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Special Section: Erosion and Lateral Surface Processes

Core Ideas

- Runoff and erosion were continuously monitored for 7 yr at four nested spatial scales.
- A small number of runoff and erosive events were very important.
- There was a seasonal time lag between runoff and erosive response due to shrink–swell processes.
- Topsoil erosion processes and high hydro-erosive connectivity were predominant.
- A simple linear model explained mean annual erosion rates between spatial scales.

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A Spatiotemporal Multiscale Analysis of Runoff and Erosion in a Mediterranean Marly Catchment

Nesrine Inoubli, Damien Raclot,* Insaf Mekki, Roger Moussa, and Yves Le Bissonnais

An intensive monitoring investigation was conducted within the Kamech catchment (Cape Bon, Tunisia) to document the effect of scale on the hydrological and erosive responses in a Mediterranean climate regime with cracking soils. An evaluation of the individual contribution of topsoil and gully processes to total sediment delivery at different spatial scales was also conducted. Continuous runoff and erosion measurements during a 7-yr period at four nested gauging stations were examined in a plot (P1; 1.32 ha), a plot + gully (P2+G; 1.37 ha), a micro-catchment (MC; 15.2 ha), and the entire catchment (C; 263 ha). Data analysis showed a high interannual variability and contribution of only a few events; three major events generated approximately 20% of the total runoff and at least 50% of the total erosion at all of the monitoring stations. The results also highlight the strong seasonal dynamics of both water and sediment delivery, with a time lag observed between runoff and sediment responses. The presence of open cracks in autumn was identified as a main factor explaining this time lag. A constant monthly and annual specific runoff response was observed among the stations, and a simple linear model combining a single rate of topsoil ($17 \text{ t ha}^{-1} \text{ yr}^{-1}$) and gully ($80 \text{ t ha}^{-1} \text{ yr}^{-1}$) erosion (weighted by their contribution area) successfully predicted the differences observed in the mean annual erosion rates among the stations. Finally, this work showed the predominance of topsoil erosion processes in a Mediterranean catchment that is characterized by cracking soils and very high hydro-sedimentary connectivity.

Currently, soil erosion is a recognized environmental problem (Kefi and Yoshino, 2010). Scientists have devoted considerable efforts to understanding and predicting erosion occurrence. They have intensively studied soil erosion processes throughout the last decades using monitoring systems at the plot scale and at larger scales (Boix-Fayos et al., 2006; Cerdan et al., 2010; Vanmaercke et al., 2012). If water erosion factors are now well known (Wischmeier and Smith, 1978), the complex hierarchy of erosion processes across a wide range of spatial and temporal scales still needs to be studied (Wei et al., 2012; Cerdà et al., 2013). The study of the scale effect on soil erosion processes requires different methods to understand the redistribution of sediments along the source–transport–sink continuum (Lane et al., 1997). Research on soil erosion has established that factors controlling the movement of water and sediment vary according to a range of spatial scales, and the connections are complex (Cammaraat, 2002; de Vente and Poesen, 2005). Furthermore, the evolution of the system and the subsequent hydrological and erosion response are highly dependent on the interactions and dominant processes at the studied scales (Cammaraat, 2002; Boix-Fayos et al., 2007). Indeed, different factors and processes controlling runoff and erosion may act at distinct scales in different ways. Therefore, we need to understand the fact that processes at one spatial or temporal scale interact with processes at another scale, and the results of such cross-scale interactions often have nonlinear dynamics (Zehe and Sivapalan, 2009).

Several previous studies have shown a scale effect on runoff. Actually, hydrological studies that focus on the scale issue have documented that runoff per unit area generally decreases with increasing area (Le Bissonnais et al., 1998; Cerdan et al., 2004; Yair

and Raz-Yassif, 2004; Raclot et al., 2009; González-Hidalgo et al., 2011; Sadeghi et al., 2013). Some research has identified a dramatic effect of spatial scale on measured runoff and sediment yields under semiarid conditions, which has been attributed to scale-dependent variation in the importance and number of control factors and dominant processes (Yair and Raz-Yassif, 2004; Sadeghi et al., 2013). However, in terms of erosion, this trend of decreasing erosion with increasing area may be more complex (Cammeraat, 2002), absent (Mathier and Roy, 1996), or even inverse (de Vente and Poesen, 2005; de Vente et al., 2007; Vanmaercke et al., 2011). In most cases, the lack of a simple relationship demonstrates the complex and scale-dependent process domination throughout a catchment and emphasizes our uncertainty and the poor conceptual basis for direct extrapolation from the plot to catchment scale of erosion rates and sediment yields (Cammeraat, 2002; Parsons et al., 2006). At local scales (m^2), where sediment mobilization is dominated by rain splash, rills, and interrill erosion, sediment yield is likely to be relatively low. However, the variability in runoff and soil loss from plot measurements is reasonably high because local conditions can vary strongly (Nearing et al., 1999; Gómez et al., 2001). When moving from small (m^2) to larger areas (km^2), sediment yield is expected to increase due to an increase in connectivity and active erosion processes such as gully processes (de Vente et al., 2007). However, more sediment deposition and storage is expected with increasing basin area at parcel boundaries, in depressions, and at footslopes (Walling, 1983). For example, Lane et al. (1997), by describing the dominant processes controlling sediment yield at various scales, showed that sedimentation may be a dominant process. Reducing sediment exports from catchments requires an understanding of the source–transport–sink continuum to be able to develop efficient management strategies to combat soil erosion (Valentin et al., 2005). An alternative explanation that might account for scale dependency of runoff coefficients and erosion rates under natural rainfall is the effect of temporal variations in rainfall amount and/or intensity. Natural rainfall exhibits considerable temporal variation in intensity so that even if the average rainfall intensity exceeds the infiltration rate for the soil, this might not be the case for some parts of the storm (Wainwright and Parsons, 2002). Thus, the temporal variability of rainfall may greatly affect runoff production (Assouline et al., 2007) and consequently the erosion response.

The Mediterranean basin is very prone to erosion in all of its forms. Several Mediterranean countries are threatened by this phenomenon because of the climate, topography, soil sensitivity, and a very long history of human presence and intense cultivation (De Franchis and Ibanez, 2003; Raclot et al., 2017). The dominance of gully erosion as the main source of sediment in Mediterranean environments has been underlined by many researchers from the local (Roose et al., 2000; Collinet and Zante, 2005) to the catchment or watershed scale (Poesen et al., 1996, 2002, 2003). As a major sediment source, gully erosion is thus suspected to generate

subsequent downstream impacts in terms of reservoir siltation (de Vente et al., 2006, 2008). However, accurate field measurements of the respective contribution of individual erosion processes at the catchment scale are rarely available (Porto et al., 2014), and existing estimations are often given for a limited number of erosive events. As a result, there is a lack of information regarding the contribution of gully erosion to sediment fluxes from the local to the catchment scale in the mid- to long-term. Runoff and erosion processes are also very dependent on the soil type (Morgan, 2005). Vertisols or vertic intergrades are present in many parts of the Mediterranean landscape, especially where fine material with a high content of swelling clay minerals has accumulated (Yaalon, 1997). The highly dynamic shrink–swell processes of such soil types due to changing soil water storage conditions (Messing and Jarvis, 1993; Favre et al., 1997; Bagarello et al., 1999) generates soil cracks and extreme seasonal variations in soil infiltration rates. As a result, their hydro-erosive response appears to be very complex at the micro, plot, and catchment scales (Gómez et al., 2014; Inoubli et al., 2016). Therefore, studying the impact of cracking soils on runoff and erosion at different temporal and spatial scales also represents a specific and interesting challenge. To our knowledge, only a few studies have analyzed the scale effect on long-term hydrological and erosive responses under a Mediterranean climate regime (e.g., Yair and Raz-Yassif, 2004), and no such study has been performed in the context of Mediterranean cracking soils.

The major purposes of this study were (i) to provide a quantitative analysis of the surface runoff and the soil erosion response at different timescales (event to year) and at different spatial scales (1–250 ha) inside a small Mediterranean marly catchment under natural rainfall conditions and (ii) to evaluate and discuss the processes explaining the temporal and spatial variability of runoff and erosion in such a context.

Materials and Methods

The Study Area

The Kamech experimental catchment belongs to a long-term environmental research observatory called OMERE (Mediterranean Observatory of Water and Rural Environment), which aims to study the anthropogenic impacts on water and sediment budgets at the catchment scale. It is a 263-ha catchment upstream from a hill reservoir built in 1994 with an initial storage capacity of 140,000 m^3 (Fig. 1). The climate is semiarid, with a mean interannual temperature of 14°C , a mean annual reference evapotranspiration of 1100 mm estimated from daily evaporation pan measurements using a pan coefficient of 0.8, and a mean interannual rainfall of 650 mm (Fig. 2). A very dry period is generally observed between June and August, whereas the rainy season occurs from mid-September to mid-May. Rainstorm events with a short duration and high rainfall intensity predominantly occur during the hot months from May to October.

Fig. 1. The location of the Kamech catchment and the different runoff and sediment monitoring stations: the downstream hillslope reservoir (C, 263 ha) that drains all of the catchment area and the three intermediate hydrological stations—the experimental field (P1, 1.32 ha), the gully (P2+G, 1.37 ha), and the micro-catchment (MC, 15.2 ha).

Land use has been stable from 1994 until now. Annually plowed croplands occupy 70% of the catchment area and mainly occur on slopes of <15%. They are predominantly composed of cereal crops, which are sometimes cultivated in rotation with leguminous crops. The remaining 30% of the catchment area consists of dwellings, low Mediterranean scrublands, and sparsely vegetated gully or channel features. The only feature prone to sediment deposits inside the Kamech catchment consists of the outlet reservoir. According to the FAO classification (2006), the soil types observed within the catchment are vertisols or assimilated (65%), regosols (25.5%), and eutric regosols (9.5%). Approximately two-thirds of soils in Kamech are consequently very sensitive to cracking and exhibit numerous wide and deep open cracks during the entire dry summer period (Inoubli et al., 2016).

Instrument Setup and Available Data

The catchment has been instrumented for the automatic acquisition of instantaneous climatic and hydrologic data since 1995. The climatic instrument set was installed at the catchment outlet and includes tipping bucket rain gauges (0.5 mm), an evaporation pan, and an air thermometer. A water-level gauge was installed in the reservoir at the catchment outlet to provide continuous data on variations in the water level within the reservoir with 1-cm accuracy (5-min time steps). The spillway and bottom drain valves were shaped in such a way that output discharge could be continuously estimated through precise level-discharge curves. Bathymetric surveys of the reservoir are made every 3 yr on average, which makes it possible to update the level-volume and level-surface curves for reservoir storage. Sediment concentrations during rare overflow through the spillway or emptying through the bottom drain valve were also evaluated using a few manual sampling surveys. These

sets of measurements enable the evaluation of the continuous runoff input into the reservoir by drawing a hydrologic budget of the reservoir, as explained by Albergel et al. (1999, 2007), Mekki et al. (2006), and Raclot and Albergel (2006). In addition, the sediment input into the reservoir was computed between successive bathymetric surveys by adding the sediment trapped in the reservoir and the sediment outputs through the spillway and bottom drain valve (Raclot and Albergel, 2006).

To describe the spatial and temporal variability of erosion, three hydrological stations following a hierarchical and nested design (Fig. 1) were installed between 2004 and 2005 at the outlet of (i) an experimental field (P1, 1.32 ha), (ii) a gully (P2+G, 1.37 ha) composed of a ravine (G) of 1700 m² that drains a cultivated field (P2) of 1.2 ha, and (iii) a micro-catchment (MC, 15.2 ha) that consists

Fig. 2. Mean monthly rainfall, temperature, and mean annual reference evapotranspiration (ET0) at the Kamech catchment for the period 2005 to 2012 as measured at the catchment outlet.

of 10.9 ha of cultivated land, 3.6 ha of scrublands, and 0.7 ha of gullies. Both P1 and P2+G are located within MC.

All of the gauging stations were equipped with two water level sensors that continuously monitored discharge (at a 1-min time step) using a combination of a rectangular channel (MC) or Venturi flume (P1 and P2+G) for high runoff rates and a V-shape weir for low runoff rates. An automatic water sampler with 24 glass bottles (Sigma Model 900P, Hach Company) was programmed to sample water in relation to the runoff level variations during floods. Automatic sampling was often complemented by manual sampling to increase the number of samples. The suspended sediment concentration for each water sample was determined using a high-precision weighing scale (Gibertini Crystal 500 CAL). The total mass of the transported suspended sediment for an event was then calculated by integrating the sediment concentrations with the flow discharge.

In addition, vegetation cover development for the different crops and the evolution of topsoil cracks were documented every 15 or 30 d on 25 fields in Kamech that were selected to represent the main crop and soil type associations. Intra-annual evolution of vegetation cover was reported by Raclot and Albergel (2006) and intra-annual evolution of cracks by Inoubli et al. (2016). An intensive topographic monitoring of a representative permanent gully (included in P2+G) was also conducted in Kamech between 2002 and 2009 using a total station (El Khalili et al., 2013). This monitoring enabled us to estimate the long-term mean interannual gully erosion rate in Kamech at $80 \text{ t ha}^{-1} \text{ yr}^{-1}$.

As a result, the Kamech monitoring system enabled us to derive the following variables:

- continuous runoff (mm) and the long-term mean interannual soil erosion rate ($\text{t ha}^{-1} \text{ yr}^{-1}$) at the catchment outlet (C) during the 1994 to 2012 period;
- rainfall amount (mm), maximum rainfall intensity in 15 min (mm h^{-1}), runoff amount (mm), soil erosion (t ha^{-1}), and mean sediment concentration (g L^{-1}) at P1, P2+G, and MC for all runoff events that occurred between 1 Sept. 2005 and 31 Aug. 2012;
- vegetation cover and topsoil crack evolution during the same period;
- the long-term mean interannual gully erosion rate in Kamech.

Multiple uncertainty sources can impact measurement precision: instrumental uncertainties or failure, water-height to flow-rate calibration curve quality, suspended sediment concentration measurements, and an integration method to derive soil erosion from the flow rates and suspended sediment concentrations. Moreover, hydrological data precision varies greatly from one event to another because most of these uncertainties are not linear (e.g., water level

and discharge measurement precision depends on the water level). Repeated sensor calibration and control were performed to avoid bias in the data measurement and to derive good expertise of data quality for the Kamech monitoring system. Using this expertise, a global uncertainty value was evaluated to <15% for event runoff and <30% for event soil erosion.

Methodology Approach

Pretreatment of Experimental Data at the Event Scale

Every event that occurred between 1 Sept. 2005 and 31 Aug. 2012 with a rainfall amount $>2 \text{ mm}$ and separated by at least 1 h was considered, but only those that have generated runoff at least at one monitored station were included in the database. For some rare events, both water level and sediment concentration values were missing because of battery failures. For a few other events, only sediment concentration values were missing because the water level didn't reach the minimum sampling water level or because of automatic water sampler malfunctioning.

A pretreatment was proposed to identify and reconstruct the missing values. For missing runoff data, the inter-station relationships (Fig. 3a; Table 1) were used because the rainfall–runoff relationships didn't provide satisfactory prediction capacities. The inter-station relationship with the best determination coefficient was finally used:

- P1 vs. MC relationship for missing runoff at P1.
- P2+G vs. MC relationship for missing runoff at P2+G or MC.

For the missing erosion values, the runoff parameters measured at the same station were used each time they were available to limit the impact of a possible shift in response between stations. In other cases, inter-station relationships (Fig. 3b) were used.

All of these relationships (Table 1) showed good prediction performance for the three stations. Depending on the considered station, 6 to 11% of the total runoff amount and 20 to 27% of the total erosion amount were reconstructed. The limited contributions of the reconstructed values and the strong prediction ability of the relationships used to evaluate missing values enabled us to elaborate on a reconstructed database that can be used with acceptable confidence.

Spatial and Temporal Multiscale Analysis

The analysis of the temporal and spatial scale effect on runoff and erosion was based on the temporal pattern of rainfall–runoff–erosion rates at the four nested spatial scales for 7 consecutive yr. This pattern was first described using plots of cumulative event-based values for each of the seven monitored hydrological years and for each of the four nested gauging stations. A synthetic view of the temporal pattern of these variables was also provided at monthly and annual timescales for the different stations. The main factors and processes explaining the runoff and erosion response at

Fig. 3. (a) Runoff and (b) soil erosion inter-correlations among the gauging stations for the downstream hillslope reservoir (C) that drains all of the catchment area, the experimental field (P1), the gully (P2+G), and the micro-catchment (MC) (blue circles represent event measured data and red crosses represent event reconstructed data).

different spatial and temporal scales were then discussed based on these plots.

A simple linear model that combines sediment sources due to gully erosion processes and sediment sources due to topsoil erosion processes was tested to explain the relative long-term contribution of erosive processes at the P2+G, MC, and C outlets:

$$E_{\text{Total}} = \alpha E_{\text{Topsoil}} + \beta E_{\text{Gullies}} \quad [1]$$

where E_{Total} is the long-term total erosion load ($t\ ha^{-1}\ yr^{-1}$), E_{Topsoil} is long-term erosion ($t\ ha^{-1}\ yr^{-1}$) of topsoil in the

cultivated field, E_{Gullies} is long-term erosion ($t\ ha^{-1}\ yr^{-1}$) from gullies, α is the percentage of cultivated topsoil surface, and β is the percentage of gully surface.

In this model, α and β were quantified through a combination of aerial photograph interpretation and field surveys, whereas long-term topsoil (E_{Topsoil}) and gully (E_{Gullies}) erosion rates were assumed to be known and representative of any parts of the Kamech catchment. In this study, E_{Topsoil} was set to the measured erosion rate between 2005 and 2012 at the experimental plot P1 (Es_P1 , this study) and E_{Gullies} was set to $80\ t\ ha^{-1}\ yr^{-1}$ as estimated by El Khalili et al. (2013). Erosion from dwellings

Table 1. The reconstruction procedure for runoff (R , mm) or soil erosion (Es , $t\ ha^{-1}$) used to evaluate missing data in the event database at the three monitoring stations: the experimental field (P1), the gully (P2+G), and the micro-catchment (MC); Q_{max} is the maximum specific runoff discharge ($m^3\ s^{-1}\ ha^{-1}$) and R^2 is the linear determination coefficient.

Event with missing data	Only erosion data missing	Both runoff and erosion data missing
Runoff amount evaluation relationships		$R_P1 = R_MC/1.12$ ($R^2 = 0.92$)
		$R_P1 = R_P2+G/1.06$ ($R^2 = 0.90$)
		$R_MC = 1.01R_P2+G$ ($R^2 = 0.95$)
Erosion rate evaluation relationships	$Es_P1 = 21.29Q_{\text{max}}^{1.16}R^{0.45}$ ($R^2 = 0.74$)	$Es_P1 = Es_MC/0.54$ ($R^2 = 0.62$)
	$Es_MC = 5.96Q_{\text{max}}^{1.27}R^{0.45}$ ($R^2 = 0.92$)	$Es_P1 = Es_P2+G/1.24$ ($R^2 = 0.58$)
	$Es_P2+G = 269.78Q_{\text{max}}^{1.76}R^{0.45}$ ($R^2 = 0.90$)	$Es_MC = 0.50Es_P2+G$ ($R^2 = 0.65$)

and scrublands was assumed to be negligible because (i) dwellings are located in the northeastern part of the catchment, which is a very flat area, (ii) most of dwellings are sedimentologically disconnected from the catchment by dense vegetative hedges at their edges that are used as animal enclosures and field fence, and (iii) the scrublands are located on sandstone outcrops that represent <10% of the catchment area and don't show field evidence of important erosion activities.

Results

Rainfall and Runoff Event Distribution

A total of 346 distinct rainfall events were recorded between 1 Sept. 2005 and 31 Aug. 2012 in the Kamech catchment (Table 2). Most rainfall events occurred during the winter and spring seasons (September–April), whereas a smaller number of rainfall events occurred in May, and the summer months were nearly always dry. Only approximately 50% of these rainfall events generated runoff at least at one of the gauging stations during the same period. These runoff events predominantly occurred between October and March.

Runoff and Erosion at an Event-Based Timescale

The cumulative rainfall curves for the 7 hydrological yr between 2005 and 2012 (Fig. 4a) show similar inter-annual trends, with a clear distinction between the rainy and dry seasons. However, some inter-annual differences related to the rainfall distribution and amount occurred. For example, annual rainfall varied

between 518 mm for the hydrological year 2006–2007 and 793 mm for 2008–2009. For 2 yr with similar annual amounts (for example 2006–2007 and 2009–2010), the rainfall regime can also be quite different.

Compared with rainfall, the runoff patterns show a higher inter-annual variability (Fig. 4b). Both the runoff regime and annual runoff amounts were considered. For example, annual runoff amounts varied from 25 mm for 2006–2007 to 216 mm for 2008–2009. The runoff regime is more compressed in time than the rainfall regime, as only a few rainfall events contributed to the annual runoff amounts. Runoff generally began only after October (2009–2010) and more generally in the beginning of December, with a significant increase until April. An exceptional 80-mm rainfall event was recorded in December 2006, contributing up to 40% of the total runoff amount in 2006–2007 at MC (37% at P1 and 40% at P2+G).

Intra-annual soil erosion patterns were approximately similar to intra-annual runoff patterns, but some significant increments in the cumulative erosion curve have disappeared compared with the cumulative runoff curves. Time compression was higher for erosion than for runoff because fewer erosion events than runoff events contributed to the total annual amounts. The contribution of a single event in December 2006 (Fig. 4c) represents approximately 65% of the total soil erosion at MC in 2006–2007 (56% at P1 and 63% at P2+G) and 38% of the total soil erosion at MC during the 7 yr (30% at P1 and 36% at P2+G). The contribution of the three largest events represented >50% of the total eroded material at the three stations. Because the three largest erosive events

Table 2. The distribution of monthly rainfall and runoff events for the seven hydrological years between 2005 and 2012.

Year	Sept.	Oct.	Nov.	Dec.	Jan.	Feb.	Mar.	Apr.	May	June	July	Aug.	Annual
<u>Number of rainfall events</u>													
2005–2006	5	1	10	5	11	4	4	1	2	1	0	1	45
2006–2007	5	4	5	8	1	10	9	2	1	0	0	2	47
2007–2008	2	10	10	8	2	7	9	2	3	0	0	0	53
2008–2009	7	4	3	3	11	7	6	7	1	0	0	1	50
2009–2010	11	11	6	5	7	6	3	5	4	0	0	0	58
2010–2011	6	6	6	5	5	6	3	4	3	0	0	0	44
2011–2012	1	7	6	6	9	8	5	4	2	0	1	0	49
Sum	37	43	46	40	46	48	39	25	16	1	1	4	346
<u>Number of runoff events</u>													
2005–2006	0	0	0	4	12	3	0	0	0	0	0	0	18
2006–2007	1	2	0	6	0	5	4	0	0	0	0	0	18
2007–2008	1	7	6	7	0	1	1	0	0	0	0	0	23
2008–2009	0	3	0	1	6	8	5	4	0	0	0	0	26
2009–2010	0	11	5	1	4	4	3	1	1	0	0	0	30
2010–2011	2	3	3	4	3	5	2	0	0	0	0	0	22
2011–2012	0	3	3	5	5	8	3	1	0	0	0	0	28
Sum	4	29	17	28	30	34	18	6	1	0	0	0	167

Fig. 4. Mean cumulative rainfall, runoff, and soil loss, for the different hydrological years (2005–2012) at the micro-catchment (MC) station.

occurred consecutively in the same month (December 2006), the 2006–2007 hydrological year clearly appears as an exceptional erosive year with $>58 \text{ t ha}^{-1}$ at MC (64 t ha^{-1} at P1 and 96 t ha^{-1} at P2+G), whereas the total erosion rate in 2010–2011 was only 0.65 t ha^{-1} at MC (0.41 t ha^{-1} at P1 and 0.31 t ha^{-1} at P2+G).

The comparison of runoff and erosion responses at the different stations is illustrated in Fig. 5 by considering two hydrological years with a similar rainfall amount but a different rainfall regime. The comparison of the runoff response at the different stations shows a very low variability. The inter-station runoff response variability appears to be much smaller than the inter-annual runoff response variability at one station. Very similar runoff responses for MC and C were always observed, whereas the runoff response

Fig. 5. Mean cumulative rainfall, runoff, and soil loss at the gauging stations for the downstream hillslope reservoir (C) that drains all of the catchment area, the experimental field (P1), the gully (P2+G), and the micro-catchment (MC) for the hydrological years 2006–2007 and 2009–2010. Note that the right-hand axes are scaled differently.

at a fine spatial scale can differ from the runoff response at large scales (P1 in 2006–2007 or P2+G in 2009–2010). In terms of soil loss, similar patterns were observed, with generally higher erosion rates at P2+G than at P1 or MC.

Runoff and Erosion at a Monthly Timescale

The mean monthly rainfall, runoff, and sediment concentration distributions at the four gauging stations were analyzed and compared for the period from 2005 to 2012 (Fig. 6).

Runoff rates at the four stations exhibited approximately similar evolution throughout the year: low runoff from September to November and high runoff from December to March or April. Approximately 70% of the total annual runoff occurred between December and February, whereas only 40% of the total annual rainfall occurred during these 3 mo. In contrast, $<8\%$ of the runoff occurred from September to November, whereas these 3 mo contributed $>35\%$ of the annual rainfall. If topsoil moisture plays a role in runoff generation in Kamech, the monitoring of several piezometers showed that the soil profile was never completely saturated

Fig. 6. Mean monthly rainfall, runoff, and sediment concentrations at the gauging stations for the downstream hillslope reservoir (C) that drains all of the catchment area, the experimental field (P1), the gully (P2+G), and the micro-catchment (MC) for the period from 2005 to 2012.

except very locally along the main channel features (results not shown). Consequently, an infiltration excess (Hortonian) overland flow process dominated the runoff generation process in Kamech, and the intra-annual evolution of runoff cannot be explained by the saturation excess overland flow process.

In terms of sediment concentrations, some very similar seasonal variations were observed at the three stations, with the exception of September values that are based on nonsignificant runoff amounts. Very high mean monthly sediment concentrations (generally up to 20 g L^{-1}) were observed from October to December. During these 3 mo, quite similar sediment concentration values were recorded for P1 and P2+G, whereas the sediment concentration at MC was lower. From January to May, sediment concentrations were similar for the three stations and clearly lower (approximately 5 g L^{-1}) except for April, where the mean monthly sediment concentrations increased to approximately 10 g L^{-1} for MC and P1 and to 20 g L^{-1} for P2+G.

A time lag between runoff and soil erosion responses is apparent, with maximum sediment concentrations occurring in autumn while the maximum runoff response occurred in winter (Fig. 6).

As a result, the month of December clearly appears to be the critical month for soil erosion because of the convergence of high runoff rates and still high sediment concentration values. October appears to be the second most significant month in terms of the soil erosion contribution and despite its low runoff amount.

Runoff and Erosion at the Annual Scale

The mean annual runoff at the different stations was not significantly different (Fig. 7a) when considering the range of measurement uncertainties. The annual runoff coefficients varied from approximately 5% for 2010–2011 to about 30% for 2006–2007 and differed slightly from one station to another.

Only erosion at the P2+G station was significantly higher than erosion at the other stations (Fig. 7b). This higher soil budget recorded for P2+G compared with P1 may be attributed to the high contribution of gully erosion processes.

Contributions of Erosion Processes

Table 3 presents the repartitioning among cultivated topsoil, gullies, and other areas (dwellings and scrublands) at different scales.

Fig. 7. Mean annual (a) runoff and (b) soil erosion calculated between 2005 and 2012 at the downstream hillslope reservoir (C) that drains all of the catchment area, the experimental field (P1), the gully (P2+G), and the micro-catchment (MC) (the soil erosion rate at C was calculated for the period between 1994 and 2012). A systematic range of 15% for runoff measurements and 30% for soil erosion measurements was considered to depict measurement uncertainties.

Table 3. Percentages of cultivated topsoil, gullies, and other areas at the spatial scales of the downstream hillslope reservoir (C) that drains all of the catchment area, the experimental field (P1), the gully (P2+G), and the micro-catchment (MC).

Scale	Cultivated topsoil surface	Gully surface	Other area
	%		
P1	100	0	0
P2+G	87.6	12.4	0
MC	71.7	4.6	23.7
C	70	4.7	25.3

Based on these values, the results of the linear model (Fig. 8) showed that the values for long-term erosion rates calculated by the model fit well with measured erosion data whatever the considered station ($R^2 = 0.97$). In other words, the erosion amount in the Kamech catchment seems to linearly depend on the relative percentage of the upstream cultivated and gully areas.

The determination of the sediment sources in the Kamech catchment using this simple linear model showed that the surface topsoil delivered approximately 75% of the sediment at the MC and C stations (Fig. 9). This result implies that surface erosion processes, including rill and interrill erosion, are the dominant processes at the catchment scale and that gully erosion processes contribute only 25% of the sediment entering into the lake reservoir.

Discussion

Temporal Variability of Runoff and Soil Erosion

Our study, which is based on seven continuous years of measurements, showed important temporal variability for both surface runoff and soil erosion. First, runoff and soil erosion showed a strong inter-annual variability that is quite typical in a Mediterranean context (Rodrigo Comino et al., 2016; Raclot et al., 2017) and can be directly related with rainfall regime variability. Second, our results highlighted a strong seasonality of runoff and soil erosion responses that resulted in a quite unusual time lag between runoff and soil erosion responses. Both the presence of cracks, especially in autumn, and the vegetation cover growth in spring were identified as major factors explaining the intra-annual evolution of the runoff and erosion response. Four characteristic periods were identified:

- September–November: Heavy rainfall on bare soil with open cracks induced very low runoff with very high sediment concentrations. The low runoff that occurred when topsoil cracks were present can be related to bypass flow, which enters the soil through open topsoil cracks (Inoubli et al., 2016). The high sediment concentration in surface runoff in autumn may be the

Fig. 8. Erosion calculated with the linear model as a function of measured erosion at the different stations: the downstream hillslope reservoir (C) that drains all of the catchment area, the experimental field (P1), the gully (P2+G), and the micro-catchment (MC).

result of soil shrinkage processes that occurred during the dry season and destroyed topsoil crusting and aggregation, which favored future soil particle detachment and export. Another hypothesis is that the shrink–swell process generates microdepressions close to the cracks that can channelize runoff and locally activate micro-rill processes. Agricultural tillage operations that generally occurred in November when a cumulated rainfall amount had reached 150 mm may also contribute to maintain both high infiltration rates and topsoil conditions that favor the detachment process. Finally, the cumulated soil loss amount during this period reached approximately 15% of the total soil loss, whereas cumulated runoff reached only approximately 7%. In fact, significant sediment yields occurred only during a very few runoff events. This period could be considered a high sediment yield risk because of very high sediment fluxes as soon as runoff occurred.

- December–February: High rainfall amounts with low vegetation cover and recently closed cracks produced the majority of the runoff (approximately 70% of the total runoff) and soil losses (approximately 75% of the total soil losses). December alone appeared to be by far the riskiest month in terms of soil

Fig. 9. Percentages of contributions of processes to total sediment load at the different gauging stations: the downstream hillslope reservoir (C) that drains all of the catchment area, the experimental field (P1), the gully (P2+G), and the micro-catchment (MC).

losses, with a contribution to total soil loss of 60% despite a contribution to total runoff of <30%. The very significant decrease in sediment concentration that occurred between December and January was probably the result of (i) the combined effects of high particle flushing that occurred in December, (ii) the evolution of soil surface characteristics induced by successive rainfall events, and (iii) progressive vegetation growth.

- March–May: Medium to dense vegetation cover with low rainfall amounts produced low runoff, low sediment concentration, and low soil erosion.
- June–August: Very low rainfall amounts on bare soil and newly opened cracks generated no runoff and no soil erosion.

Numerous previous studies have already identified a seasonality of the infiltration–runoff response in the Mediterranean context. For example, Cerdà (1997, 2002) described the evolution of the infiltration rate in relation to soil moisture and showed greater soil erosion rates on eroded surfaces in eastern Spain during the wet season. Jorgensen and Gardner (1987) found the importance of the surface morphology for seasonal infiltration and runoff rates on disturbed soils, as already mentioned by Schümm and Lusby (1963) on badland slopes. In Kamech, the presence of open cracks in autumn was clearly the main factor affecting both runoff and erosion in the beginning of the hydrological year. This is the first time to our knowledge that a time lag between the erosion response and the runoff response has been so clearly highlighted as a result of the complex temporal interactions between several factors and processes.

In some cases, soil erosion can be dominated by a few extreme events. Therefore, monitoring and simulation studies need to be long enough to capture these erosive events. For our study and despite the variation in rainfall–runoff–sediment responses at different scales, the contribution of the major events to the annual balance was remarkably high. The data show a strong sediment time compression, with the contribution of a few events representing a very high proportion of the total soil losses. A similar result was found by González-Hidalgo et al. (2007) in western Mediterranean areas, where three daily erosive events per annum represented >50% of the soil erosion and regardless of the total amount of soil eroded. Analyses of the extreme events already suggested that a single event (or just a few events) may produce a high amount of sediment transport, and they can exceed the annual or interannual geomorphic activity (Lane and Kidwell, 2003; Lenzi et al., 2003). This explains why research attention has focused on major events and their effects (Hudson, 2003; Tomkins et al., 2007). In Kamech, the high availability of soil particles eroded during extreme events seems to be promoted by the shrink–swell processes occurring during the dry season. This behavior is different of what occurs in other catchments, where the availability of soil particles can result from successive and previous low-intensity events, especially in the case of catchments with low sedimentological connectivity where intra-catchment deposition is favored (Raclot et al., 2009).

Spatial Variability of Runoff and Soil Erosion

In Kamech, runoff slightly increased from the field to the micro-catchment and then was constant from the micro-catchment to the catchment outlet. Previous studies in Mediterranean areas (Cammeraat, 2004; Boix-Fayos et al., 2007) and in other semi-arid areas throughout the world (Estèves and Lapetite, 2003; Yair and Raz-Yassif, 2004; Parsons et al., 2006) generally described a decrease of runoff with an increase in the spatial scale. These researchers explained this decrease in runoff with increasing spatial area by an increase in spatial heterogeneity in the soil surface characteristics and the emergence of other processes such as re-infiltration in downslope areas of a catchment or transmission losses in stream channels. In the studied catchment, the runoff response didn't follow this downward trend but was found to be similar between the monitored spatial scale at an event (Fig. 3) or monthly (Fig. 6) timescale. Such a linear runoff response across the spatial scale can be explained by the specific context of the Kamech catchment, which combines (i) a large predominance of a Hortonian overflow process, (ii) a very high hydrological connectivity induced by a dense permanent gully network, and (iii) a very dry summer season that limits vegetation cover inside the gully and channel features. Some minor deviations from this linear behavior may be noticed, however. For example, the runoff amount at the catchment outlet is slightly higher than at other station outlets during March and April, which is in line with the observation of short periods of wadi baseflow that occurred in spring at the catchment outlet and indicated a small contribution of shallow groundwater to wadi flow. Another example is the small increase in runoff between the field and the micro-catchment that can be explained by the presence of a compacted track inside the micro-catchment that behaved as a low-infiltration area.

Sediment loads usually depend on a wide range of variables and reflect the combined effect of all active and interacting runoff, erosion, and sediment deposition processes and the time duration of the measurements (de Vente and Poesen, 2005). Surprisingly, soil erosion dynamics in the studied area showed very similar patterns from field to catchment outlets. The difference in mean inter-annual soil erosion amounts when moving from one station to another was well explained using a linear model based on the percentage of gully and cultivated areas. Again, the very high hydro-sedimentary connectivity of the Kamech catchment, which is confirmed by field evidence of the absence of significant permanent sediment deposition areas upstream of the outlet reservoir, is likely to be the main reason of such a linear behavior. In fact, the dense network of permanent gullies and wadis in the Kamech catchment acts as both sediment sources and connectivity features that facilitate sediment transport from field to catchment outlet.

Erosion Processes and Sediment Sources

The determination of the sediment sources in the Kamech catchment using a simple linear model showed that the topsoil erosion processes delivered approximately 75% of the sediment at the

Kamech catchment lake during a period of 7 yr. This result is consistent with results of Ben Slimane et al. (2013), who found, using a mixing model based on radionuclides, that 80% of the soil inputs into the Kamech reservoir during the 1994 to 2012 period came from topsoil sources. This result is also in agreement with the results obtained by Jebari et al. (2010), who used a rainfall erosivity analysis to calculate that interrill processes produced 83% of the erosion within the Kamech catchment. However, the results presented here differ from the conclusions of several other erosion studies conducted in the Mediterranean region that indicate a predominance of gully erosion (Poesen et al., 1996, 2002, 2003). One explanation for this difference may be that the low rates of sheet and rill erosion and the relatively large importance of gully erosion have often been attributed to the shallow depth and high extent of stoniness of many Mediterranean soils (Poesen et al., 1994; Poesen and Hooke, 1997; Cerdan et al., 2010), which is not the case in the study catchment.

Conclusions

This study documented the temporal dynamics of the runoff and erosion response during a 7-yr period at the outlet of four nested gauging stations in a cultivated Mediterranean catchment dominated by cracking soils. The presence of open cracks was identified as a major factor controlling both the runoff and soil loss response in autumn whatever the spatial scale. December was identified as a very risky month in terms of soil losses, and this single month contributed 60% of the total soil loss in the year. The autumn period was also identified as an erosion-risk period, with very high average sediment concentrations (up to 50 g L⁻¹). The contribution of a very few major events to the annual balance was remarkably high for the whole study period, which revealed an extreme sediment time compression in this catchment. In terms of spatial variations, the runoff amounts increased with an increase in the drainage area or scale so that the specific runoff rate was quite similar across scales. This behavior has been explained by a high predominance of the Hortonian runoff process and very high hydrological connectivity. For soil erosion, a slight decrease in the erosion budgets was observed between the field and the catchment scale. A simple linear model combining a single rate of topsoil erosion and a single rate of gully erosion and weighted by their respective surface areas successfully reproduced the differences observed in the mean annual erosion rates between the spatial scales. The numerous permanent gullies in Kamech have to be considered both as sediment contributing sources and as connectivity components. Finally, our results showed that major erosion processes within the Kamech catchment were related to sheet and rill erosion, which contributed 75% of the total soil erosion budget at the catchment outlet, and that most of sediment yields occurred from October to December. Thus, this study contributed to our understanding of hydrological and erosive processes in a cultivated marly Mediterranean catchment and provided important information to help identify effective sediment control techniques adapted to such a context.

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References

- Albergel, J., S. Nasri, and M. Boufaroua. 1999. Small dams water balance: Experimental conditions, data processing and modeling in arid and semiarid areas. In: R. Berndtsson, editor, *Proceedings of the International Seminar on Rain Water Harvesting and Management of Small Reservoirs in Arid and Semiarid Areas*, Lund, Sweden. 29 June–2 July 1998. Dep. of Water Resour. Eng., Lund Inst. of Technol., Lund, Sweden. p. 45–48.
- Albergel, J., S. Nasri, and J.-M. Lamachère. 2007. Les lacs et barrages collinaires: Infrastructures hydrauliques d'aménagement et de conservation des terres agricoles en zone semi-aride méditerranéenne. *Sols Tunisie* 2:7–20.
- Assouline, S., J.S. Selker, and J.-Y. Parlange. 2007. A simple accurate method to predict time of ponding under variable intensity rainfall. *Water Resour. Res.* 43:W03426. doi:10.1029/2006WR005138
- Bagarello, V., M. Iovino, and W. Reynolds. 1999. Measuring hydraulic conductivity in a cracking clay soil using the Guelph permeameter. *Trans. ASAE* 42:957–964. doi:10.13031/2013.13276
- Ben Slimane, A., D. Raclot, O. Evrard, M. Sanaa, I. Lefèvre, A. Mehdi, et al. 2013. Fingerprinting sediment sources in the outlet reservoir of a hilly cultivated catchment in Tunisia. *J. Soils Sediments* 13:801–815. doi:10.1007/s11368-012-0642-6
- Boix-Fayos, C., M. Martínez-Mena, E. Arnau-Rosalén, A. Calvo-Cases, V. Castillo, and J. Albaladejo. 2006. Measuring soil erosion by field plots: Understanding the sources of variation. *Earth Sci. Rev.* 78:267–285. doi:10.1016/j.earscirev.2006.05.005
- Boix-Fayos, C., M. Martínez-Mena, A. Calvo Cases, E. Arnau Rosalén, J. Albaladejo, and V. Castillo. 2007. Causes and underlying processes of measurement variability in field erosion plots in Mediterranean conditions. *Earth Surf. Processes Landforms* 32:85–101. doi:10.1002/esp.1382
- Cammeraat, L.H. 2002. A review of two strongly contrasting geomorphological systems within the context of scale. *Earth Surf. Processes Landforms* 27:1201–1222. doi:10.1002/esp.421
- Cammeraat, L.H. 2004. Scale dependent thresholds in hydrological and erosion response of a semi-arid catchment in southeast Spain. *Agric. Ecosyst. Environ.* 104:317–332. doi:10.1016/j.agee.2004.01.032
- Cerdà, A. 1997. Seasonal changes of the infiltration rates in a Mediterranean scrubland on limestone. *J. Hydrol.* 198:209–225. doi:10.1016/S0022-1694(96)03295-7
- Cerdà, A. 2002. The effect of season and parent material on water erosion on highly eroded soils in eastern Spain. *J. Arid Environ.* 52:319–337. doi:10.1006/jare.2002.1009
- Cerdà, A., R. Brazier, M. Nearing, and J. de Vente. 2013. Scales and erosion. *Catena* 102:1–2. doi:10.1016/j.catena.2011.09.006
- Cerdan, O., G. Govers, Y. Le Bissonnais, K. Van Oost, J. Poesen, N. Saby, et al. 2010. Rates and spatial variations of soil erosion in Europe: A study based on erosion plot data. *Geomorphology* 122:167–177. doi:10.1016/j.geomorph.2010.06.011
- Cerdan, O., Y. Le Bissonnais, G. Govers, V. Lecomte, K. van Oost, A. Couturier, et al. 2004. Scale effect on runoff from experimental plots to catchments in agricultural areas in Normandy. *J. Hydrol.* 299:4–14. doi:10.1016/j.jhydrol.2004.02.017
- Collinet, J., and P. Zante. 2005. Analyse du ravinement de bassins versants à retenues collinaires sur sols à fortes dynamiques structurales (Tunisie). *Geomorph.: Relief, Processus, Environ.* 1:61–74.
- De Franchis, L., and L. Ibanez. 2003. Threats to soils in Mediterranean countries. Document review. Plan Bleu Pap. 2. Plan Bleu UNEP, Sophia Antipolis, Valbonne, France.
- de Vente, J., and J. Poesen. 2005. Predicting soil erosion and sediment yield at the basin scale: Scale issues and semi-quantitative models. *Earth Sci. Rev.* 71:95–125. doi:10.1016/j.earscirev.2005.02.002
- de Vente, J., J. Poesen, M. Arabkhedri, and G. Verstraeten. 2007. The sediment delivery problem revisited. *Prog. Phys. Geogr.* 31:155–178. doi:10.1177/0309133307076485
- de Vente, J., J. Poesen, P. Bazzoffi, A. Van Ropae, and G. Verstraeten. 2006. Predicting catchment sediment yield in Mediterranean environments: The importance of sediment sources and connectivity in Italian drainage basins. *Earth Surf. Processes Landforms* 31:1017–1034. doi:10.1002/esp.1305
- de Vente, J., J. Poesen, G. Verstraeten, A. Van Ropae, and G. Govers. 2008. Spatially distributed modelling of soil erosion and sediment yield at regional scales in Spain. *Global Planet. Change* 60:393–415. doi:10.1016/j.glopla-

- El Khalili, A., D. Raclot, H. Habaieb, and J.M. Lamachère. 2013. Factors and processes of permanent gully evolution in a Mediterranean marly environment (Cape Bon, Tunisia). *Hydrol. Sci. J.* 58:1519–1531. doi:10.1080/02626667.2013.824086
- Estèves, M., and J.M. Lapetite. 2003. A multi-scale approach of runoff generation in a Sahelian gully catchment: A case study in Niger. *Catena* 50:255–271. doi:10.1016/S0341-8162(02)00136-4
- Favre, F., P. Boivin, and M. Wopereis. 1997. Water movement and soil swelling in a dry, cracked Vertisol. *Geoderma* 78:113–123. doi:10.1016/S0016-7061(97)00030-X
- Gómez, J.A., M.A. Nearing, J.V. Giraldez, and E.E. Albert. 2001. Analysis of sources of variability of runoff volume in a 40 plots experiment using a numerical model. *J. Hydrol.* 248:183–197. doi:10.1016/S0022-1694(01)00402-4
- Gómez, J.A., T. Vanwalleghe, A. De Hoces, and E.V. Taguas. 2014. Hydrological and erosive response of a small catchment under olive cultivation in a vertic soil during a five-year period: Implications for sustainability. *Agric. Ecosyst. Environ.* 188:229–244. doi:10.1016/j.agee.2014.02.032
- González-Hidalgo, J.C., M. Brunetti, and M. de Luis. 2011. A new tool for monthly precipitation analysis in Spain: MOPREDAS database (monthly precipitation trends December 1945–November 2005). *Int. J. Climatol.* 31:715–731. doi:10.1002/joc.2115
- González-Hidalgo, J.C., M. de Luis, and J.L. Peña. 2007. A review of daily soil erosion in western Mediterranean areas. *Catena* 71:193–199. doi:10.1016/j.catena.2007.03.005
- Hudson, P.F. 2003. Event sequence and sediment exhaustion in the Lower Pánuco basin, Mexico. *Catena* 52:57–76. doi:10.1016/S0341-8162(02)00145-5
- Inoubli, N., D. Raclot, R. Moussa, H. Habaieb, and Y. Le Bissonnais. 2016. Soil cracking effects on hydrological and erosive processes: A study case in Mediterranean cultivated vertisols. *Hydrol. Processes* 30:4154–4167. doi:10.1002/hyp.10928
- Jebbari, S., R. Berndtsson, A. Bahri, and M. Boufaroua. 2010. Spatial soil loss risk and reservoir siltation in semi-arid Tunisia. *Hydrol. Sci. J.* 55:121–137. doi:10.1080/02626660903529049
- Jorgensen, D.W., and T.W. Gardner. 1987. Infiltration capacity of disturbed soils: Temporal change and lithologic control. *J. Am. Water Resour. Assoc.* 23:1161–1172. doi:10.1111/j.1752-1688.1987.tb00868.x
- Kefi, M., and K. Yoshino. 2010. Evaluation of the economic effects of soil erosion risk on agricultural productivity using remote sensing: Case of watershed in Tunisia. In: K. Kahiwaru et al., editors, *Networking the world with remote sensing: ISPRS Technical Commission 8 Symposium*, Kyoto, Japan. 9–12 Aug. 2010. ISPRS Arch., Vol. 38, Part 8. Copernicus, Göttingen, Germany. p. 930–935.
- Lane, L.J., M. Hernandez, and M. Nichols. 1997. Processes controlling sediment yield from watersheds as functions of spatial scale. *Environ. Modell. Softw.* 12:355–369. doi:10.1016/S1364-8152(97)00027-3
- Lane, L.J., and M.R. Kidwell. 2003. Hydrology and soil erosion. In: M.P. McClaran et al., editors, *Santa Rita Experimental Range: 100 Years (1903 to 2003) of Accomplishments and Contributions: Conference Proceedings*, Tucson, AZ. 30 Oct–1 Nov. 2003. USDA For. Serv. Proc. RMRS-P-30. Rocky Mtn. Res. Stn., Fort Collins, Co. p. 92–100. https://www.fs.fed.us/rm/pubs/rmrs_p030_p030_092_100.pdf
- Le Bissonnais, Y., C. Montier, J. Daroussin, and D. King. 1998. Cartographie de l'aléa "érosion des sols" en France. Collection Etudes et Travaux 18. Inst. Français de l'environnement, Orléans, France. http://www.side.developpement-durable.gouv.fr/EXPLOITATION/Infodoc/ged/viewportalpublished.ashx?eid=IFD_FICJOINT_0012020
- Lenzi, M.A., L. Mao, and F. Comiti. 2003. Interannual variation of suspended sediment load and sediment yield in an alpine catchment. *Hydrol. Sci. J.* 48:899–915. doi:10.1623/hysj.48.6.899.51425
- Mathier, L., and A.G. Roy. 1996. A study of the effects of spatial scale on parameters of sediment transport equation for sheet wash. *Catena* 26:161–169. doi:10.1016/0341-8162(95)00038-0
- Mekki, I., J. Albergel, N. Ben Mechlia, and M. Voltz. 2006. Assessment of overland flow variation and blue water production in a farmed semi-arid water harvesting watershed. *Phys. Chem. Earth* 31:1048–1061. doi:10.1016/j.pce.2006.07.003
- Messing, I., and N. Jarvis. 1993. Temporal variation in the hydraulic conductivity of a filled clay soil as measured by tension infiltrometers. *J. Soil Sci.* 44:11–24. doi:10.1111/j.1365-2389.1993.tb00430.x
- Morgan, R.P.C. 2005. *Soil erosion and conservation*. 3rd ed. Blackwell Publ., Oxford, UK.
- Nearing, M.A., G. Govers, and L.D. Norton. 1999. Variability in soil erosion data from replicated plots. *Soil Sci. Soc. Am. J.* 63:1829–1835. doi:10.2136/sssaj1999.6361829x
- Parsons, A.J., R.E. Brazier, J. Wainwright, and D.M. Powell. 2006. Scale relationships in hillslope runoff and erosion. *Earth Surf. Processes Landforms* 31:1384–1393. doi:10.1002/esp.1345
- Poesen, J., and J. Hooke. 1997. Erosion, flooding and channel management in Mediterranean environments of southern Europe. *Prog. Phys. Geogr.* 21:157–199. doi:10.1177/030913339702100201
- Poesen, J., J. Nachtergaele, G. Verstraeten, and C. Valentin. 2003. Gully erosion and environmental change: Importance and research needs. *Catena* 50:91–133. doi:10.1016/S0341-8162(02)00143-1
- Poesen, J., D. Torri, and K. Bunte. 1994. Effects of rock fragments on soil erosion by water at different spatial scales: A review. *Catena* 23:141–166. doi:10.1016/0341-8162(94)90058-2
- Poesen, J., K. Vandaele, and B. Van Wesemael. 1996. Contribution of gully erosion to sediment production on cultivated lands and rangelands. *IAHS Publ.* 236:251–266.
- Poesen, J., L. Vandekerckhove, J. Nachtergaele, D.O. Wijdene, G. Verstraeten, and B. Van Wesemael. 2002. Gully erosion in dryland environments. In: L.J. Bull and M.J. Kirkby, editors, *Dryland rivers: Hydrology and geomorphology of semi-arid channels*. John Wiley & Sons, Chichester, UK. p. 229–262.
- Porto, P., D.E. Walling, C. Alewell, G. Callegari, L. Mabit, N. Mallimo, et al. 2014. Use of a ^{137}Cs re-sampling technique to investigate temporal changes in soil erosion and sediment mobilisation for a small forested catchment in southern Italy. *J. Environ. Radioact.* 138:137–148. doi:10.1016/j.jenvrad.2014.08.007
- Raclot, D., and J. Albergel. 2006. Runoff and water erosion modelling using WEPP on a Mediterranean cultivated catchment. *Phys. Chem. Earth* 31:1038–1047. doi:10.1016/j.pce.2006.07.002
- Raclot, D., Y. Le Bissonnais, M. Annabi, M. Sabir, and A. Smetanova. 2017. Main issues for preserving Mediterranean soil resources from water erosion under global change. *Land Degrad. Dev.* doi:10.1002/ldr.2774
- Raclot, D., Y. Le Bissonnais, X. Louchart, P. Andrieux, R. Moussa, and M. Voltz. 2009. Soil tillage and scale effects on erosion from fields to catchment in a Mediterranean vineyard area. *Agric. Ecosyst. Environ.* 134:201–210. doi:10.1016/j.agee.2009.06.019
- Rodrigo Comino, J., J.D. Ruiz Sinoga, J.M. Senciales González, A. Guerra-Merchán, M. Seeger, and J.B. Ries. 2016. High variability of soil erosion and hydrological processes in Mediterranean hillslope vineyards (Montes de Málaga, Spain). *Catena* 145:274–284. doi:10.1016/j.catena.2016.06.012
- Roose, H., R. Chebbani, and L. Bourougaa. 2000. Ravinement en Algérie: Typologie, facteurs de contrôle, quantification et réhabilitation. *Sécheresse* 11:317–326.
- Sadeghi, S.H.R., M.B. Seghaleh, and A.S. Rangavar. 2013. Plot sizes dependency of runoff and sediment yield estimates from a small watershed. *Catena* 102:55–61. doi:10.1016/j.catena.2011.01.003
- Schumm, S.A., and G.C. Lusby. 1963. Seasonal variations of infiltration capacity and runoff on hillslopes of western Colorado. *J. Geophys. Res.* 68:3655–3666. doi:10.1029/JZ068i012p03655
- Tomkins, K.M., G.S. Humphreys, M.T. Wilkinson, D. Fink, P.P. Hesse, S.H. Doerr, et al. 2007. Contemporary versus long-term denudation along a passive plate margin: The role of extreme events. *Earth Surf. Processes Landforms* 32:1013–1031. doi:10.1002/esp.1460
- Valentin, C., J. Poesen, and Y. Li. 2005. Gully erosion: Impacts, factors and controls. *Catena* 63:132–153. doi:10.1016/j.catena.2005.06.001
- Vanmaercke, M., W. Maetens, J. Poesen, B. Jankauskas, G. Jankauskiene, G. Verstraeten, and J. de Vente. 2012. A comparison of measured catchment sediment yields with measured and predicted hillslope erosion rates in Europe. *J. Soils Sediments* 12:586–602. doi:10.1007/s11368-012-0479-z
- Vanmaercke, M., J. Poesen, G. Verstraeten, and J. de Vente. 2011. Sediment yield in Europe: Spatial patterns and scale dependency. *Geomorphology* 130:142–161. doi:10.1016/j.geomorph.2011.03.010
- Wainwright, J., and A.J. Parsons. 2002. The effect of temporal variations in rainfall on scale dependency in runoff coefficients. *Water Resour. Res.* 38:1271. doi:10.1029/2000WR000188
- Walling, D.E. 1983. The sediment delivery problem. *J. Hydrol.* 65:209–237. doi:10.1016/0022-1694(83)90217-2
- Wei, W., L. Chen, L. Yang, B. Fu, and R. Sun. 2012. Spatial scale effects of water erosion dynamics: Complexities, variabilities, and uncertainties. *Chin. Geogr. Sci.* 22:127–143. doi:10.1007/s11769-012-0524-2
- Wischmeier, W.H., and D.D. Smith. 1978. Predicting rainfall erosion losses: A guide to conservation planning. *Agric. Handb.* 537. USDA Science and Educ. Admin. Hyattsville, MD.
- Yaalon, D.H. 1997. Soils in the Mediterranean region: What makes them different? *Catena* 28:157–169. doi:10.1016/S0341-8162(96)00035-5
- Yair, A., and N. Raz-Yassif. 2004. Hydrological processes in a small arid catchment: Scale effects of rainfall and slope length. *Geomorphology* 61:155–169. doi:10.1016/j.geomorph.2003.12.003
- Zehe, E., and M. Sivapalan. 2009. Threshold behaviour in hydrological systems as (human) geoecosystems: Manifestations, controls, implications. *Hydrol. Earth Syst. Sci.* 13:1273–1297. doi:10.5194/hess-13-1273-2009