

How to meet nutritional recommendations and reduce diet environmental impact in the Mediterranean region? An optimization study to identify more sustainable diets in Tunisia

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1 Title:

- 2 How to meet nutritional recommendations and reduce diet environmental impact in the
- 3 Mediterranean region? An optimization study to identify more sustainable diets in Tunisia
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Abstract

32 Tunisia is a typical country of the Mediterranean region where high prevalence of overweight,

obesity and non-communicable diseases co-exist with some micronutrient deficiencies, and

diet-related environmental issues must be addressed. Individual food choices may influence

both health and environment. The aim of this study was to identify diets that are nutritionally

adequate, culturally acceptable, and with low environmental impact for Tunisian adults.

37 Individual dietary data from a national Tunisian survey on food consumption (n=7209, 35-70

years) and the national food composition table were used to estimate the food and nutritional

content of the mean observed (OBS) diet. The diet environmental impact was assessed through

40 seven metrics: water deprivation, land-use, land-use potential impacts on biodiversity loss,

erosion resistance, mechanical filtration, groundwater replenishment, and biotic production.

Quadratic optimization models were implemented to obtain diets that met the nutritional

recommendations, and concomitantly respected increasingly stringent environmental

constraints and minimized the departure from the OBS diet.

Without environmental constraints, the nutritional recommendations were met by increasing

46 the amount of dairy, starch and vegetables, and decreasing foods high in fat/salt/sugar (HFSS)

47 and added fat. Compared with the OBS diet, the environmental impact of this diet increased: +32 % for water deprivation and +46-48 % for land use and its impacts. 48 When a moderate environmental impact reduction (\le 30\%) was added to the nutritional 49 50 constraints, the dietary changes at the food group level were similar to those required to reach 51 nutritional adequacy, except for a progressive decrease in meat/fish/egg quantities. Animal-52 based product contributions to the total energy and protein content were close or slightly lower 53 than in OBS diet, but a redistribution of sources was required: less meat in favor of dairy, egg and fish products. Stronger reductions (≥40%) required substantial changes that might 54 55 compromise the optimized diet acceptability. 56 Targeting a nutritionally adequate diet without considering its environmental impact might 57 increase water deprivation, land use and its impacts on biodiversity and soil quality. In Tunisia, 58 moving towards healthy diets with lower environmental impact relied more on redistributing 59 the sources of animal-based products rather than on reducing their total contribution, together 60 with a decrease of HFSS and added fats, and an increase of vegetables. Actions to favor the 61 adoption of such dietary changes by consumers should be explored to promote more sustainable 62 diets in the Mediterranean region. 63 64 Keywords: nutrition; sustainability; diet; optimization; mathematical programming; multi-65 criteria analysis; Tunisia; water footprint; biodiversity; land-use; food consumption; dietary

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shifts; Mediterranean region

1 Introduction

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During the last decades, several countries in the Mediterranean region underwent an epidemiological and nutritional transition that has resulted in a major increase of the prevalence of overweight, obesity and non-communicable diseases (NCD), while some micronutrient deficiencies persist (Gartner et al., 2014; NCD Risk Factor Collaboration (NCD-RisC), 2016a, 2016b). Besides these public health challenges, the Mediterranean area is also facing climate and environmental issues, especially water deprivation and biodiversity loss, particularly in the Near East and North Africa (CIHEAM/FAO, 2015). The current food system has a major environmental impact by contributing between 19-29% of global greenhouse gas emissions (GHGE) (Vermeulen et al., 2012) and by representing ~70% of global freshwater use (Whitmee et al., 2015). Therefore, changes in food consumption and production patterns are needed to ensure more sustainable food systems and achieve food and nutrition security in the Mediterranean region. As individual food choices can influence public health and also the environment, there is an urgent need to promote sustainable diets, defined by the Food and Agriculture Organization (FAO) as nutritionally adequate, safe and healthy, culturally acceptable, financially affordable, and with low environmental impacts (Food and Agriculture Organization, 2010). Previous studies have explored the potential of dietary shifts towards more sustainability by assessing the environmental impact of the existing diets (Perignon et al., 2017) or of dietary scenarios, such as the Mediterranean-type diet, New Nordic diet, and diets with reduced levels of animal products, compared with the national average diet (Berners-Lee et al., 2012; Risku-Norja et al., 2009; Sáez-Almendros et al., 2013; Saxe et al., 2013; Temme et al., 2013; van Dooren et al., 2014). However, these approaches do not allow identifying diets that improve simultaneously all dimensions of diet sustainability. Indeed, high nutritional quality is not necessarily associated with affordability or lower environmental impact (Biesbroek et al., 2017;

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Perignon et al., 2017; Vieux et al., 2013). Moreover, some dietary scenarios may be too different from the dietary habits in the studied countries, which limits their acceptability. Mathematical diet optimization (herein referred as "diet optimization") can be used to find the optimal combination of foods to fulfil a set of constraints, and is a unique and powerful tool for studying simultaneously the multiple dimensions of diet sustainability (Gazan et al., 2018). For instance, when applied to study sustainability issues, diet optimization can be used to meet nutrient recommendations and reduce environmental impacts, while maximizing the similarity with the current diets. Moreover, previous studies essentially assessed the environmental impact in terms of GHGE (Payne et al., 2016; Perignon et al., 2017). However, it is well known that GHGE are not a proxy for the full range of environmental impacts associated with a diet. Indeed, among the 169 targets of the Sustainable Development Goals, water deprivation, land degradation, and biodiversity loss have been identified as environmental areas of concerns that need to be addressed (IPBES, 2019; Ridoutt et al., 2017; United Nations General Assembly, 2015). Yet, a recent review reported that very few studies investigated dietary changes to reduce these impacts by using metrics that can be applied in a life cycle context (Ridoutt et al., 2017). The authors concluded that the available evidence on dietary patterns and water deprivation, land degradation and biodiversity loss is very limited, and did not identify generalizable findings. In addition, the few existing studies were all conducted in Europe, and only one explored the effect of shifting to a healthier diet on water deprivation (Hess et al., 2015). Therefore, more analyses of the compatibility between nutritional and environmental goals using appropriate metrics are needed, especially in the Mediterranean region where water deprivation is critical. Tunisia is a typical country of the Mediterranean region that is undergoing a nutrition transition and where high prevalence of overweight, obesity and NCD co-exist with some micronutrient deficiencies (Atek et al., 2013; Traissac et al., 2016). Tunisia has a marked climatic north-south

gradient, from a Mediterranean region in the north to a semi-arid and then desert area in the south. This puts the country especially at risk to climate change effects on land, coastal zones, water, and agriculture (Thiébault et al., 2016; Verner and World Bank. Middle East and North Africa Region. Sustainable Development., 2013).

The objective of the present study was to identify, using diet optimization models, the dietary changes that allow fulfilling the World Health Organization (WHO) nutritional recommendations, reducing the diet environmental impact, and respecting the Tunisian population's dietary habits.

2 Methods

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among whom 52.9 % were women.

2.1 Dietary and food composition data

Dietary data were derived from a nationally representative cross-sectional survey performed among 35-70-year-old adults of both sexes in Tunisia in 2005, as part of the Transition and Health Impact in North Africa (TAHINA) project (Atek et al., 2013). This survey collected retrospective data on food consumption during one month using a validated semi-quantitative food frequency questionnaire (El Ati et al., 2004). For the purpose of our study, the 138 food items declared to be consumed by the participants were classified in 8 food groups [fruits & vegetables, starch, meat/fish/eggs (MFE), dairy, foods high in fat/salt/sugar (HFSS), mixed dishes, added fat & seasoning, drinks], and 23 food sub-groups (Supplemental Table 1). A specific Tunisian food composition database (El Ati et al., 2007), completed by the USDA table (US Department of Agriculture, 2008), additional laboratory analyses and the Food Processor software, version 8.3 (ESHA-Research-Inc, 2003) were used to estimate the energy and nutritional content (macro- and micronutrients) of the identified food items and diets. The 138 food items were also classified as animal- or plant-based products to estimate the animal-based product contributions to the diet total energy and protein content. Energy intake under-reporters were identified using Black's equations (Black, 2000). As the prevalence of overweight and obesity was high in the studied population (71% and 37% among women) (Atek et al., 2013), the basal metabolic rates used to calculate Black's cutoffs were estimated using Mifflin equation (Mifflin et al., 1990). The mean observed (OBS) diet was

estimated using data from a final sample that included 6279 adults, aged 49.2 ± 9.5 years,

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2.2 Environmental impact of diets

The environmental impact of food items was estimated using seven metrics: water deprivation, land use, land use potential impacts on erosion resistance, mechanical filtration, groundwater replenishment, biotic production, and biodiversity. Impacts were computed with a life cycle vision using a hybrid method that combined trade statistics and production data, in order to estimate the impact in the countries of production (Tunisia and/or other countries, if imported) of the food items consumed in Tunisia. The methodology used to estimate the seven metrics (expressed by kg of food) for each of the 138 foods declared to be consumed by the Tunisian population in the nationally representative cross-sectional study has been described elsewhere (Sinfort et al., 2019). Briefly, national dietary survey data were matched with the UNComtrade and the FAOstat databases to obtain the quantity of food produced per production country, for each food item consumed in Tunisia. Yield per crop and per country were used to compute the occupied surfaces, and blue water consumption was extracted from the Water Footprint Network datasets. The potential impacts were then obtained by multiplying the amounts of consumed water and land use surface with characterization factors. The characterization factor used to estimate water deprivation impacts was the Water Stress Indicator provided by Pfister et al. for each country (Pfister et al., 2009). Land use impacts were computed from the occupied surface (including land occupied by animal feed crops), from land use types, and from the main biome of the production country. Then LANCA characterization factors (Beck et al., 2011; Bos et al., n.d.) were used to compute land use potential impacts. The land use impacts on biodiversity were calculated using country-specific global characterization factors estimated by Chaudhary et al. with the countryside species-area relationships (SAR) model and average approach (Chaudhary et al., 2015). The developed methodology assessed the impact at a global scale, which is necessary when studying complete diets that include food items from many countries.

The environmental impact of the OBS and optimized diets was then estimated for each metric by summing the impact of all food items weighed by their quantity in the diet. For each metric, a positive value indicates a detrimental impact, and a negative value a beneficial impact.

2.3 Diet optimization

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Quadratic optimization models were used to obtain nutritionally adequate diets that departed the least from the food content of the OBS diet. The model variables were the 138 food items consumed by the population. For each model, the objective function to be minimized was the quadratic deviation from the mean observed intake for each food item and food group, in order to promote minimal variations on all foods and penalize large variations in the diet composition.

182 The objective function was expressed as follows:

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$$Minimize f = \frac{1}{138} \sum_{i=1}^{138} \left(\frac{Q_i^{opt} - Q_i^{obs}}{Q_i^{obs}} \right)^2 + \frac{1}{8} \sum_{j=1}^{8} \left(\frac{Q_j^{opt} - Q_j^{obs}}{Q_j^{obs}} \right)^2$$

where i represents the 138 food items and j the eight food groups (starch, vegetables; fruits; MFE, dairy, mixed dishes, added fat & seasoning, and drinks), Q^{obs} is the mean observed quantity, and Q^{opt} the optimized quantity. The minimization function was applied at the food item level to deviate as little as possible from the OBS diet, but also at the food group level to respect the meal structure habits and favor substitutions by foods from the same meal component.

The total energy intake of the OBS diet was imposed in all models, as well as nutritional constraints in order to meet the WHO recommendations for 30 nutrients (list of nutritional constraints in **Table 1**). In addition, the fish subgroup was constrained to a maximum intake of two portions per week, to avoid high exposure to contaminants (ANSES, 2010).

Models with increasingly stringent environmental constraints were defined: a model without constraints on the environmental metrics (Nut-Env_{free} model), a model with constraints that

limited the environmental metrics to the observed level (Nut-Env_{obs}), and models with 196 197 constraints to decrease the environmental indicators by 10% at each step (Nut-Env-10, Nut-Env-198 20, etc... until mathematical infeasibility). 199 Finally, realism constraints were included in all models to avoid implausible changes relative 200 to the diet consumed by the general Tunisian adult population. Specifically, the total diet weight 201 could vary only by $\pm 20\%$ relative to the mean observed intake. Moreover, the quantities of food items, groups and subgroups could range between the 5th and 95th percentiles of the observed 202 203 intakes (percentiles were calculated for consumers in the case of food items, and for the whole 204 population in the case of food groups and subgroups). All models were run using the GAMS 205 software package (version 23.8.2). **Results** 206 3 207 The food group and subgroup quantities in the OBS diet and in the optimized diets are detailed 208 in Supplemental Table 2. Food composition, nutritional content, and environmental impact of the mean 209 210 observed diet 211 The food group composition and nutritional content of the OBS diet are presented in Figure 1 212 and Table 1, respectively. The OBS diet did not meet the nutritional constraints for calcium, 213 copper, iron, magnesium, potassium, vitamin D, and vitamin E (all below the WHO 214 recommendations), as well as sodium and total fat (both above the WHO recommendations). 215 Animal products represented 15.4% of the total energy, and 42% of the total proteins. 216 The environmental impacts (per person and per day) of the OBS diet were 0.32 m³ eq of water deprivation (Figure 2A), 262 m² of land use (Figure 2B), 16538 m³ of water infiltration loss 217 from land occupation (Figure 2C), 2.6 m³ of groundwater regeneration loss from land 218 occupation (Figure 2D), and 2.2.10⁻¹³ species lost due to land use (Figure 2E). The impact was 219

220 beneficial for two metrics: -274.3 kg of soil loss due to erosion from land occupation and -139.5 221 kg of biotic production loss from land occupation (data not shown). Therefore, the subsequent 222 analyses focused on the five environmental metrics that showed a detrimental impact of the diet 223 on the environment. 224 3.2 Dietary changes needed to reach nutritional adequacy and consequences on the diet 225 environmental impact (Nut-Env_{free} vs. OBS diets) 226 Compared with the OBS diet, dairy (+98%), fruit (+13%), vegetables (+23%), and starch 227 (+33%) quantities were increased, and HFSS (-58%), and added fat & seasoning (-21%) were 228 reduced in the Nut-Env_{free} diet (**Figure 1A**). The total MFE quantity did not change between 229 OBS and Nut-Env_{free} diets, but intra-group substitutions occurred (Figure 1B): the quantity of 230 eggs (+49%), red meat (+47%), fish/seafood (+15%) and offal (+13%) increased, while poultry 231 decreased (-72%). 232 All five environmental metrics that showed a detrimental impact increased in the Nut-Env_{free} 233 diet compared with the OBS diet (Figure 2): water deprivation (+32 %), land use (+46%), 234 biodiversity loss (+48%), groundwater regeneration loss (+47%), and mechanical filtration 235 (+47%).236 Dietary changes needed to reduce the diet environmental impact and reach 237 nutritional adequacy (Nut-Env_{free} vs Nut-Env₋₁₀, ... diets) 238 At the food group level, reducing by up to 30% each of the environmental metrics (Nut-Env-10. 239 Nut-Env-20, and Nut-Env-30 models) did not require any additional change in food group 240 quantities than those present in the Nut-Env_{free} diet, except for a progressive decrease in MFE 241 quantities (Figure 1A, "moderate impact reductions" section). Conversely, for reducing the 242 environmental impact by more than 40% (Nut-Env-40 to Nut-Env-70 models), major changes in food group quantities were needed (Figure 1A, "strong impact reductions" section): higher 243 244 vegetable and fruit quantities, and progressive reduction of the amount of starch and dairy.

Thereafter, "moderate impact reductions" and "strong impact reductions" will be used to define 245 246 environmental impact reductions up to 30% and equal/higher than 40%, respectively. 247 At the food subgroup level, the changes within the MFE group were different from those 248 induced by the Nut-Env_{free} model when the environmental impact constraints were imposed 249 (Figure 1B). Specifically, red meat quantity was increased in the Nut-Env_{free} diet (+47% vs. OBS diet), whereas it was reduced by 50% in the Nut-Envobs diet (vs. OBS diet) and even more 250 251 for moderate impact reductions. For stronger impact reductions, red meat was nearly (impact 252 reductions of 40%) or totally suppressed (impact reductions ≥50%). Egg quantity increased for 253 moderate impact reductions ($\leq 30\%$), whereas it progressively decreased for stronger reductions 254 $(\geq 40\%)$. 255 In terms of dietary energy content (**Figure 1C**), moderate impact reductions (≤30%) required a 256 decrease of meat contribution to the total energy that was compensated by a progressive 257 increase of the egg and starch contributions. For strong impact reductions (≥40%), the energy 258 contributions of dairy products and egg decreased, and were balanced by higher contributions 259 of fruits and starch. 260 The greatest achievable environmental impact reduction while respecting all nutritional 261 recommendations was 70%. For 80% reduction, there was no feasible solution (i.e., no 262 combination of foods) to fulfill the whole set of nutritional and realism constraints. The 263 constraints on vitamin D and calcium made not feasible the diet optimization associated with 264 80% reduction of the environmental impact. Although mathematically possible, reaching 265 nutritional adequacy while reducing by 70% the environmental impact required an extreme shift 266 from the OBS diet (Figure 1), particularly very high intakes of fruits and vegetables (almost 267 1.1 kg/day). Considering that vitamin D can primarily be obtained from sun exposure, 268 sensitivity analyses were performed with the constraint that vitamin D level should not be 269 reduced relative to the level in the OBS diet (~3 µg/day), instead of imposing to fulfill the

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recommendation of 5 µg/day (data not shown). With this new constraint, changes in food group quantities were very similar, except for the model with the highest environmental impact reduction (Nut-Env₋₇₀). In the Nut-Env₋₇₀ diet with the new constraint on vitamin D, the quantity of eggs (an important contributor to vitamin D content) did not increase, unlike in the Nut-Env- $_{70}$ model with the vitamin D >5 μg/d constraint (**Figure 1B**), while that of vegetables increased to compensate for the egg contribution to vitamin A. 3.4 Changes in animal-based product contributions Reaching nutritional adequacy (Nut-Env_{free}) induced an increase of the animal-based product contribution to the total energy (from 15.4% in the OBS diet to 18.5% in the Nut-Env_{free} diet) (Figure 3A). The share of proteins from animal origin was 42% in the OBS diet and 40% in the Nut-Env_{free} diet (**Figure 3B**). When moderate environmental impact reductions ($\leq 30\%$) were added to the nutritional constraints, the total contribution of animal products to the dietary energy (approximately 1/6 of the total energy) remained similar, but the fraction of animal proteins was lower (approximately one third of the total protein content), compared with the OBS diet. For stronger environmental impact reductions (>40%), the total contribution of animal products to dietary energy and protein content progressively decreased.

Beyond the total contribution, the contribution of each animal-based product changed. For moderate environmental impact reductions (\leq 30%), the contribution of the dairy and egg subgroups to the total energy and protein content increased, while that of red meat and poultry decreased compared with the OBS diet. For stronger reductions (>40%), the meat contribution was <0.5% and the dairy group contribution progressively decreased.

4 Discussion

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Based on individual dietary data from a national survey, the present study i) estimated the environmental impact (water deprivation, land use, land use potential impacts on biodiversity, erosion resistance, mechanical filtration, groundwater replenishment, and biotic production) of the average diet consumed by the adult Tunisian population, and ii) identified the main dietary shifts required to meet the nutritional recommendations, and concomitantly reduce the environmental impact and minimize the departure from the observed average diet for respecting eating habits and cultural acceptability. We estimated the water deprivation impact of the average Tunisian diet to 316 Leg/person per day. This value is twice higher than the 160 L/person per day estimated for the current UK food consumption (Hess et al., 2015). We did not find any literature data to compare the biodiversity impact due to land use. Our study revealed that the soil impact related to land occupation was beneficial, for two of the four indicators. This highlights the complexity of land use impact on the environment, especially in semi-arid regions, such as the south of Tunisia. Indeed, landsaving measures are needed to reduce biodiversity loss and protect ecosystem services (Foley et al., 2011). On the other hand, developing farming in semi-arid areas could be beneficial in terms of biomass production and resistance to erosion, if livestock management is adequate. However, the land use impact estimations should be interpreted with caution because they are computed from annual and country-level averages that do not take into account local specificities. We showed that fulfilling the WHO nutritional recommendations induced an increase of the diet environmental impact: by approximately 30% for water deprivation and by nearly 50% for indicators of land use impact, particularly biodiversity loss. A previous study reported that the dietary scenario designed to conform with the "Eatwell plate" guidelines led to a modest change

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in the water-scarcity footprint of UK food consumption (-3%), with a large impact variability depending on the production countries (from -18% for the impact in Belgium to +30% for the impact in Pakistan) (Hess et al., 2015). Our results are consistent with the study by Tom et al. (Tom et al., 2015) who found that the blue water footprint increased by 16% when shifting from the current US diet to a healthier diet. The increased environmental impacts found in our study were primarily driven by the increase in dairy products (for water footprint and land use) and starch and fruits (for water footprint). This diet change was probably driven by the low intake of calcium, vitamin D and magnesium in the mean observed diet. Our results highlight the challenge of developing more sustainable diets, with trade-offs between health and environmental goals. Similarly, previous studies observed that healthier diets were associated with higher GHGE (Biesbroek et al., 2017; Perignon et al., 2016; Vieux et al., 2018). However, our diet optimization study also showed that some dietary shifts (increasing the amount of vegetables, dairy and starch products, decreasing HFSS and fats, and reducing meat in favor of fish and eggs) could reconcile nutritional adequacy and a lower environmental impact, while minimizing the departure from the average Tunisian diet. For a 30% reduction of the environmental impact, the magnitude of dietary changes was similar to that required to reach nutritional adequacy alone. However, for higher environmental impact reductions ($\geq 40\%$), more substantial dietary shifts are required that might compromise the cultural acceptability of such optimized diets. Reaching nutritional adequacy induced an increase of animal-based products (from approximately 1/6 of the total energy in the observed diet to 1/5 in the optimized diets). When environmental impact reductions were imposed in addition to the nutritional adequacy goal, their energy contribution was decreased to similar levels as in the mean observed diet, but a redistribution within animal-based products occurred with a reduction of meat contribution in favor of dairy products, fish and eggs. Therefore, although reducing the consumption of animal

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products is often suggested as a key strategy to lessen the environmental impact of diet (Ridoutt et al., 2017), recommendations targeting total animal products may not be appropriate in some Mediterranean countries where the current intake of animal-based products can be, in some contexts, already low. Our optimization study showed that in Tunisia, moving towards a more sustainable diet relied more on redistributing the sources of animal-based products (increase in dairy, fish and eggs vs. reduction of meat products) rather than on reducing their total contribution. Our results underline the importance of context-specific recommendations and confirm that the regional realities need to be carefully considered when examining the role of animal-source foods in achieving more sustainable diets (Willett et al., 2019). The first strength of our study is the assessment of the diet environmental impact based on several water and land use indicators, and estimated using a life cycle approach that considers the impacts in the food-producing countries. By taking into account international trade and weighing water use with Water Stress Index factors and land use with country-specific characterization factors, the present study assessed sustainability concerns on a global scale. Moreover, our study is based on dietary data from a national survey using a specific and validated food frequency questionnaire, and a Tunisian food composition table. Moreover, our study took into account simultaneously several dimensions of diet sustainability (nutrition, environment, and cultural acceptability) using diet optimization. Accordingly, our study identified not only the dietary shifts required to reach a healthy diet that fulfils a whole set of nutritional recommendations, but also the shifts needed to move towards a healthy diet with a lower environmental impact. Furthermore, by minimizing the changes from the observed diet, the optimized diets were more realistic and potentially culturally acceptable (Gazan et al., 2018). The present study has some limitations. It could be improved by taking into account the bioavailability of key nutrients, such as iron and zinc, that is influenced by the presence of absorption enhancers and inhibitors in the diet (Barré et al., 2018). Moreover, fish consumption

has important effects on biodiversity that are not taken into account in this study due to the lack of data. The studied population (35-70 years) did not include younger adults and this can also be seen as a limitation. In addition, using an individual diet optimization approach (rather than optimizing the population diet as done in the present study) would better integrate individual food preferences and eating habits (Gazan et al., 2018). Moreover, although several sustainability dimensions were taken into account, this study could be improved by integrating the diet cost in the models. Finally, although minimizing the departure from the observed diet and introducing realism constraints allowed avoiding extremely theoretical diets, such method cannot guarantee that the resulting dietary shifts would be acceptable to the consumer.

5 Conclusion

This diet optimization study showed that designing a nutritionally adequate diet without considering its environmental impact might increase diet-related land use, water deprivation, and land use impacts on biodiversity and soil quality. However, nutritional adequacy and moderate reductions of the environmental impacts (-30%) might be achieved through dietary shifts different in type but of similar magnitude than those required to meet the nutritional recommendations alone. In Tunisia, moving towards healthy diets with lower environmental impact relied more on redistributing the sources of animal-based products rather than on reducing their total contribution (less meat in favor of dairy, egg and fish products), together with an increase of vegetables and a decrease of fat and sweet products. The dietary changes identified in this study can be translated into action proposals to target food consumption and production in order to promote more sustainable diets in the Mediterranean region. The implementation of actions to favor the adoption of the identified dietary changes by consumers should be investigated.

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7 References

- 396 ANSES, 2010. Rapport de l'Anses relatif à la consommation des poissons, mollusques et
- 397 crustacés Aspects nutritionnels et sanitaires pour l'Homme.
- 398 Atek, M., Traissac, P., El Ati, J., Laid, Y., Aounallah-Skhiri, H., Eymard-Duvernay, S.,
- Mézimèche, N., Bougatef, S., Béji, C., Boutekdjiret, L., Martin-Prével, Y., Lebcir, H.,
- Gartner, A., Kolsteren, P., Delpeuch, F., Romdhane, H. Ben, Maire, B., 2013. Obesity and
- 401 Association with Area of Residence, Gender and Socio-Economic Factors in Algerian and
- 402 Tunisian Adults. PLoS One 8, e75640. https://doi.org/10.1371/journal.pone.0075640
- Barré, T., Perignon, M., Gazan, R., Vieux, F., Micard, V., Amiot, M.-J., Darmon, N., 2018.
- Integrating nutrient bioavailability and co-production links when identifying sustainable
- diets: How low should we reduce meat consumption? PLoS One 13, e0191767.
- 406 https://doi.org/10.1371/journal.pone.0191767
- 407 Beck, T., Bos, U., Wittstock, B., Baitz, M., Fischer, M., Seldbauer, K., 2011. Land Use
- 408 Indicator Value Calculation in Life Cycle Assessment Method Report, Cycle. Fraunhofer
- Verlag.
- Berners-Lee, M., Hoolohan, C., Cammack, H., Hewitt, C.N., 2012. The relative greenhouse gas
- 411 impacts of realistic dietary choices. Energy Policy 43, 184–190.
- 412 https://doi.org/10.1016/j.enpol.2011.12.054
- Biesbroek, S., Verschuren, W.M.M., Boer, J.M.A., van de Kamp, M.E., van der Schouw, Y.T.,
- Geelen, A., Looman, M., Temme, E.H.M., 2017. Does a better adherence to dietary
- guidelines reduce mortality risk and environmental impact in the Dutch sub-cohort of the
- European Prospective Investigation into Cancer and Nutrition? Br. J. Nutr. 118, 69–80.
- 417 https://doi.org/10.1017/S0007114517001878

418 Black, A.E., 2000. Critical evaluation of energy intake using the Goldberg cut-off for energy 419 intake:basal metabolic rate. A practical guide to its calculation, use and limitations. Int. J. 420 Obes. Relat. Metab. Disord. 24, 1119-30. Bos, U., Horn, R., Beck, T., Lindner, J.P., Fischer, M., Fraunhofer IBP, S., Fraunhofer IRB-421 422 Verlag, n.d. LANCA. Characterization Factors for Life Cycle Impact Assessment, Version 423 2.0. 424 Chaudhary, A., Verones, F., de Baan, L., Hellweg, S., 2015. Quantifying Land Use Impacts on 425 Biodiversity: Combining Species-Area Models and Vulnerability Indicators. Environ. Sci. 426 Technol. 49, 9987–9995. https://doi.org/10.1021/acs.est.5b02507 427 CIHEAM/FAO, 2015. Mediterranean food consumption patterns: diet, environment, society, 428 economy and health. 429 El Ati, J., Béji, C., Farhat, A., Haddad, S., Chérif, S., Trabelsi, T., Danguir, J., Gaigi, S., Le 430 Bihan, G., Landais, E., Eymard-Duvernay, S., Maire, B., Delpeuch, F., 2007. Table de 431 composition des aliments tunisiens. Tunis, Tunisia. El Ati, J., Le Bihan, G., Haddad, S., Eymard-Duvernay, S., Cherif, S., Holdsworth, M., Traissac, 432 433 P., Ben Rayana, C., Delpeuch, F., 2004. Food Frequency Questionnaire for Tunisian 434 dietary intakes: development, reproducibility and validity. Arab J. Food Nutr. 5, 10–30. 435 ESHA-Research-Inc, 2003. Food Processor Software Version 8.3. FAO/WHO, 2007. FAO/WHO scientific update on carbohydrates in human nutrition: 436 437 conclusions. Eur. J. Clin. Nutr. 61 Suppl 1, S132-7. 438 https://doi.org/10.1038/sj.ejcn.1602943 439 FAO, 2010. Fats and fatty acids in human nutrition, Report of an expert consultation, FAO, 440 Food and Nutrition Paper. https://doi.org/978-92-5-106733-8

441 Foley, J.A., Ramankutty, N., Brauman, K.A., Cassidy, E.S., Gerber, J.S., Johnston, M., Mueller, 442 N.D., O'Connell, C., Ray, D.K., West, P.C., Balzer, C., Bennett, E.M., Carpenter, S.R., 443 Hill, J., Monfreda, C., Polasky, S., Rockstrom, J., Sheehan, J., Siebert, S., Tilman, D., Zaks, 444 2011. Solutions for a cultivated planet. Nature 478, D.P.M., 337–342. 445 https://doi.org/10.1038/nature10452 446 Food and Agriculture Organization, 2010. Sustainable diets and biodiversity - Directions and 447 solutions for policy, research and action, in: International Scientific Symposium 448 "Biodiversity and Sustainable Diets United Against Hunger." 449 Gartner, A., El Ati, J., Traissac, P., Bour, A., Berger, J., Landais, E., El Hsaïni, H., Ben Rayana, 450 C., Delpeuch, F., 2014. A Double Burden of Overall or Central Adiposity and Anemia or 451 Iron Deficiency Is Prevalent but with Little Socioeconomic Patterning among Moroccan 452 and Tunisian Urban Women. J. Nutr. 144, 87–97. https://doi.org/10.3945/jn.113.178285 453 Gazan, R., Brouzes, C.M.C., Vieux, F., Maillot, M., Lluch, A., Darmon, N., 2018. Mathematical 454 Optimization to Explore Tomorrow's Sustainable Diets: A Narrative Review. Adv. Nutr. 455 9, 602–616. https://doi.org/10.1093/advances/nmy049 456 Hess, T., Andersson, U., Mena, C., Williams, A., 2015. The impact of healthier dietary 457 scenarios on the global blue water scarcity footprint of food consumption in the UK. Food 458 Policy 50, 1–10. https://doi.org/10.1016/J.FOODPOL.2014.10.013 459 Institute of Medicine, 1997. Dietary reference intakes: for calcium, phosphorus, magnesium, 460 vitamin D, and fluoride. National Academy Press. 461 IPBES, 2019. Global assessment report on biodiversity and ecosystem services of the 462 Intergovernmental Science- Policy Platform on Biodiversity and Ecosystem Services. 463 IPBES Secretariat, Bonn, Germany.

Mifflin, M.D., St Jeor, S.T., Hill, L. a, Scott, B.J., Daugherty, S. a, Koh, Y.O., 1990. A new 464 465 predictive equation for resting energy expenditure in healthy individuals. Am. J. Clin. Nutr. 466 51, 241–247. 467 NCD Risk Factor Collaboration (NCD-RisC), 2016a. Trends in adult body-mass index in 200 468 countries from 1975 to 2014: a pooled analysis of 1698 population-based measurement 469 studies with 19.2 million participants. Lancet 387, 1377-1396. 470 https://doi.org/10.1016/S0140-6736(16)30054-X 471 NCD Risk Factor Collaboration (NCD-RisC), 2016b. Worldwide trends in diabetes since 1980: 472 a pooled analysis of 751 population-based studies with 4.4 million participants. Lancet 473 387, 1513–1530. https://doi.org/10.1016/S0140-6736(16)00618-8 474 Payne, C.L., Scarborough, P., Cobiac, L., 2016. Do low-carbon-emission diets lead to higher 475 nutritional quality and positive health outcomes? A systematic review of the literature. 476 Public Health Nutr. 1–8. https://doi.org/10.1017/S1368980016000495 477 Perignon, M., Masset, G., Ferrari, G., Barré, T., Vieux, F., Maillot, M., Amiot, M.-J., Darmon, 478 N., 2016. How low can dietary greenhouse gas emissions be reduced without impairing 479 nutritional adequacy, affordability and acceptability of the diet? A modelling study to 480 guide sustainable food choices. **Public** Health Nutr. 19, 2662–2674. 481 https://doi.org/10.1017/S1368980016000653 482 Perignon, M., Vieux, F., Soler, L.-G., Masset, G., Darmon, N., 2017. Improving diet 483 sustainability through evolution of food choices: review of epidemiological studies on the 484 environmental impact of diets. Nutr. Rev. 75, 2–17. 485 Pfister, S., Koehler, A., Hellweg, S., 2009. Assessing the environmental impacts of freshwater 486 consumption in LCA. Environ. Sci. Technol. 43, 4098–104.

487	Ridoutt, B.G., Hendrie, G.A., Noakes, M., 2017. Dietary Strategies to Reduce Environmental	
488	Impact: A Critical Review of the Evidence Base. Adv. Nutr. 8, 933-946.	
489	https://doi.org/10.3945/an.117.016691	
490	Risku-Norja, H., Kurppa, S., Helenius, J., 2009. Dietary choices and greenhouse gas emissions	
491	– assessment of impact of vegetarian and organic options at national scale. Pr	
492	Ind. Ecol. An Int. J. 6, 340. https://doi.org/10.1504/PIE.2009.032323	
493	Sáez-Almendros, S., Obrador, B., Bach-Faig, A., Serra-Majem, L., 2013. Environmental	
494	footprints of Mediterranean versus Western dietary patterns: beyond the health benefits of	
495	the Mediterranean diet. Environ. Health 12, 118 (8pp). https://doi.org/10.1186/1476-	
496	069X-12-118	
497	Saxe, H., Larsen, T.M., Mogensen, L., 2013. The global warming potential of two healthy	
498	Nordic diets compared with the average Danish diet. Clim. Change 116, 249-262.	
499	https://doi.org/10.1007/s10584-012-0495-4	
500	Sinfort, C., Perignon, M., Drogue, S., Darmon, N., El Ati, J., Amiot, M., the MEDINA study	
501	Group, 2019. Potential environmental impacts of water deprivation and land use for food	
502	consumption in France and Tunisia. J. Environ. Qual. Under revi.	
503	Temme, E.H.M., van der Voet, H., Thissen, J., Verkaik-Kloosterman, J., van Donkersgoed, G.,	
504	Nonhebel, S., 2013. Replacement of meat and dairy by plant-derived foods: estimated	
505	effects on land use, iron and SFA intakes in young Dutch adult females. Public Health	
506	Nutr. 16, 1900–1907. https://doi.org/10.1017/s1368980013000232	
507	Thiébault, S., Moatti, JP., Ducrocq, V., Gaume, E., Dulac, F., Hamonou, E., Shin, YJ., Guiot,	
508	J., Cramer, W., Boulet, G., Guégan, JF., Barouki, R., Annesi-Maesano, I., Marty, P.,	
509	Torquebiau, E., Soussana, J.F., Aumeeruddy-Thomas, Y., Chotte, JL., Lacroix, D., 2016.	
510	The Mediterranean Region under Climate Change. A Scientific Update. IRD Éditions.	

511	Tom, M.S., Fischbeck, P.S., Hendrickson, C.T., 2015. Energy use, blue water footprint, and	
512	greenhouse gas emissions for current food consumption patterns and dietary	
513	recommendations in the US. Environ. Syst. Decis. https://doi.org/10.1007/s10669-015-	
514	9577-у	
515	Traissac, P., El Ati, J., Gartner, A., Ben Gharbia, H., Delpeuch, F., 2016. Gender inequalities	
516	in excess adiposity and anaemia combine in a large double burden of malnutrition gap	
517	detrimental to women in an urban area in North Africa. Public Health Nutr. 19, 1428–1437.	
518	https://doi.org/10.1017/S1368980016000689	
519	United Nations General Assembly, 2015. Transforming our world: The 2030 agenda for	
520	sustainable development.	
521	https://sustainabledevelopment.un.org/content/documents/7891Transforming%20Our%20	
522	World. pdf 1-5. https://doi.org/10.1007/s13398-014-0173-7.2	
523	US Department of Agriculture, 2008. National nutrient database for standard reference. Release	
524	21. Nutrient data laboratory.	
525	van Dooren, C., Marinussen, M., Blonk, H., Aiking, H., Vellinga, P., 2014. Exploring dietary	
526	guidelines based on ecological and nutritional values: A comparison of six dietary patterns.	
527	Food Policy 44, 36–46. https://doi.org/10.1016/j.foodpol.2013.11.002	
528	Vermeulen, S.J., Campbell, B.M., Ingram, J.S.I., 2012. Climate Change and Food Systems.	
529	Annu. Rev. Environ. Resour. 37, 195-222. https://doi.org/10.1146/annurev-environ-	
530	020411-130608	
531	Verner, D., World Bank. Middle East and North Africa Region. Sustainable Development.,	
532	2013. Tunisia in a changing climate: assessment and actions for increased resilience and	
533	development. World Bank.	

534	Vieux, F., Perignon, M., Gazan, R., Darmon, N., 2018. Dietary changes needed to improve diet			
535	sustainability: are they similar across Europe? Eur. J. Clin. Nutr. 1.			
536	https://doi.org/10.1038/s41430-017-0080-z			
537	Vieux, F., Soler, LG., Touazi, D., Darmon, N., 2013. High nutritional quality is not associated			
with low greenhouse gas emissions in self-selected diets of French adults. Am				
539	Nutr. 97, 569–83. https://doi.org/10.3945/ajcn.112.035105			
540	Whitmee, S., Haines, A., Beyrer, C., Boltz, F., Capon, A.G., de Souza Dias, B.F., Ezeh, A.,			
541	Frumkin, H., Gong, P., Head, P., Horton, R., Mace, G.M., Marten, R., Myers, S.S., Nishtar,			
542	S., Osofsky, S.A., Pattanayak, S.K., Pongsiri, M.J., Romanelli, C., Soucat, A., Vega, J.,			
543	Yach, D., 2015. Safeguarding human health in the Anthropocene epoch: report of The			
544	Rockefeller Foundation–Lancet Commission on planetary health. Lancet 386, 1973–2028.			
545	https://doi.org/10.1016/S0140-6736(15)60901-1			
546	WHO-FAO, 2004. Vitamin and mineral requirements in human nutrition - Second Edition,			
547	Report of a joint FAO/WHO expert consultation. World Health Organization.			
548	WHO-FAO, 2003. Diet, Nutrition, and the Prevention of Chronic Diseases: Report of a Joint			
549	WHO/FAO Expert Consultation, WHO Technical Report Series, No. 916 (TRS 916).			
550	World Health Organization, Geneva.			
551	WHO/FAO/UNU, 2007. Protein and amino acid requirements in human nutrition: Report of a			
552	joint FAO/WHO/UNU expert consultation. World Health Organization.			
553	WHO, 2012. Sodium intake for adults and children. Guidel. Potassium Intake Adults Child. 1-			
554	46.			
555	WHO, 2011. Guideline: Potassium intake for adults and children 52.			
556	WHO, 1996. Trace elements in human nutrition and health World Health Organization. World			

557	Heal. Organ. 1–360.
558	Willett, W., Rockström, J., Loken, B., Springmann, M., Lang, T., Vermeulen, S., Garnett, T.,
559	Tilman, D., DeClerck, F., Wood, A., Jonell, M., Clark, M., Gordon, L.J., Fanzo, J., Hawkes
560	C., Zurayk, R., Rivera, J.A., De Vries, W., Majele Sibanda, L., Afshin, A., Chaudhary, A.,
561	Herrero, M., Agustina, R., Branca, F., Lartey, A., Fan, S., Crona, B., Fox, E., Bignet, V.,
562	Troell, M., Lindahl, T., Singh, S., Cornell, S.E., Srinath Reddy, K., Narain, S., Nishtar, S.,
563	Murray, C.J.L., 2019. Food in the Anthropocene: the EAT-Lancet Commission on healthy
564	diets from sustainable food systems. Lancet 393, 447-492. https://doi.org/10.1016/S0140-
565	6736(18)31788-4
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8 Tables and Figures

nutrient content in the mean observed diet.

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Table 1: Nutritional constraints implemented in the diet optimization models, and

Nutrient	Constraint applied in the modeled diets ¹	Content in the mean observed diet ²
Proteins (g*kg of body weight³/d)	> 0.83 (WHO/FAO/UNU, 2007)	83.7
Carbohydrates (%E)	[50-75] (FAO/WHO, 2007)	50.7
Total fat (%E)	[15-35] (FAO, 2010)	35.6
Saturated fatty acids (%E)	< 10 (FAO, 2010)	7.4
Total PUFA ⁴ (%E)	[6-11] (FAO, 2010)	9.4
n-6 PUFA (%E)	[2.5-9] (FAO, 2010)	8.0
n-3 PUFA (%E)	[0.5-2] (FAO, 2010)	1.1
Cholesterol (mg/d)	<300 (WHO-FAO, 2003)	237.5
Fibers (g/d)	> 25 (WHO-FAO, 2003)	31.8
Free sugars (%E)	< 10 (WHO-FAO, 2003)	6.7
Vitamin A (μg RE/d)	[550-3000] (WHO-FAO, 2004)	751.6
Vitamin B1 (mg/d)	> 1.15 (WHO-FAO, 2004)	2.4
Vitamin B2 (mg/d)	> 1.2 (WHO-FAO, 2004)	2.2
Vitamin B3 (mg/d)	[15-35] (WHO-FAO, 2004)	27.7
Vitamin B5 (mg/d)	> 5 (WHO-FAO, 2004)	5.0
Vitamin B6 (mg/d)	[1.3-100] (WHO-FAO, 2004)	1.9
Folates (μg DFE/d)	[400-1000] (WHO-FAO, 2004)	706.3
Vitamin B12 (μg/d)	> 2.4 (WHO-FAO, 2004)	5.0
Vitamin E (mg α-tocopherol/d)	> 15 (WHO-FAO, 2004)	10.5
Vitamin C (mg/d)	[45-1000] (WHO-FAO, 2004)	168.3
Vitamin D (μg/d)	[5-50] (WHO-FAO, 2004)	3.1
Calcium (mg/d)	[1000-3000] (WHO-FAO, 2004)	723.6
Magnesium (mg/d)	[242-350] (WHO-FAO, 2004)	206.9
Zinc (mg/d)	[5.95-45] (WHO-FAO, 2004)	9.7
Selenium (µg/d)	[30-400] (WHO-FAO, 2004)	122.5
Iron (mg/d)	[21.5-45] (WHO-FAO, 2004)	18.2
Sodium (g/d)	< 2 (WHO, 2012)	4.7
Copper (mg/d)	[1.25-11] (WHO, 1996)	1.0
Potassium (mg/d)	> 3510 (WHO, 2011)	3146.8
Phosphorus (mg/d)	[700-4000] (Institute of Medicine, 1997)	1147.4
Total energy (kcal/d)	Equal to the total energy of the mean observed diet	2702

^{571 &}lt;sup>1</sup>Mean of the recommended dietary allowances for men and women

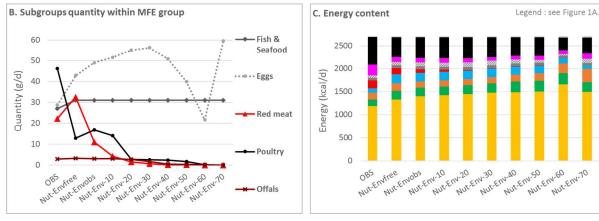
²Bold values indicate when a nutrient content does not fulfill the constraint.

^{573 &}lt;sup>3</sup>Mean body weight was estimated using national Tunisian dietary survey data

^{574 &}lt;sup>4</sup>Polyunsaturated fatty acids

Figure 1: Food groups quantity (A), Subgroup quantity within the Meat/Fish/Egg group (B), and Energy content (C) in the observed and optimized diets.

A. Food groups quantity -Starch Vegetables 700 -Fruits 600 Quantity (g/d) Dairy 500 -MFE 400 300 -HFSS 200 -Mixed dishes 100 Added fats & Seasonings -Drinks



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MFE: Meat/Fish/Egg; OBS: observed diet; Nut-Envfree: model without environmental constraints; Nut-Envobs: model with constraints limiting the environmental metrics to the observed level; Nut-Env₋₁₀, Nut-Env₋₂₀, ... etc: models with constraints imposing a 10% decrease of the environmental indicators at each step.

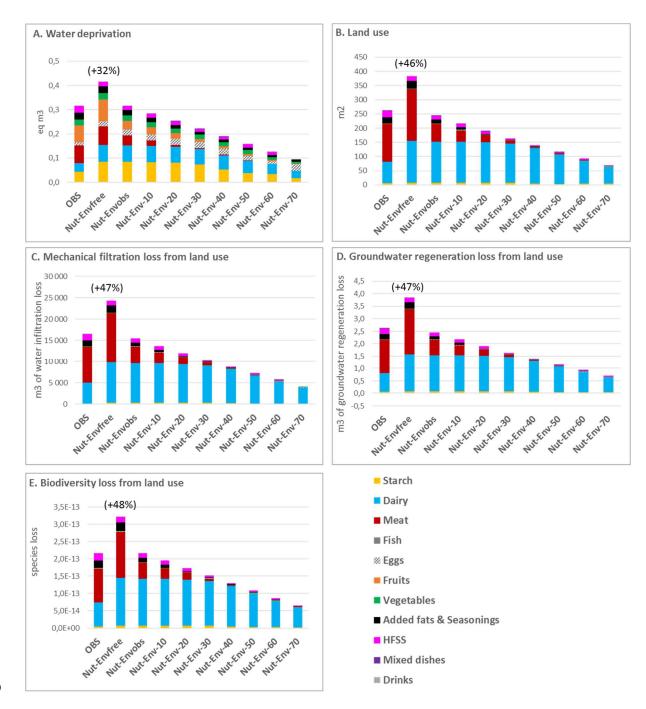
Nut-Env-60

Figure 2: Food group contributions to water deprivation (A), land use (B), land use impacts on mechanical filtration (C), groundwater regeneration (D), and biodiversity (E) in the observed and optimized diets ¹.



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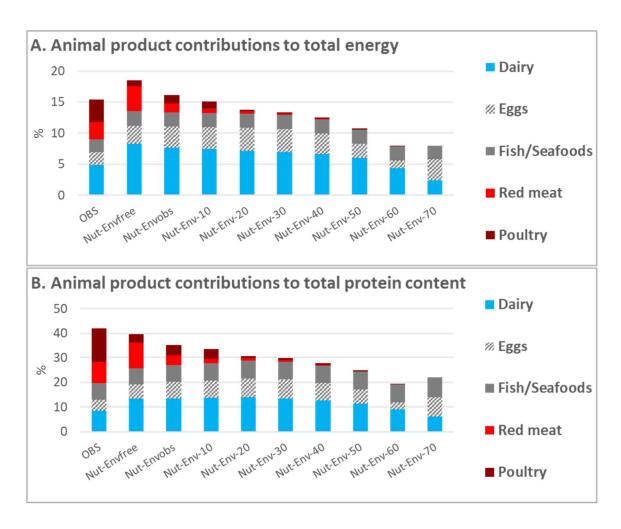
¹ Percentages between brackets show the increase of the Nut-Env_{free} diet impact (vs. impact of the observed diet). HFSS: foods high in fat/salt/sugar; OBS: observed diet; Nut-Env_{free}: model

without environmental constraints; Nut-Env_{obs}: model with constraints limiting the environmental metrics to the observed level; Nut-Env₋₁₀, Nut-Env₋₂₀, ... etc: models with constraints imposing a 10% decrease of the environmental indicators at each step.

Figure 3: Animal-based product contributions to total energy (A), and total protein content (B) in the observed and optimized diets.



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OBS: observed diet; Nut-Env_{free}: model without environmental constraints; Nut-Env_{obs}: model with constraints limiting the environmental metrics to the observed level; Nut-Env₋₁₀, Nut-Env₋₂₀, ... etc: models with constraints imposing a 10% decrease of the environmental indicators at each step