Chapitre 10. Conclusions générales

Deux définitions de l'eutrophisation ont été proposées à partir de l'analyse de la littérature : l'une traitant de l'eutrophisation dite « naturelle » ou géologique, et l'autre de l'eutrophisation anthropique. Il est apparu important de les dissocier car la description de l'eutrophisation naturelle ne doit pas occulter la menace que font peser les activités humaines sur les cycles du phosphore et de l'azote, au même titre que sur celui du carbone. Aujourd'hui, c'est bien le développement accéléré de l'eutrophisation d'origine anthropique qui concentre l'essentiel des inquiétudes sociétales et des préoccupations de gestion. C'est donc de l'eutrophisation d'origine anthropique qu'il s'agit dans cette expertise scientifique collective.

10.1. Les constats

Les écosystèmes aquatiques sont des systèmes complexes, dont le fonctionnement est régi par des équilibres dynamiques. L'eutrophisation est un déséquilibre de fonctionnement, déclenché par un changement dans les quantités, les proportions relatives ou les formes d'azote et de phosphore entrant dans les systèmes. La nature et l'intensité des réponses dépendent également de facteurs environnementaux tels que la lumière, le temps de résidence de l'eau et la température.

Le mécanisme général de réponse des écosystèmes à ces changements d'apports de nutriments est commun aux écosystèmes d'eaux continentales et marines : une augmentation de l'azote et du phosphore entraîne une augmentation de la biomasse végétale, générant progressivement une diminution de la pénétration de la lumière dans la colonne d'eau. Les écosystèmes aquatiques passent ainsi d'un système avec des apports limités de nutriments à un système progressivement saturé en nutriments, dans lequel le nouveau facteur limitant devient la lumière. Ce mécanisme induit une cascade de réactions en chaîne, avec notamment une modification de la structure des communautés biologiques et des réseaux trophiques, ainsi que des changements dans les cycles biogéochimiques. Ces changements peuvent s'opérer de manière progressive, proportionnellement aux forçages, ou au contraire de manière brutale. Les effets les plus notables de l'eutrophisation sont les proliférations végétales, parfois toxiques, la perte de biodiversité et les anoxies qui peuvent se traduire par la mort massive d'organismes aquatiques.

Chaque écosystème est unique et possède son histoire et sa dynamique propre, elle-même liée aux conditions géologiques, géomorphologiques, hydrologiques, écologiques et climatiques locales, mais aussi aux pressions anthropiques passées et présentes et à leur nature, ainsi qu'aux contextes sociologiques et économiques dans lesquelles elles se sont inscrites. Ainsi, si les mécanismes sont génériques, les trajectoires de l'eutrophisation sont diverses, en lien avec la diversité des situations locales, avec des effets de seuils très dépendants des contextes passés et présents. Des conditions aggravantes existent : des liens sont ainsi suspectés entre eutrophisation, toxicité et invasions biologiques. Les apports d'éléments toxiques modifient les chaînes trophiques, ceci pouvant créer des conditions plus favorables à l'eutrophisation. Les modifications des chaînes trophiques peuvent offrir des conditions favorables aux invasions biologiques. La vulnérabilité à l'eutrophisation varie avec ces propriétés intrinsèques de l'écosystème aquatique récepteur. La vulnérabilité est donc à définir en prenant en compte toute la chaîne de causalités directes et indirectes qui les relie.

Il résulte de cette complexité qu'il est très difficile de prédire l'évolution écologique et biogéochimique des écosystèmes aquatiques. Il est ainsi très difficile d'extrapoler avec précision des résultats obtenus sur un type d'écosystème à d'autres. Par ailleurs, les changements climatiques vont probablement aussi jouer un rôle favorisant l'eutrophisation. Si le rôle du climat sur les flux de nutriments semble encore en débat dans la littérature scientifique, car il dépend de l'interaction avec les usages des sols et les activités humaines, l'effet sur les communautés biologiques fait l'objet de plus de consensus : l'élévation de la température est clairement un facteur aggravant le développement de la biomasse végétale. L'allongement du temps de stratification des milieux lents, la consommation plus précoce des nutriments par le phytoplancton et la modification induite des équilibres stœchiométriques favoriseraient en outre la fréquence de développement des proliférations algales nocives, en particulier à cyanobactéries.

Les usages du sol et les activités humaines (urbanisation, industrialisation, agriculture) de ces cent dernières années ont radicalement amplifié les pressions et transformé les paysages. Ils ont impacté la qualité des sols, des eaux de surface et des eaux souterraines. La majorité des nappes phréatiques est polluée par les nitrates tandis que les sols et les sédiments sont souvent enrichis en phosphore. Le temps de transfert de l'azote depuis les bassins-versants vers les écosystèmes aquatiques se chiffre en dizaines d'années et la biodisponibilité en phosphore des sols et des sédiments s'est accrue. Cela explique en partie la faible diminution des flux d'azote, et, dans une moindre mesure, de phosphore aux exutoires des bassins-versants malgré les efforts engagés de diminution des intrants. Les trajectoires de restauration doivent donc se construire en fonction de leurs contextes locaux.

Un consensus croissant existe sur la nécessaire réduction conjointe des apports en azote et en phosphore, même si certains écosystèmes sont plus sensibles à l'azote ou au phosphore. En effet, tous les systèmes (traitement des eaux, des déchets, systèmes agricoles), les données biologiques, et les approches économiques soulignent l'importance de considérer ensemble N et P dans la remédiation et la prévention. Les stratégies de limitation des flux d'azote et de phosphore passent par différents niveaux d'action selon la vulnérabilité des écosystèmes aquatiques :
(i) Une utilisation raisonnée des intrants, fonction des milieux (sol, nappe) ; des gains d'efficacité de l'utilisation de l'azote et du phosphore semblent en effet encore possibles. Ces gains sont à rechercher tant dans le traitement des déchets et des eaux usées que dans la gestion agricole des fertilisants (composition, outils de pilotage et de contrôle dans les sols, la plante, etc.) et de l'alimentation animale (composition adaptée, etc.). Des marges de manœuvre existent à ce niveau et doivent être mises en œuvre. Cependant, dans un certain nombre de cas, elles ne suffiront pas, et d'autres leviers, touchant à la conception même des systèmes (par exemple un changement de type de production), devront être envisagés.
(ii) Une couverture végétale des sols la plus continue possible, dans l'espace et le temps, qu'il s'agisse d'interculture, de cultures associées, de prairies, d'agroforesterie, etc. qui permette l'assimilation, en continu, des nutriments par la biomasse végétale terrestre. Cette biomasse contribuera à la séquestration du carbone dans les sols, renforçant par là même celle de l'azote et du phosphore. Les densités animales doivent être revues et limitées à la capacité des sols à recevoir des effluents (directement ou après traitement). Un maintien des conditions aérées des sols, des nappes, des sédiments favorisera l'adsorption du phosphore.
(iii) Le maintien ou la restauration de la diversité des paysages (haies, zones humides, ripisylves, etc.) qui limitent les fuites des nutriments vers les écosystèmes aquatiques par différents processus (adsorption, dénitrification, etc.). À l'inverse, il faut éviter toute transformation augmentant la vitesse d'écoulement de l'eau, comme le drainage, la rectification, l'endiguement et le dragage des cours d'eau, l'imperméabilisation des sols (urbanisation, dégradation et tassement des sols).

L'action publique en matière de gestion de la qualité de l'eau s'est organisée en trois périodes : celle de l'assainissement, en grande partie réalisée dans les pays industrialisés, mais qui demeure une urgence dans les pays à fort développement ; celle, toujours actuelle, du traitement des pollutions industrielles et domestiques ; et celle désormais prioritaire du traitement des pollutions diffuses agricoles. Dans les pays industrialisés, les évolutions observées en eau douce ont été plutôt positives depuis quelques décennies, plus fortement sur le phosphore que sur l'azote, tandis qu'en milieu marin les phénomènes d'eutrophisation semblent peu diminuer depuis le début du XXI^{e} siècle.

L'eutrophisation est un processus encadré par plusieurs textes réglementaires aux logiques différentes. Ainsi coexistent des directives « usages » morcelées, datant des années 1980, ciblées sur l'encadrement d'un domaine, avec des directives à ambition plus globale comme la DCE ou la DCSMM des années 2000 : la Directive Nitrates est axée sur l'origine agricole des nitrates, avec un seuil de 50 mg/L défini par rapport à la norme de potabilité ; la sensibilité des écosystèmes doit être explicitement prise en compte dans la caractérisation des masses d'eau pour la définition des zones vulnérables, sachant que le processus d'eutrophisation en milieu aquatique continental dépend également fortement du phosphore assimilable et des rapports relatifs entre nutriments ; la DERU encadre la collecte, le traitement et les rejets d'eaux résiduaires, avec des normes d'émission ponctuelle, mais pas de norme pour le milieu récepteur ; la DCE et la DCSMM exigent, quant à elles, la mise en place des mesures nécessaires pour maintenir ou atteindre l'objectif de bon état écologique des masses d'eau, *via* notamment un recensement régulier de l'état de santé général intégré des hydrosystèmes. À l'exception de la DCSMM, les directives ne donnent pas de préconisation précise concernant l'eutrophisation, considérée dans un ensemble de pressions potentiellement dégradantes. À chacun de ces textes répondent des dispositifs de suivi ciblés, qui sont essentiellement utilisés pour vérifier la conformité aux normes de leur domaine et qui sont insuffisamment mis en relation.

Dans le domaine de la caractérisation de l'eutrophisation, deux considérations fortes ressortent : d'une part, des dispositifs mis en place antérieurement à la DCE étaient plus ciblés pour encadrer le processus d'eutrophisation (par exemple diagnose rapide des plans d'eau ; suivi eutrophisation de l'Agence de l'eau Rhône-Méditerranée-Corse) : certaines variables, la fréquence de leurs relevés ou leur spatialisation étaient riches d'enseignement (cycles de 24 heures pour l'oxygène et le pH, à différentes hauteurs de la colonne d'eau pour les milieux profonds, successions saisonnières des communautés de phytoplancton ou de macrophytes, etc.). Cette assertion doit être nuancée pour le milieu marin, souvent régi par des conventions internationales et pour lequel des protocoles dédiés à l'eutrophisation ont été conservés. D'autre part, les indices biologiques de chaque élément de qualité de la DCE rendent compte d'un état général des masses d'eau sous l'effet de multiples pressions ; c'est cette propriété qui a été optimisée lors de leurs récentes mises en conformité (indices multimétriques répondant à la variété des dégradations possibles). L'examen conjoint des différents indices renforce l'ambition portée par la DCE de réaliser une évaluation plus holistique et lisible du grand public et des gestionnaires ; en revanche, cela rend plus difficile l'extraction d'un signal individualisé, comme l'eutrophisation et plus encore l'information relative à certains symptômes susceptibles de survenir (proliférations algales ou anoxies temporaires).

La quasi-absence de modèles bioéconomiques rend difficile l'accompagnement de la remédiation. Ces modélisations seraient à construire très tôt dans le processus de restauration pour intégrer dès le départ des aspects biophysiques, écologiques et économiques. Les études existantes indiquent que les coûts associés aux impacts qui ont pu être évalués sont élevés, renforçant l'importance de la remédiation et de la prévention. Les études concernant les plans de remédiation indiquent qu'il n'existe pas de solution idéale, mais seulement des politiques ciblées, conçues pour des situations particulières, avec des instruments développés de manière souvent *ad hoc*, une fois les problèmes correctement identifiés, analysés et les différentes solutions possibles évaluées.

De même, la sociologie en environnement est actuellement peu développée en France. Des pans entiers ne sont pas étudiés, tels que les mobilisations environnementales et les représentations des milieux. Le cas des marées vertes fait exception. Les transformations des territoires ne sont plus uniquement perçues comme d'ordre biophysique. Les dimensions sociologiques commencent à être prises en compte, appelant à une gestion différenciée selon les socioécosystèmes et leurs différentes échelles spatiales, mettant en jeu des acteurs différents autour de l'eutrophisation.

10.2. Les verrous et les pistes d'investigations futures

10.2.1. Vers des approches de recherche systémiques

Les travaux de recherche très intégrateurs, à l'échelle territoriale, sont encore peu présents pour répondre à des enjeux de gestion différenciée des têtes de bassin-versant, du corridor fluvial, des espaces côtiers. La remédiation de l'eutrophisation doit donc aller vers des approches systémiques intégrant les hydrosystèmes, les espaces urbains et agricoles, les modes de production, d'alimentation et de recyclage. La question des transitions agricoles est de manière générale étroitement liée à celle de l'eutrophisation. Les modèles mêlant simultanément les aspects biophysiques et économiques devraient être développés pour servir de base de discussion pour la définition de programmes de remédiation, malgré les incertitudes liées à une connaissance imparfaite mais en constante progression des phénomènes biophysiques, écologiques et socio-économiques. L'évolution de l'eutrophisation doit également être mieux mise en regard avec les évolutions des socioécosystèmes, dépassant des focalisations sectorielles comme celle sur l'agriculture de ces dernières décennies, le partage de savoir pouvant recréer du lien entre des groupes sociaux et des secteurs d'activité s'inscrivant dans des mondes sociaux aujourd'hui disjoints.
Des sites d'investigations interdisciplinaires, où s'étudieraient sur le long terme les dynamiques biophysiques et sociétales doivent donc être plus nombreux et plus divers (lacs, rivières, littoral), et ceux existants pérennisés. Le renforcement des coopérations entre sciences biophysiques et sciences humaines et sociales doit se faire en prêtant une attention particulière à la bonne intégration des disciplines travaillant à partir de données quantitatives et qualitatives.

10.2.2. Utilisation des données existantes et pistes d'amélioration des cadres et réseaux de suivi

Les résultats d'expériences françaises de remédiation, pourtant nombreuses et anciennes, restent souvent enfouis dans une littérature grise difficile d'accès et de qualité inégale. Elles risquent de sortir de la mémoire collective si on n'y prête pas intérêt rapidement. Un enjeu est d'analyser ces expériences et d'en publier des synthèses dans des revues internationales. Il faut aussi encourager la numérisation des données anciennes, leur bancarisation et leur mise à disposition, compte tenu de l'importance des séries de longue durée. L'analyse des trajectoires des problèmes publics, selon la diversité des systèmes, les gouvernances mises en place, notamment à l'échelle européenne et internationales font partie intégrante des analyses de ces expériences. Enfin, il faut tirer parti des séries ayant pu être constituées et les interpréter (par exemple les données de proliférations de cyanobactéries mesurées à l'échelle nationale au sein des réseaux santé, etc.).

La réflexion sur le champ réglementaire, ses limites d'efficacité et d'applicabilité et sa possible harmonisation entre milieux, mais aussi entre pays voisins, constitue un axe de recherche pour des investigations futures : comment mieux tenir compte du continuum terre-mer dans lequel le processus d'eutrophisation se manifeste à différents degrés ? Mais aussi comment mieux tirer parti de l'information recueillie au sein des réseaux de surveillance, tout en distinguant bien quatre fonctions distinctes pour l'utilisation des données : (1) conformité aux normes en vigueur, (2) états statistiques réguliers des milieux et de leurs évolutions spatiales et temporelles, (3) approfondissement des connaissances pour un thème donné et (4) suivi de l'efficacité des actions de remédiation. Il est possible que certaines confusions et incompréhensions perdurent quand les trois premières fonctions sont assignées aux mêmes jeux de données. La qualité des données mobilisées (nature, précision, représentativité spatiale et temporelle), tant pour les variables d'état chimique que pour les variables biologiques est essentielle : données de pressions (passées et actuelles), données chimiques et leur contextualisation géographique. Ces réflexions permettront de dépasser les approches actuelles définissant des seuils d'eutrophisation sur une base essentiellement physicochimique. S'agissant de la norme de 50 mg/L de nitrates, elle est clairement relative à la potabilité de l'eau et non à la prévention des milieux vis-à-vis du processus d'eutrophisation. Avant la mise en œuvre de la DCE, une grille de consensus avait été proposée avec différentes valeurs guides dans la gamme de 2 à 50 mg/L de nitrates. Il serait intéressant d'analyser la trajectoire historique de ces valeurs et leur déclinaison territoriale. Dans tous les cas, la transparence sur les critères de jugement et la pédagogie qui y est associée sont primordiales pour établir des gammes de valeurs seuils. Des conseils scientifiques, des interfaces sciences-société doivent être mis en place de manière à créer des espaces de discussion dans la mise en place de ces démarches de diagnostic et de remédiation. Si des démarches normatives peuvent encadrer les problèmes d'eutrophisation, leur adaptabilité aux caractéristiques des bassins-versants doit aussi être pensée dans une démarche de progression des connaissances et des diagnostics et de gestion adaptative. Les sciences citoyennes sont encore peu visibles dans le domaine de l'eau en France à l'heure actuelle. Ces initiatives pourraient représenter un apport supplémentaire de connaissances, notamment pour des crises fugaces qui échappent à la surveillance actuelle des réseaux, mais pourraient aussi renfermer des vertus pédagogiques, ouvrant et dynamisant le dialogue entre spécialistes et observateurs potentiels.

L'analyse des impacts sociaux, sanitaires et culturels de l'eutrophisation appelle la réalisation d'études de cas et la production de connaissances très contextualisées, aujourd'hui presque inexistantes. Du point de vue des sciences sociales, il s'agit de prendre en compte les mobilisations sociales, les conflits et les formes de problématisation multiples qui accompagnent l'émergence de l'eutrophisation comme problème public, au travers d'une part de ses conséquences les plus dommageables, d'autre part des transformations profondes des représentations sociales associées aux écosystèmes qui caractérisent nos sociétés contemporaines. Les travaux existants montrent en effet que le spectre des instruments d'action possibles et l'efficacité des politiques en dépendent.
L'évaluation monétaire des impacts de l'eutrophisation reste encore une vraie question de recherche. Peu de références sont disponibles, et les impacts à long terme sont difficiles à évaluer.

10.2.3. Vers une méthodologie d'analyse du risque d'eutrophisation

Concernant les dispositifs d'étude et de suivi, une réflexion devrait s'engager pour dégager les espaces à risques de demain et définir quels devraient être leurs accompagnements opérationnels et scientifiques, complémentaires des dispositifs existants. Cela devient de plus en plus nécessaire compte tenu d'une possible exacerbation des phénomènes d'eutrophisation dans le futur sous l'effet conjugué des changements globaux. L'analyse de la littérature préconise qu'une méthodologie d'analyse de risque devrait combiner 1) les transferts et les transformations hydrobiogéochimiques le long du continuum terre-mer, 2) les facteurs de vulnérabilité des écosystèmes, ainsi que 3) les aléas climatiques.
La prise en compte de la variabilité des apports, des temps de résidence et des transferts de nutriments dans les têtes de bassins-versants et plus généralement le long du continuum terre-mer est un enjeu à la fois théorique et appliqué. Cela nécessite de renforcer la surveillance des sols et des nappes phréatiques, de développer pour cela des indicateurs de la mobilité du phosphore et des transferts d'azote dans les têtes de bassins-versants. Le suivi des évolutions des sols et des eaux, au plus près des actions, est une nécessité pour s'assurer de leur efficacité. Le renforcement des méthodologies de surveillance et d'estimation des flux, en analysant l'effet des fréquences, des périodes et zones optimales de mesures, des incertitudes, en combinant des acquisitions par des capteurs haute fréquence, constitue également un enjeu. Les futures technologies *in situ* ou aériennes ouvrent des champs d'investigation par rapport aux situations actuelles, soit à la haute fréquence temporelle, soit à une plus large emprise spatiale, qu'il s'agisse des sols ou des eaux. Il s'agira de les intégrer à de nouvelles stratégies de suivi, et de revoir la stratégie d'échantillonnage des eaux. Par sa prise en compte de plus en plus réaliste des facteurs physiques, chimiques et biologiques et l'enchaînement possible d'amont en aval de modèles de l'eutrophisation de divers milieux aquatiques allant des têtes de bassins jusqu'à l'océan, la modélisation déterministe restera un outil clé pour l'exploration de scénarios d'apports nutritifs et climatiques, et devra assortir ses prévisions d'intervalles de confiance.

La vulnérabilité est à définir en prenant en compte toute la chaîne de causalités directes et indirectes qui influencent les propriétés intrinsèques de l'écosystème aquatique récepteur, en lien avec la diversité des situations locales, des contextes passés et présents. Les voies de progrès pour caractériser cette vulnérabilité portent sur plusieurs points. D'une part, à partie de l'extraction et de la calibration de l'information portée par les indicateurs biologiques : (1) mieux comprendre comment certaines métriques déjà utilisées dans les indicateurs biologiques ou des nouvelles métriques, signent des basculements significatifs vers des situations eutrophes, (2) repérer ces basculements dans les trajectoires de situations suivies de façon approfondie dans le temps et les interpréter en termes fonctionnels (flux et concentrations de nutriments versus relations entre les groupes biologiques). En parallèle, certains paradigmes doivent être revisités, notamment la relation entre nutriments C et Si et la production de biomasse végétale, qui semble plus complexe quand on cherche à mieux comprendre le déterminisme des proliférations algales, les changements de communautés et les ajustements trophiques. Pour ce faire, des travaux couplés en hydrologie, en géochimie, en physiologie et en écologie seront nécessaires.

Les changements globaux impliquent des changements dans les dynamiques hydrologiques, sédimentaires et thermiques ainsi que des changements d'usage des sols et des activités humaines. La prise en compte de l'aléa climatique est donc essentielle dans une méthodologie de caractérisation du risque d'eutrophisation. Déterminer les rôles respectifs du climat et des activités humaines est une question de recherche centrale que la modélisation peut contribuer à instruire, en complément et à partie de l'observation long terme. Les interactions entre le climat et les processus écophysiologiques, biogéochimiques (biotransformations dans les sols, les zones humides de transition et les milieux aquatiques), hydrologiques (connectivité, distribution des temps de résidence et transferts) et écologiques (réseaux trophiques) devront être éclairées. L'augmentation des concentrations en CO_2 atmosphérique devra être prise en compte, celle-ci pouvant par exemple favoriser la productivité des écosystèmes terrestres et aquatiques, et dans le même temps, intensifier les proliférations de cyanobactéries. Ces éléments créent une grande incertitude, dans les pays de l'hémisphère nord en particulier, sur les évolutions futures des écosystèmes aquatiques. Le rôle des espèces invasives vis-à-vis de l'eutrophisation en eau douce devrait être intégré aux investigations ; à titre d'exemple, la Loire, mais aussi d'autres cours d'eau de l'ouest de la France, sujets à ces invasions, sont des sites d'études à privilégier, dès lors qu'ils bénéficient de séries suffisamment riches (flux de nutriments, autres compartiments biotiques) pour y réaliser une étude systémique fonctionnelle.

Annexe : Cahier des charges de l'ESCo eutrophisation

I - Contexte et enjeu

Les phénomènes d'eutrophisation et le rôle qu'y jouent notamment les apports de nutriments et leur interaction avec d'autres facteurs sont actuellement peu ou mal perçus, leurs connaissances ne sont ni consolidées et encore moins partagées. C'est pourquoi, les ministères en charge de l'écologie et de l'agriculture, ont demandé une expertise scientifique collective (ESCo)[1].

Le besoin de réaliser cette expertise scientifique collective sur l'eutrophisation se situe dans un contexte réglementaire large comprenant la ***Directive cadre sur l'eau (DCE)*** (directive 2000/60/CE), la ***Directive cadre stratégie pour le milieu marin (DCSMM)*** (directive 2008/56/CE), qui intègrent la notion d'eutrophisation dans les critères de bon état écologique respectivement des masses d'eau continentales, de transition et marines (descripteur 5 pour l'eutrophisation), la ***Directive nitrates*** (directive 91/676/CEE) dans les critères de classement en zones vulnérables, ainsi que la ***Directive pour les Eaux résiduaires urbaines (DERU)*** (directive 91/271/CEE) pour la délimitation des zones sensibles.

Compte-tenu de ces enjeux et en raison des délais impartis liés aux calendriers de mise en œuvre des directives, l'expertise scientifique collective apportera des informations validées, permettant d'éclairer ces politiques publiques.

II - Objet et champ de l'Expertise Scientifique collective

L'eutrophisation a pour particularité qu'il s'agit d'une notion utilisée à la fois par la communauté scientifique et par les politiques publiques, dont les définitions sont multiples. L'expertise clarifiera la définition de l'eutrophisation en prenant en compte les besoins et enjeux opérationnels pour l'action publique. L'analyse produira un état des lieux critique des connaissances scientifiques certifiées au plan européen et mondial sur les causes et mécanismes d'eutrophisation des eaux et identifiera les verrous scientifiques nécessitant l'acquisition de nouvelles connaissances qui pourraient faire l'objet de recherches supplémentaires.

Elle prendra en compte le continuum terrestre-aquatique, c'est à dire le système de transfert des bassins versants aux écosystèmes aquatiques inclus dans la notion de bassin versant dans la mesure où ils permettent de caractériser le risque d'eutrophisation. Sont exclus du champ de l'ESCo l'analyse détaillée de l'impact des activités humaines (systèmes agricoles, modalités de traitements de l'eau, …) sur l'eutrophisation.

L'expertise servira de base scientifique en vue d'améliorer la cohérence des modalités de mise en œuvre des différentes directives concernées (Directive nitrates, DCE, DCSMM et DERU).

III – Questions et thématiques scientifiques

1-Etat des lieux

Sur la base des publications scientifiques, il sera dressé un état des lieux dans le monde des formes et manifestations d'eutrophisation, des enjeux associés pour se recentrer sur le cas de la France (métropolitaine et DOM). Ce constat permettra de poser le contexte et de relier l'analyse théorique avec les enjeux identifiés en France. Sur la base de l'analyse de la littérature scientifique, l'ESCo dressera, sur la base des connaissances disponibles :

a) Un état des lieux des formes d'eutrophisation
- Les formes de manifestations de l'eutrophisation

[1] Cette expertise a notamment été annoncée lors d'une réunion avec les représentants de la profession agricole le 24 juillet 2014 dans le cadre de la directive nitrate.

- Le recensement et la description des impacts environnementaux, économiques et sociaux
    - Impacts environnementaux (approches ACV et multicritères)
    - Impacts économiques et sociaux des phénomènes d'eutrophisation et politiques de lutte contre l'eutrophisation (coûts des dommages, analyse coûts-bénéfices, ...)

b) Un état des lieux en France

- Les formes d'eutrophisation observées en France : les manifestations de l'eutrophisation (lieux et périodes, conditions, …) et les enjeux associés (toxicité, déséquilibres trophiques, …)
- Les modalités de surveillance de l'eutrophisation en France : état des programmes de surveillances, données disponibles, degré de connaissance concernant l'état d'eutrophisation des eaux et critères utilisés.
    - Données mobilisables (nationales, régionales et locales), sur la qualité de l'eau, les suivis biologiques, la structure des réseaux hydrographiques, l'occupation du sol…, leur pertinence et leur qualité. Analyse des lacunes et difficultés, y compris spatiales et temporelles.

2- Définition de l'eutrophisation, causes et symptômes

L'expertise fera une analyse des mécanismes biotiques et abiotiques responsables de l'eutrophisation et identifiera les facteurs qui lui sont liés. L'expertise donnera une définition scientifique de l'eutrophisation en soulignant le caractère complexe et multiforme de ce processus et de ses manifestations dans les grands types d'écosystèmes aquatiques continentaux et marins. Cette analyse sera mise en parallèle d'une analyse juridique sur la définition de l'eutrophisation dans les directives. Cette analyse devra permettre aux pouvoirs publics de proposer une définition de l'eutrophisation pertinente d'un point de vue scientifique, adaptée aux enjeux identifiés et permettant de répondre aux exigences réglementaires liées à l'application des directives européennes.

a. Définition scientifique

- Définitions scientifiques de l'eutrophisation : les différentes formes d'eutrophisation (marées vertes à ulves ; phytoplancton ; …), description des mécanismes associés, critères (communautés d'algues, facteurs abiotiques, …) caractérisant les perturbations écologiques
- Facteurs à l'origine de l'eutrophisation : identification des facteurs du milieu prédominants, variation en fonction du contexte
- Symptômes et manifestations de l'eutrophisation : conséquences sur les différents écosystèmes aquatiques et risques associés

b. Cadre réglementaire ou juridique

Sur la base des documents réglementaires qui seront notamment fournis par les Ministères en charge de l'Ecologie et de l'Agriculture, l'ESCo analysera le(s) cadre(s) juridiques relatif(s) à l'eutrophisation : à partir des publications, comment la notion d'eutrophisation est définie dans les réglementations, quelle est leur complémentarité, et quels sont les critères pris en compte.

3 - Caractérisation de la vulnérabilité des milieux à l'eutrophisation : outils et méthodes pertinents

L'objectif est d'éclairer les pouvoirs publics dans la définition des critères permettant de caractériser la vulnérabilité des milieux aquatiques à l'eutrophisation et dans le choix des méthodes appropriées. Ces questions scientifiques seront traitées séparément pour les cas particuliers de l'eutrophisation continentale et de l'eutrophisation littorale et marine.

3-1 Vulnérabilité des milieux aquatiques

a) Etat des lieux des connaissances scientifiques relatives aux critères les plus pertinents pour caractériser la vulnérabilité des milieux aquatiques vis-à-vis de l'eutrophisation (eutrophisation continentale, littorale et marine)
b) Etat des lieux des méthodes ou modèles existants pour évaluer la vulnérabilité des milieux à l'eutrophisation, domaine de validité et pertinence
c) Identification des critères et méthodes pertinents pour caractériser la vulnérabilité des milieux continentaux et marins à l'eutrophisation en relation avec la richesse des milieux en nutriments

- Rôle des nutriments dans les phénomènes d'eutrophisation, notion de risque lié aux nutriments présents dans le milieu et de gradation du risque, prise en compte des flux de nutriments dans le réseau hydrographique et dans le milieu marin
- Changements et évolutions probables avec le changement climatique
- Etudes et modèles existants, seuils proposés par la littérature : analyse des variabilités et pertinence
- Identification des critères les plus adaptés pour mesurer les teneurs en nutriments dans le milieu (quand et où mesurer, à quelle fréquence, …), en lien avec le fonctionnement biologique des écosystèmes et le comportement des nutriments dans le milieu

3-2 Risques liés aux transferts dans les bassins versants
Comment caractériser le risque d'eutrophisation des milieux en lien avec les propriétés des bassins versants vis-à-vis des transferts de nutriments ?
Estimation des stocks et des temps de résidence des nutriments dans les bassins versants.
Analyse critique des modèles couplés et non couplés de carbone, azote et phosphore selon différentes échelles et contextes.

4 – Bilan et identification des besoins en termes de connaissances

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