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Characterization of Novel Ligands of ERα, ERβ, and PPARγ: The Case of Halogenated Bisphenol A and Their Conjugated Metabolites

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The capability of the flame retardants tetrabromobisphenol A (TBBPA) and tetrachlorobisphenol A (TCBPA) to activate peroxisome proliferator–activated receptors (PPARs) α, β, and γ and estrogen receptors (ERs) α and β has been recently investigated, but the activity of their biotransformation products and of their lower molecular weight analogues formed in the environment remains unexplored. The aim of this study was to investigate the relationship between the degree of halogenation of BPA analogues and their affinity and activity towards human PPARγ and ERs and to characterize active metabolites of major marketed halogenated bisphenols. The biological activity of all compounds was studied using reporter cell lines expressing these nuclear receptors (NRs). We used NR-based affinity columns to rapidly evaluate the binding affinity of halogenated bisphenol analogs for PPARγ and ERs and to trap active metabolites of TBBPA and TCBPA formed in HepG2 cells. The agonistic potential of BPA analogs highly depends on their halogenation degree: the bulkier halogenated BPA analogs, the greater their capability to activate PPARγ. In addition, PPARγ-targeted affinity column, HGELN-PPARγ reporter cell line and crystallographic analysis clearly demonstrate that the sulfation pathway, usually considered as a detoxification process, leads for TBBPA and TCBPA, to the formation of sulfates conjugates which possess a residual PPARγ-binding activity. Our results highlight the effectiveness NR-based affinity columns to trap and characterize biologically active compounds from complex matrices. Polyhalogenated bisphenols, but also some of their metabolites, are potential disrupters of PPARγ activity.

Key Words: halogenated BPA; affinity column; nuclear receptor; metabolism; obesity; endocrine disrupter.

Over the past decades, growing concerns have been raised about human and wildlife exposure to endocrine disrupting chemicals (EDC). Bisphenol A (BPA), a major EDC used for the production of polycarbonates and epoxy resins, is among the highest volume chemicals produced worldwide (Jimenez-Diaz et al., 2010). The occurrence of BPA in the environment, wildlife, and in humans has been demonstrated. In comparison, available data regarding BPA analogs is much scarcer. On the one hand, this group comprises bisphenols which differ at the level of the central part of the molecule bridging between the two phenols. All these bisphenols (BPF, BPC, BPS, etc.) are produced in much lower amounts than BPA and are weak estrogens (Perez et al., 1998). On the other hand, analogs of BPA where the phenolic moieties are substituted with halogens (Br or Cl) deserve a broader attention because such substitution may significantly modify both their fate (Schauer et al., 2006; Zalko et al., 2006) and their capability to interact with nuclear receptors (NRs).

Tetrabromobisphenol A (TBBPA) is mainly used as a flame retardant to protect computer motherboards and other electronic equipment. Its production is currently estimated around 200,000 tons/year, and its presence has been reported in the environment (de Wit et al., 2009) and in wildlife (Darnerud, 2003). In the environment, TBBPA can be debrominated into tri-, di-, and monoBPA (Jasur-Kruh et al., 2010). Its presence in human samples has only been monitored in a limited number of studies but was clearly demonstrated in maternal serum and mother milk (Cariou et al., 2008; Johnson-Restrepo et al., 2008).

Tetrachlorobisphenol A (TCBPA) has also been reported to be used as a flame retardant, but in much lower quantities than TBBPA (< 10,000 tons/year) (Chu et al., 2005). The presence of TCBPA, as well as that of lower chlorinated analogs (monoCBPA, diCBPA, and triCBPA) in environmental samples has been unequivocally demonstrated (Fukazawa et al., 2001; Gallart-Ayala et al., 2010; Liu et al., 2009). Chlorinated BPA analogs can be considered as an emerging group of contaminants, which is also supported by recent reports demonstrating their presence in human (Fernandez et al., 2007; Jimenez-Diaz et al., 2010). Contrary to TBBPA,
and given the low levels of production of TCBPA, the origin of most chlorinated BPA in the environment is more likely the chlorination of BPA than the dechlorination of TCBPA. Indeed, BPA is readily chlorinated by reacting with sodium hypochlorite, which is commonly used as a bleaching agent in paper factories and also as a disinfectant in sewage treatments plants (Gallart-Ayala et al., 2010; Yamamoto and Yasuhara, 2002). Like many other phenolic compounds (Deborde et al., 2004), BPA easily chlorinates in aqueous media, and chlorinated BPA have also been found in seawater (Liu et al., 2009).

Despite its weak estrogenicity compared with the natural ligand 17β-estradiol, BPA is a model EDC. Not all halogenated BPA analogs are estrogen receptor (ER) agonists. With an increasing number of bromine atoms, brominated BPAs are decreasingly potent ER agonists (Meerts et al., 2001; Riu et al., 2011). Conversely, the estrogenic activity of some chlorinated congeners exceeds that of BPA (Takemura et al., 2005). Likewise, TBBPA could act as an antiandrogene ligand (Christen et al., 2010) and both TBBPA and TCBPA are candidate thyroid hormone disrupting chemicals (Kitamura et al., 2002; Sun et al., 2009). Recently, we demonstrated that brominated bisphenols (di-, triBPA, and TBBPA) as well as TCBPA are potent peroxisome proliferator-activated receptor-PPARγ agonists (Riu et al., 2011).

PPARγ is a member of the NR superfamily, and the disruption of regulation pathways under the control of these NR may be involved in the onset of diabetes, obesity, and reproductive disorders (Swedenborg et al., 2009). BPA analogs with a lower degree of chlorination, but also the biotransformation product of halogenated BPA analogs, may also target PPARγ.

In this study, we first examined the effect of the degree of halogenation of chlorinated BPA on both their affinity and activity toward human ERs and PPARs, using ERα, ERβ, PPARα, PPARβ, and PPARγ reporter cell lines. Then, we took advantage of the difference observed between the ERα and PPARγ activities of BPA and its halogenated derivatives to explore the capability of ERα-based affinity columns (Riu et al., 2008) and newly developed PPARγ-based affinity columns and to retain these different molecules. Affinity columns were further used in combination with high-performance liquid chromatography (HPLC) and mass spectrometry, to isolate and characterize the active metabolites of major halogenated BPA (TCBPA, TBBPA) formed in vitro after their biotransformation by human hepatoblastoma cells (HepG2).

MATERIALS AND METHODS

Chemicals. 17β-Estradiol, BPA, and TBBPA (2,2-Bis(3,5-dibromo-4-hydroxyphenyl)propane) were purchased from Sigma-Aldrich (Saint-Quentin Fallavier, France). TCBPA (2,2-Bis(3,5-dichloro-4-hydroxyphenyl)propane) was purchased from TCI Europe (Zwijndrecht, Belgium). CD4700, a PPARγ-selective agonist, either unlabelled or tritiated (specific activity: 1100 GBq/nmol), were gifts from Michel Rivier (Galderma Research, Sofia-Antipolis, France). 3H-rosiglitazone (2000 GBq/mmol), 14C-17β-estradiol (14C-E2; 1.9 GBq/mmol) and 3H-17α-estradiol (3H-E2; 1.528 GBq/mmol), and 14C-BPA (7.4 GBq/mmol) were purchased from PerkinElmer (Courtaboeuf, France), NEN Life Sciences Products (Paris, France), and Moravek Biochemicals (CA), respectively. All solvents (analytical grade) were purchased from Scharlau Chemie SA (Barcelona, Spain). Materials for cell culture were obtained from Invitrogen (Cergy-Pontoise, France) and luciferin from Promega (Charbonnieres, France).

Synthesis of halogenated BPA analogs. Radiolabeled TBBPA was synthesized from ring-14C-BPA as previously described (Zalik et al., 2006; Fig. 1). TBBPA radio-purity was checked by HPLC (>99.8%). Its specific activity was 3.64 kBq/μmol, and its structure was confirmed by electrospray ionization-mass spectrometry (ESI-MS) and nuclear magnetic resonance (NMR).

Radiolabeled brominated BPA (3-monobromobPA, 3,3'-dibromobPA and 3,3',5'-tribromobPA) were similarly synthesized from ring-14C-BPA, using 2.2 equivalents of bromine, added to 1.48 MBq 14C-BPA, and 2 mg unlabelled BPA in 80 μl methanol/water (1:1 vol/vol). They were individually isolated and purified using the HPLC system coupled to a Gilson 202 fraction collector (Gilson France, Villiers-Le-Bel, France). The radio-purity of all brominated analogs was checked by HPLC (>99.8%). Their structures were confirmed by ESI-MS and NMR. Their specific activity was 168 Bq/nmol.

The synthesis of radiolabeled chlorinated BPA (3-monochlorobPA, 3,3'-dichlorobPA, 3,3',5-trichlorobPA and TCBPA) was adapted from Gallard et al. (2004). BPA, 1mM (3.7 MBq of 14C-BPA) in buffered ultrapure water (phosphate buffer 10mM, pH 8) was chlorinated with NaOCl (2.5 μg/ml). After 10 min shaking at room temperature, reaction was quenched with 400 μl of Na2S2O4 (0.1g/ml). Radiolabeled 14C-chlorinated BPA analogs were individually isolated and purified by HPLC fractionation. Radiochemical purity of all chlorinated congeners was checked by radio-HPLC (>99.8%). Their specific activity was 3.7 KBq/μmol, and their structures were confirmed by ESI-MS and NMR.

The monosulfate conjugate of TBBPA (TBBPA-sulf) was synthesized from TBBPA (adapted from Lindholst et al., 2003). Chlorsulfonic acid (2.76 mmol), from Sigma-Aldrich, was added drop by drop to a constantly stirred ice-cooled solution of 1.84 mmol TBBPA dissolved in 3 ml pyridine. After 2 h at room temperature, the reaction was quenched with concentrated KOH (1:1). The solid portion was filtered off by suction and allowed to dry. Purification of TBBPA-sulf was obtained by solid-phase extraction (SPE) followed by HPLC. SPE chromatobond 1 g C8 glass cartridges (Macherey Nagel, Hoerdt, France) were washed with 4 ml methanol and equilibrated with 4 ml of solution A ("A"):20mM ammonium acetate, pH 3.5/acetonitrile (95/5, vol/vol). The powder was dissolved in 5 ml "A" and loaded on cartridge. Elution was performed successively with 8 ml "A", 4 ml "A":acetonitrile (85:15, vol/vol), 4 ml "A":acetonitrile (50:50, vol/vol), and 4 ml acetonitrile. TBBPA-sulf was eluted in the penultimate fraction. Purity and chemical structure were checked by HPLC-diode array detector (DAD), ESI-MS, and NMR.

Analytical procedure. HPLC analyses were performed on a Spectra system P1000 (Thermo Electron, Les Ulis, France) equipped with a Rhodyne injector and an UV-DAD set at 254 nm (Spectra system UV6000LP, Thermo Electron), connected to a Flo-one/B A500 radioactivity detector (Radiomatic 610TR, PerkinElmer, Villebon-sur-Yvette, France). Radio-HPLC metabolic profiles of TBBPA and TBBPA incubated with HepG2 cells were carried out with an HPLC system based on a Nucleodur C8 column (250 × 4 mm, 5 μm; Macherey-Nagel) in the following conditions: mobile phases: A:"A"; B: 100% acetonitrile; gradient: 0–4 min, A: 100% isocratic; 4–6 min, linear gradient from A: 100% to A:B 70:30; 6–21 min, A:B 70:30 isocratic; 21–23 min, linear gradient from A:B 70:30 to A:B 50:50; 23–33 min, A:B 50:50 isocratic; 33–34 min, linear gradient from A:B 50:50 to B: 100%; 34–50 min, B: 100% isocratic.

Cell incubation. HepG2 human hepatoblastoma cell (Aden et al., 1979) (ATCC No. HB-8065) were cultured in minimum essential medium, 10% fetal calf serum, vol/vol, penicillin (100 U/ml), and streptomycin (100 μg/ml) in a 5% CO2 atmosphere at 37°C.

HepG2 cells (3.5 × 105 per well) were seeded in 12-well plates. After 16 h, the medium was replaced by fetal calf serum-free medium supplemented with 1μM 14C-TBBPA (or 14C-TCBPA) (3700 Bq/incubation)
Control incubations were carried out in the same conditions but without cells. At the end of the incubation period, media were removed and the cells were recovered by washing each well with 1 ml H2O/acetone (50:50, vol/vol). Supernatant and cells were processed, that is, (1) radioactivity was quantified by direct counting in a Packard scintillation counter (Model Tri-Carb 200CA; PerkinElmer) using Ultima Gold as the scintillation cocktail and (2) analyzed by radio-HPLC. For all vials, sample quenching was compensated by the use of quench curves and external standardization. All experiments were made in triplicate.

**Reporter cell lines and stable gene expression assays.**  Generation of HGELN, HGELN-ERα, HGELN-ERβ, HGELN-GAL-PPARγ, HGELN-GAL-PPARβ, and HGELN-GAL-PPARγ reporter cell lines was performed as described in Riu et al. (2011). Briefly, reporter cells were seeded at a density of 20,000 cells per well, in 96-well white opaque tissue culture plates, and maintained in Dulbecco’s modified eagle medium (DMEM) without phenol red, supplemented with 5% dextran-coated, charcoal-treated FCS; 24 h later, culture medium was replaced by DMEM containing tested compounds. Assays were performed in absence of serum in order to avoid ligand capture by serum proteins; 16 h after exposure, media was replaced with media containing 20% glycerol and the corresponding ligand at a concentration of 1 mM. Three volumes of methanol were added to the pooled eluted fractions in order to precipitate and denaturate the receptor. Supernatants, recovered after a centrifugation step at 5000 rpm (2 min), were evaporated to eliminate organic solvent and concentrated on C18 glass cartridges (Macherey Nagel) previously washed with 4 ml methanol and equilibrated with 1 ml EB. Elution was performed with 2 ml acetonitrile. Aliquots from eluted fractions were analyzed by radio-HPLC.

**Crystallization.** Prior to crystallization trials, the purified PPARγ-LBD was concentrated to 8.5 mg/ml in a buffer containing 20 mM Tris-HCl, pH 8.5, 250 mM NaCl, 5 mM dithiothreitol, and 1 mM ethylenediaminetetraacetic acid. Crystals with TBBPA-sulf were obtained by vapor diffusion in hanging drops at 293 K by mixing 1 μl of protein solution with 1 μl of well solution containing 1 M trisodium citrate, 100 mM HEPE, pH 7.5, 3.5%, 1.2-propanediol, and TBBPA-sulf at 0.2 mM in order to have a molar ratio protein:ligand of 1:2 in the molar solution. Crystals appeared after 1 day and grew to about 100 μm within few days. Crystals were transferred to a cryoprotectant comprised of the well solution containing 20% glycerol and the corresponding ligand at a concentration of 1 mM and frozen in liquid nitrogen.

**Crystallographic data collection, processing, and structure refinement.** Diffraction data were collected using an ADSC Quantum Q315r detector at the ID14-4 beamline of the European Synchrotron Radiation facility (ESRF, Grenoble, France) at 2.30 Å. Diffraction data were processed using MOSFLM (Leslie, 1992) and scaled with SCALA from the CCP4 program suite (CCP4 1994). The structure was solved by using the previously reported structure 2ZVT (Leslie, 1992) and scaled with SCALA from the CCP4 program suite (Afonine et al., 2005). Additional experiments were carried out to trap the metabolites formed in incubations of TCBPA or TBBPA with HepG2 cells that would behave as PPARγ ligands. An aliquot of each cell incubation (0.5 nmol of equivalent of the parent compound), diluted in 500 μl washing buffer (WB, final volume 1 ml, with 5% organic solvent) was directly applied on 2 nmol of immobilized PPARγ-LBD. WB and elution buffer (EB) volumes were 1 ml and 500 μl, respectively. Three volumes of methanol were added to the pooled eluted fractions in order to precipitate and denature the receptor. Supernatants, recovered after a centrifugation step at 5000 rpm (2 min), were evaporated to eliminate organic solvent and concentrated on C18 glass cartridges (Macherey Nagel) previously washed with 4 ml methanol and equilibrated with 1 ml EB. Elution was performed with 2 ml acetonitrile. Aliquots from eluted fractions were analyzed by radio-HPLC.

**RESULTS**

Activity of Chlorinated BPAs on ERs and PPARs

The occurrence of chlorinated BPA derivatives in the ecosystem, likely originating from the chlorination of BPA,
prompted us to evaluate the activities of these BPA analogs on ERs and PPARs, using stably transfected HGELN-ERα and -ERβ and HGELN-PPARα, -PPARδ, and -PPARγ cell lines. Dose-responses curves obtained with ERs and PPARγ cells (Fig. 2A) demonstrated the sensitivity of the assays for the natural ER ligand E2 (EC50 of 18.9 and 81.5pM for ERα and ERβ, respectively) and for pharmaceutical ligands of PPARγ, namely CD4700 and rosiglitazone (EC50 of 1.32nM and 29.6nM, respectively). All chlorinated BPAs were first tested for nonspecific modulation of luciferase expression on the HGELN parental cell line containing the same reporter gene as HGELN-ER or HGELN-PPAR cells but devoid of ERs or PPARs, then on HGELN-ERα and -ERβ and HGELN-PPARα, -PPARδ, and -PPARγ cell lines. All compounds except BPA were cytotoxic above 10μM, in the six HGELN cell lines. As shown in Figure 2B, BPA and all chlorinated BPA analogs displayed a full agonistic activity toward ERα with a 1,000 to 10,000-fold lower potency than E2. However, as observed for brominated BPAs (Riu et al., 2011), the capability of these compounds to transactivate luciferase gene expression in HELN-ERβ cell lines was found to depend on their chlorination level (Fig. 2C). BPA and monoCBPA displayed the highest transactivation efficiency, followed by diCBPA and triCBPA, and finally TCBPA. Figure 2D shows that in contrast to HGELN-ERβ cells, the potency to transactivate luciferase gene expression in the HGELN-PPARγ cell line was reversed, the highest agonistic activity being observed for BPA analogs with the highest degree of halogenation (TCBPA and triCBPA). Compared with the reference agonist, rosiglitazone, these compounds were found to be 50- to 500-fold less potent toward PPARγ. Halogenated BPAs were also tested on HGELN-PPARα and HGELN-PPARβ cell lines, revealing no significant agonist activity on these PPAR subtypes. For each tested compound, the EC50 values were calculated and reported in Table 1.

Establishment and Validation of NR-Based Affinity Columns

ERα-LBD affinity columns were previously demonstrated to efficiently trap estrogenic compounds from complex matrices (Riu et al., 2008). We developed the same tool based on the human PPARγ-LBD, in order to characterize new compounds able to disrupt PPARγ activities. The functionality of ERα- and

![FIG. 2. Dose-response curves for reference high-affinity ligands, BPA, and its halogenated analogs in reporter cell lines. (A) HELN (ERα and ERβ) and HGELN-PPARγ luciferase assays of natural estrogen and pharmaceutical high-affinity ligands. (B) HELN-ERα luciferase assays of BPA and halogenated BPA. (C) HELN-ERβ luciferase assays of BPA and halogenated BPA. (D) HGELN-PPARγ luciferase assays of BPA and halogenated BPA. Results are expressed as a percentage of luciferase activity measured per well (mean ± SEM, n = 4). The values obtained in the presence of 10nM E2 and 100nM CD4700, respectively, were taken as 100.](http://toxsci.oxfordjournals.org/issue/3375)
TABLE 1
Chlorinated BPA Potency to Activate ERs and PPARγ

<table>
<thead>
<tr>
<th>Compound</th>
<th>HELN-Era EC_{50 \pm SE}</th>
<th>HELN-ERa EC_{50 \pm SE}</th>
<th>HGELN-PPARγ EC_{50 \pm SE}</th>
</tr>
</thead>
<tbody>
<tr>
<td>E2</td>
<td>18.9 μM ± 1.8</td>
<td>81.5 μM ± 8.8</td>
<td></td>
</tr>
<tr>
<td>BPA</td>
<td>0.29 μM ± 0.03</td>
<td>0.28 μM ± 0.05</td>
<td></td>
</tr>
<tr>
<td>monoCBPA</td>
<td>0.28 μM ± 0.04</td>
<td>0.35 μM ± 0.12</td>
<td>ND</td>
</tr>
<tr>
<td>diCBPA</td>
<td>0.54 μM ± 0.09</td>
<td>0.63 μM ± 0.27</td>
<td>ND</td>
</tr>
<tr>
<td>triCBPA</td>
<td>0.34 μM ± 0.17</td>
<td>0.88 μM ± 0.36</td>
<td>2.34 μM ± 0.86</td>
</tr>
<tr>
<td>TCBPA</td>
<td>0.53 μM ± 0.25</td>
<td>1.15 μM ± 0.51</td>
<td>1.28 μM ± 0.16</td>
</tr>
<tr>
<td>Rosiglitazone</td>
<td>29.6 nmM ± 0.62</td>
<td>1.32 nmM ± 0.13</td>
<td></td>
</tr>
</tbody>
</table>

Note. ND, not determined.

PPARγ-LBDs was verified by measuring their affinities for reference ligands as previously described (Pillon et al., 2005). The dissociation constant (Kd) determined by Scatchard analysis and using the human recombinant ERα-LBD was of 0.31 nM for 3H-E2, a value very similar to that obtained with the full-length receptor. Similarly, the binding affinities of the pharmaceutical ligands 3H-rosiglitazone and 3H-CD4700 loaded on the PPARγ-LBD were determined as 25 and 0.56 nM, respectively. Having demonstrated the functionality of the bacterially expressed human receptors, we then examined the ability of NR-based affinity columns to retain high-affinity radiolabeled ligands (Fig. 3). With 91.6 ± 1.3% of the radioactivity loaded on immobilized ERα-LBD recovered in eluted fractions, the ERα-LBD column showed a high effectiveness to capture 14C-E2. Likewise, as shown in Figure 3A, 89.9 ± 0.9% of the 3H-C4D700 loaded on the PPARγ-LBD column was trapped, whereas no radioactivity was recovered in eluted fractions in the absence of PPARγ-LBD (data not shown). Together, these data validate affinity columns both as novel tools for the identification and characterization of ERα and PPARγ ligands.

Radiolabeled BPA and each of the halogenated analogs (1 nmol) were loaded separately on columns after the immobilization of 2 nmol of either ERα or PPARγ on the Ni-NTA phase. Using the ERα-based affinity column (Fig. 3B), most of the radioactivity was recovered in eluted fractions, for BPA (98.0 ± 1.9%), monoBBPA (94.6 ± 0.2%), diBBPA (96.2 ± 1.1%), and triBBPA (59.1 ± 7.3%). As observed in stably transfected cell line assays with which we showed that TBBPA was not able to activate ERα (Riu et al., 2011), TBBPA did not bind to the immobilized ERα-LBD, with most of the loaded radioactivity recovered in washing fractions. Conversely, using the PPARγ-LBD affinity column, most of the radioactivity was recovered in eluted fractions for TBBPA (97.5 ± 0.6%), triBBPA (94.9 ± 1.2%), and diBBPA (83.5 ± 2.1%). Neither monoBBPA nor BPA were trapped on the PPARγ-LBD affinity column. The same binding profile was observed for chlorinated BPA analogs (Fig. 3C). All of them were trapped on ERα-based affinity columns, with more than 90% of the radioactivity recovered in elution fractions for mono-, di-, and triCBPA, whereas only 63.1 ± 4.7% of the loaded radioactivity was recovered in elution fractions for TCBPA. In full agreement with cell-based assays (Fig. 2D), only the most chlorinated analogs were efficiently trapped on the PPARγ-LBD affinity column but to a lesser extent than the corresponding brominated BPA analogs: 27.7 ± 0.7%, 72.2 ± 0.4%, and 84.7 ± 0.5% of the radioactivity recovered in elution fractions for diCBPA, triCBPA, and TCBPA, respectively. As shown in Figure 3A, monoCBPA was not able to bind PPARγ-LBD immobilized on agarose columns, with almost all the radioactivity being recovered in the loading and washing fractions.

The nonspecific binding to the NR-LBD columns was evaluated for each tested compound using a large amount of non-radiolabeled high-affinity ligand (E2 or CD4700) trapped on the columns prior to the assays with the radiolabeled compound of interest. In additional controls carried out, none of the halogenated compounds was retained on the phase when no NR-LBD was immobilized on it. All together, these results demonstrate that NR-LBD affinity columns are able to capture both high- and low-affinity ligands and then could allow the identification and the characterization of unknown ligands from complex matrices.

Metabolites of Halogenated BPAs Are New PPARγ Ligands

In the HPLC system developed for this study, TCBPA and TBBPA were eluted at 36.5 and 37.5 min, respectively. After 24-h incubations of HepG2 cells with either 14C-TBBPA or 14C-TCBPA, more than 95% of the radioactivity was recovered in incubation media. As observed on the radio-chromatograms obtained from incubations with TCBPA (Fig. 4A) or TBBPA (Fig. 4B), both molecules were completely metabolized after 24 h into two metabolites, namely TC1/TC2 for TCBPA and TB1/TB2 for TBBPA. These metabolites accounted for 14.1 ± 5.6% (TC1), 85.9 ± 5.6% (TC2), 13.5 ± 2.9% (TB1), and 86.5 ± 2.9% (TB2) of the recovered radioactivity. ESI-MS analyses performed on incubation media extracts allowed to characterize TC1 and TB1 as glucuronic acid conjugates of TCBPA (Glc-TCBPA, Rf: 16.8 min) and TBBPA (Glc-TBBPA, Rf: 19.9 min), respectively, whereas TC2 and TB2 were identified as the sulfate conjugates of TCBPA (TCBPA-sulf; Rf: 25.8 min) and TBBPA (TBBPA-sulf; Rf: 27.4 min), respectively.

To further characterize these metabolites, TCBPA and TBBPA incubation media were separately loaded on the PPARγ-based column. As shown in Figures 4C and 4D, more than 80% of the radioactivity was retained on the column and recovered in elution fractions, which were subsequently processed and analyzed by radio-HPLC (Figs. 4E and 4F).

A single peak was observed on each radiochromatogram with TBBPA incubation media were separately loaded on the PPARγ-LBD column, neither TBBPA nor TCBPA would behave as PPARγ ligands. Conversely, none of the glucuronide conjugates was retained on the PPARγ-LBD column, neither
glucuronide nor sulfate conjugates were trapped. To confirm and better characterize the agonistic activity of TCBPA-sulf and TBBPA-sulf, both ligands were chemically synthesized and purified. As shown in Figure 5, the two metabolites displayed residual agonistic activity on HGELN-PPARγ cell lines. Not only these sulfate-conjugated metabolites were able to bind PPARγ-LBD immobilized on affinity columns but they were also found to be able to activate this NR.

It is worth notifying that similar experiments performed with BPA provided monoglucuronide and monosulfate conjugates after 24-h incubation with HepG2 cells, none of them being trapped on ERα-LBD or PPARγ-LBD columns (data not shown).

**Structural Analysis of PPARγ in Complex with TBBPA-sulf**

In order to compare the binding mode of this metabolite with that of the parent compound TBBPA (Riu et al., 2011), we...
solved the crystal structure of PPAR\textsubscript{c}-LBD bound to TBBPA-sulf (TB2) (Atomic coordinates and structure factors for the reported crystal structure has been deposited with the Protein Data Bank under accession code 3 pba.). Overall, the structure of PPAR\textsubscript{c}-LBD is identical when bound to TBBPA or TBBPA-sulf, with an RMSD value of 0.24 Å for superimposed alpha carbons (Fig. 6A). Contrary to TBBPA, TBBPA-sulf is not a symmetric molecule, and in principle, the sulfate group could have been positioned either on the H11/H12 side or near the \(\beta\)-sheet S1/S2. Inspection of the omit \(F_o - F_c\) difference map (Fig. 6B) revealed density for the ligand which could be placed in a position almost identical to that of TBBPA, the sulfate group pointing toward H11/H12 (Fig. 6A). Interestingly, accommodation of the sulfate group by PPAR\textsubscript{c} does not require any conformational change of the surrounding residue side chains. In this orientation, the sulfate moiety fills a portion of the ligand-binding pocket occupied by water molecules in the TBBPA structure and is potentially involved in hydrogen bonds with Tyr327, Lys367, His449, Gln286, and Ser289 (Fig. 6C). However, the electron density of the sulfate group appears as a blob, thus precluding the precise positioning of individual oxygen atoms. Finally, the presence of a conserved water molecule mediates the indirect interaction between TBBPA-sulf and PPAR\textsubscript{c} helix H12 (Fig. 6C). The rather poorly defined density of the sulfate moiety suggests a dynamic behavior of this chemical group. Overall, the crystallographic data validate the notion that this TBBPA metabolite is a PPAR\textsubscript{c} ligand with partial agonistic profile.

**DISCUSSION**

Most of the reported effects of EDC are attributed to their interference with hormonal signaling mediated by NR. Even
low doses of contaminants can trigger biological effects when exposure occurs during critical windows of exposure (Newbold et al., 2009; Vandenberg et al., 2009). An extensive variety of man-made contaminants exist. Many of these chemicals undergo biotic or abiotic structural changes in the environment, and biotransformations at the level of target organisms can also modulate EDC’s activity. These facts greatly complicate the picture and require the use of adequate approaches for the screening of environmental contaminants able to interfere with NR.

Environmental as well as human exposure data clearly indicate that halogenated bisphenols are emerging contaminants, which fate and cellular targets deserve to be better characterized (Cariou et al., 2008; de Wit et al., 2009; Fernandez et al., 2007; Johnson-Restrepo et al., 2008). The occurrence of brominated bisphenols mainly results from the extensive use of TBBPA as a flame retardant. Conversely, a growing body of evidence suggests that chlorinated BPA analogs arise from the abiotic chlorination of BPA residues (Fukazawa et al., 2001; Gallart-Ayala et al., 2010). In this study, we investigated both the ER and PPAR activities of BPA and related halogenated analogs and characterized their binding affinity using newly developed NR-LBD affinity columns, combined with radio-HPLC and MS analyses. The potential of these columns to be used as powerful tools to identify metabolites of interest was further validated using reference ligands, then applied to the study of TBBPA and TCBPA metabolites formed after 24-h incubation with HepG2 cells.

For chlorinated BPA analogs, the results obtained with HeLa cells stably transfected with ERα showed that these compounds are at least as potent agonists as BPA, consistent with a previous study (Takemura et al., 2005). Contrary to brominated BPA (Riu et al., 2011), the halogenation degree of chlorinated bisphenols did not decrease their ERα potency. Likely due to the lower ligand-binding pocket size of ERβ compared with ERα, chlorinated analogs with the lowest molecular weight (mono- and diCBPA) exhibited the highest agonistic activity in the HELN-ERβ cell line. As previously observed for brominated BPA analogs, in HGELN-PPARγ cells, the order of potency was reversed, with TCBPA > triCBPA > diCBPA, whereas monoCBPA as well as BPA showed no PPARγ activity at all. In binding affinity assays using NR-based affinity columns, we confirmed the ranking order observed with stably transfected cells for brominated BPA, as well as for chlorinated BPA, demonstrating a good correlation between binding affinities and agonistic activities. Taken together, these data demonstrate that the ERβ and PPARγ agonistic activities of BPA analogs highly depend on their halogenation degree, contrary to their ERα activities.

Up to now, very few environmental ligands of PPARγ have been characterized (Feige et al., 2007; Grun and Blumberg, 2006; Takeuchi et al., 2006). In fact, only mono(2-ethyl-hexyl) phthalate and tributylin have been shown to significantly activate PPARγ, whereas organotins activate RXR/PPARγ, primarily by interacting with RXR (le Maire et al., 2009). The use of PPARγ-LBD–based affinity columns clearly demonstrates a direct interaction of halogenated BPA derivatives with...
PPARγ. The fact that bisphenol analogs bearing at least two halogens are also ligands of PPARγ considerably strengthens the hypothesis that environmental ligands could be involved in the disruption of energy balance regulation. Like for other EDC, perinatal exposure could play a critical role. Indeed, a large body of data regarding EDCs nowadays underlines the importance of exposure during early stages of the development, which can later of data regarding EDCs nowadays underlines the importance of exposure during early stages of the development, which can later result in adverse effects (Newbold et al., 2009). For BPA, this includes reproductive targets and obesity. As demonstrated by Somm et al. (2009), a perinatal exposure to BPA alters early adipogenesis in rat, a process known to be mediated by PPARγ. According to Cariou et al. (2008), significant levels of TBBPA are found in human cord blood (200 pg/g fresh weight) and maternal milk (0.1–37.4 ng/g lipid weight), demonstrating both prenatal and postnatal exposure in a large fraction of the tested population. Recently, a study conducted in Spain demonstrated the presence of mono-, di-, and triCBPA in human placenta (5.1–21.4 ng/g, 12.7–58.8 ng/g, and 4.0–31.2 ng/g fresh tissue, respectively), suggesting as well a perinatal exposure to chlorinated BPA (Jimenez-Diaz et al., 2010).

We previously showed that ERz-LBD affinity columns are able to trap estrogenic compounds present in environment or in food samples (Pillon et al., 2005; Riu et al., 2008). In this study, the newly developed PPARγ-LBD columns were demonstrated to efficiently trap both parent compounds and compounds present in complex matrices, namely the metabolites formed in in vitro incubations. Human hepatoblastoma cells metabolized both TBBPA and TCBPA into the corresponding glucuronide and sulfate conjugates. BPA is known to be predominantly metabolized through conjugative pathways. Likewise, in vivo, TBBPA is mainly metabolized into TBBPA-monoalcohol (Schauer et al., 2006; Van der Ven et al., 2008). The metabolic fate of TCBPA and other chlorinated BPA analogs in vitro and in vivo is still unknown, but conjugative metabolism is expected to be predominant, consistent with the results obtained in this study using incubations with HepG2 cells. Phase II metabolism, with the production of conjugated metabolites, is almost always a detoxification process. The occurrence of bioactive conjugates is not a common finding. It has previously been reported only for a very limited number of compounds, including morphine and estrogens (Glatt, 2000; Ritter, 2000). In this study, the development of PPARγ-LBD–based columns allowed to identify, for the first time, that the sulfation of halogenated BPA produces bioactive metabolites. Using bio-luminescent cell lines assays expressing PPARγ and X-ray crystallography, we showed that the sulfate conjugates of TBBPA and TCBPA are able to bind PPARγ but also display an agonistic activity toward this NR. Chlorinated bisphenols have already been demonstrated to be present in human tissues. The occurrence of brominated bisphenols has been demonstrated in maternal milk, cord blood, and milk, suggesting direct fetal exposure during pregnancy (Cariou et al., 2008), as well as a neonatal exposure to TBBPA residues (parent molecule and/or conjugates) through milk. The fact that halogenated analogs of BPA are selective PPARγ ligands highlights this NR as a new target for endocrine disruption. Taking into account their structure, their biotransformation process (biotic or abiotic), and their capability to activate different NR (ERs and PPARγ), these halogenated BPA analogs may contribute to the onset of metabolic disorders in human. Their additive effects could further increase such adverse outcomes and should therefore be examined as well.

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### TABLE 2

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<th>Complex</th>
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<td>PDB code</td>
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*Values in parentheses are for highest resolution shell.*
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REFERENCES


