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Maximum Residual Levels of Pesticides and Public Health

Best Friends or Faux Amis?

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Abstract: *The purpose of this article is to analyse the relation between public health and the regulations of Maximum Residue Limit (MRL) of pesticides. Many authors underline the role of trade protectionism in fixing these limits whereas these regulations should be intended for public health protection. We first establish the link between the MRL for a given chemical in plant products and its level of toxicity. In order to perform this analysis we cross the FAS USDA MRL database and the classification of the long term toxicological effects (LTE) for active substances provided by SAgE pesticide. We then compute a synthetic and polyvalent tool namely "Health Score" which provides a first overview of the link between LTE and MRL by country. Then this score is regressed in a logit model in order to identify the relationship between the countries' Health Score and the socio-economic and political characteristics of such areas. Results highlight the importance of public health expenditures in determining the settings of MRL towards stricter levels.*

Keywords: food safety, pesticide residues, fruits and vegetables, scoring, logit model

JEL codes: C51, I18, Q18

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Data Appendix Available Online

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1. Introduction

According to Henson and Caswell (1999), food safety regulations are *the outcome of a complex trade-off between (...) consumers, food manufacturers, food retailers, farmers, (...) the government itself and taxpayers*. Concerning regulations on Maximum Residual Level (MRL) of pesticides, it is widely recognized that they have been designed in order to protect the health of consumers and ensure that farmers adopt good practices. But even if an international regulation has been designed by the Codex Alimentarius (hereafter Codex), its application is not compulsory and nations keep their sovereignty in fixing these limits. Some countries appear as very severe while others, and particularly those deferring to the Codex, are considered as much laxer. This regulatory heterogeneity is at the core of a growing literature (Winchester et al., 2012; Drogué and Demaria 2012; Li and Beghin, 2014; etc.). These authors try to measure the differences in regulations by comparing the limits set by countries' authorities. This literature generally tries to measure the trade impact of setting stricter MRL for a product or a group of products. It draws a picture of the trade-restrictiveness of MRL regulations around the world. National regulations are compared to the international regulation used as the reference point above which national standards may be considered as protectionist.

Recently, a new literature emerges to analyse the “formation” of the MRL. Referring to the model of Grossman and Helpman (1994), Foletti (2012) has opened the debate on *disentangling protectionism from health protection as drivers for MRL regulations* inserting the Acceptable Daily Intake (ADI) in its empirical analysis. Her analysis also includes the impact on human health of the various chemicals referenced in her database. She draws some results on the motivations of countries in setting their MRL which can be either set for health

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protection or for trade protectionism. More recently Li, Xiong and Beghin (2014) propose a model which *contributes to the understanding of the policy formation of food safety standard*. These authors use a political support function approach to represent the decision of policy makers in setting the MRL and measure econometrically the level of trade protectionism induced by the MRL of pesticides and veterinary drugs. They introduce in their model economic, social and political variables and measure the weight of each other. They find that MRL regulations and tariffs are substitute trade protectionism tools and that, *countries with higher regulatory quality set tougher standards*. For these authors MRL are the outcome of the complex trade-off described by Henson and Caswell (1999). But claiming that MRL of pesticides are the results of a political compromise denies that they are set independently, based on sound science, and therefore suggest that they do not respond to legitimate social objectives. The WTO and particularly the SPS committee declares that *SPS measures should be based on (i) international standards as those of the Codex, (ii) science, including assessment of risk, (iii) a temporary principle of precaution in the absence of international standards or scientific evidence. Countries are free to set their own standards based on science*. As a consequence, each country can establish MRL independently on national conditions and on the basis of agricultural practice. This means that for a given product, countries may regulate various substances. A pesticide authorized in a country may not be regulated or even prohibited in another one.

But how are the MRL actually established? According to the FAO, the MRL is the maximal concentration of a residue legally permitted or considered acceptable in or on a food product, a farm produce or a product intended for animal feeding. When a chemical is approved, studies on residues are made in order to determine the residue level which could remain in the product after harvest in the worst case (worst case scenario). For this purpose, the substance

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studied is applied at the maximum of its recommendation for a given crop. During the field experiment a theoretical maximum daily intake (TMDI) is computed taking into account all the vegetables and animal products a human being can consume given his diet pattern ($TMDI = \sum_i MRL_i(mg.kg^{-1}) * Consumption_i(kg)$), being i , the product. This theoretical maximum daily intake is then compared to an admissible daily intake (ADI) which is the maximum quantity of a substance someone can consume during its lifetime without adverse effects. The MRL should be fixed such that:

$$TMDI(mg) \leq ADI(mg.kg^{-1}) * Body\ mass(kg)$$

Each country establishes a list of MRL and sometimes provides a default value for those pesticides that are not explicitly listed (see Table 1).

[Insert Table 1]

It seems thus, that MRL are efficient to protect human's health. Despite this, consumers' protection non-governmental organisations still claim for a reduction in the use of pesticides or even the ban of other ones and declare that the setting of MRL can be called into question. For instance, regularly the Pesticide Action Network (PAN) tries to alert the public opinion that current MRL would not be protective enough regarding consumers' health (PANE, 2014).

In this article we try to investigate the strength of the relationship between the setting of MRL and their primary objective that should be the consumer's health protection. We reconcile the link between MRL regulations and public health and investigate, more thoroughly, the social drivers of regulations on limits of pesticides residues. We restrict our analysis to plant products (i.e. fruits, vegetables, seeds, nuts, spices, tea, etc.). In order to perform this analysis we compute a synthetic and polyvalent tool hereafter Health Score (HS). This Health Score takes into account the level of long term toxicity of the regulated pesticides on consumer

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health and ranks their MRL in order to establish if they are consistent with health protection (and have legitimate purpose) or if it is much laxer than the required level of protection it should provide.

We use the database on Maximum Residual Level of pesticide provided by the Federal Agricultural Service of the United State Department of Agriculture (FAS USDA MRL database <http://www.fas.usda.gov/maximum-residue-limits-mrl-database>). We cross the MRL with the classification of long term toxicological risk effects (hereafter Long Term Effects or LTE) provided by the Quebec Ministries of Agriculture, and Environment and the National Institute of Public Health of Quebec on the SAgE pesticides website (<http://www.sagepesticides.qc.ca/Default.aspx>). The information provided in this website was used to compute the Quebec Pesticide Risk Indicator (Samuel *et al*, 2012). The combination of these two sets of information provides an original database that establishes the link between long term toxicology and MRL by product and country. Intuitively these two variables should be negatively correlated, as a low LTE should imply a high MRL and *vice-versa*. At first, we cross data on MRL and LTE and compute the HS in order to identify how cautious a country is when it fixes the limits of pesticides residue that plant products should contain in order to be suitable for consumption. We cluster the HS to draw a world picture of consumers' health protection. Then, we model the relation between the HS and certain variables that take into account not only socio-economic patterns but also political and health characteristics to understand the formation of this specific standard.

We analyse the link between pesticides MRL regulations and public health as this link has been somehow neglected in previous literature which mainly focus on the trade protectionism aspect of the MRL. The literature on political economy on food standards, particularly the works of Swinnen and Vandemoortele (2008, 2009 and 2011) provide our analysis for

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theoretical foundations. Following these authors a standard should be the result of a social optimum which is determined by comparing the gains and losses of both producers and consumers. The determinants can be summarized from the supply side by the difference between the prices and the costs of compliance and from the demand side by the difference between the price and the “consumption effect” which is the benefit for a consumer to consume pesticide-free goods. But given the existence of conflicting interests they model the standards as the results of a political optimum by introducing in the analysis the weight of the lobbies and the taxpayers. Our analysis addresses two issues: (i) are the MRL of pesticides driven by legitimate precautionary motives? (ii) What is the role of these objectives (proxied by public health expenditures) in the setting of the MRL? We deal with the first issue by trying to give a measure of the “precautionary MRL” via the computation of a Health Score; the second issue is addressed thanks to an econometric modelization of the relation between this score and selected variables to understand the role of certain determinants (and particularly public health) in its formation. The results show that public health expenditures play a stronger role in the setting of stricter MRL for hazardous substances than the level of income.

This article is structured as follows. Section 2 presents the motivation of the analysis performed in this article. Section 3 describes the data and the clustering approach we implemented. The model is specified in Section 4 and results are developed in Section 5. Section 6 is devoted to the discussion of the results and conclusion.

2. Building a Health Score

At first we draw our analysis using information on MRL and on LTE which include three different dimensions: countries, products and substances. For more clarity, we will use

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hereafter the term “product” to designate the plant products and the word “substance” to designate the chemicals. Information on the level of MRL is from the FAS USDA international MRL database which provides the level of pesticides and veterinary drug for 661 plant and animal products commodities, 90 markets (the European Union constitutes one market) and 352 pesticides. The database does not include processed food products. Our analysis focuses only on plant products (fruits, vegetables, cereals, etc.); meat, fish and dairies are not considered.

The toxicity of the pesticides is captured through the Long Term Effects provided by the SAgE pesticides website maintained by the government of Quebec (Canada). This website describes for a list of 441 substances their impacts on health and on environment. The impacts on human health are disaggregated in acute toxicity and long term toxicity. Long term toxicity of a substance is graded according to its risk level (extremely high, high, moderate and low, see Table 2 and Table A1 in Appendix).

[Insert Table 2]

The intuition is that the MRL of a given substance should vary inversely with the level of its toxicity. Here, we are mainly interested in the chronic toxicity (or long term effect) which affects the consumer (acute toxicity affects above all the producers). We can summarize this proposition in the following way: a low LTE should be associated with a high MRL for a harmless substance while a high LTE should be associated with a low MRL for a hazardous substance. Thus we can grade the substances listed by the SAgE pesticides by level of toxicity. We attribute a grade of 1 if the substance has a low level of risk, to 4 if the risk for human health is extremely high. Then we match the list of substances contained in the FAS database with the corresponding one in the SAgE pesticides database. This merging led us to

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keep a list of 252 substances, 116 countries (for the need of the estimation we disaggregated the EU in 27 countries) and 585 plant products.

Combining LTE and MRL allows the performance of a score analysis to classify countries regarding their sensitivity to health concerns. In some way we measure the level of precaution of countries with respect to human food exposure to pesticides and we can rank countries depending on their precautionary status.

Before computing the HS we study the distribution of MRL (see Table A2 in Appendix). The MRL range between 0.0005 and 1500 mg/kg with a kurtosis equal to 535.5. The peak of the distribution is around zero. To get an accurate overview of the sample we first focus on the values of the MRL ranging between 0 and 500, and then on the values between 0 and 50; and subsequently between 0 and 10 to finally zoom in on the MRL between 0 and 1 which are the most frequent values. As far as the MRL values above 500 are marginal (they represent 0.004% of the observations), the kurtosis is still high; in the second we lose 1.1% of the observations and the kurtosis is still high; in the third one, we lose 5.14% of observations and finally in the fourth 21%.

Looking at the distribution of MRL for LTE equal to 3 or 4 (see Figures A1 to A3 in Appendix), we define 7 classes of scoring for each triplet country (c)/product (p)/substance (s) based on the following rule of decision:

$HS_{cps} = 7$ if $LTE = 1$ or 2 and $0 \leq MRL \leq 1500$

$HS_{cps} = 6$ if $LTE = 3$ or 4 and $0 \leq MRL < 0.02$

$HS_{cps} = 5$ if $LTE = 3$ or 4 and $0.02 \leq MRL < 0.05$

$HS_{cps} = 4$ if $LTE = 3$ or 4 and $0.05 \leq MRL < 0.1$

$HS_{cps} = 3$ if $LTE = 3$ or 4 and $0.1 \leq MRL < 1$

$HS_{cps} = 2$ if $LTE = 3$ or 4 and $1 \leq MRL < 10$

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$HS_{cps} = 1$ if $LTE = 3$ or 4 and $10 \leq MRL \leq 1500$

Table 3 summarizes the distribution of the $HS_{cps} = 7$

[Insert Table 3]

Then, we calculate the score by country as follows:

$$HS_c^7 = \frac{\sum_p \sum_s HS_{cps}}{7 * P.S}$$

where c denotes the countries, p denotes the product, s the substance; $P.S$ is the number of products and substances pairs. The denominator is multiplied by 7 to take into account the number of classes for each product and substance pair. Due to its formalization, the HS_c is lower and upper bounded between 0 and 1. This score should be interpreted as a precaution level index. The higher the score, the more a country is concerned with human dietary exposure to pesticides in food. A score equal to 1 means that the country has ruled out all pesticides which LTE is high or extremely high.

We observe that whatever a given MRL the HS_{cps} is decreasing with LTE for a given pair of product/substance. Reciprocally, the HS_{cps} is decreasing or constant with MRL whatever the level of LTE. This index allows us to rank countries combining the level of the toxicity of a substance and the enforced regulation. Our previous considerations about the monotonicity along the LTE (respectively MRL) are transmitted to the HS_c at the country level. As a matter of fact, the HS_c can be considered as a mean of the HS_{csp} for a specific country. Thus, if almost all the substances regulated by a country displayed a high LTE, it would correspond to an over representation of low HS_{cps} , which sum would tend to remain low (and vice-versa). The same argumentation is also true for the MRL.

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Finally, the HS_c is invariant to the number of the regulations by averaging the sum of HS_{cps} by the total number of products and substances pairs. In other words, even changing the number of products or substances referenced by a given country, the computation of the HS_c does not change.

This score ranges between 0.43 and 0.74 (see Map 1, Figure A4 and Table A3 in Appendix). For instance, a country with a score equal to 0.74 can be considered as more cautious than another with a score equal to 0.43. Globally the score is not very high (between 0.4 and 0.7). But when we cluster it some rough observations can be made.

[Insert Map 1]

The lowest scores appear for countries referring to the rules of the Gulf Cooperation Council. Then, come two groups of countries where we find China, New Zealand, Mexico, India, Australia, Brazil and the USA, among the main actors on the world agricultural markets. The fourth group gathers the countries which defer to the Codex. Finally, the European Union and Russia close the series. It seems that for the chemicals explicitly listed (we do not take into account missing MRL), those with the highest toxicological effects are not those with the more stringent limits.

3. The model

In order to establish the determinants of the HS_c we complete our analysis with an econometric estimation accounting for legitimate public health objectives. Following Swinnen and Vandemoortele (2008, 2009 and 2011), we consider the HS_c as the level of standard s . They consider two levels of standards: $s^\#$ corresponds to the social optimum resulting from the maximization of a welfare function and s^* is the political optimum, a result of the maximization of the same welfare function taking the lobbies influence into account. s^*

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is driven by a trade-off between social, economic and political interests while $s^\#$ is driven by legitimate social objectives.

We are interested in econometrically investigating the relationship between certain determinants (and particularly public health) and the variation of the Health Score. Today, the Codex is considered as the reference by many countries. But we observe in our data that countries which adopt more restrictive standards than the Codex have a HS_c greater than the countries which follow the Codex. Moreover as underlined by Josling and al. (2004), page 43: *the SPS agreement has politicized decision making within the Codex more than in other standards organizations.*² Thus we suppose that standards higher than the Codex's HS (hereafter HS_{Codex}) tend to be more influenced by public health considerations than the others. Countries with a $HS_c > HS_{Codex}$ may be considered as countries which standards are more driven by precautionary legitimate motives than countries with a $HS_c \leq HS_{Codex}$. We thus take the HS_{Codex} as the threshold (denoted $\bar{s}^*$). We define a binary variable $\tilde{s}^*$ to discriminate against these two classes in order to understand what drives the HS_c beyond the reference level defined by the Codex (HS_{codex}).

For this purpose we use a logit model by re-defining the dependent variable as follows:

$$\begin{cases} \tilde{s}^* = 0 & \text{if } s^* \leq \bar{s}^* \\ \tilde{s}^* = 1 & \text{if } s^* > \bar{s}^* \end{cases}$$

Let y the observed dependent variable and y^* the latent variable satisfying the single index model

$$y_j^* = x_j' \beta + \varepsilon_j \quad (1)$$

Even if y^* is not observed, we do observe

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² We thank the anonymous referee for this valuable suggestion.

$$y_j = \begin{cases} 1 & \text{if } y_j^* > 0 \\ 0 & \text{if } y_j^* \leq 0 \end{cases} \quad (2)$$

In their model Swinnen and Vandermoortele (2009) considered that the level of a food safety standard is driven on the supply side by the producers' cost of compliance, and on the demand side by the consumers' preferences, their perception bias (also called consumption effect) and the externality effect (also called warm glow effect). For the authors public food safety standards do not induce any warm glow effect (Swinnen and Vandermoortele, 2009; page 515) thus we do not include it in our empirical analysis. Increasing the stringency of a food safety standard is a cost for both domestic and foreign producers, consequently the impact will depend on the relative competitiveness of both industries. Concerning the consumers, increasing the stringency of the standard, the cost for the consumer increases but it increases its utility because it guarantees a safer product particularly if consumers trust in the capacity of the government to enforce the standard. We test some of these indicators on the standard setting. Our empirical model is written as follows:

$$\begin{aligned} \widetilde{S}_c^* = & \beta_0 + \beta_1 Lngdp\_cap_c + \beta_2 Corruption_c + \beta_3 Autoc\_bin_c + \beta_4 Lnpubhexp_c \\ & + \beta_5 RCA\_veg_c + \beta_6 MFN\_veg_c + \varepsilon_c \end{aligned}$$

On the demand side $Lngdp\_cap_c$ is the logarithm of the gross domestic product (GDP) per capita of country c . It takes the level of consumption into account. $Corruption_c$ is the World Bank's Control of Corruption indicator. It captures perceptions of the extent to which public power is exercised for private gain, including both petty and grand forms of corruption, as well as "capture" of the state by elites and private interests. Following Onyango et al. (2007) the trust in institutions, particularly those in charge of food safety, has a positive influence on

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the consumer's perception of food safety, thus this variable accounts for the perceived risk probability. Both indicators come from the World Bank Development Indicators (WBDI) database.

On the supply side, The RCA_Veg_c is the revealed comparative advantage of Balassa for vegetable products which accounts for the competitiveness of the domestic industry. MFN_veg_c is the weighted average most-favoured-nation tariff rate for agricultural products. Both indicators have been retrieved from the World Bank World Integrated Trade Solution (WITS) database.

We also add two variables which may explain the formation of stringent/loose standards and have an impact on the cost for both consumers and producers. $Lnpubhexp_c$ is the logarithm of public health expenditures in percentage of GDP. Public expenditure consists in the *recurrent and capital spending from (central and local) government budgets, external borrowings and grants and social (or compulsory) health insurance funds*. We suppose that countries with high percentage of public health expenditures will be more inclined to set stringent standards.

This indicator comes from the World Health Organization National Health Account database. On the contrary, following Besley and Kudamastu (2006) whose results suggest that there is a positive link between health and democracy, we suppose that countries with a lower level of democracy are less able to set and enforce stricter standards. In order to test this relation, we add the Autocracy score from the Polity IV Projects (www.systemicpeace.org/polity/polity4.htm) which measures the degree of a government authority. Where "autocracy" defines *the political systems whose common properties are a lack of regularized political competition and concern for political freedoms*. It ranges from 0 (weak autocracy) to 10 (strong autocracy). We transform this variable in a binary variable. $Autoc\_bin_c$ is equal to 0 for a value of the original variable lower than 2; and 1 otherwise.

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MLR variables are measured for 2010; all other variables are for 2011 (except when they were not available in this case the closest values were substituted).

4. Results

Our model includes economic, social and political factors that may influence the $\widetilde{S}_c^*$. This section presents the results from the specification of a logit regression³ (see Table 4).

[Insert Table 4]

The model is consistent in the sense that the impacts of most regressors are statistically significantly different from zero but *Lngdp_cap* and *Corruption*. The confusion matrix gives the performance of the model in the classification of the observations. Looking at the confusion matrix (see Table 5) we can tell that this model fits quite well regarding the sample size in terms of its overall and correct classification rates. It also correctly predicts more than 86% of the countries positioning. The confusion matrix highlights also that the model performs better (88%) the negative predicted values (that is to say, countries with the lower score) than the others (81%). These slightly unbalanced predictive performances can be imputed to the fact that there are more countries which score is lower or equal than the Codex (corresponding to 2/3 of the sample).

[Insert Table 5]

In the estimation, the coefficient of *Lngdp_cap*, the economic dimension of the countries is positive but not significant. However, the coefficient of *Lnpubhexp*, the share of public health expenditures is positive and significant. This model highlights also the non-neutrality of the political regime as the lack of democratic institutions (*Autoc_bin*) has a negative and significant coefficient. The perception of the quality of institutions by consumers

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³ The reduced number of observations does not allow performing an ordered logit regression.

(*Corruption*) is positive but not significant. Concerning the remaining variables, the coefficient of *RCA_veg* is negative and significant which means that countries less competitive tend to set stricter standards. And the coefficient of *MFN_veg* is also negative and significant meaning that countries with higher tariff rates have less restrictive standards.

In order to quantify the impact of each variables involved in the model, we compute the odds ratio and marginal effects (see Table 4). The odds ratios give information on the multiplicative effect of the variation, for a unit change, on the probability for the dependent variable to switch from zero to one. That is to say, odds ratios specify the determinants (independent variables) which make a country move from the group of countries less concerned with pesticide exposure to the group of countries more concerned with pesticides exposure.

The marginal effects measure discrete change for categorical variables while they measure the instantaneous rate of change for continuous variables; when all other variables are at mean. But interpreting the marginal effects is only straightforward for categorical variable because they change for each level of the independent variable, thus we will mainly focus on the interpretation of the odds ratios.

Focusing on odds ratios, a 1 percentage point increase in the share of public health expenditures (*Lnpubhexp*) increases by 69% the odds of setting a stringent standard, all other variables being constant. On the contrary, if *Autoc_bin* switch from 1 to 0 the odds of $\tilde{s}^* = 1$ diminish by 94%. Finally, countries more competitive tend to resort less on the stringency of MRL and a unit increase in *RCA_veg* decreases by 62% the odds of a precautionary standard, in a similar way countries with higher MFN import duty rates are less likely to set stricter MRL (-8%).

5. Conclusion and discussion

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Our analysis highlights some factors involved in the formation of MRL of pesticides. We first compute a Health Score reconciling data on MRL and data on chronic toxicity for several chemicals. Then we transform this score in a binary variable in order to test with a logistic model the relation between this score and some economic, social and political variables. We are interested in analysing the factors that drive the setting of MRL for substances with the highest long term toxicological impacts on human health and thus the level of health protection a nation wants to achieve. We want to know first if standards on MRL are driven by precautionary legitimate objectives or if it is the outcome of a political trade-off and what influences their divergence. In contrast with other research (particularly the work of Li and al; 2014) we find no statistical evidence on a positive link between the Health Score and the level of income or the consumer's risk perception but with the importance of public health expenditures. The results show here that it is not the absolute wealth of a nation that influences the level of health protection but rather the share of its wealth devoted to health. If the perception by consumers of the quality of the institutions does not matter (*Corruption* is not significant), the political regime matters, in the sense that a restriction in the openness and competitiveness in executive recruitment and participation diminishes the chance to set stricter standards. Finally, in line with other works, economic factors also matter. The competitiveness of the domestic industry limits the stringency in standards (the odds of setting stricter standards decreases by 62% when *RCA_veg* increases) while freeing trade has instead a positive impact on it (+8%).

We underline here, the important role of public health in the setting of MRL. These results are interesting as they show that at a same level of wealth, population, democracy, a 1 unit increase in the share of public health expenses in GDP multiplies the probability of setting stricter MRL regulations by 69%. These results are not fully consistent with the recent

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outcomes of the literature dealing with this issue. Li, Xiong and Beghin (2014) show that in higher per capita income countries, consumers' economic and socio-demographic characteristics are the main drivers of MRL. We shed light on the fact that public health expenditures also matter in limiting the residue of highly hazardous substances. Those authors measure the absolute stringency of the MRL while adding the level of toxicology (via the LTE), we provide a measure of their relative stringency. We are thus able to determine the way countries arbitrate between the level of the regulation and the health impact of pesticides.

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8. Tables and figures

Table 1: List of countries in the analysis and MRL regulation rules

Rule	Countries
Countries set their own standards	Argentina, Australia, Brazil, Canada, Chile, China, India, Indonesia, Israel, Japan, Korea, Russia, Singapore, South Africa, Taiwan, Thailand, Turkey, New Zealand, Vietnam, United Arab Emirates, Malaysia, Switzerland
Countries defer to Codex	Algeria, Angola, Barbados, Bermuda, Bahamas, Bangladesh, Colombia, Costa Rica, Cuba, Dominican Republic, Ecuador, Egypt, El Salvador, Guatemala, Hong Kong, Honduras, Jordan, Kenya, Lebanon, Morocco, Netherlands Antilles, Nicaragua, Pakistan, Panama, Peru, Philippines, Trinidad and Tobago, Tunisia, Venezuela
Countries defer to EU standards	Belgium, Denmark, Finland, French Pacific Islands, France, French West Indies, Germany, Greece, Ireland, Italy, Jamaica, Netherlands, Norway, Poland, Portugal, Spanish, Sweden, United Kingdom
Countries defer to Gulf Cooperation Council standards	Bahrain, Kuwait, Saudi Arabia, Oman, Qatar
Countries defer to exporting countries standards	Albania, Antigua and Barbuda, Cayman Island, Haiti, St. Kitts and Nevis, Sri Lanka, St. Lucia
Countries defer to US standards	United States, Mexico

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Table 2: Long term effect of active ingredients

Long term effect	Severity of the effects					
	Extremely high risk		High risk	Moderate risk		Low risk
	Indicators value					
Carcinogenicity	Human carcinogen	Probable human carcinogen	Possible human carcinogen	Data inadequate for assessment of human carcinogen potential		Not likely to be carcinogen to humans
Genotoxicity		Genotoxic for human	Potential genotoxic for humans		No or inadequate data	No evidence of human genotoxicity
Endocrine disruption		Evidence of endocrine disruption	Potential endocrine disruption		No or inadequate data	No evidence of endocrine disruption
Reproductive effects	Confirmed human effects	Suspected human effects	Confirmed animal effects	Suspected animal effects	No or inadequate data	No effects
Development	Confirmed human effects	Suspected human effects	Confirmed animal effects	Suspected animal effects	No or inadequate data	No effects

Source: Samuel et al. (2012) & Sage Pesticide

<http://www.sagepesticides.qc.ca/Infos/SignificationSymbole.aspx#link-TraitementSanteTableau1>

Table 3: Distribution of $HS_{cps}=7$

HS_{cps}	Freq.	Percent	Cum.
1	210,856	25.85	25.85
2	19,533	2.39	28.24
3	29,695	3.64	31.88
4	7,543	0.92	32.8
5	25,578	3.14	35.94
6	15,805	1.94	37.88
7	506,804	62.12	100
Total	815,814	100	

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Map 1: World distribution of the HS_c^7 **Table 4:** Estimations results

	Coefficients	Odds-ratios	Marginal effects
$\widetilde{S}_c^*$			
$Lngdp\_cap_c$	0.006	1.006	0.0006
	(0.602)		(0.062)
$Lnpubhex_c$	1.688*	5.413	0.174*
	(0.950)		(0.909)

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<i>RCA_veg_c</i>	-0.976***	0.376	-0.100***
	(0.409)		(0.037)
<i>Corruption_c</i>	0.797	2.219	0.082
	(0.767)		(0.077)
<i>Autoc_bin_c</i>	-2.821***	0.059	-0.291***
	(1.027)		(0.086)
<i>MFN_veg_c</i>	-0.080*	0.919	-0.008*
	(0.049)		(0.004)

Standard errors in parentheses; significant at level: * $p < 0.10$, ** $p < 0.05$, *** $p < 0.01$

Table 5: Confusion matrix

True			
Classified	D	~D	Total
+	22	5	27
-	7	53	60
Total	29	58	87
Classified + if predicted $\Pr(D) \geq .5$ True D defined as $HSc \neq 0$			
Sensitivity	$\Pr(+ D)$ 75.86%		
Specificity	$\Pr(- \sim D)$ 91.38%		
Positive predicted value	$\Pr(D +)$ 81.48%		
Negative predictive value	$\Pr(\sim D -)$ 88.33%		
False + rate for true ~D	$\Pr(+ \sim D)$ 8.62%		
False – rate for true D	$\Pr(- D)$ 24.14%		
False + rate for classified +	$\Pr(\sim D +)$ 18.52%		
False – rate for classified -	$\Pr(D -)$ 11.67%		
Correctly Classified	86.21%		

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