated from
233 available observations across over 50 gauges (figure 2) and potential evapo-
234 transpiration (PET , mm/day) interpolated from 180 MODIS-derived 1 km²
235 monthly tiles. These MOD16 datasets exploit global weather data sets com-
236 bined with MODIS derived land cover types, leaf area index and albedo (Mu
237 et al., 2011) to provide monthly PET estimates, at a higher resolution than
238 the 0.5 ° Climate Research Unit products.

239 Models were calibrated using an objective function of maximal Nash Sut-
240cliffe Efficiency (NSE) on runoff. Q_{obs} was estimated based on stage mon-
241 itoring (figure 3) and a simplified water balance equation, as diffuse sheet
242 runoff and subsurface runoff prevent direct observations (Albergel et al.,
243 2003; Lacombe et al., 2008). Several fluxes can be neglected (groundwater

inflow, leaks) or assumed null (e.g. withdrawals) during the violent Horto-
nian runoff events resulting from limited vegetation, low soil water holding
capacities, prominent topography and high rainfall intensity characteristic
of Mediterranean climates (Lacombe et al., 2008). The other water balance
fluxes (P , E , I , releases, overflows) were assessed based on local monitoring
and observations as described previously. The airGR code (Coron et al.,
2017) which allowed for integrated numerical modelling and remote sensing
processing in R, as well as superior results thanks to the HBAN optimisation
function, was used.

2.4. Combining remote sensing observations and hydrological modelling

2.4.1. Landsat surface water observations

The remote sensing observations employed in the Ensemble Kalman Filter
were Landsat-derived surface water areas for each lake over 1999-2014. 526
Landsat 5-8 images available freely from USGS were corrected to surface
reflectance and filtered to remove acquisitions with excessive clouds, shadows
and inactive Scan Line Corrector (SLC-off) pixels over each lake. Flooded
areas were extracted using the Modified Normalised Difference Water Index
(Xu, 2006) calibrated against extensive field data. Full details of the approach
are available in Ogilvie et al. (2018) and led to a mean surface area RMSE
of 9 300 m². Surface areas were converted to volumes using the available
rating curves and values were linearly interpolated to provide a continuous
time series and allow comparisons with field data (V_{field}) and the Ensemble
Kalman Filter (V_{ENKF}) outputs. Alternate interpolation approaches (Forkel
et al., 2013) to gap fill and smooth daily time series failed here to provide
significant benefit, partly due to the abrupt fluctuations observed contrasting

269 with gradual seasonal flood pulses in larger water bodies (Leauthaud et al.,
270 2013; Ogilvie et al., 2015).

271 2.4.2. Ensemble Kalman Filter

272 Ensemble Kalman Filtering (ENKF, Evensen (2003)) is a stochastic data
273 assimilation method suited to smaller scale non-linear systems, including
274 where initial states are highly uncertain (Gillijns et al., 2006) as may be the
275 case due to poor rainfall-runoff modelling of intense rainfall events. It also
276 reduces the difficulties associated with developing a tangent linear model and
277 deriving its *adjoint* counterpart (Vermeulen and Heemink, 2006), required in
278 variational data assimilation (e.g. 3D-Var, 4D-Var), widely used in the mod-
279 elling of large systems such as atmospheric circulation models, oceanography,
280 and more recently in hydrology and hydraulics applications (Oubanas et al.,
281 2018).

282 With the Kalman filter, an initial forecast is updated using the Kalman
283 gain when an external observation is available, based on the following equa-
284 tions:

$$V_{update} = V_{forecast} + K_k * [Y_{obs} - H(V_{forecast})] \quad (2)$$

$$Y_{obs} = H(V_{obs}) + v_k \quad (3)$$

285 where K_k is the Kalman gain defined as:

$$K_k = Cy * H^T * (H * Cw * H^T + Cv)^{-1} \quad (4)$$

286 $V_{forecast}$ is here the lake volume outputted by our daily hydrological model
287 $f(V)$ with a random error w_k .

$$V_{forecast} = f(V) + w_k \quad (5)$$

288 H , called the observation operator, is the model to convert observed state
289 variables to observations. Where observations are directly inputted, as in
290 this case, $H = \text{Id}$ (identity) and equations simplify as below (equations 6
291 and 7). The external observation is the remotely sensed lake volume based
292 on Landsat imagery (V_{RS}) which have an associated random error v_k .

$$K_k = Cy * (Cw + Cv)^{-1} \quad (6)$$

$$V_{update} = V_{forecast} + K_k * [V_{obs} + v_k - V_{forecast}] \quad (7)$$

293 which can here be rewritten as:

$$V_{ENKF} = K_k * (V_{RS} + v_k) + (1 - K_k) * V_{WB+GR4J} \quad (8)$$

294 The forecast step is repeated on a daily basis and $V_{forecast}$ is updated
295 when acceptable Landsat observations are available (equation 8). The up-
296 dated volume (V_{ENKF}) is then fed back into the daily hydrological model
297 and sequentially updated over 1999-2014 with the valid remote sensing (RS)
298 observations (figure 5).

299 Cv is the observation error covariance matrix, Cw is the forecast error
300 covariance matrix and Cy is the cross covariance matrix between the state
301 variable and the forecast. As the state variable used is the volume and not

an intermediary state variable, Cy is equivalent to Cw . Cv and Cw values were estimated using the covariances of errors between stage observations and remote sensing observations and between stage and model outputs respectively. Stage related volumes include their own element of error (ladder readings, rating curve imprecisions and evolving flood bed topography) but here these are neglected compared to the errors from remote sensing (incl. radiometric corrections, detection errors) and hydrological modelling. Cw variance was 20 times greater than Cv variance and contributed to attributing greater confidence to the Landsat values over the model outputs in the Kalman filtering. Alternate combinations were tested but these did not lead to performance improvements. Cw remained constant as recommended by Clark et al. (2008), allowing the method to be used on periods and lakes with non-continuous ground truth data.

In the Ensemble version of the Kalman filter, n values of the initial state are generated and each ensemble member is run through the forecast and update step. The n values of the initial state are generated based upon a random synthetic error y so that values have mean value initial state and predefined covariance Cy . Initial states are the same as V (equation 5) and not an intermediary variable, so y was taken to be w_k , the forecast error (Moradkhani et al., 2005). The n ensemble of external observations are generated randomly to obtain a normal (Gaussian) distribution with error v_k , i.e. centred on the observation value and with predefined covariance Cv (Reichle et al., 2002). Here $n = 100$, as Gillijns et al. (2006) reveal marginal benefits above 100 and greater errors for n values below 40.

326 2.5. Performance and sensitivity of the ENKF approach

327 The performance of the Ensemble Kalman Filter (V_{ENKF}) was assessed
328 against available field data (V_{field}) and compared with the performance of
329 using only hydrological model ($V_{WB+GR4J}$) and only remote sensing (V_{RS})
330 data. NSE values were calculated but considering their sensitivity to tim-
331 ing of outputs and ability to disguise certain errors (Moussa, 2010), RMSE
332 values were provided. The performance in terms of individual daily volumes
333 was investigated as well as on annual water availability, considering their
334 importance to local users.

335 The method's performance, as inputs and parameters were degraded, was
336 then tested on four lakes to study its sensitivity and identify the ability of
337 RS observations to correct for greater uncertainties. The influence of re-
338 duced rainfall observation networks was considered, based on rainfall time
339 series interpolated after artificially removing gauges in the catchment. In-
340 formation gathered across 15 gauged reservoirs was also used to consider
341 the applicability of the approach to nearby ungauged catchments based on
342 average infiltration rules, transposing GR4J parameters and modeling an av-
343 erage surface volume power relation adapted for silting over time detailed in
344 Ogilvie et al. (2016a).

345 3. Results and discussion

346 3.1. Hydrological modelling of small water bodies

347 Figure 7 illustrates the daily volume dynamics on lake Gouazine simulated
348 by the hydrological model. Compared to the long term field observations,

349 results highlight the ability of the model to reproduce coherent flood dy-
350 namics and declines rates, for floods of varying amplitudes. Flood peaks
351 were however, in some cases, under and over estimated as in 2003 (-54%) or
352 2007 (+198%) according to field data on lake Gouazine. Difficulties occurred
353 due to the low performance of the GR4J rainfall-runoff model, where NSE
354 reached values around 0.5-0.6 on Gouazine and Dekikira, but nearer 0.2-0.3
355 on other lakes (table 2), notably on lakes with less extensive and reliable field
356 data (rainfall, stage and rating curves). Though low, these are comparable
357 to previous GR4J results in the basin (Lacombe, 2007) and due largely to
358 insufficient rainfall gauge densities which fail to capture the high intensities
359 of very localised rainfall events (Neppel et al., 1998).

360 *3.1.1. Rainfall-runoff modelling limitations*

361 To simulate the uncertainties from the absence of upstream rainfall gauges
362 in the catchments of small reservoirs, rainfall was interpolated for the Gouazine
363 catchment after artificially excluding its upstream gauge data. IDW interpo-
364 lated rainfall was then underestimated by 20.3% on 44 out of 53 events over 20
365 mm. The performance of the GR4J model decreased marginally (NSE=0.55),
366 however for the combined WB+GR4J model NSE declined from 0.57 to 0.24,
367 due to the knock on effect of errors during the flood decline (e.g. floods
368 missed in 2012 and 2014 in figure 12). Conversely, rainfall underestimation
369 forced the model during calibration to increase runoff coefficients, leading to
370 overestimation on other events which had been accurately detected due to
371 their larger spatial extent. These results highlight the importance of reli-
372 able upstream gauges to detect orographic rainfall intensities and the order
373 of magnitude of uncertainties in catchments where upstream stations are

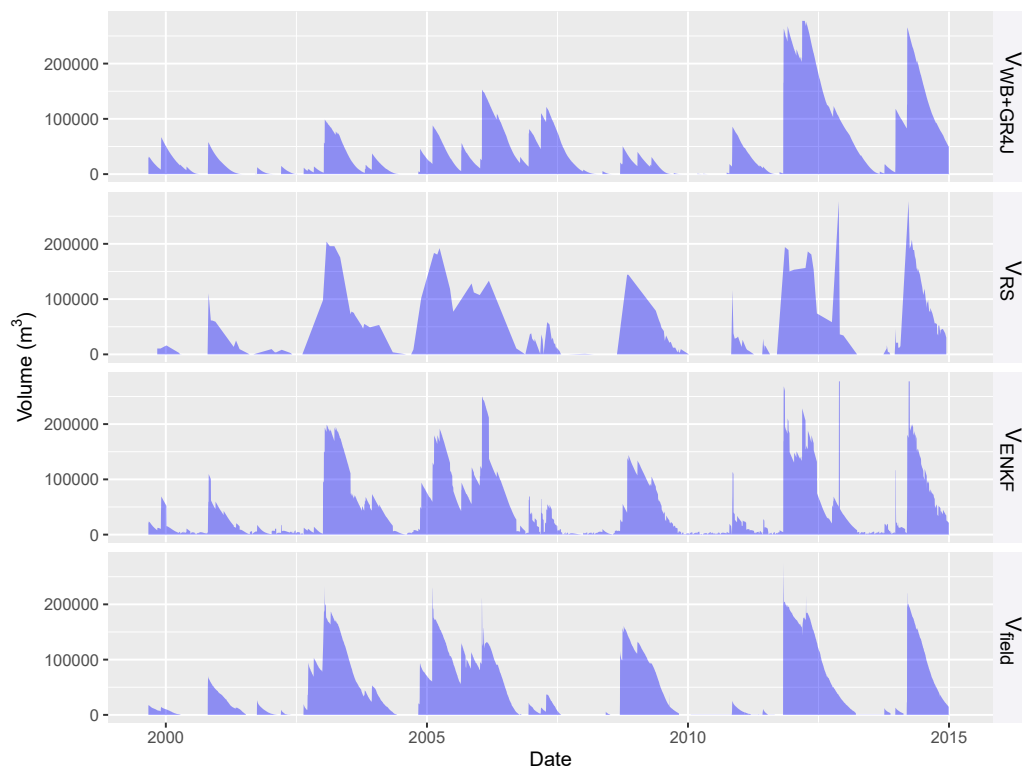


Fig. 7. Comparing observed daily volumes (V_{field}) for Gouazine lake, 1999-2014 with values obtained by the hydrological model ($V_{WB+GR4J}$), the remote sensing observations (V_{RS}) and their combination through Ensemble Kalman Filtering (V_{ENKF})

374 unavailable (i.e. all here, except Gouazine).

375 Even with an upstream station, certain events were underestimated on
376 Gouazine (in 2003, 2005, and 2009 on figure 7) due to undetected localised
377 storm cells. At the event scale, altitude variograms (e.g. KED) were not suffi-
378 cient either to correctly modulate over space the amplitude of events. Though
379 meteorological satellite observations (e.g. TRMM) do not provide reliable es-
380 timates at the event scale on such small catchments ($< 20 \text{ km}^2$), these or even
381 phone signal networks (Doumounia et al., 2014; Overeem et al., 2013) may
382 help define variograms and improve geostatistical interpolation. In larger
383 catchments or where the density of observations is greater, distributed mod-
384 els may also help account for space-time rainfall variability (Aouissi et al.,
385 2018).

386 Errors from the limited rainfall gauge density were further exacerbated
387 by inherent measurement gaps and errors due to equipment malfunctions
388 (obstructions, low maintenance) and the absence of sub-daily time series to
389 capture the flood peak accurately. Though 93% of storms over 10 mm were
390 separated by 24 hours (Lacombe, 2007), certain large events were poorly
391 modelled as substantial rains scattered over successive days, led to very high
392 runoff on the third day only, due to saturated soils and delayed subsurface
393 flows, causing calibration difficulties. Furthermore, the volume of the first
394 storms can be overestimated due to silting, and because ladders rarely moni-
395 tor the lowest stage levels, due to logistical reasons of installation and regular
396 access.

397 3.1.2. *Heterogeneous catchment responses*

398 The low GR4J performance partly translated difficulties to model the
399 catchment's response. The intensity but also land cover, antecedent soil
400 humidity or conservation works such as contour benches can significantly in-
401 fluence runoff coefficients in these catchments as discussed in (Ogilvie et al.,
402 2016b). Model parameter $X1$ notably seeks to account for the soil humidity
403 and the threshold effect, leading to greater runoff once $X1$ is saturated. The
404 lumped (i.e. not spatialised) nature of the GR4J model makes accounting
405 for localised changes in catchment behaviour (water conservation works, land
406 cover and cropping) difficult however. Model choice guided by limited data
407 availability precluded the selection of a more data intensive semi-distributed
408 and/or physical model capable of accounting for discrete changes over time
409 in land cover and land use. Changing model parameters over time can al-
410 ternatively indirectly account for this but only at the catchment scale. On
411 Gouazine, where numerous studies discuss the possible reduction in runoff
412 from the development of contour benches on 43% of its catchment area (Nasri,
413 2007), calibrating over 1997-2003 led to a routing store capacity ($X3$ param-
414 eter) 5 times greater than over the whole period, possibly pointing to the
415 greater retention capacity from water soil and conservation works. Model
416 performance improved (NSE rose to 0.67) but only marginally as it remained
417 affected by the other difficulties discussed above.

418 3.2. *Combining remote sensing and hydrological modelling*

419 3.2.1. *Ensemble Kalman filter performance on daily volumes*

420 Figure 7 compares the daily volume dynamics on lake Gouazine based on
421 outputs from the hydrological model, the remote sensing observations, and

422 their combination through the Ensemble Kalman filter. Remotely sensed
423 volumes provided greater accuracy in the estimations of flood peaks than
424 the hydrological model however outliers remained present (e.g. 2006 and
425 2013). Furthermore, the low frequency of acceptable observations (on average
426 1.5/month) led to poor representation of the rapid flood rises as in 2003
427 (Ogilvie et al., 2018).

428 The Ensemble Kalman filter improved the performance of the site-specific
429 hydrological models, with Landsat observations notably modulating the ini-
430 tial $V_{WB+GR4J}$ forecast and usefully correcting the flood peaks under and
431 overestimated by the model (figure 7). These errors were carried through the
432 decline phase of the hydrological models and figure 8 clearly illustrates the
433 correction from the satellite observation which draws volumes closer to the
434 1:1 line, raising the NSE value, for instance from 0.57 to 0.81 on Gouazine.
435 This effect was pronounced on larger lakes that do not dry out, as overes-
436 timations in the model outputs led to a progressive drift, which the ENKF
437 usefully corrected (figure 9).

438 Accordingly, RMSE (table 2) reduced thanks to the Landsat corrections
439 on 5 of the lakes (Dekikira, Gouazine, Fidh Ali, Morra and Guettar). Mean
440 RSME reduced by 54% to 31 200 m³ across all lakes and 21 400 m³ when
441 excluding the much larger dam (Morra). Compared to the range of flood val-
442 ues experienced by these lakes, NRMSE reached an acceptable 0.26. Greater
443 errors were observed due in part to reduced model performance, preponder-
444 ant remote sensing uncertainties (e.g. Hoshas) and less reliable hydrometric
445 field data (HSV on Morra and Guettar). The lower NSE on the smallest
446 reservoirs (Hoshas, Fidh Ben Nasseur) were to be expected here considering

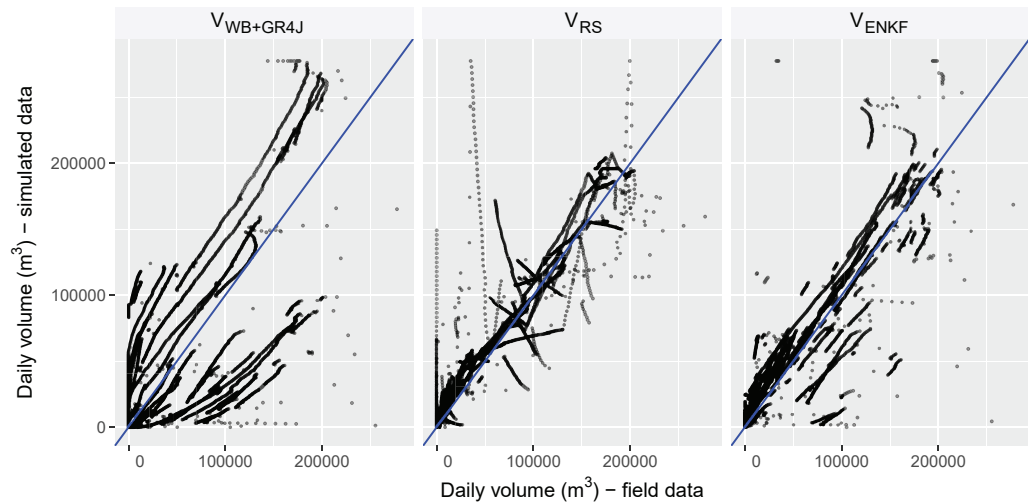


Fig. 8. Scatterplot between modelled and observed daily volumes on lake Gouazine, 1999-2014

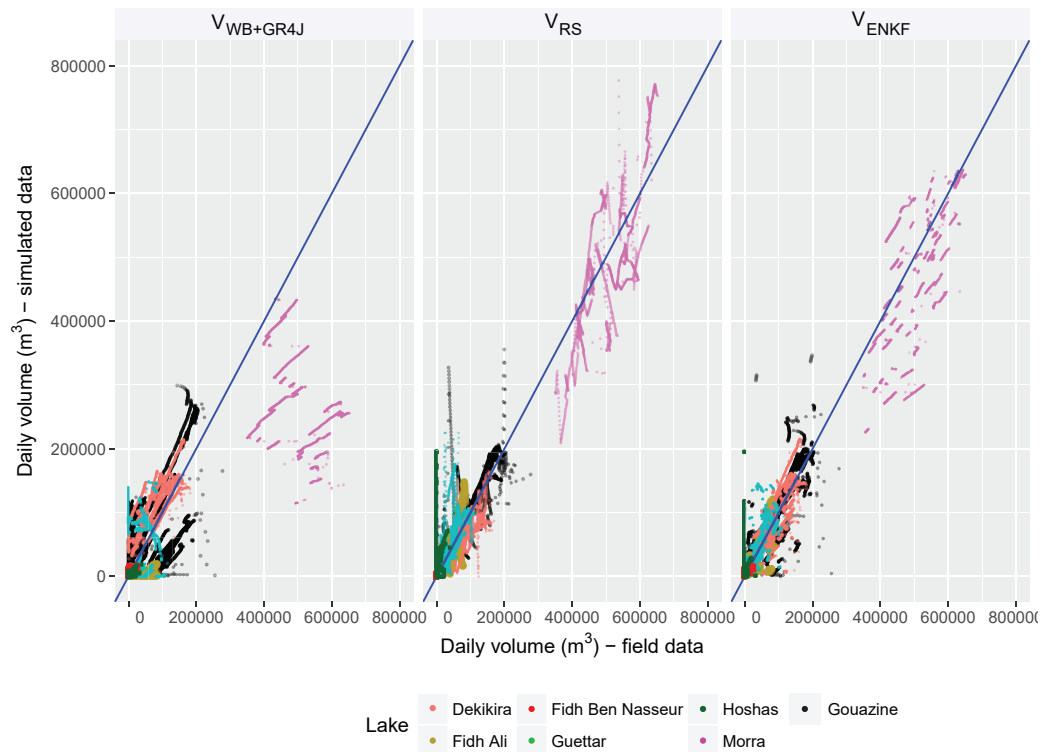


Fig. 9. Scatterplot between modelled and observed daily volumes for all 7 lakes, 1999-2014

the spatial resolution (30 m) of satellite imagery used here and the mean flooded surface area around 1000 m² (Ogilvie et al., 2018).

Remote sensing observations are capable of representing flood dynamics with low RMSE but suffer from overclassifications due to undetected clouds and from the reduced temporal resolution of Landsat imagery (on average 1.5 image/month due to clouds) (Ogilvie et al., 2018). The ENKF approach developed here enabled remote sensing outliers to be rapidly corrected here, as seen on Gouazine in 2012 (figures 7 and 8) and Morra (figure 9) for instance. The combination with rainfall-runoff modelling also reduced interpolation errors resulting from insufficient observations close to the flood peak as seen on figure 7. Similarly, the ENKF also helped identify additional flood peaks as in 2006. Over long periods, ENKF led to a reduction of RMSE near 10% on Dekikira and Guettar. Large errors in the initial forecast led to marginally higher RMSE with ENKF than V_{RS} on some lakes. However, as seen in figure 7, the ENKF approach enabled a more coherent and accurate reproduction of daily flood dynamics even on these lakes. Over a single hydrological year, the reduction in RMSE from ENKF over interpolated remote sensing observations also reached up to 46% on Gouazine.

3.2.2. Ensemble Kalman Filter performance on annual water availability

The method's performance in assessing annual water availability rather than fine flood dynamics (i.e. individual observations) is shown in figures 10 and 11, and summarised in table 3. The ENKF method improved on the initial $V_{WB+GR4J}$ results (NSE=0.62), except on the smallest lakes (Hoshas and Fidh Ben Nasseur). Nevertheless, the orders of magnitude of the ENKF estimated volumes on Hoshas (figure 10) remain correct in comparison to



Fig. 10. Modelled and observed mean daily water volumes per year for all 7 lakes. Years with no field observations between 1999-2014 were excluded here.

Table 2: Ensemble Kalman Filter performance on daily volumes

Lakes (modelled period)	Initial capacity (10 ³ m ³)	NSE			RMSE (m ³)			NRMSE
		$V_{WB+GRAJ}$	V_{RS}	V_{ENKF}	$V_{WB+GRAJ}$	V_{RS}	V_{ENKF}	V_{ENKF}
Gouazine (1999-2014)	237	0.57	0.84	0.81	45200	25300	25900	0.09
Dekikira (1999-2008)	219	0.69	0.73	0.78	44000	25800	23700	0.13
Fidh Ali (1999-2005)	134	0.17	0.70	0.55	39200	20900	20900	0.24
Fidh Ben Nasseur (1999-2001)	47	0.45	0.11	0.44	6500	1500	6600	0.21
Morra (1999-2014)	705	0.12	0.62	0.46	274300	76400	90000	0.30
Hoshas (2001-2014)	130	0.48	0.02	0.02	3000	23400	23100	0.56
Guettar (2003-2014)	150	0.18	0.50	0.49	62500	31800	28300	0.25

much larger volumes on nearby lakes. Again, by modelling the decline between two Landsat observation and reducing certain outliers, the ENKF also improved upon V_{RS} on certain lakes (e.g. Gouazine, Dekikira) but on others, the poor initial forecast degraded the ENKF performance (e.g. Guettar and Morra). Overall, ENKF displayed superior results than on individual values due to the annual smoothing of observations, leading to very high levels of NSE (0.99 across all lakes) and a mean RMSE (excluding the larger Morra dam) reduced here to 10 500 m³.

On Hoshas, $V_{WB+GRAJ}$ continued to perform better than V_{ENKF} despite underestimating all events, due to the small and short floods experienced which lead to a drastic, incorrect increase in water availability from single remote sensing outliers. These were removed here through cloud and shadow filtering and capping volume outputs to the known maximum capacities, however residual outliers due to undetected cirrus clouds or shadows remain.

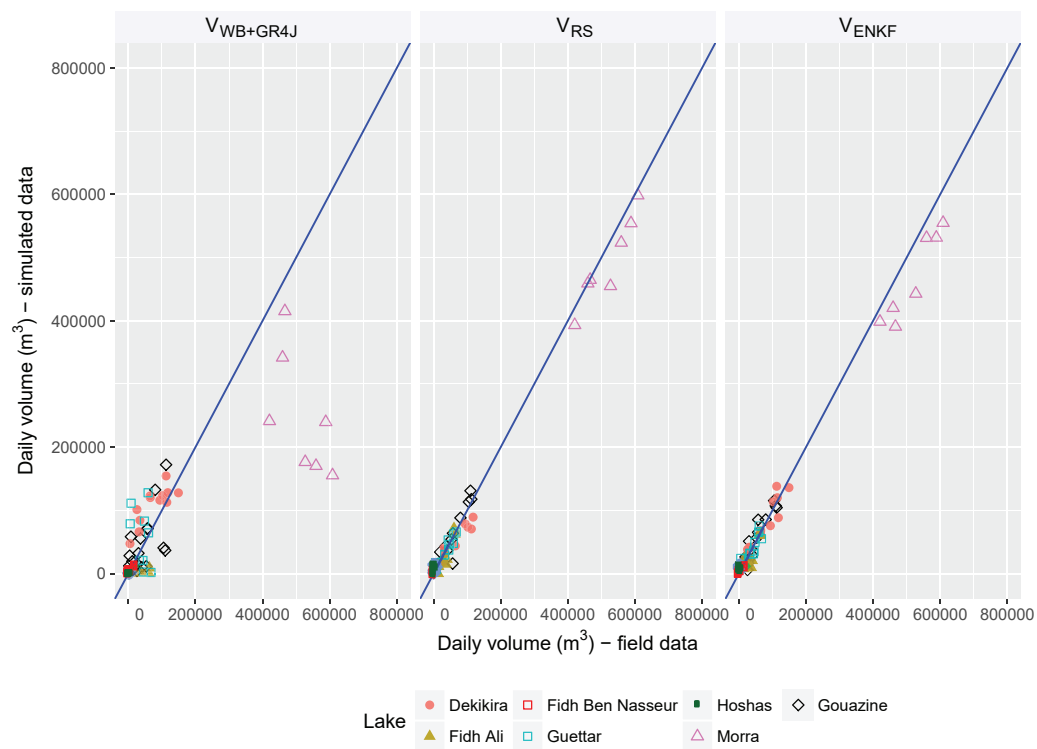


Fig. 11. Scatterplot between modelled and observed mean daily volumes per year for 7 lakes, 1999-2014

486 Improvements in the way clouds (especially cirrus clouds) are detected as well
487 as increased temporal and spatial accuracy will help reduce remote sensing
488 errors. Higher spatial resolution will increase precision, while more frequent
489 images will allow outliers to be corrected faster, reducing water availability
490 errors which depend on the lag between subsequent correct observations.
491 Improvements in the method may also be gained by defining specific Kalman
492 gain values for each RS observation to reflect for the presence of clouds at the
493 image level and the associated greater uncertainty over specific observations.
494 Interestingly, remote sensing uncertainties affected the smaller lakes (e.g.
495 Hoshas) where errors are proportionally more important but also lakes with
496 limited variation in surface area. On Morra for instance, the variations are
497 contained within the % error of surface area estimates from our MNDWI
498 method. Accordingly, on Morra ENKF outputs for individual observations
499 were heavily affected (NSE=0.46), but mean annual availability performed
500 well (NSE=0.89).

501 *3.3. Ensemble Kalman filter performance as data uncertainties rise*

502 Figure 12 and table 4 illustrate the difficulties in modelling daily flood
503 dynamics as uncertainties in the data inputs rise. In the absence of upstream
504 rainfall gauges, the performance of the hydrological model degraded (cf. sec-
505 tion 3.1.1) and RMSE rose by 28%. The ENKF however continues to improve
506 performance and correct for these errors, with NSE on daily volumes reduc-
507 ing marginally from 0.81 to 0.75. RMSE values for the daily observations
508 increase by 21% but remain 46 % lower than the initial WB+GR4J forecast
509 thanks to the remote sensing corrections. Using an average locally derived
510 infiltration rule based on 13 small reservoirs (Ogilvie, 2015) prevented the

Table 3: Ensemble Kalman Filter performance on mean annual water availability

Lakes (modelled period)	Mean daily volume (m ³)	NSE			RMSE (m ³)		
		$V_{WB+GRAJ}$	V_{RS}	V_{ENKF}	$V_{WB+GRAJ}$	V_{RS}	V_{ENKF}
Gouazine (1999-2014)	42800	0.38	0.87	0.89	36700	13500	11500
Dekikira (1999-2008)	59000	0.65	0.71	0.89	41600	20500	14500
Fidh Ali (1999-2005)	32200	0.09	0.52	0.60	35300	13500	12400
Fidh Ben Nasseur (1999-2001)	1000	0.66	0.67	NA	4300	2400	4400
Morra (1999-2014)	448900	0.40	0.89	0.89	304500	31700	55700
Hoshas (2001-2014)	800	0.37	0.11	0.09	2400	8700	9500
Guettar (2003-2014)	29000	0.09	0.88	0.74	56500	7100	10700

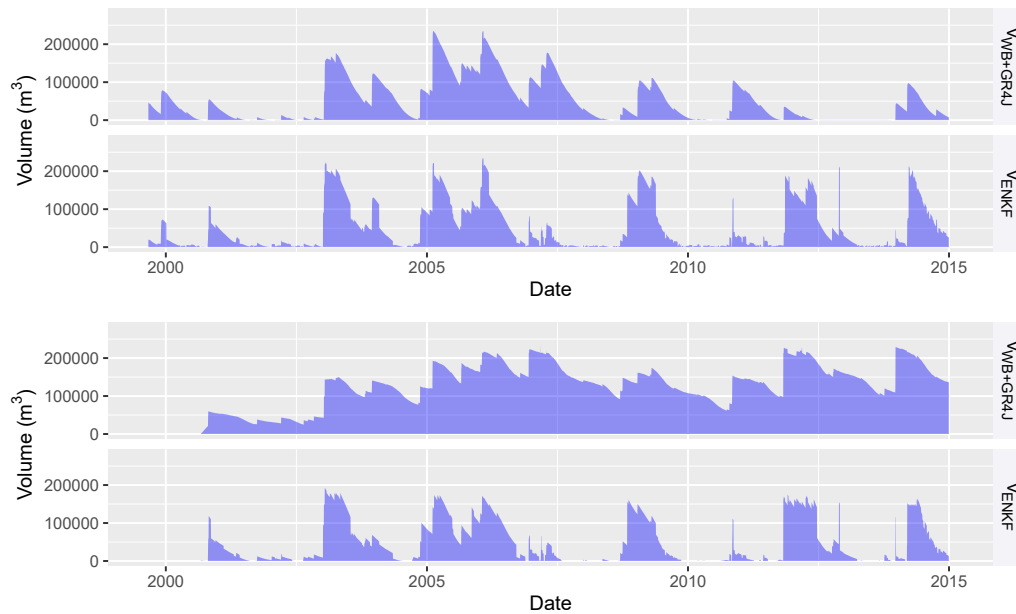


Fig. 12. Modelled daily volumes on Gouazine lake when degrading inputs. Top: with no rainfall gauge in the catchment. Bottom: with the average infiltration value from 15 reservoirs

511 model on Gouazine to reproduce the emptying of the lakes between succes-
512 sive events. This lead to a rising drift in volumes and RMSE values of the
513 $V_{WB+GR4J}$ initial forecast rising drastically to 97 000 m³. Again, the Kalman
514 filter using Landsat observations provided valuable corrections and RMSE
515 values on individual observations remained close (+ 15%) to those with the
516 site specific model. As the confidence in the model inputs & parameters (on
517 $I, P,$) degrades or significant additional fluxes can not be modelled reliably
518 (releases, withdrawals), the benefit of assimilation with remote sensing ob-
519 servations as expected increases. However the benefit of V_{ENKF} over simply
520 exploiting interpolated V_{RS} values also declines, due to the initial forecast
521 becoming so uncertain. RMSE on V_{RS} remains on average 18% lower than
522 V_{ENKF} in these four examples (table 4).

523 When considering ungauged catchments with no locally calibrated GR4J
524 parameters and no site specific HSV relation, the Kalman gain continues to
525 valuably correct the hydrological model's initial forecast, reducing RMSE by
526 30% (table 4). The increase in RMSE for V_{ENKF} is however amplified by
527 the uncertainties in V_{RS} , due to the surface-volume power relations used. A
528 locally derived power relation was shown to increase errors to near 40% on
529 Dekikira due to the difficulty in accounting for local lake morphologies and
530 the abrupt changes from silting (Ogilvie et al., 2016a). New techniques based
531 on high spatial resolution sensors open up increased possibilities to acquire at
532 lower costs (time, equipment) sufficient topographic detail to render surface
533 volumes rating curves (Avisse et al., 2017; Baup et al., 2014; van Bemmelen
534 et al., 2016; Massuel et al., 2014a) of multiple reservoirs of different geo-
535 morphology. Similarly, data assimilation with Landsat observations could be

Table 4: Kalman Filter performance on daily volumes when degrading model inputs and parameters

Lake	Degraded inputs	RMSE (m ³ /day)			RMSE increase (%)		
		$V_{WB+GRAJ}$	V_{RS}	V_{ENKF}	$V_{WB+GRAJ}$	V_{RS}	V_{ENKF}
Gouazine	Rainfall data	57800	25300	31400	+28%	+0%	+21%
Gouazine	Infiltration data	97000	25300	29800	+115%	+0%	+15%
Dekikira	GR4J parameters and HSV	52800	35800	38500	+20%	+39%	+63%
Fidh Ali	GR4J parameters and HSV	43000	20100	29000	+10%	-3%	+39%

used to calibrate over time the GR4J parameters, notably $X1$, based on the estimated runoff. This approach was not explored here due to the rainfall uncertainties observed at this sub basin scale and the temporal resolution of Landsat imagery, which would lead to incorrect quantification of daily runoff and thus calibration of the parameters.

4. Conclusions

Landsat surface water estimates coupled with an Ensemble Kalman Filter showed their potential to improve hydrological modelling of small reservoirs. Remote sensing observations provided vital corrections to the flood amplitudes incorrectly estimated by the GR4J model which suffered notably from rainfall detection issues. Conversely, site specific rules on depletion fluxes (infiltration, withdrawals, etc.) led to an accurate modelling of the flood decline, improving over interpolated Landsat observations, limited by reduced temporal resolution. Overall performance reached high skill levels (NSE rose from 0.64 to 0.94 on daily values) and RMSE reduced by two thirds down to 10 500 m³ when considering annual water availability.

552 Uncertainties from limited data availability (rainfall, infiltration, stage
553 data to calibrate P-Q models) were seen to increase the benefit of the ENKF
554 approach, but can also degrade the hydrological model to a point where it
555 becomes preferable to rely exclusively on interpolated Landsat surface area
556 observations. These performed well except on the smallest lakes, coherent
557 with the medium resolution imagery used here, and due to certain outliers
558 whose interpolation can reduce skill values over short periods. Time series
559 from the new generation of high temporal and spatial resolution satellite
560 imagery (e.g. Sentinel-2) are expected to further improve the accuracy of
561 remote sensing and associated data assimilation approaches on these smaller
562 reservoirs.

563 The Kalman filter approach may also be varied to seek to correct not
564 individual observations, but rather to estimate model inputs (e.g. rainfall) or
565 model parameters. This could notably be developed to improve hydrological
566 models on ungauged lakes, but would require frequent satellite observations,
567 close to flood peaks to provide sufficient accuracy in estimating daily runoff.
568 Similarly, over decline phases, where sufficient confidence in infiltration and
569 evaporation exists, the remote sensing observations could be used to identify
570 withdrawal rates. The ENKF method may also be enhanced by fine tuning
571 (Moradkhani et al., 2005) the covariances to compose with both sources of
572 uncertainty and provide greater confidence to remote sensing observations
573 over field data based on additional criteria (lake size, cloud presence across
574 image, etc.).

575 By drastically improving the performance of hydrological modelling in
576 data scarce semi-arid catchments, the Ensemble Kalman filter may improve

577 local water availability assessments (Wisser et al., 2010) but also provides
578 much needed data on the runoff captured by multiple reservoirs. These may
579 then serve as multiple runoff gauges to be integrated into larger scale models
580 (Gal et al., 2016; Liebe et al., 2009) and feed into the growing discussions
581 over their influence for downstream water users and uses.

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591 Appendix A. Supplementary materials

Table A.1: Infiltration values (mm/day) for small reservoirs in and around the Merguellil upper catchment. Values for Fidh Ali, Fidh Ben Nasseur and Morra were adapted from Lacombe (2007)

Lake	Mean infiltration	Infiltration for Z_{min} (i_0)	Infiltration for Z_{max}	Infiltration rise per m ($a \cdot 1000$)
Dekikira	2.7	2.70	2.7	0
Hoshas	28	3.62	77.1	24.50
Guettar	10	10.00	10.0	0
Gouazine	9	13.00	7.5	-1.38
Fidh Ali	3.6	3.60	3.6	0
Fidh Ben Nasseur	7.8	3.06	12.5	3.14
Morra	2	1.48	2.5	0.53

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