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Do biodiversity offsets achieve No Net Loss? An evaluation of offsets in a French department

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Key words

Biodiversity offsets, no net loss, monitoring, mitigation hierarchy, offset efficiency

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1 **Do biodiversity offsets achieve No Net Loss? An evaluation of offsets in a French**
2 **department**

3
4 **Abstract**

5 Biodiversity offsetting is a policy approach that compensates for the ecological losses from
6 development projects affecting biodiversity with equivalent gains through offsets, aiming at
7 “No Net Loss” (NNL). Although offsets seem appealing in theory, several concerns have
8 been raised about the difficulties reaching NNL in practice. While most of the discussion
9 about offsets improvement is based on principles and strategies, we evaluated empirical
10 evidence of offsets implemented, both from the procedure files (protected species and
11 wetlands) and field surveys. Our objective was to evaluate whether offsets achieve NNL
12 based on 91 procedure files in the Isère department in France. We identified that necessary
13 data for assessing offsets gains, such as the location and offset sites’ initial state, were not
14 available in part (location) or all (initial state) procedure files investigated. We evaluated 59
15 offsets implemented for 22 development projects and where minimum data for monitoring
16 offsets were available; we surveyed the presence or absence of target species and habitat
17 from the offset site. The type of offsets (restoration, creation or maintenance of target habitat)
18 was one of the characteristics that helped to explain both species and habitat absence,
19 implying offset failure. Based on our analysis, we suggest three principal angles for
20 progressing in NNL achievement: (i) collecting and publishing a set of essential information
21 on offsets, (ii) requiring a management plan for each offset, and (iii) accumulating empirical
22 evidence of offsets failure and success.

23 **1. Introduction**

24

25 Biodiversity offsets have become a widespread conservation policy tool (McKenney &
26 Kiesecker 2010) in response to the massive biodiversity loss caused by human activities
27 (Naeem *et al.* 2012). Offsets aim to compensate for unavoidable negative impacts from
28 development projects on biodiversity. The overall result should be a “No Net Loss” (NNL) of
29 biodiversity (BBOP 2012), or even a net gain (but see Bull & Brownlie 2015). This NNL
30 objective is set to stop biodiversity erosion. To reach this goal, offsets must be actions (such
31 as ecosystem restoration) that provide gains of biodiversity equivalent to related losses due
32 to impacts (Quétier & Lavorel 2011).

33 Although offsets seem appealing in theory, several concerns have been raised about the
34 difficulties implementing offsets and achieving NNL (Bull *et al.* 2013; Gibbons & Lindenmayer
35 2007; Quetier *et al.* 2014). Firstly, the term “NNL” has various meanings depending on the
36 national policy and is applied only on a few biodiversity features (Maron *et al.* 2018). In the
37 European offset policy, NNL is applied for protected species and habitats (Regnery *et al.*
38 2013). In addition, in France legislation requires that losses of wetlands be compensated
39 according to a specific procedure (Bezombes *et al.* 2018). Secondly, offset implementation
40 and monitoring by developers and authorities face many uncertainties regarding (i) the
41 measurement of losses and gains of the targeted biodiversity features (Moreno-Mateos *et al.*
42 2015) (ii) the time until the expected gains are provided by offsets (Curran *et al.* 2013) and
43 (ii) the risk that offsets fail to provide the expected gains (Maron *et al.* 2012).

44 Most of the discussion about offsets improvement is based on the principles and strategies of
45 offsetting (Bull *et al.* 2013; Gibbons & Lindenmayer 2007; Quetier *et al.* 2014). Only little
46 empirical evidence of whether offsets actually succeed in achieving NNL exists (Quigley &
47 Harper 2006b; May *et al.* 2017) and even less is validated by field monitoring (Pickett *et al.*
48 2013; Lindenmayer *et al.* 2017). In this paper we provide an evaluation of offsets
49 implemented in the Isère department in France, with empirical evidence from the procedures
50 files and field surveys.

51 According to the French law on biodiversity no. 2016-1087, developers must demonstrate
52 that their offsets result in NNL of the targeted biodiversity features (protected species and
53 habitats including wetlands). This implies measuring that gains (i) are equivalent in quality
54 and quantity compared to losses (Bezombes *et al.* 2017), (ii) are only attributable to the
55 offset actions (Bull *et al.* 2014) and (iii) would not have happened otherwise (Maron *et al.*
56 2013). For all the above, and in particular for (i), monitoring is necessary, but offset
57 monitoring is still very heterogeneous in France (Quetier *et al.* 2014).

58 To evaluate the efficiency of offsets in achieving NNL, we first analysed the availability of
59 data necessary to measure biodiversity gains in the procedure files of the Isère department

60 of France. Based on available data, we surveyed for each offset whether the target
61 biodiversity features were present or absent from the offset site after offset implementation.
62 Although the presence of the target biodiversity features alone is not sufficient to measure
63 gains, their absence is considered as an indicator of failure to achieve NNL. Finally, we
64 analysed the characteristic of the offsets where target biodiversity features were absent and
65 provide perspectives for better monitoring and NNL achievement.

66 **2. Material and methods**

67 2.1. Minimum information, study area and time span covered

68 Prior to the field survey and characterisation of offsets, we elaborated a list of minimum
69 information necessary to include the offsets in our analysis: the targeted biodiversity
70 features, offset location, type of offsets and contact details of the manager in charge of
71 monitoring. We also identified that the information necessary to determine whether offsets
72 provided gains was the offset site's initial state (before offset implementation) in quality
73 (presence or absence of target habitat and/or species) and quantity (surface for habitats and
74 number of individuals for species).

75 We reviewed all procedure files available in the archives of the Isère department
76 administration in charge of the procedure directive. Only the files fulfilling the requirements
77 for the minimum information were selected. In these files, all projects required offsets for
78 protected species, wetlands or both following the designated procedures (Bezombes *et al.*
79 2018), and all were approved by the Isère department's prefect. Offsets were implemented
80 between 2001 and 2015 on sites that were all located in the Isère department in South-East
81 France (Figure 1). The projects were mainly situated around urban areas and along
82 motorways, in areas dominated by agriculture and small natural patches.

83 Each offset site was surveyed between spring and summer 2016 during a 1-day visit.
84 Managers in charge of monitoring were always present to allow access to the sites and
85 provided complementary information on offsets that were not available in the files. During the
86 survey, we visited all offsets implemented and assessed the presence or absence of the
87 protected species and their habitat (protected species procedure), or wetland vegetation and
88 the related fauna (wetland procedure). The target habitat was considered present when the
89 necessary hydrologic conditions (for wetlands) and vegetation structure (for wetlands and
90 species habitat) were observed. When the vegetation was not mature, the habitat was
91 considered absent. The target species was considered present when individuals or tracks of
92 individuals (nest, egg-laying) were observed.

93 2.2. Offset characteristics influencing the presence or absence of the target species
94 and habitats

95 To analyse whether certain offset characteristics influence the presence or absence of the
96 target species and habitats, we described in Table 1 the offsets with eight characteristics:
97 (1) the offset procedure, (2) the type of offsets, (3) the type of developer, (4) the distance
98 between the impacted and the offset site, (5) the distance between the offset site the nearest
99 landscape ecological corridor, (6) the time since offset implementation, (7) the number of
100 offsets implemented on the site and (8) five variables for the target species group (birds,
101 reptiles, amphibians, mammals, and “others” including bats, insects and flora (see Table 1
102 for the reasons for these choices).

103 We classified offsets into five types, from the least to the most intensive actions and
104 uncertainties about the outcomes (Tischew *et al.* 2010): maintenance of a favourable habitat;
105 management of a terrestrial ecosystem; restoration of a wetland or river; and creation of
106 linear plantation, micro-habitat or pond (see Appendix 1 for details). The distance between
107 offset sites and either impacted sites or the nearest ecological corridor was calculated as the
108 shortest distance from one perimeter to another.

109 2.3. Data analysis

110 We assessed the influence of the offset characteristics on the presence or absence of (i) the
111 target habitat and (ii) the target species at the offset site using a generalised linear mixed
112 model (GLMM; R package MASS, Ripley *et al.* 2013). Due to the nature of our response
113 variables (i.e. the presence/absence of the target species or the habitat) we used a binomial
114 error distribution. The hierarchical structure of our sampling design (59 offsets allocated on
115 30 offset sites), imposed that we treated the variable “offset site” as a random effect, while
116 considering the other explanatory variables as fixed effects (Zuur *et al.* 2009).

117 Following a multi model inference (Burnham *et al.* 2011; Grueber *et al.* 2011) we generated a
118 set of candidate models containing all possible variable combinations and ranked them by
119 corrected Akaike information criterion (AICc) using the dredge function (R package MuMIn;
120 Barton 2015). We only integrated the models complying with the following conditions: (i)
121 models do not include simultaneously correlated covariates ($R^2 > 0.7$) and (ii) models do not
122 include more than three variables to avoid over-parameterisation due to the limited data set.
123 This resulted in a total model set of 251 models for both targeted habitat and targeted
124 species (see Appendix 2 and 3). We restricted this set of models using a cut-off of 3.8 AICc
125 for targeted habitat and 7 AICc for targeted species corresponding to the AICc values of a
126 model with no predictor variables. We finally obtained a total of 45 and 31 top models,
127 respectively (see Appendix 2 and 3).

128 **3. Results**

129 3.1. Missing data necessary to monitor offsets

130 Out of the 91 offset procedure files reviewed, only 33% contained the minimum information
131 necessary to include the offsets in our analysis. Among the 30 projects selected, eight had to
132 be excluded because the offset site manager did not respond to our inquiry. Finally, 22
133 procedures files were included in the analysis (see Table 2 for details on the projects). Since
134 between one and five offsets were implemented on one or two sites per project (Table 2), a
135 total of 59 offsets were surveyed.

136 Nonetheless, information on the offset sites' initial state (both in quality and quantity) was
137 absent from all 22 projects selected. Therefore, we were not able to verify whether actual
138 gains were provided by the offsets. Only offset failure (absence of the target species and
139 habitat on the offset site after offsets were implemented) could be analysed.

140 In addition, some data expected to be found due to regulatory requirements were absent.
141 Indeed, for seven out of 22 projects (33.9% of the offsets surveyed) the impacted site area
142 was not found in the files. For six out 22 projects (23.7% of the offsets surveyed), no mention
143 of a management plan was found in the files (even though an offset manager was in charge
144 of the offsets).

145 3.2. Offset characteristics influencing the presence or absence of the target species
146 and habitats

147 The target habitats were found absent for 27.1% of the offsets and target species absence
148 represented for 39% of the 59 offsets surveyed in 2016 (i.e. 1–15 years after offset
149 implementation depending on the project; Table 2).

150 In the set of the top models for the target habitat (Table 3), only the type of offset exhibited
151 notable relative importance (0.73, see Appendix 2) as it appeared in every top models. The
152 type of offset influenced the occurrence of target habitats, which was greater when offsets
153 consisted in creation of a micro-habitat and maintenance of the already favourable habitat
154 (Figure 2a).

155 In the set of the top models for the target species (Table 3), the type of offset and the number
156 of offsets were both variables with the greatest sum of the Akaike weights (respectively, 0.96
157 and 0.73, see Appendix 3). Note, however, that the confidence interval for the estimate of the
158 number of offsets on the offset site (conditional average $\beta = 7.22 \pm 13.45$ SE) includes zero,
159 so there is little evidence in the data set that this affects the occurrence of target species.
160 However, the type of offset influenced the occurrence of the target species, which was lowest
161 when offsets consisted in management of a terrestrial ecosystem (Figure 2b).

162 **4. Discussion**

163 4.1. Information gaps when evaluating NNL

164 By evaluating offset implementation in the Isère department in France, we showed that the
165 implementation of offsets lags behind the regulatory requirements of NNL policy (Gibbons &
166 Lindenmayer 2007; Quétier *et al.* 2014; Lindenmayer *et al.* 2017). We established two levels
167 of information gaps in current practices: i) access to procedures files and ii) missing data in
168 the procedures files.

169 First, because of organization difficulties within the regional authorities' services, we could
170 only obtain information on offset implementation at the department level (even though this
171 information should be publicly available). Once access was granted by the Isère
172 departmental authorities, 66% of the projects could not be selected for the survey due to the
173 absence of the minimum necessary data, mostly the offset site location. Lack of human and
174 financial resources provided for public environmental authorities (Quétier *et al.* 2014), which
175 impacts the archiving of procedure files, may contribute to these difficulties to access the
176 files.

177 We observed further that 25% of the procedure files did not contain a management plan,
178 which we interpreted as an indication that no monitoring (and therefore no control) was
179 foreseen by the developer. However, to reach NNL, international recommendations (BBOP
180 2012) and the national legislation (law on biodiversity no. 2016-1087) stipulate that offset
181 sites have to be maintained and managed by the developer until the gains are effective

182 Second, the data found in the procedure files was of heterogeneous quality (i.e. missing
183 data, confusion between minimisation and offsets), also noted for another department in
184 France (Bigard *et al.* 2017). This restrained our analysis to potentially 33% of all offsets
185 (assuming 100% were accessible). Moreover, given that we depended on the managers for
186 access to the offset sites, our data set did not necessarily reflect the reality of the offset
187 practices as there is a higher probability that managers did only grant access to the sites if
188 they were confident in the success of their offset. Consequently, our results presented here
189 might underestimate offset failure.

190 Even for projects that had been surveyed, available information was not sufficient to verify if
191 the offsets provided the expected gains. The initial state of the offset site is the essential data
192 missing, even in its simplest form: presence or absence of the target habitats and species.
193 This information could be used in future analysis to measure gains of the target biodiversity
194 features. These gains would be provided if the biodiversity features were absent at the offset
195 site at the start of implementation but were present during the survey. To achieve NNL,
196 quantitative criteria related to the dynamics of populations and counterfactual scenarios
197 should also be measured (Bull *et al.* 2014). Other combinations such as no changes of the

198 biodiversity features on the offset site, or even losses if they disappear from the site (e.g.
199 meadow succession towards forest) could be considered as failures to achieve NNL.

200 4.2. Characteristics of offsets that fail in providing gains

201 Characteristics of offsets that fail to provide gains can be discussed based on the absence of
202 the target habitats or species from the offset site during the survey. However, their presence
203 alone does not provide enough information to discuss the characteristics of offsets that
204 succeed in providing gains (Bull *et al.* 2014). There could still be a bias in the measurement
205 of the absence of species (we could have missed individuals or tracks during the 1-day
206 survey), but we consider our measurement as valid since (i) we were assisted by the site
207 manager (with deeper knowledge of the offset site) and (ii) we were specifically searching for
208 the target species.

209 Our analysis should be seen as preliminary results since the data set is relatively small from
210 a statistical point of view. That being said, we showed that among the eight identified offset
211 characteristics (Table 1), only the type of offsets influenced the absence of the protected
212 species and their habitat or wetland and the related fauna community.

213 Two out of six types of offset contributed to a high occurrence of the target habitat (creation
214 of micro-habitats and maintenance of an already favourable habitats; Figure 2a). These
215 actions, often implemented by developers (Regnery *et al.* 2013) are indeed simple to realize.
216 In most cases, rocks or wood piles are placed in a relevant spot for micro-habitats and
217 human disturbances are prevented from the area. Reasons for unsuccessful establishment
218 of the target habitat are manifold: absence of a persistent management effort over time
219 leading to recolonization of grasslands by shrubs (Gibbons & Lindenmayer 2007), or non-
220 adapted techniques that do not achieve the estimated effect when setting up the offset, as
221 demonstrated for invasive plant removal (Early *et al.* 2016).

222 Similarly, management of terrestrial ecosystem never contributed to any occurrence of the
223 target species (Figure 2b). In most cases, the target species required a particular vertical
224 habitat structure (e.g. no shrub for birds of open areas), but as observed for the target
225 habitat, the management was often inappropriate (e.g. too much shrubs), resulting in the
226 absence of the target species.

227 4.3. Perspectives for better monitoring and NNL achievement

228 Even despite the small sample size of the offsets surveyed, the principal lesson learnt from
229 our study is that an analysis of offsets can be conducted in order to verify NNL achievement
230 (equivalence between losses and gains, effectiveness of gains in time, etc.; BBOP 2012).
231 Based on the above-mentioned limitations concerning accessibility and availability of data of
232 offset and the high failure rates for some types of offsets, we suggest three principal angles

233 for progress: (i) collecting and publishing minimum information on offsets, (ii) requiring a
234 management plan for each offset, and (iii) accumulating empirical evidence of offsets failure
235 and success.

236 First, systematically collecting the minimum necessary information, most particularly a
237 comprehensive initial state of the offset site, allows evaluating the gains provided by offsets
238 and is essential to assess NNL accurately (Pickett *et al.* 2013). Better organisation of offset
239 data would substantially improve offset monitoring at the regional and national scale (Quigley
240 & Harper 2006a; May *et al.* 2017). In France, since the law on biodiversity (2016), a national
241 database is being developed with the aim of centralising information on biodiversity offsets,
242 including the localisation of offset sites. Making such a database publicly available should
243 allow more transparency in offset implementation and facilitate standardised monitoring.

244 Second, according to international recommendations, a management plan should be
245 required to (i) monitor offset effectiveness and (ii) adapt management practices if necessary
246 to reach the expected gains. The management plan must last long enough to record the
247 succession of all biodiversity features, mostly those with slow vegetation succession (such as
248 forest; Maron *et al.* 2012).

249 Third, accumulating empirical evidence following the principles of evidence-based
250 conservation (Sutherland 2003) is suggested as a final axis of improvement. The offset
251 characteristics leading to failure (Lindenmayer *et al.* 2017) or success (Tischew *et al.* 2010)
252 in providing gains can be identified and either avoided or developed. Furthermore, offset
253 implementation could include an efficiency rate with more attention given to risky offsets. We
254 encourage that feedback from analyses of offsets such as provided herein is transferred from
255 scientists to stakeholders (developers, authorities, etc.) to make offsets efficient in actually
256 achieving NNL.

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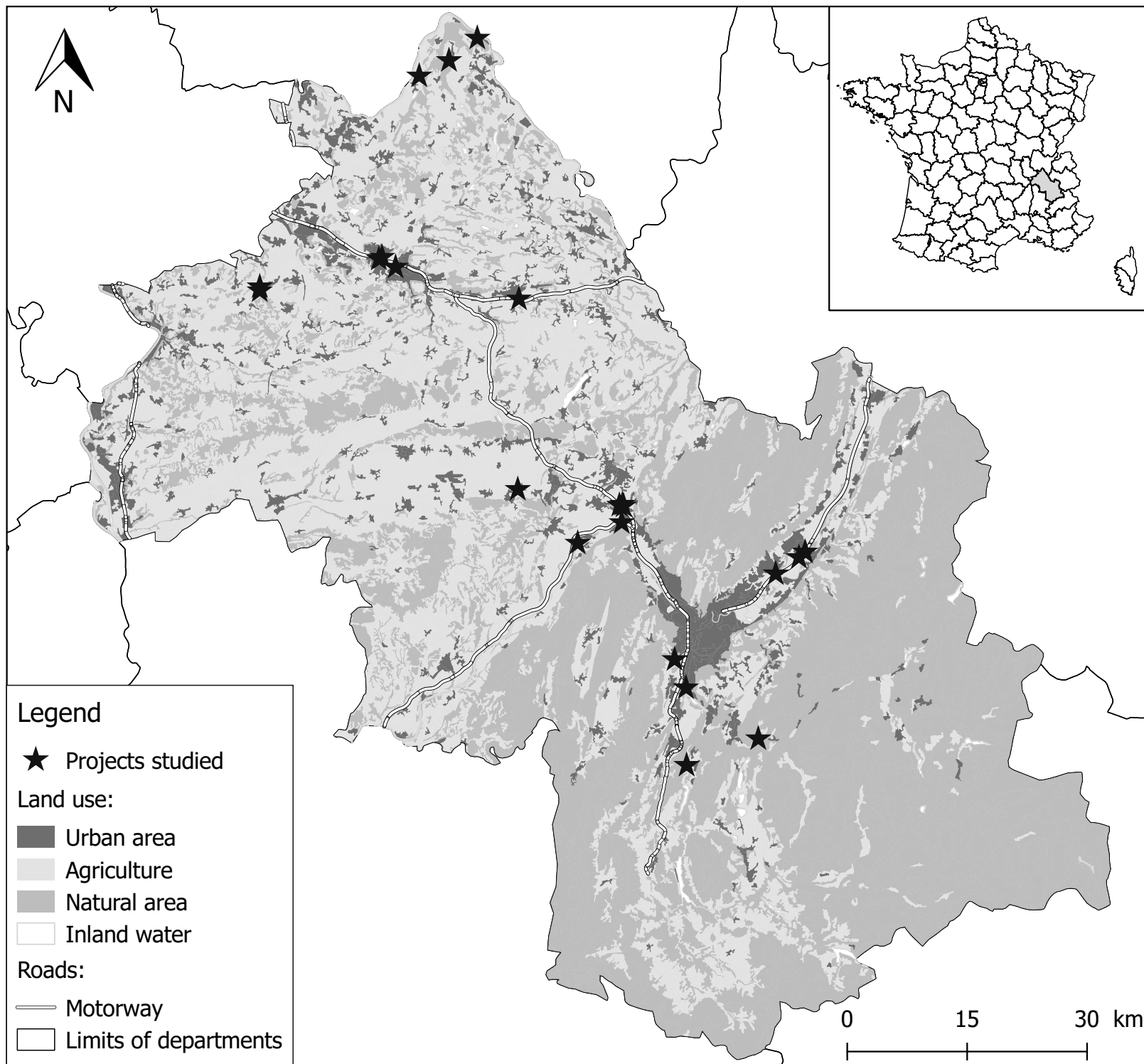


Figure 1 caption: Location of the 22 projects surveyed in the Isère department in France

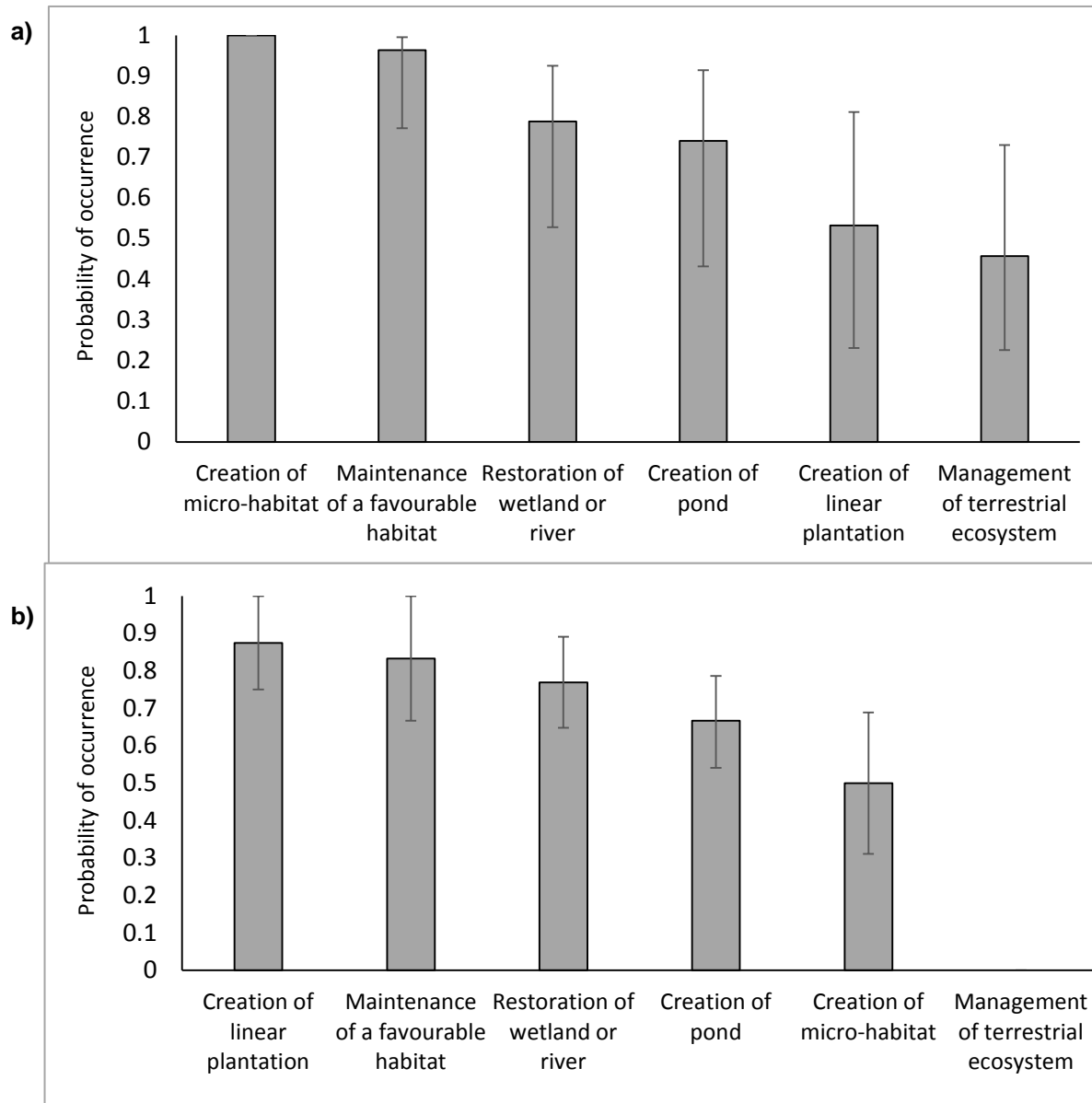


Figure 2: Probability of occurrence, i.e. backtransform estimate from the selected set of model, of a) the target habitats and b) the target species on offset site depending on the type of offset implemented.

Table 1: Offset characteristics that could influence the presence or absence of the target habitat or species.

Offsets characteristics	Variable description	Reasons for variable choice
Offsets procedure (Procedure)	Categorical: Protected Species (n=18), Wetland (n=24), Both (n=17)	There may be different expectations from the administration towards offsets implementation and monitoring depending on the procedure
Type of offsets (Type offsets)	Categorical: Management of terrestrial ecosystem (n=5), Restoration of wetland or river (n=13), Creation of pond (n=16), Creation of micro-habitat (n=10), Creation of linear plantation (n=9), Maintenance (n=6).	Using different ecological engineering techniques can lead to various outcomes of offset actions (Maron <i>et al.</i> 2012)
Type of developer (Type dev)	Categorical: Private (n=31), Public (n=28)	There may be different expectations from the administration towards offsets implementation and monitoring depending on the status of the developer.
Distance between impacted and offsets site (Dist. Imp. offset)	Quantitative continuous (km)	There is a greater chance that the environment on offset site is similar to impacted site when they are nearby (Kiesecker <i>et al.</i> 2009)
Distance between offsets site and nearest landscape ecological corridor (Dist. offset. Corr.)	Quantitative continuous (km)	Landscape ecological corridors are important areas for biodiversity (Taylor <i>et al.</i> 1993)
Time since offsets implementation (Time)	Quantitative continuous (years)	The outcomes of offset actions can be visible after a certain number of years (Maron <i>et al.</i> 2012)
Number of offsets on the site (Nb. offsets)	Quantitative continuous (number of offset measures)	The combination of several offsets on a compensatory site could be more favourable for the target biodiversity features (Tischew <i>et al.</i> 2010)
Group of species targeted by offsets: birds, mammals, reptiles, amphibians, "others" (bats, insects and flora) (Target species)	Categorical: Yes /No	Groups of species have different requirements (home range, dispersal ability) and life cycles: some might be easier to maintain, restore or create.

Table 2: Details on projects studied (impacted and offset sites)

Project	Procedure W=wetland PS=protected species	Impacted site area (ha)	Type of developer	Type of impacts	Target taxonomic groups	Year of offsets	Number of offset sites	Number of offsets on each site
Project A	W	0,95	Public	Urbanization	Wetland fauna	2012	1	1
Project B	W	0,45	Private	Transport infrastructure	Birds, bats, reptiles, odonates, amphibians	2013	2	3 - 3
Project C	W	0,3	Private	Hydraulic construction	Birds, bats, reptiles, odonates, amphibians	2013	1	1
Project D	W PS	0,005	Public	Transport infrastructure	Odonates, amphibians	2014	1	1
Project E	PS	NA	Private	Quarry	Amphibians, reptiles, birds, mammals	2012	2	1 - 1
Project F	PS	5,58	Private	Quarry	Amphibians, reptiles, birds	2013	2	2 - 1
Project G	PS	NA	Private	Quarry	Amphibians, reptiles, birds	2013	1	5
Project H	W	7	Public	Urbanization	Wetland fauna	2001	1	1
Project I	W	NA	Private	Transport infrastructure	Amphibians, reptiles, birds	2012	1	2
Project J	W PS	0,55	Private	Transport infrastructure	Birds, reptiles, odonates, amphibians	2013	1	4
Project K	W	0,4	Public	Hydraulic construction	Wetland fauna	2011	1	6
Project L	W	0,21	Public	Urbanization	Wetland fauna	2009	1	1
Project M	PS	13,37	Private	Hydraulic construction	Amphibians, reptiles, birds, mammals, bats, insects, flora			
Project N	W	0,974	Public	Urbanization	Wetland fauna	2014	1	2
Project O	W	1,8	Public	Transport infrastructure	Wetland fauna	2012	1	1
Project P	W PS	NA	Public	Urbanization	Amphibians, reptiles, birds, mammals	2011	2	3 - 1
Project Q	PS	NA	Private	Electricity network	Reptiles, flora	2013	2	1 - 1
Project R	W PS	NA	Public	Hydraulic construction	Birds, bats, reptiles, amphibians, fishes	2013	1	2
Project S	W PS	27,5	Private	Hydraulic construction	Amphibians, reptiles, birds, mammals, bats, flora	2012	2	2 - 1
Project T	W	0,2	Public	Water treatment	Wetland fauna	2013	2	1 - 5
Project U	W	0,5	Public	Urbanization	Wetland fauna	2015	2	1 - 2
Project V	W PS	NA	Public	Urbanization	Amphibian	2007	1	3

Table 3: Models set for target habitat and species

Models set with df, log-likelihood, AICc, Δ AICc and Akaike weights for target habitat and species. Only the first models (cut-off Δ AICc < 3.8 for target habitat and 7.0 for target species) are provided. Abbreviations: the type of offsets (Type offset), the number of offsets implemented on the site (Nb. Site), the distance between the offset site the nearest landscape ecological corridor (Dist. Cor.), the time since offsets implementation (Year), and variables for the target species group “others” including bats, insects and flora (Others).

Models	df	log-likelihood	AICc	delta	weight
Target Habitat					
Others+Proced.+Type offset	10	-24.28	68.57	0	0.04
Proced.+Type offset	9	-25.93	69.87	1.30	0.02
Type offset	7	-28.14	70.27	1.71	0.02
Others+Type offset	8	-27.20	70.41	1.84	0.02
Amph.+Proced.+Type offset	10	-25.29	70.59	2.02	0.02
Target Species					
Dist. Cor.+Nb. site+Type offset	9	-18.89	59.87	0.00	0.13
Nb. site+Type offset	8	-20.54	60.28	0.42	0.11
Others+Type offset	8	-20.83	60.87	1.00	0.08
Nb. site+Type offset+Year	9	-19.39	60.87	1.00	0.08
Nb. site+Others+Type offset	9	-19.81	61.70	1.84	0.05