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1 **Comparison of direct and indirect genetic methods for estimating**
2 **seed and pollen dispersal in *Fagus sylvatica* and *Fagus crenata*.**

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20
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1

2 Abstract

3 The comparison between estimates of historical gene flow, using variance in allelic
4 frequencies, and estimates of contemporary gene flow, using parentage assignment, is
5 expected to provide insights into ecological and evolutionary processes at work within and
6 among populations. Genetic variation at microsatellite loci was used to quantify genetic
7 structure in two wind pollinated, gravity and animal-dispersed tree species (*Fagus sylvatica* L.
8 and *Fagus crenata* Blum.) and to derive historical estimates of gene flow. The gene dispersal
9 distances estimated assuming effective population density to be $\frac{1}{4}$ of the observed density
10 were ~ 77 m in European beech and ~ 40 m in Japanese beech. Parentage analyses and a
11 neighbourhood model approach were used to estimate contemporary patterns of seed and
12 pollen dispersal. Our results suggest restricted seed dispersal abilities in both European beech
13 ($\delta_s = 10.5$ m) and Japanese beech ($\delta_s = 12.4$ m), with an exponential shaped seed dispersal
14 kernel. A non-negligible rate of seed immigration ($m_s = 27\%$) was detected in European beech
15 sites but not in Japanese beech site. Pollen dispersal within studied sites also appeared limited
16 ($\delta_p = 41.63$ m in European beech and $\delta_p = 79.4$ m in Japanese beech), despite high rate of
17 pollen immigration ($m_p = 68\%$ in European beech and $m_p = 40\%$ m in Japanese beech).
18 Interestingly, contemporary and historical estimates of gene flow were within the same order
19 of magnitude (a few tens of meters).

1

2 Introduction

3 The impact of predicted climatic changes on forest ecosystems is expected to be acute,
4 resulting in notable changes in species' ranges, ecosystem function and species interactions.
5 Because trees are long-lived they can disappear, disperse to other places or adapt *in situ* to
6 climate change over a few generations. For several tree species bioclimatic models predict
7 geographical shifts in the range of suitable habitat of several hundreds of kilometres by 2100
8 (without altitudinal compensation, Jump et al., 2009). The range of European beech (*Fagus*
9 *sylvatica* L.) for example is predicted to strongly contract in the west while expanding
10 northwards, due to inadequate chilling in the west as winter warms more than summer (Sykes
11 et al., 1996). These important predicted distribution shifts partly account for the intense
12 research effort dedicated to species dispersal in recent decades (Bullock and Nathan, 2008;
13 Ronce, 2007).

14 A primary concern in plants is whether seed dispersal capacities will allow species to
15 track 21st-century global warming. Analyses of paleoecological records are classically
16 interpreted to suggest that tree populations are capable of rapid migration when climate
17 warms, with rates of range expansion as high as 1000 m/yr during the early Holocene. Such a
18 fast migration rate is possible by rare long-distance migration events (Clark, 1998) or by high
19 latitude refugia reseeded the landscape (McLachlan et al., 2005). The latter means that post-
20 glacial re-colonization may have been as much as an order of magnitude slower than
21 previously thought (< 100 m/yr). Another major issue related to plant dispersal is whether
22 gene migration either through pollen or seeds can support local adaptation to changing
23 climatic conditions by mixing or adding well adapted genetic material to poorly-adapted gene
24 pools (Davis and Shaw, 2001). The impact of migration on local adaptation depends on other
25 important biological parameters (genetic effects, life cycle, mating system), and is

1 complicated to predict when realistic population demography estimates are included
2 (Lenormand, 2002). For instance, while gene flow tends to oppose the effects of local
3 selection and thus limits adaptation, intermediate rates of migration have been advocated to be
4 optimal for local adaptation in low density populations at the range margin (Alleaume-
5 Benharira et al., 2006). A more dynamic view of distribution ranges is now emerging in the
6 literature which accounts for demographic and evolutionary processes associated to species
7 expansions and retraction, and to marginal populations (Hampe and Petit, 2005).

8 Thus it is clear that predicting populations' responses to ongoing climate change
9 requires fine characterisation of individual and gene dispersal processes at local, ecological
10 scales as well as large, species-range scales. A way to address this issue is to use the rich
11 toolbox available to estimate pollen and seed dispersal from genotype data at these different
12 scales (Manel et al., 2005; Smouse and Sork, 2004). These approaches are generally termed
13 "indirect" when using gene pool data or "direct" when using exact genotypes. The most
14 indirect approaches estimate historical dispersal parameters from the observed genetic
15 structure by fitting them to theoretical population models generally assuming drift– migration
16 equilibrium. By contrast, the most direct approaches use genetic markers to monitor
17 contemporary movements of individuals or propagules by reconstructing parentage
18 relationships (Manel et al., 2005), as in paternity/parentage analyses which are model-free
19 beyond assuming Mendelian inheritance of the markers.

20 Indirect approaches to assess historical gene flow within continuous populations employ
21 the principle that Spatial Genetic Structure (SGS) displayed by neutral markers is essentially
22 caused by local genetic drift, the effect of which is counterbalanced by gene dispersal. Under
23 this isolation by distance model, the decay rate of genetic relatedness between individuals
24 with distance has been shown to be inversely proportional to $d_e \sigma_e^2$, with σ_e^2 equal to half of
25 the mean-squared parent-offspring distance and d_e equal to the effective density of individuals

1 (Rousset, 2000). Intuitively, the product $d_e\sigma_e^2$ expresses the degree of overlap between
2 individual “gene shadows” (the spatial distribution of gene dispersal events around each
3 parent). It implies that the intensity of genetic structuring decreases both with increasing
4 dispersal and increasing individual density. Assuming that the genetic markers used in these
5 analyses are not affected by natural selection, the summary parameter $d_e\sigma_e^2$ reflects the
6 importance of both gene flow and genetic drift in shaping local differentiation on a historical
7 time-scale. When reliable estimate of effective population size is available refined methods
8 have been proposed to estimate the gene dispersal distance σ_e from patterns of SGS
9 (Vekemans and Hardy, 2004).

10 Alternatively, patterns of contemporary gene flow between the parental and offspring
11 cohorts can be estimated by using genetic fingerprint and parentage analyses to assign
12 seedlings to their parent trees (parentage inference), and to infer the pollen and seed dispersal
13 curves (Oddou-Muratorio and Klein, 2008). In the case of plant populations, parentage
14 analysis consists of genotyping a sample of dispersed seeds or established seedlings and all
15 the reproductive plants within a circumscribed area for a set of shared polymorphic markers to
16 detect the parent pair of each seedling (Meagher, 1986). To discriminate between male and
17 female parentage of seed and seedlings, one can genotype maternally inherited tissues
18 collected on dispersed seeds (Jones et al., 2005; Jordano et al., 2007). When dealing with
19 established seedlings, where purely maternal tissues are no longer available, average effective
20 pollen/seed dispersal distance can be directly estimated from parent-offspring genotype data
21 using model fitting, such as the Neighbourhood model (Burczyk et al., 2006). These spatially
22 explicit mating models have gained a broad acceptance among population geneticists and
23 ecologists because they allow characterisation the seed and pollen dispersal processes and the
24 heterogeneity in male/female fertility at an ecological time scale (Burczyk et al., 2006;
25 Morgan and Conner, 2001), as well as the ecological factors that are likely to influence these

1 patterns, such as parental phenotypic traits (Gonzalez-Martinez et al., 2006), seed disperser
2 behaviour (Jordano et al., 2007), or spatial environmental heterogeneity (Jones et al., 2005).

3 The comparison of historical versus contemporary estimates of gene flow is expected to
4 provide interesting insights into the balance between and the time-scale of evolutionary forces
5 at work within and among populations (Oddou-Muratorio and Klein, 2008). It has been
6 argued that in undisturbed populations the consistency between historical and contemporary
7 estimates of gene flow supports the drift-migration equilibrium hypothesis (Dunphy and
8 Hamrick, 2005; Otero-Arnaiz et al., 2005), while divergence among real-time and effective
9 migration rates has been used to highlight the consequence of sex-biased dispersal and social
10 structure in small mammal species (Schweizer et al., 2007). In very recently disturbed
11 populations, only contemporary estimates of gene flow should be affected because of
12 temporal inertia of historical estimates for few (<10) generations (Leblois et al., 2004) ; the
13 discrepancy among historical and contemporary estimates is thus expected to reflect the
14 magnitude of the perturbation (Dutech et al., 2005). By contrast, consistent contemporary and
15 historical estimates of gene flow in a recently perturbed area may indicate that compensatory
16 mechanisms allow the maintenance of gene movement (Bacles et al., 2005).

17 European and Japanese beech (resp. *Fagus sylvatica* L. and *Fagus crenata* Blum.) are
18 two related, deciduous, broad-leaved tree species that cover large portions of Europe and
19 Japan., where they tend to grow in high density stands. Despite the high economic and
20 ecological importance of forests dominated by European or Japanese beech, the processes
21 shaping their ecological and evolutionary response in face of global changes are only starting
22 to be understood. In particular, contemporary pollen and seed dispersal patterns in beech have
23 begun to be studied only recently (Kramer et al 2008), showing that 90% of seeds disperse
24 within 25m. Patterns of SGS show a tendency for strong family structure up to 30-40 m,
25 which suggests restricted migration rates of both pollen and seeds (Chybicki et al., 2009;

1 Jump and Penuelas, 2007; Leonardi et al., 1996; Vornam et al., 2004). However, to our
2 knowledge, the most recent available methods to estimate contemporary patterns of gene flow
3 in these species have not been used, and historical and contemporary estimates have not been
4 compared to date.

5 In the present study, we applied direct and indirect approaches to investigate both
6 patterns of contemporary and historical gene flow in 4 sites (3 for European beech and 1 for
7 Japanese beech). This study has several objectives: (1) to provide robust historical estimates
8 of gene dispersal distance σ_e based on SGS, and contemporary estimates of seed and pollen
9 dispersal based on spatially explicit mating models in European and Japanese beech ; (2)
10 based on these estimates to investigate variation in the estimated dispersal abilities between
11 sites (for European beech) and species and (3) to compare historical and contemporary
12 estimates of dispersal within each site, with regard to the statistical specificities of both
13 estimates and the expected effects of recent population demography within and among sites.
14 Based on previous studies of SGS in beech, we expect rather restricted contemporary
15 estimates of dispersal abilities in beech. However, pollen dispersal abilities are usually
16 assumed important in wind-pollinated tree species. Also, our sampling design include recently
17 disturbed forest stands, in which discrepancies between contemporary and historical gene
18 flow estimates are susceptible to arise. Confronting contemporary and historical dispersal
19 estimates will thus allow us to investigate how pollen and seed mediated gene flow
20 respectively shape patterns of SGS in beech.

21 **Material and methods**

22 **Studied species and sampling design**

23 European and Japanese beech are described as highly outcrossing wind-pollinated trees
24 (Asuka et al., 2005; Hanaoka et al., 2007; Merzeau et al., 1994). Reproductive trees (typically
25 older than 40-50 years) produce beech nuts in irregular mast years (with an average interval

1 between mast years of ~5 or ~7 years; Tomita et al., 2002). Mature seed dissemination begins
2 in September, primarily through gravity and secondarily assisted by various animals including
3 rodents such as voles and wood mice that have been reported to scatter-hoard seeds (Jensen,
4 1985).

5 For European beech three plots were studied within two different locations (Table 1). FS1 and
6 FS2 are located on Mont Ventoux, South-Eastern France, where an expanding population of
7 European beech mixed with *Abies alba* is re-colonising a pine forest (Bontemps et al.,
8 unpublished). FS3 is located in the ancient Haye Forest, North-Eastern France (Oddou-
9 Muratorio et al., unpublished). Within each plot, adult trees (i.e. with diameter > 10 cm or
10 producing seeds) were exhaustively mapped (Table 1) and their diameter was measured (Fig.
11 1). Within each site, both seedlings (seedlings, age <3 years) and saplings (< 1 m high, app.
12 younger than 20 years, roughly estimated from scars counts) were sampled non exhaustively,
13 and mapped (Fig 1).

14 For Japanese beech, the 170×170 m study plot was located in a primary beech forest in
15 northern Japan on the southern foot of Mt Kurokima (see also Takahashi et al., 2000). All of
16 the 287 adult trees were sampled, mapped, and their diameter was measured (Fig 1). 1387
17 seedlings emerged in 2001 and were also mapped and sampled.

18 **Genotyping**

19 All Japanese beech individuals were genotyped at 7 highly variable microsatellite markers
20 (Table 2), while in European beech, all individuals were genotyped at a set of either 5 (FS3)
21 or 11 (FS1 and FS2) microsatellite markers (Asuka et al., 2004; Pastorelli et al., 2003; Tanaka
22 et al., 1999).

23 High prevalence of null alleles has been reported for microsatellite markers developed for
24 European beech (Chybicki and Burczyk, 2009; Oddou-Muratorio et al., 2009). Therefore, we
25 carefully checked for presence of null alleles both in adult and seedling cohorts using the

1 maximum-likelihood method accounting for genotyping error implemented in ML-NullFreq
 2 (Kalinowski and Taper, 2006).

3 **SGS and historical estimates of gene flow**

4 SGS was characterised by analysing variation of genetic relatedness, as measured by the
 5 kinship coefficient (F_{ij}) of Loiselle et al. (1995), as function of the logarithm of geographic
 6 distance d_{ij} between every pair of adults/seedlings i and j .

7 To visualize SGS, F_{ij} values were averaged over a set of distance classes (d), and plotted
 8 against the logarithm of the distance. Approximate standard errors for the multilocus
 9 estimates of F_{ij} within each distance class were obtained through a jackknife procedure that
 10 consisted of deleting each locus one at a time. This assumes that the different loci provide
 11 independent replicates of the genetic structure process. To test SGS, the F_{ij} values were
 12 regressed on $\ln(d_{ij})$, where d_{ij} is the spatial distance between individuals i and j , to provide the
 13 regression slope b . Then, the spatial positions of the individuals were permuted 5,000 times in
 14 order to get the frequency distribution of b under the null hypothesis that F_{ij} and d_{ij} were
 15 uncorrelated. Following Vekemans and Hardy (2004), the SGS intensity was quantified by Sp
 16 $= b/(F_1 - 1)$, where F_1 is the average kinship coefficient between individuals of the first
 17 distance class (< 50 m).

18 An indirect estimate of the product $d_e\sigma_e^2$, where d_e is the effective population density, was
 19 obtained through:

$$20 \quad d_e\sigma_e^2 = -(1-F_N)/4\pi b, \quad (1)$$

21 where F_N is the kinship coefficient between neighbouring individuals, estimated here by F_1
 22 (Rousset, 2000). Ideally, b should be computed as the restricted regression slope of F_{ij} on
 23 $\ln(d_{ij})$ in the range $\sigma_e > d_{ij} > 20\sigma_e$ for relationship (1) to be valid. However, in most data sets
 24 investigated here the iterative procedure required for estimating this restricted regression
 25 slope failed to converge. Therefore, all results reported here were based on regression over the

1 whole range of geographical distances between individuals. All computations were performed
2 with the SPAGeDi software(Hardy and Vekemans, 2002).

3 **Neighbourhood model and contemporary estimates of gene flow**

4 We used a slightly modified version of the neighbourhood mating model described in
5 Burczyk et al. (2006), where all potential parents are encapsulated within a single
6 neighbourhood corresponding to the study site (Oddou-Muratorio and Klein, 2008). We
7 assumed that each sampled seedling i could be mothered either (1) by a mother tree located
8 outside the study site due to seed immigration (with probability m_s) or (2) by a mother tree
9 located within the studied site (with probability $(1- m_s)$). In the latter case offspring i may be
10 the result of self-pollination (with probability s), pollen flow from outside the study site (with
11 probability m_p), or pollen from a sampled male (with probability $1-s-m_p$). The probability of
12 observing a multilocus diploid genotype G_i among the seedlings is thus:

$$13 \quad P(G_i) = m_s T(G_i|BAF) + (1-m_s) \sum_{j \in ISP} \Psi_{ij} \left[s T(G_i|M_j, M_j) + m_p T(G_i|M_j, BAF) + (1-s-m_p) \sum_{k \in ISP} \Phi_{j,k} T(G_i|M_j, F_k) \right]$$

14 (2)

15 where $T(G_i|BAF)$ is the transition probability that an offspring immigrating from a mother
16 tree located outside of the study site has genotype G_i knowing the Background Allelic
17 Frequencies (BAF) ; $T(G_i|M_j, M_j)$, $T(G_i|M_j, BAF)$, $T(G_i|M_j, F_k)$ are the transition probabilities
18 that an offspring has diploid genotype G_i when its mother tree of genotype M_j is respectively
19 (1) self-pollinated (2) pollinated by a father drawn at random in a population with allelic
20 frequencies BAF or (3) pollinated by a father of genotype F_k .

21 Parameter $\Psi_{i,j}$ is the proportion of seeds from mother tree j at the location of seedling i among
22 the virtual seed pool originating from all known mother trees from the neighborhood N :

$$23 \quad \Psi_{ij} = \frac{\tau_{ij}}{\sum_{l \in N} \tau_{il}} \quad (3)$$

1 where τ_{ij} is a function of one or more factors influencing female reproductive success,
 2 including the seed dispersal kernel (see below).

3 Parameter Φ_{jk} is the proportion of pollen from tree k in the pollen pool of mother tree j in the
 4 total pollen pool from the known father trees in the neighborhood N :

$$\Phi_{jk} = \frac{\pi_{jk}}{\sum_{l \in N} \pi_{jl}} \quad (4)$$

5 where π_{jk} is a function of one or more factors influencing male reproductive success,
 6 including the pollen dispersal kernel (see below).

7 *Models for pollen and seed dispersal kernel-* As in Oddou-Muratorio et al. (2005), pollen
 8 dispersal was modelled using a dispersal kernel $p_p(.,x,y)$ describing the probability for a
 9 pollen grain emitted at position (0,0) to participate to the pollen cloud at position (x,y).

10 Following this scheme, we modelled seed dispersal using a dispersal kernel $p_s(.,x,y)$
 11 describing the probability for a seed emitted at position (0,0) to establish as a seedling at
 12 position (x,y). We considered here the family of exponential power functions:

$$p(a,b;x,y) = \frac{b}{2\pi a^2 \Gamma(2/b)} \exp\left(-\left(\frac{\sqrt{x^2+y^2}}{a}\right)^b\right), \quad (5)$$

13 where Γ is the classically defined gamma function (Abramowitz and Stegun, 1964). The
 14 parameter b is the shape parameter affecting the tail of the dispersal function and a is a scale
 15 parameter homogeneous to distance. The mean distance (δ) traveled by a pollen grain/seed
 16 under the kernel $p(a,b;.)$ and the root-mean-square axial distance are given by Austerlitz et al.
 17 (2004) and Rousset (2004):

$$\delta = a \frac{\Gamma(3/b)}{\Gamma(2/b)} \quad \text{and} \quad \sigma = a \sqrt{\frac{\Gamma(4/b)}{2\Gamma(2/b)}} \quad (6)$$

18 Both increase with a and decrease with b . When $b < 1$, the dispersal kernel is fat-tailed (Clark,
 19 1998), i.e. the long-range decrease is slow, increasing the probability of long-distance

1 dispersal events. Conversely, when $b > 1$ (for instance the Gaussian model, for which $b = 2$)
 2 the dispersal is thin-tailed, with a rapid decrease of the dispersal function, implying fewer
 3 long-distance dispersal events than when $b < 1$.

4 *Model for female/male reproductive success (the τ_{ij} 's and the π_{jk} 's):* individual female
 5 reproductive success τ_{ij} of female j on seedlings location i was determined by two kind of
 6 factors interacting multiplicatively: (1) the distance of seedling i from mother j through the
 7 dispersal kernel described above and (2) adult tree diameter as a surrogate for female
 8 fecundity. Here τ_{ij} is expressed as:

$$9 \quad \tau_{ij} = e^{\gamma_f D_j} \times p_s(a_s, b_s; x_i - x_j, y_i - y_j) \quad (7)$$

10 where D_j is the diameter of mother j , and a_s and b_s the parameters of the seed dispersal kernel
 11 p_s described by equation (5). Classically, we chose an exponential function of parameter γ_f to
 12 relate the fecundity surrogate (here D_j) to reproductive success because it assures a positive
 13 value for the reproductive success parameter. Additionally, this model of the selection
 14 gradient is well-suited for a continuously distributed variable such as diameter.

15 Individual male reproductive success π_{jk} of tree k in the local pollen cloud of mother j
 16 depended on two independent factors: (1) the distance of mother j to father k and the pollen
 17 dispersal parameters and (2) the diameter of tree k as a surrogate for male fecundity.

18 Therefore, π_{jk} was expressed as:

$$19 \quad \pi_{jk} = e^{\gamma_m D_k} \times p_p(a_p, b_p; x_j - x_k, y_j - y_k) \quad (8)$$

20 where D_k is the diameter of father k , γ_m is the male fecundity parameter, and a_p and b_p are the
 21 parameters of the pollen dispersal kernel p_p .

22 *Parameters inference-* We obtained maximum likelihood estimates for the seed and pollen
 23 immigration rates (m_s and m_p), selfing rate (s), kernel parameters (a_s and b_s for seed dispersal

1 kernel; a_p and b_p for pollen dispersal kernel) and contextual gradients (γ_f and γ_m), by
 2 maximising the log-likelihood functions:

$$\log L(a_p, b_p, a_s, b_s, s, m_p, m_s, \gamma_f, \gamma_m) = \sum_{i=1}^S \log [P(G_i)]. \quad (9)$$

4 where S is the total number of genotyped seedlings.

5 Parameters were fitted with the Newton-Raphson algorithm using NM+ software (Chybicki
 6 and Burczyk unpublished). NM+ allowed us to estimate simultaneously the frequencies of
 7 null alleles at each marker, and to eliminate markers significantly affected by null alleles (i.e.
 8 with frequency of null alleles >0.10).

9 Results

10 *Genetic markers and null alleles*- The frequency of null alleles was overall lower in the
 11 Japanese beech data set than in any of the three European beech data sets (Table 2).
 12 Frequencies of null alleles ranged from 0 (in 12 site/loci combination out of 33) to 0.125.

13 *SGS and historical estimates*- SGS was significant in the four data sets, with both adults and
 14 seedlings more related than expected by chance for distances up to 30-40 metres (Figure 2).
 15 The strength of SGS, as depicted by the Sp -statistics, significantly differed among sites: the
 16 ancient Japanese beech stand showed the weakest pattern of SGS ($Sp_{FC} = 0.0032$ in adults),
 17 followed by the ancient European beech stand ($Sp_{FS3} = 0.0087$ in adults), then by the high-
 18 density recolonising European beech stand ($Sp_{FS1} = 0.0218$ in adults), and finally by the low-
 19 density recolonising European beech stand ($Sp_{FS2} = 0.0354$ in adults). Within site, stronger
 20 SGS was detected in adults compared to seedlings in the recolonising European beech stand
 21 FS1 (FS1: $Sp_A = 0.0218$ versus $Sp_S = 0.0163$, Table 3) while in FS2 SGS did not differ among
 22 adults and seedlings ($Sp_A = 0.0354$ versus $Sp_S = 0.0205$). In mature populations of Japanese
 23 beech, SGS was significantly weaker among adults compared to seedlings (FS1: $Sp_A = 0.0032$
 24 versus $Sp_S = 0.0124$). In FS3, the sampling design of seedlings did not allow estimation of SGS
 25 (Figure 1).

1 Using observed density $d_{obs}/2$ to estimate d_e , SGS converted into axial variance of gene
 2 dispersal (σ_e) between 31 m (adult, FS2) and 73 m (seedlings, FS2, Table 3) for the European
 3 beech, and between 19 m (seedlings) and 37 m (adult) for the Japanese beech. Fixing $d_e =$
 4 $d_{obs}/10$, σ_e increased markedly, ranging between 70 m and 163 m for the European beech and
 5 between 42 m and 82 m for the Japanese beech. In Japanese beech, $\sigma_{e\text{-seedlings}} < \sigma_{e\text{-adults}}$
 6 whatever the assumed d_e/d_{obs} ratio (Table 3)

7 *Neighbourhood model and contemporary estimates*- Direct estimates of mating and dispersal
 8 parameters are summarised in Table 4. For these analyses, the marker sfc1063 was removed
 9 due to high frequency of null alleles (up to 12.5%). For the European beech, important seed
 10 migration ($21\% < m_s < 36\%$) and very high pollen migration ($63\% < m_p < 72\%$) from outside
 11 the studied site were detected. In the Japanese beech by contrast, seed immigration rate was
 12 not different from zero, and pollen immigration rate was estimated to be 40%.

13 The average distance of seed dispersal within the study site fell within the same order of
 14 magnitude across sites and species with $\delta_s = 10.5$ m (for European beech on average on FS1
 15 and FS3) and $\delta_s = 12.4$ m (for Japanese beech). Average distance of pollen dispersal were also
 16 not significantly different among species and sites with $\delta_p = 38$ m on average in European
 17 beech and $\delta_p = 79$ m in Japanese beech. Some of the estimated values for δ_s and δ_p suffered
 18 from low precision (large standard errors); this is an expected behaviour when the slope of the
 19 dispersal kernel (b_s and b_p) are well below 1 (Oddou-Muratorio et al., 2005). For both pollen
 20 and seed dispersal, we observed a general tendency of slightly fat-tailed dispersal kernels (i.e.
 21 $b \leq 1$). In FS3, the b -parameters could not be simultaneously estimated with the other
 22 parameters due to lack of convergence, but the best fits were obtained with an exponential
 23 kernel for seeds and pollen. The estimated seed dispersal kernel was exponential in FC, only
 24 slightly more fatter-tailed than the exponential in FS3 ($b_s = 0.72$), while it was strongly thin-

1 tailed in FS2 ($b_s = 0.15$). For the pollen dispersal kernel the estimated shape ranged from 0.30
2 (FC) to 1.03 (FS1).

3 The effect of diameter on female and male fertility was highly significant, with both male and
4 female fecundities increasing with diameter ($\gamma_s=0.86$ and $\gamma_p=0.62$). The selection gradient
5 tended to be steeper for female fecundity than for male fecundity. Selection gradients could
6 not be estimated for FS3 (no data) or FS2 (due to lack of convergence).

7 **Discussion**

8 This study provided historical and contemporary estimates of gene flow in European and
9 Japanese beech based on common, robust methods applied to four different parent-offspring
10 genotype data sets. In the following we first discuss methodology related to each estimate
11 (historical or contemporary) and investigate their variation in the observed range of species
12 and sites. Finally, we discuss the implication of our results for the understanding of gene
13 dispersal in beech, and the consequences for management of the genetic resources of this
14 species.

15 **Patterns of SGS and historical estimates of gene flow**

16 In both European and Japanese beech seedlings and adults showed a significant pattern of
17 SGS associated with historical dispersal rates ranging from ~47 m to ~105 m depending on
18 the assumed ratio of effective versus observed density (respectively $\frac{1}{2}$ and $\frac{1}{10}$). For long-
19 lived organisms such as trees it is notoriously difficult to estimate d_e/d_{obs} , or N_e/N (the ratio of
20 effective versus census population size), as these quantities depend on the variation in lifetime
21 reproductive success among individuals. Following Hardy et al. (2006) and assuming
22 $d_e=d_{obs}/4$ in our long-lived hermaphrodite tree species the estimated gene dispersal distances
23 were ~77 m in European beech and ~40 m in Japanese beech. Despite a slightly lower σ_e
24 estimate for the adults in FS1 ($\sigma_{eFS1}=44.35$ m, significantly inferior to estimates from FS2 and
25 FS3), gene dispersal distances were similar across sites and cohorts in European beech.

1 Moreover, our results overall reflect a trend towards restricted dispersal abilities in beech
2 compared to tropical (Hardy et al., 2006) or temperate, low-density (Oddou-Muratorio and
3 Klein, 2008) tree species, where σ_e ranged from 150 to 1200 m.

4 Interestingly, the comparison of SGS intensity between cohorts and between expanding and
5 ancient populations highlights the interaction between population history, effective population
6 density and dispersal in the development of SGS. First, the increasing SGS (and conversely
7 decreasing $d_e\sigma_e$ product) observed from adult to seedlings in Japanese beech contrasted with
8 the more similar SGS patterns observed among cohorts in European beech. The reduction in
9 the $d_e\sigma_e$ product from adults to seedlings observed in the Japanese beech may be due to
10 several causes : (1) a recent reduction in effective gene dispersal (σ_e), (2) a recent reduction in
11 effective population density (d_e) in Japanese beech, and/or (3) a reduction of a substantial part
12 of coancestry between nearest neighbours during stand maturation from seedlings to adults
13 (Epperson, 1992). The fact that all the analysed Japanese beech seedlings correspond to a
14 single year of reproduction (as compared to European beech seedlings, recruited across
15 several year of reproduction) support hypothesis (2) above: single-year d_e is indeed expected
16 to be lower (due to high inter-individual variance in reproductive success) as compared to
17 average d_e accumulated over several reproductive events (Krouchi et al., 2004).

18 Second, SGS was significantly weaker in ancient stands (FS3 and FC) compared to re-
19 colonising stands (FS1 and FS2). This variation was independent of observed adult density, as
20 FS3 and FS1 showed roughly the same density (Table 1). Marked patterns of SGS have also
21 been found in recently founded populations (Parker et al., 2001). In our case, it is likely that
22 the strongly spatially aggregated distribution of adults in FS1, compared to the more regular
23 distribution in FS3 (Figure 1), accounts for the higher intensity of SGS due to restricted
24 overlap of seed shadows in FS1. By contrast, the higher SGS in FS2 compared to FS1 can be

1 explained by the lower population density in FS2, which also reduces the seed shadow
2 overlap.

3 **Contemporary estimates of seed and pollen dispersal**

4 Direct estimates of seed and pollen dispersal curves using the spatially explicit mating model
5 also highlight restricted seed dispersal abilities in European and Japanese beech, despite non-
6 negligible long-distance dispersal events, and potentially important pollen-mediated gene
7 flow.

8 Considering first seed dispersal patterns within studied sites, seed dispersal distances were
9 consistent across sites and species, with a mean δ_s value of 10.5 m for European beech and
10 12.4 m for Japanese beech. Moreover the best fits were obtained either with exponential or
11 fat-tailed kernels. The different values obtained for seed dispersal parameters in site FS2 may
12 be partially explained by the age distribution of seedlings, because old seedlings (> 4 years
13 old) were over-represented in this site compared to the 3 other sites. Estimates of the seed
14 dispersal parameter for the sub-sample of young seedlings in FS2 fell within the same order
15 of magnitude as the 3 others ($\delta_s=7.68$ and $b_s=1$, Bontemps et al., unpublished). Thus, the seed
16 rain generated within a single or few reproductive events and observed at the early stage of
17 seedling recruitment reflects rather restricted seed dispersal abilities in beech compared to
18 other species where such estimates are available (e.g. $\delta_s \approx 135$ m in *S. torminalis*, Oddou-
19 Muratorio and Klein, 2008); $\delta_s = 277$ m in *Prunus mahaleb*, Robledo-Arnuncio and Garcia,
20 2007). These discrepancies are consistent with life history traits of the species and suggest
21 that seed dispersal in beech is primarily mediated by gravity and small rodents, as compared
22 to fleshy-fruit species where dispersal is assisted primarily by birds but also bigger mammals
23 (Jordano et al., 2007).

24 In contrast to seed dispersal within site, seed immigration rates in European beech ($m_s=27\%$
25 on average) indicate that “long-distance” seed dispersal events (i.e. from outside the study

1 site) are not negligible in the investigated sites. This was not the case for Japanese beech
2 where we found a null seed migration rate. High rates of seed immigration could be
3 associated with disturbances, as the three European beech sites all have a more recent history
4 of perturbation than the Japanese beech sites, with recent expansion in FS1 and FS2, but also
5 strong canopy opening in FS3 due to a storm occurred in 1999. However, the estimated values
6 for immigration rates should be considered with caution, because they are also probably
7 affected by different experimental biases: (1) mortality/logging of adults within the studied
8 site could result in false migration events, as supported by high m_s -values observed in the
9 saplings ($m_s= 48\%$) compared to young seedlings ($m_s=16\%$ for see Bontemps et al.,
10 unpublished); (2) genotyping error and null alleles could also result in false parentage
11 exclusion and overestimated migration rates.

12 Regarding pollen dispersal, the results for Japanese beech suggest greater abilities of pollen
13 versus seed dispersal, with both a higher mean dispersal distance ($\delta_p=79.04$ m versus $\delta_s=12.44$
14 m), and a fatter-tailed kernel ($b_p = 0.31$ versus $b_s = 1.06$). Pollen dispersal parameters (δ_p and
15 b_p) for European beech should be considered with caution as they were estimated based only
16 on seedling for whom at least one compatible parent pair was found within the study site. Few
17 seedlings met this requirement, as shown by the high m_s and m_p -values.

18 **Comparison of historical and contemporary estimates of gene flow**

19 We computed axial root mean squared σ_{p-rt} and σ_{s-rt} from our contemporary estimates of mean
20 distance of seed and pollen dispersal (δ_s and δ_p , Table 4) using equation (6). This provided
21 real-time, total gene flow estimates (σ_{rt}) ranging between 19 m and 142 m in European beech
22, and equal to 71 m in Japanese beech (Table 4). These are probably underestimates as they do
23 not account for immigration from outside the study site. Interestingly, contemporary (σ_{rt}) and
24 historical (σ_e) estimates of gene flow fell within the same order of magnitude (a few tens of
25 meters). In FS2 and FC σ_{rt} fell within the confidence interval of σ_e estimated when assuming

1 $d_e = d_{obs}/10$. In FS1 and FS3 contemporary estimates of gene flow tended to be lower than
2 historical estimates, suggesting a possible recent reduction in effective dispersal abilities.
3 However it is difficult to establish whether the observed discrepancy (1) is truly significant, (2)
4 is due to the assumptions and statistical specificities of the two approaches or (3) whether the
5 range of gene dispersal has actually changed. These results highlight the caution required
6 when drawing quantitative conclusions from the comparison of contemporary and historical
7 estimates because they both rely on different assumptions that affect the accuracy and
8 precision of estimates in an incomparable way. Contemporary estimates may suffer from low bias at
9 the scale of the study site, but they may underestimate long-distance dispersal (Oddou-
10 Muratorio and Klein, 2008). By contrast, historical estimates have minimum precision, and
11 suffer from high general uncertainty for the value of the effective density d_e value.

12 The comparison of historical versus contemporary estimates sheds some light on the different
13 components of gene flow contributing to the development of SGS. For instance, the
14 increasing SGS/ decreasing $d_e \sigma_e$ product observed from adults to seedlings in Japanese beech
15 is consistent with the restricted dispersal distances estimated using the contemporary
16 approach, and in particular with the low pollen and seed migration rates ($m_s \approx 0$, $m_p = 40\%$). By
17 contrast, in European beech, higher pollen and seed migration rates ($m_s \approx 27\%$, $m_p = 68\%$) may
18 explain why SGS does not increase in the seedling cohort despite restricted seed and pollen
19 dispersal distances within the study site. Thus, besides variation in the effective population
20 density discussed above, our results suggest that long-distance seed and/or pollen flow can
21 have a strong impact in the building of SGS within beech populations.

22 Finally, the roughly similar measures of gene flow obtained using historical or contemporary
23 estimates in this study and others (Oddou-Muratorio and Klein, 2008) may provide guidance
24 for researchers interested in measuring gene flow to select one of the two approaches,
25 depending on the resources or constraints of the study. Historical approaches are clearly less

1 costly in term of sampling effort since they do not require exhaustive sampling and mapping.
2 They can thus be useful in providing some preliminary estimates of total gene flow,
3 particularly if the sampling design includes different cohorts among which estimates can be
4 compared. By contrast, contemporary estimates are clearly more costly to obtain since they
5 require exhaustive sampling of the parental population within a given area, as well as, ideally,
6 measures of covariates of fecundity for these trees. Conversely, they yield a much more
7 detailed description of gene flow, and in particular they allow the relative importance of seed
8 versus pollen mediated dispersal and of long versus short distance dispersal to be gauged. A
9 main conclusion of this study is that the combination of both historical and contemporary
10 approaches may be the only solution when the objective is to detect recent changes in
11 effective dispersal ability or to gauge the influence of the different components of gene flow
12 in the building of SGS.

13 **Perspectives on beech's response to climate change and management of genetic** 14 **resources**

15 Estimates of seed dispersal in European and Japanese beech (a few tens of meters/ generation)
16 obtained in this study are roughly one order of magnitude below what would be needed for
17 beech to track 21st-century global warming (hundreds of meters for the next generation). This
18 substantial discrepancy also supports recent re-interpretation of post-glacial expansion rates
19 during the early Holocene in the light of chloroplast DNA surveys (McLachlan et al., 2005;
20 Magri 2008). Rather than rapid range expansions from well-established refuges at rates of
21 100-1000 m/yr, molecular studies suggested that for different tree species including beech
22 post-glacial recolonisation relied on a mosaic of sparse, small populations scattered in
23 multiple regions closer to modern range limit than previously thought. These authors
24 proposed that average post-glacial migration rates of beech may have been slower than those
25 inferred from fossil pollen, and closer to those estimated in this study. However, one can not

1 rule out that rare events of long-distance seed dispersal still contributed to post-glacial
2 recolonisation, and may contribute to future response to climate change. The importance of
3 these long-distance dispersal events is notoriously complicated to quantify, and requires to be
4 addressed at spatial and temporal scales larger than in the present study.

5 Assuming that gene migration may favour a rapid adaptive response to the ongoing climate
6 change by mixing or adding well adapted genetic material to poorly-adapted gene pools, this
7 study suggests several ways in which forest management may enhance the beneficial effect of
8 migration. Stand regeneration is clearly a crucial phase: to increase genetic diversity at local
9 scale, management should (1) favour a regular spatial distribution of selected seed-trees (i.e.
10 avoid clumping), (2) spread the recruitment process across several different years, even in
11 masting species, (3) favour regeneration plots both close to and far from the seed trees. Note
12 however that the final outcome of the interaction between gene flow and selection in the
13 response of Beech population to global change can not be addressed without additional
14 studies, comparing core and marginal populations, either at the front or at the rear edge of
15 distribution range, and evaluating simultaneously the potential for genetic adaptation and gene
16 flow.

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16

1 Figures legends

2

3 **Figure 1:** Sampling design within each studied site (FS1, FS2 and FS3 for *European beech*

4 and FC for *F. crenata*) . Adult trees (▲) were exhaustively mapped within the studied site.

5 Saplings (+, age ≥ 3) and/or seedlings (*, age < 3) were sampled preferentially in the central

6 part of studied site.

7 **Figure 2:** Correlograms of kinship coefficient for A) adult trees and B) seedlings in the four

8 sites. Abscise values correspond to the upper limit of the distance intervals. Filled symbols

9 represent values significantly different from the expected value under a random distribution of

10 genotypes (95% confidence level). Confidence intervals around each F_{ij} -value were obtained

11 through a jackknife procedure over loci.

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1 **Table 1: Information on study sites for each species**

Species	Site	Coordinates (UTM)		Number of genotyped adults	Observed adult density (tree.ha ⁻¹)	Number of genotyped seedlings
<i>Fagus sylvatica</i>	FS1 (Mont Ventoux)	44°10'45"N	5°14'01"E	96	50	256
<i>Fagus sylvatica</i>	FS2 (Mont Ventoux)	44°11'03"N	5°16'57"E	50	19	356
<i>Fagus sylvatica</i>	FS3 (Haye Forest)	48°45'30"N	6°20'29"E	350	44	221
<i>Fagus crenata</i>	FC (Kurokima)	38°55'30"N	140°47'50"E	287	379	1387

2

1

2 **Table 2:** Number of alleles (Na) and null allele frequencies (Fnull) per species, site and
 3 microsatellite locus. Detailed information on microsatellite loci can be found in: ¹ Pastorelli et
 4 al., (2003) ; ² Tanaka et al. (1999); ³ Asuka et al., (2004) and ⁴Vornam et al. (2004)

Species	<i>Fagus sylvatica</i>						<i>Fagus crenata</i>	
	FS1		FS2		FS3		FC	
Locus	Na	Fnull	Na	Fnull	Na	Fnull	Na	Fnull
FS1-03a ¹							17	0.000
FS1_15 ¹	13	0.037	14	0.006				
FS4-46 ¹					24	0.094	14	0.000
FS3-04 ¹	3	0.000	4	0.000	5	0.000		
mfc5 ²					24	0.084		
mfc7 ²	7	0.057	7	0.058				
sfc0007 ³	7	0.000	8	0.019				
sfc1105 ³	14	0.000	9	0.031			33	0.002
sfc1143 ³	12	0.012	11	0.000			23	0.000
sfc0036 ³	10	0.000	11	0.000			24	0.000
sfc0018 ³	9	0.006	11	0.013			19	0.002
sfc0378 ³							27	0.010
sfc0161 ³	14	0.003	14	0.000	18	0.000		
sfc1063 ³	10	0.125	8	0.107				
mfs11 ⁴	8	0.103	9	0.045				

5

1 **Table 3:** Estimates of SGS and gene dispersal parameters for each species by site and cohort
 2 (A = Adult, S=Seedlings): average kinship coefficient between individuals separated by less
 3 than 10 m (F1), SGS intensity (Sp) and its standard error (SE), neighbourhood size, (Nb),
 4 gene dispersal distance (σ_e) and 95% confidence interval (CI) for three effective densities
 5 (De) estimated from the density of adults (D).
 6

Species	Site	Stage	SGS parameter		Gene dispersal parameters			
			F1	Sp (SE)	Nb	σ_e (CI) De=D/2	σ_e (CI) De=D/4	σ_e (CI) De=D/10
<i>Fagus sylvatica</i>	FS1	A	0.0821	0.0218 (0.0046)	45.80	31.36 (26.2-41.5)	44.35 (37.1-58.7)	70.12 (58.7-92.8)
		S	0.0328	0.0163 (0.0018)	61.43	40.62 (36.3-47)	57.44 (51.3-66.4)	90.82 (81.2-105)
	FS2	A	0.0785	0.0354 (0.0076)	28.22	64.69 (54.3-84.7)	91.48 (76.8-119.8)	144.64 (121.5-189.4)
		S	0.051	0.0205 (0.0026)	48.76	72.98 (66-82.9)	103.21 (93.3-117.2)	163.19 (147.5-185.3)
	FS3	A	0.0577	0.0087 (0.0012)	114.33	65.97 (58.4-77.4)	93.30 (82.6-109.5)	147.52 (130.7-173.2)
<i>Fagus crenata</i>	FC	A	0.0343	0.0032 (0.0007)	312.15	36.84 (30.6-49.8)	52.10 (43.2-70.4)	82.38 (68.4-111.3)
		S	0.0255	0.0124 (0.0012)	80.67	18.64 (17.1-20.7)	26.37 (24.2-29.3)	41.69 (38.2-46.4)

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1 **Table 4:** Selfing rate (s), seed and pollen migration rate (m_s and m_p), average distance of seed
 2 and pollen dispersal (δ_s and δ_p), shape parameter of exponential-power kernel for seeds
 3 (b_s) and pollen (b_p) and selection gradients (γ_s and γ_p) estimated by the direct approach
 4 in the different species and ISP. See text for details on real-time, total gene flow
 5 estimates (σ_{rt}).

Species	<i>Fagus sylvatica</i>						<i>Fagus crenata</i>	
	FS1		FS2		FS3		FC	
Site	Estimate	SE	Estimate	SE	Estimate	SE	Estimate	SE
m_s	0.23	0.03	0.36	0.04	0.21	0.04	0.007	0.0035
m_p	0.63	0.04	0.68	0.04	0.72	0.06	0.40	0.019
s	0.01	0.01	0.04	0.02	0.03	0.02	0.00*	NE
δ_s (m)	10.42	1.39	62.42	73.68	10.89	0.90	12.14	0.36
δ_p (m)	28.12	4.39	30.52	18.56	55.14	17.99	79.04	30.88
b_s	0.72	0.15	0.15	0.10	1.00*	NE	1.06	0.07
b_p	1.03	0.33	0.68	0.23	1.00*	NE	0.30	0.07
γ_s	0.86	0.08					0.85	0.06
γ_p	0.60*	NE					0.65	0.07
σ_{rt} (m)	19.66		142.61		35.06		71.21	

6 * fixed value for the parameter, NE= corresponding standard error not estimated

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