The impact of biogas digestate typology on nutrient recovery for plant growth: Accessibility indicators for first fertilization prediction

Julie Jimenez, Marco Grigatti, Elisa Boanini, Dominique Patureau, Nicolas Bernet

HAL Id: hal-02921497
https://hal.inrae.fr/hal-02921497v1

Submitted on 22 Aug 2022

1   **The impact of biogas digestate typology on nutrient recovery for plant growth: accessibility**

2   **indicators for first fertilization prediction**

3   Julie Jimenez[1], Marco Grigatti[2], Elisa Boanini[3], Dominique Patureau[1], Nicolas Bernet[1]

4   [1]LBE, INRAE, Univ Montpellier, 102 Avenue des Etangs, Narbonne, F-11100, France

5   [2] Department of Agricultural and Food Sciences, Alma Mater Studiorum, University of Bologna, Viale

6   G. Fanin, 40, 40127 Bologna, Italy

7   [3] Department of Chemistry "Giacomo Ciamician", Alma Mater Studiorum – University of Bologna,

8   Via Selmi 2, 40126 Bologna, Italy

9   E-mail: julie.jimenez@ inrae.fr

10   **Abstract**

11   In recent years, anaerobic digestion of organic waste (OW) is rapidly appearing as a winning waste

12   management strategy by producing energy and anaerobic digestates that can be used as fertilizers in

13   agricultural soils. In this context, the management of the OW treatment process to maximize agro-

14   system sustainability satisfying the crop nutrient demands represents the main goal. To investigate

15   these traits, two protocols to assess the plant availability of digestate nitrogen (N) and phosphorus (P)

16   were evaluated. With this aim, the N and P availability was determined on 8 digestates and 2 types of

17   digestate-based compost from different OW via sequential chemical extractions (SCE). In addition, the

18   digestates were tested in soil incubations and in plant pot tests with Italian ryegrass and compared with

19   chemical fertilizer and a non-amended control soil. The N extracted from digestates via SCE was

20   related to soil N mineralization and plant N recovery. The C: N ratio had negative impact on

21   mineralized N and its recovery in shoots ($\mathbf{Shoots_N = -0.0085 \times \frac{c}{N} + 0.1782}$, r²=0.67), whereas water

22   extractable mineral N was positevely related to the root N apparent recovery fraction (N-ARF) with

23   ($\boldsymbol{Roots_N = 5E^{-5} \times N_{solublemin} + 0.0138}$, r²=0.53). The shoot P-ARF was positively correlated with the

24   inorganic water extractable fraction of P ($\boldsymbol{Shoots_P = 0.1153 \times H_2O - P_i - 0.2777 \times H_2O - P_o +}$

25    $\mathbf{0.0249}$, r²=0.71) whereas the root P-ARF was positively correlated with the less accessible fractions

26    $(\textbf{Roots}_\textbf{P} = \textbf{0.0955} \times \textbf{NaHCO}_\textbf{3}\textbf{P}_\textbf{o} + \textbf{0.0955} \times \textbf{NaOH P}_\textbf{o} - \textbf{0.0584} \times \textbf{NaHCO}_\textbf{3}\textbf{P}_\textbf{i} + \textbf{0.0128}$, r²=0.8641).

27   Feedstock digestate typology impacted the N and P recovery results leading to a better description of

28   the typology properties and a first nutrients ARF prediction.

29   **Key words** digestates, typology, fertilizers, characterization indicators, nutrient recovery

30

31   **1. Introduction**

32   An increasing interest exists in improving the soil quality following the utilization of organic wastes

33   (OW) within the environmentally friendly bio-refinery approach (Alburquerque et al., 2012; Gissén et

34   al., 2014). In this respect, anaerobic digestion is a major building block in the circular bioeconomy.

35   Furthermore, both energy (ie. biomethane) and organic fertilizers (ie. digestate) replacing the chemical

36   fertilizers provided to crops can be recovered from anaerobic digestion of OW. From an agronomic

37   and economic point of view, digestate use as a fertilizer can be considered not only as a supplement

38   for traditional organic fertilizers (i.e. slurry) but also as an alternative fertilizer, which under certain

39   soil conditions, is more effective than using NPK fertilizer (Barlog et al., 2019). Indeed, anaerobic

40   digestion is reported as a suitable treatment to easily recover P and N fertilizers from digestate

41   (Mazzini et al., 2020). This is due to organic matter and free organic nutrients biodegraded into

42   mineral forms. However, agricultural reuse of treated OW as soil fertilizers is limited by

43   environmental constraints related to the quality of organic matter, nutrient availability and safety

44   issues. Furthermore, the chemical composition of digestate depends on the nature of feedstock and

45   digestion process (Alburquerque et al., 2012, Guilayn et al., 2019, Barlog et al., 2019). Indeed, a

46   statistical classification of 91 digestates based on their agronomic quality (i.e. C, N, P, total solids,

47   organic matter) was applied on a large range of raw digestates by (Guilayn et al., 2019) and revealed 6

48   groups of different digestate typology. By analyzing the obtained digestates groups, two criteria

49   impacted the classification: the TS concentration of the digestates associated with dry or wet anaerobic

50   digestion process and the feedstock type, as following: liquid fibrous feedstock (crop residues silage,

51   cattle slurry), liquid sewage sludge, liquid pig slurry, slurries co-digested with silage and green wastes

(dry anaerobic digestion), municipal solid wastes (dry anaerobic digestion) and fibrous feedstock (dry

anaerobic digestion of cattle manure and green waste). In this framework, mineral N can be readily

available for crops but can be temporarily immobilized by microbial biomass during OW

mineralization in soil (Lashermes et al., 2010). Therefore, it is important to know the OW

mineralization N kinetics to optimize the N supply synchronisation with plant requirement in

agricultural systems.

In this context, to better control and manage the quality of digestates, a better knowledge of the

accessibility and availability of nutrients and organic matter would improve the fertilizing potential of

these products. According to Möller et al. (2012), an accurate characterization of digestate nutrient

content and organic matter (OM) composition, combined with experiments to assess the N

mineralization and N immobilization processes after field spreading, would be essential for a better

characterization of the driving factors driving N turnover in the soil.

Total content of nutrients in OW harldy predicts their fertilizer potential. Ahmad et al. (2018) reported

more than 80% of applied P is rapidly immobilized being unavailable for plant following

adsorption/precipitation processes or conversion into organic form. To gain a better insight on this

issue, many authors proposed to characterize the available P via P fractionation. He et al. (2010)

reported that bioavailability of applied P depends on the presence of specific P forms and that labile P

includes the sum of inorganic and organic P from H_2O and $NaHCO_3$ extracted fractions. In this regard,

Grigatti et al. (2015; 2017; 2019), performed a dedicated SCE on digestates and composts in order to

measure the phosphorous potentially available for plants. The authors found good correlations

between the water- and the sodium bicarbonate-extractable P from composts with short- and medium-

term plant P uptake.

Lashermes et al. (2010) proposed an OW classification based on their chemical characteristics to

predict N availability. However, only 2% of the 273 OW studied came from similar typology of

digestates coming from fibrous feedtstock anaerobic digestion with low N availability (i.e. digestates

was classified in a group with initial N concentration under 65 g kg^{-1} and C: N ratio around 15).

However, according to Guilayn et al. (2019), the N content from digestates varies from 20 g kg^{-1} to

175 g kg^{-1} and C: N ratio varies from 2 to more than 30. It would be interesting to complete the

80   Lashermes study with different typologies of digestate to classify their fertilizer potential. To do that,

81   the first step would be to evaluate the impact of digestate typology on plant growth. Recently, a

82   methodology proposed for organic matter characterization has been applied to a large range of OW in

83   order to predict both bio-avalability and complexity/biodegradability of the organic matter (Jimenez et

84   al., 2015). This technique has been successfully used for OM biodegradation kinetics modelling

85   (anaerobic digestion, compost, soil) (Jimenez et al., 2017) and for organic N dynamics in anaerobic

86   digestion (Bareha et al., 2018; 2019). The basic idea was to find a consistent characterization method

87   in order to describe the bioprocess kinetics and to model the whole treatment chain in terms of organic

88   carbon fate, e.g. anaerobic digestion, compost and organic matter fate in soil. The methodology used is

89   based on SCE to simulate organic matter bioaccessibility for microorganisms combined with three-

90   dimensional fluorescence spectroscopy to analyze the organic matter complexity. The challenge is

91   now to transfer this technique to the characterization of nitrogen accessibility in order to predict its

92   fate after land-spreading and its availability for plant growth.

93   Apart from the previously mentioned studies, there is no literature information on the development of

94   indicators for both N and P plant availability from digestates. This is the reason this study aimed to

95   propose indicators based on chemical extractions and characterization for N and P.

96   In this light, the objectives of this work were to determine the plant-available N and P fractions based

97   on characterization of the accessibility study in order to: (i) assess the impact of typology of OW on

98   the nutrient availability, and (ii) use these indicators to predict plant's nutrients recovery. This study,

99   focussed on a wide range of anaerobic digestates from many types of OW. The results could be used

100   for rapid diagnostic of the (N, P) fertilization potential of a digestate.

101   **2. Materials and methods**

102   **2.1. Organic waste**

103   In order to ensure a representative survey, the anaerobic digestates used in this work were selected on

104   the basis of the results reported by Guilayn et al. (2019). The typology developed by (Guilayn et al.,

105   2019) was used to select the digestate samples out of a large panel of digestates. For this study, 6 main

106   groups were selected from a statistical study applied to the agronomic characteristics: sewage sludge,

107   municipal waste, cow manure (dry -anaerobic digestion), pig manure (liquid anaerobic digestion),

108   centralised (co-digestion), and crop residues. Table 1 presents the samples used with the type of OW

109   used as feedstock and the conditions of the anaerobic digestion process. Post-treated digestates

110   through phase separation (solid or liquid) and composting were also added. Indeed, phase separation is

111   the most classical digestate post-treatment (Alburquerque et al., 2012), leading to two products of

112   different quality (i.e. solid phase as soil amendment and liquid phase as fertilizer). Besides, two

113   digestate-based types of compost (FFMSW_2 and Sludge_2) were used since composting is widely

114   adopted treatment for the solid phase of urban OW digestate to meet the European fertilizing

115   regulation parameters (European Parliament and Council of the European Union, 2016). The organic

116   waste samples were freeze-dried and ground (ø 1mm), in order to reduce the particle size effect in the

117   soil incubations and plant growth experiments.

118   **2.2. Analytical measurements**

119   Freeze-dried ground samples were further ball milled and analysed for the main physico-chemical

120   parameters. The moisture was determined at $105 \pm 2°C$ until constant weight (24-48h), the volatile

121   solid (VS) was determined on the total solids (TS) at 550 °C for 4 h. The total carbon (TC) and total

122   nitrogen (TN) were determined via an elemental analyser (FlashSmart, Thermo Fisher Scientific). The

123   total nutrient and trace elements were determined by ICP (Inductively Coupled Plasma-OES, Spectro

124   Arcos, Ametek) on ≈250 mg of samples after microwave assisted digestion with 65% HNO_3 + 37%

125   HCl. All the analyses were done in duplicates.

126   Based on Jimenez et al. (2015), the sequential chemical extractions of organic matter were applied on

127   the freeze-dried and grounded samples (0.5 g). An orbital shaker was used for the extractions steps as

128   following described:30 mL of 0.01M $CaCl_2$ (twice, 1h, 30°C, 300 rpm) ; 0.01M (NaOH+NaCl) (4

129   times, 15 min, 30 °C, 300 rpm); 0.1M HCl (once, 1h, 30°C, 300 rpm), 0.1M NaOH (4 times, 1h, 30

130   °C, 300 rpm) and 72% H_2SO_4 (twice, 3h, 30°C, 300 rpm). The fractions names are respectively:

131   Soluble extracted from Particulate Organic Matter (SPOM), Readily Extractible Organic Matter

132   (REOM), Slowly Extractible Organic Matter (SEOM) and Poorly Extractible Organic Matter

133 (PEOM). The not extracted fraction is called Non Extractible Organic Matter (NEOM). Each

134 extraction was done in two replicates, centrifuged and the supernatant collected and filtered. The

135 recovered pellets were submitted to the subsequent extraction. Ammonium and total nitrogen were

136 measured in the supernatants using HachLange® kits based on colorimetry methods (LCK 303, 2-47

137 mg N L^{-1}, and LCK 238, 5-40 mg L^{-1} respectively).

138 The organic products were submitted to P fractionation via SCE according to the method of Dou et al.

139 (2000). Freeze-dried and ball-milled products were extracted for 24 h with deionized water (H$_2$O) in

140 an end-over-end shaker, and then centrifuged. The supernatants were passed through a Whatman #42

141 filter, and the recovered pellets were extracted with 0.5 M NaHCO$_3$ (pH 8.5) for 24 h. The same

142 procedure was repeated with 0.1N NaOH and 1N HCl. The residual pellets were treated with a mixture

143 of 96% H$_2$SO$_4$ and 30% H$_2$O$_2$ via hot acid digestion at 360 °C. Inorganic P (P$_i$) in the extracts was

144 determined via the molybdenum blue method of Murphy and Riley (1962). The P recovered in each

145 fraction [water extractable P (H$_2$O-P), bicarbonate extractable P (NaHCO$_3$-P), alkali extractable P

146 (NaOH-P), acid extractable P (HCl-P), Residual-P] was calculated as follows:

147 $$P_{i\ fraction\ x}(\%) = \frac{Pi_{fraction\ x}}{P_{tot\ OW}} \times 100 \qquad \textbf{Equation 1}$$

148 where $P_{i;fraction\ x}$ is the inorganic P determined in each fraction (H$_2$O, NaHCO$_3$, NaOH, HCl, Residual),

149 and P$_{tot\ OW}$ is the total-P determined in the different organic waste via ICP after microwave-assisted

150 acid digestion.

151 The total recovery was calculated as the sum of all fractions (H$_2$O-P + NaHCO$_3$-P + NaOH-P + HCl-P

152 + Residual-P) by using the Equation2:

153 $$Tot\ P_{recovery}(\%) = \frac{\sum_{H2O}^{Residual} P_{i\ fraction\ x}}{P_{tot\ OW}} \times 100 \qquad \textbf{Equation 2}$$

154 where $\sum_{H2O}^{Residual} P_{i\ fraction\ x}$ represents the sum of the single P recovery values (P$_i$) in each fraction (H$_2$O-P +

155 NaHCO$_3$-P + NaOH-P + HCl-P + Residual-P), and P$_{tot\ OW}$ is the total P determined in the different

156 products via ICP after microwave-assisted acid digestion. Organic P was calculated as the difference

157 between P$_{tot}$ and P$_i$ in the first four fractions.

158   According to He et al. (2010), H_2O-P, NaHCO3-P, NaOH-P and HCl-P are respectively water-soluble,

159   bioavailable, potential bioavailable (Fe/Al bound) and Ca-bound P.

160

161   **2.3. X-ray diffraction**

162   Samples submitted to X-ray diffraction analysis were carefully ground and placed in a flat stage

163   sample holder. Data were collected by means of a PANalytical X'Pert PRO powder diffractometer

164   equipped with a fast X'Celerator detector. Cu Kα radiation was used (40 mA, 40 kV). The 2θ range

165   was from 5° to 60° with a step size of 0.1° and time/step of 100 s. Data were processed for phase

166   identification using a HighScore Plus software package (PANalytical).

167

168   **2.4. Soil incubation**

169   **2.4.1. Soil**

170   The soil used for the incubation and pot trial was collected from the top layer in a field in the Po

171   Valley (Bologna, Italy). This soil showed the following characteristics: pH (H_2O 1:2.5), 7.90; particle-

172   size distribution; 184 mg kg^{-1}, sand, 425 mg kg^{-1}, silt, 391 mg kg^{-1} clay; total $CaCO_3$, 85 g kg^{-1}; total

173   organic carbon (TOC), 10.2 g kg^{-1} ; total Kjeldahl nitrogen (TKN), 1.60 g kg^{-1}; C:N, 8.3; exchangeable

174   K, 330 mg kg^{-1} as K_2O, CEC 27.2 meq. 100 g^{-1}. The total (extractable in aqua regia + HF) Al, Fe and P

175   were 35661, 22224 and 808 mg kg^{-1} respectively. The NH_4-oxalate (pH 3) extractable Al and Fe were

176   764 and 2158 mg kg^{-1} respectively, while the Na dithionite-citrate extractable Al and Fe were 281 and

177   2462 mg kg^{-1} respectively.

178   **2.4.2. Incubation tests**

179   Soil incubation was conducted in 500 ml plastic vessels with perforated plastic cap, on 250 g of soil

180   (TS basis). The soil was rewetted at 60 % of the water holding capacity (WHC) and pre-incubated for

181   4 weeks at 25 °C in the dark. Water content was kept at 60% of WHC by weighing the vessels every

182   2-3 days and by adjusting with water drops if needed. After this period, the organic products (i.e.

183   digestates) were added to the soil at 170 kg N ha^{-1}, defined as the maximum N load per year

184 authorized by the European Nitrates Directive (1991) and accurately mixed by hand. Chemical

185 references used as positive control (Ctrl⁺) were added by distributing N (as NH_4NO_3) at the same

186 concentration of 170 kg N ha⁻¹, and in addition P and K were added at 117 and 147 kg ha⁻¹ (as

187 KH_2PO_4). On the basis of N loading the P added to the soil with the different organic product was on

188 average 20.45 +/- 2.5 mg P kg soil⁻¹. Calculation of the amount of digestate mass added is described

189 by the Equations 3 to 5 considering 0.3 m of soil depth (d) and a bulk density of 1.33 g/cm³ for 1 ha of

190 soil.

191 $$\text{dosis}\left(\frac{\text{mgN}}{\text{kg soil}}\right) = \frac{dosis\left(\frac{mgN}{ha}\right)}{area_{1\,ha} \times d} = \frac{170 \times 10^6}{10^4 \times 0.3 \times 1.3 \times 10^3} = 44\ mgN.kg^{-1}$$ Equation 3

192 $$\text{dosis}\left(\frac{\text{gN}}{\text{kg soil}}\right) = \frac{m_{\text{digestate}} \times N_{\text{digestate}} \times TS_{\text{digestate}}}{m_{\text{soil}}}$$ Equation 4

193 From the Equation 4, digestate amount is calculated as:

194 $$m_{\text{digestate}} = \frac{dosis \times m_{\text{soil}}}{N_{\text{digestate}} \times TS_{\text{digestate}}}$$ Equation 5

195 $m_{\text{digestate}}$ is the digestate mass (kg)

196 $N_{\text{digestate}}$ is the N concentration of the digestate (gN.kgTS-1)

197 $TS_{\text{digestate}}$ is the Total solids content of the digestate (% of fresh matter)

198 m_{soil} is the considered soil mass in the test (kg)

199 Three replicates for each treatment were assessed in a completely randomized block design, in

200 addition to an unfertilized treatment control (Ctrl⁻). Two parallel soil sampling series were done to

201 follow the P and N evolution during incubation. Olsen-P, was determined according to Watanabe and

202 Olsen (1965) at days: 0, 14, 28, 56, 84.The Olsen-P data were used to calculate the relative percentage

203 extractable (RPE) P, used to normalise the P extractability for each treatment relative to Olsen-P

204 obtained for Ctrl⁺ (Grigatti et al., 2019). The mineral N course in soil was assessed by extracting soil

205 samples (1g TS basis) at days 0, 14, 28, 56 and 84 with 1M KCl for 30 min on an end-over-and end

206 shaker. The solution was filtered over a Whatman #42 filter and ammonium and nitrates were

207 determined using Hach Lange® kits (LCK 304, 0.015-2 mg N.L^{-1} and LCK 339, 0.23-13.5 mg N L^{-1}

208 respectively).

2.5. Plant pot trials

210 Based on Grigatti et al. (2014 ; 2015), the digestate samples were added to the ground soil at 170 mg

211 N kg^{-1} and thoroughly mixed by hand in 2 liter plastic pots (ø 140 mm × h 150 mm). These were filled

212 with 1 liter of inert material (agricultural light expanded clay), and 1 kg of each different treated soil in

213 3 replicates. Pots were seeded with 0.8 g of seeds of Italian ryegrass (*Lolium multiflorum* subsp.

214 Italicum), cv. Sprint, covered with a thin layer of sand to prevent drying, watered and placed in a

215 growth chamber at 16-18 h day-night photoperiod at 13–23 °C (±3 °C) day-night temperature, the light

216 was ensured by 6 Philips Master Tld 58 W-840 tubes. Besides the organic products, an unfertilized

217 control (Ctrl^{-}), and a chemical reference (at the same rate used for soil incubation) was added (Ctrl^{+}).

218 The same treatments and dosis performed in soil incubation test were applied. The use of fast growing

219 species as ryegrass in controlled conditions (moisture; temperature; light) leads to a multiple harvest

220 approach giving the opportunity to the best description of apparent N and P utilization kinetics in the

221 time frame of a growing season (Gunnarson et al., 2010; Schiemenz et al., 2010 ; Tampio et al., 2016).

222 After emergence, plants were regularly watered with tap water to keep soil at 60 % Water Holding

223 Capacity (WHC). At each harvest, ryegrass plants were cut 2 cm above ground and collected 3 times:

224 at 28, 56 and 84 days after sowing. The plant biomass was then dried in a forced air oven at 60 °C for

225 3 days and weighed, to determine dry weight (DW) per pot. Dry biomass was also ball-milled for

226 subsequent analysis. At the last harvest, the roots were divided from the soil with a combined water-

227 sieving separation, and the root biomass was then dried as above, weighed and milled. On plant tissue

228 (shoots and roots), the total P content was determined by means of ICP after (HNO$_3$, H$_2$O$_2$,)

229 microwave assisted digestion. The total N content was determined by elementar analysis as the OW

230 samples. Apparent plant N and P utilization efficiency was calculated on the basis of the Apparent

231 Recovery Fraction (ARF) approach (Gunnarsonn et al., 2010), according to the Equation 6:

232 $$\text{ARF\_X } (\%) = \sum_{i=1}^{3} \frac{\text{X uptake treatment}_{ti} - i\text{X uptake ctrl}-_{ti}}{\text{X added}} \cdot \qquad \textbf{Equation 6}$$

233   in which X uptake treatment (t_n) is the total nitrogen or phosphorus uptake (mg pot^{-1}) of a fertilizer

234   treatment at time t (i = cut 1-3.); X uptake ctrl$^-_{ti}$ is the total nitrogen or phosphorus uptake (mg pot^{-1}) of

235   the unfertilized control at time t (i = cut 1-3); X added is the total nitrogen or phosphorus added to the

236   pot (mg pot^{-1}).

237   Chemical nutrient equivalent coefficient k_{eq} is defined as the equivalent chemical nutrient dosis (i.e.

238   positive control) percentage needed to reach similar crop yield and is calculated according the

239   Equation 7.

240     $$k_{eq}X = \frac{ARF\_X}{ARF_{Ctrl+}} \qquad\qquad\qquad \text{Equation 7}$$

241   **2.6. Statistical analysis**

242   All the data from the plant growth experiment were analyzed by means of Kruskal-Wallis non

243   parametric test. Principal Component Analysis (PCA) and Hierarchical Clustering Analysis (HCA)

244   were also performed on all the data using FactomineR package from the R software. Partial Least

245   Square Regression (PLSR) was performed for variables prediction using the SIMCA® software. For

246   PCA analysis, data from Grigatti et al. (2019) were included as far as similar experiments were done

247   on P speciation and plant pot tests. The three samples were D1, D2 and BD, which are respectively

248   digestate from thermophilic wastewater sludge digestion, digestate from mesophilic winery sludge

249   treatment and digestate from mesophilic bovine slurry and energy crops treatment. Table 1 presents

250   these samples, the conditions of the anaerobic digestion process and the associated feedstocks.

251   **3.   Results and discussion**

252   **3.1. Digestates characterization**

253   The main physico-chemical characteristics of the studied digestates are reported in the Table 2. A high

254   variability of the parameters analysed was apparent. Indeed, organic matter ranged between 40 and

255   and 90% (TS basis), the TKN varied between 15 and 45 g kg^{-1} while P ranged between 4 and 20 g kg$^-$

256   ^{1}.The most N-rich samples (44-48 g kg^{-1}) were the liquid phase of the centralised digestate (Centr_2),

257   the pig manure digestate (Agri_2) and the sludge digestates (Sludge_1, D1 and D2). The poorest

258   samples were the digestates of FFMSW and its compost (between 15 and 17 g N kg^{-1}). The richest P

259   samples were the sludge digestate (Sludge_1) and its compost (Sludge_2), along with the solid phase

260   of the centralised digestate (Centr_1) and digestate D2 (sludge from winery processing) with a P

261   content ranging between 18 and 20 g kg^{-1}. The FFMSW_1 digestate and its compost FFMSW_2

262   besides the silage straw digestate (Agri_3) were the poorest P samples (4 g P kg^{-1}). PCA and HCA

263   analyses were performed on the samples characteristics for a whole sight of the digestates profiles thus

264   allowing the comparison with the typology of Guilayn et al. (2019). The results are presented in the

265   Supplementary Material (Figure A.1). The PCA was applied on nine parameters (i.e from TS to K in

266   Table 2). The first two components recovered $\approx$ 80% variance. The first component (PC1) was mainly

267   related to the TS, negatively related to OM and K.

268   PC1 clustered dried (i.e. FFMSW_1, composts Sludge_2 and FFMSW_2) and liquid AD digestates

269   (i.e. all the others). The second component (PC2) was formed by P, which was negatively related to

270   the C: N ratio related to fibrous composition. The sludge digestate samples were associated with P

271   concentration variable whereas Agri_3 (wheat straw digestate) was mainly associated to C: N ratio

272   variable. Furthermore, Sludge_1 and 2 were the samples containing the highest metals concentrations

273   (Fe, Al, Pb, Cu, Mn, Mo, and Cd).

274   These results were in agreement with the raw digestate typology found by Guilayn et al. (2019).

275   Indeed, Agri_1 and Agri_3 characteristics were in agreement with the characteristics associated to the

276   fibrous feedstock digestate group after dried digestion (i.e. cattle manure and crop residues). Agri_2

277   characteristics were consistent with those of the pig slurry digestate group from the typology.

278   Sludge_1 characteristics were consistent with those of the sludge digestate group and FFMSW_1

279   characteristics fit with the municipal solid wastes digestate group characteristics. BW_1

280   characteristics were mainly close to the biowaste and municipal solids waste digestate composition

281   after dry digestion. However, NH$_4$ concentration and NH$_4$: TN ratio had similar values than the

282   manure co-digestion (i.e. NH$_4$: TN ratio > 70%). This is probably due to the high ratio of agro-

283   industrial waste in the feedstock used in addition of biowastes. Only digestate from the wet digestion

284   of fibrous substrate was not taken into account (silage) in this study. Composts Sludge_2 and

285 FFMSW_2 and Centr_1 and 2 obtained after phase separation were not compared with the typology as

286 they were not raw digestates.

287 Attention should be paid to the results for TKN and $N\text{-}NH^+_4$ concentration obtained by the raw sample

288 analysis and to the freeze-dried sample extraction analysis. Regarding some digestates, highly

289 significant positive differences were obtained between raw and freeze-dried samples analyses (from

290 40% for Centr_2 to 61% for Agri_2 of TKN loss). This result was due to ammonia volatilization

291 during freeze-drying operation. The loss of ammonia could have an impact on soil incubation and

292 plant N recovery results. Considering the 8 others digestates, no significant differences appeared and a

293 linear relation was obtained between raw sample analysis and freeze-dried analysis (i.e. TKN_raw =

294 0.9443.TKN_freeze-dried, R^2 = 0.8832). Furthermore, the N typology of Agri_2 became similar as

295 Agri_1 and 3 (fibrous feedstock involving manures). NH_4: TN ratio evolved from 76% to 39%. This

296 ratio decreased also for Centr_2 from 46% to 9% after freeze-drying. Consequently, Agri_2 and

297 Centr_2 results were used in statistical tests by using the freeze-dried analysis. N-ARF recovery will

298 be discussed accordingly. However, concerning liquid digestates with high ammonia content as pig

299 slurry digestates, the best option would be to perform soil incubation and plant pot trials with fresh

300 matter as Rigsby et al. (2013), de la Fuente et al. (2013). A higher mass of soil would be required to

301 maintain the soil WHC with the liquid digestate moisture, according Equation 5.

302 This first characterization approach showed that the digestates selected were representative of a large

303 range of digestates (i.e. municipal waste, sludge, manure, crop residue, bio-wastes, centralised).

304 **3.1.1. X-ray powder diffraction**

305 The X-ray powder diffraction has been widely use for the study of the crystalline P species in anaerobic

306 digestate and compost (Li et al., 2019; Grigatti et al., 2017; 2019). In this light the XRD profiles from the

307 tested products can give a valuable insight to their P extractability. The XRD patterns from the digestates

308 are presented in Figure 1. The XRD patterns of samples Agri_1, Agri_2 and Agri_3 were similar

309 showing the presence of crystalline inorganic phases, namely quartz (PDF no. 01-087-2096), calcite

310 (PDF no. 01-085-1108) and sylvine (KCl, PDF no. PDF no.41-1476).

However, these three samples differed significantly in their amorphous phase content, which was definitively higher in Agri_3, followed by Agri_1, whereas sample Agri_2 was mostly constituted of crystalline material. Furthermore, Agri_2 presents also struvite peaks (PDF no. 15-0762), that could be detected on account of the higher crystallinity of this sample. Amongst the analyzed samples, Sludge_1 was the most crystalline one, with no significant presence of amorphous material. The Sludge_1 pattern showed the presence of calcite and quartz as main constituting phases, together with a low amount of vivianite ($Fe_3(PO_4)_2(H_2O)_8$, PDF no. 01-075-1186). Li et al. (2018) found also that quartz was one of the major phases for sewage sludge. Sludge_2 was somewhat similar to Sludge_1, but less crystalline so that only calcite and quartz were detected (in this case quartz is more abundant than calcite). FFMSW_1 and FFMSW_2 were very similar both regarding the amount of amorphous phase which is present together with the fine crystalline fraction of the materials. Regarding the qualitative detection of crystalline phases, it was possible to identify quartz (the most abundant), calcite, together with low amounts of kaolinite ($Al_2(Si2O_5)(OH)_4$, PDF no.01-080-0885) and sylvite. BW_1, Centr_1 and Centr_2 patterns all displayed the presence of quartz, calcite and struvite as crystalline components, but significantly differed for the amount of amorphous material that was relatively low in Centr_2, and higher and similar in BW_1 and Centr_1. The relationship between the XRD outocomes and the P extractability are dicussed in the following section.

3.1.2. Sequential chemical extractions of Nitrogen and Phosphorous

The N fractionation protocol used in this work was developed by Jimenez et al. (2015), and has been validated to describe OM and N evolution during biological degradation processes (Jimenez et al., 2017 and Bareha et al., 2018). The P fractionation has been adapted from Hedley et al. (1982), using methodology by Grigatti et al. (2015, 2017, and 2019). These authors showed the link between labile-P from organic waste samples and the plant P-uptake. The N and P speciations have been assessed applying both fractionation protocols to the selected digestates (Figure 2 a and b respectively). The results showed that the digestates performed different N and P speciations. This was related to the digestate typology and the nature of the feedstock for anaerobic digestion, according to Guilayn et al.

13

(2019). The observed variability in N and P speciation showed the significant effect of the feedstock on the tests discussed in section 3.2. In the study of N accessibility (Figure 2 a), mineral nitrogen was assessed in the water extractable fraction SPOM. The non-extractable organic nitrogen varied between 25 and 48% except for Agri_3 with only 5% N in the NEOM fraction.

HCA was applied on the N speciation data (not shown). Four groups were found, ordered by N accessibility basis as follows: (i) Group 1: High NH_4-N SPOM samples: Centr_1 and BW_1. This result is consistent since these samples had poor TS; (ii) Group 2: High organic SPOM and REOM: Agri_2 and Sludge_1. SPOM and REOM are mainly composed of accessible and readily extractable proteins; (iii) Group 3: High organic SEOM: Agri_3. SEOM is mainly composed of complex proteins and humic-like substances; (iv) Group 4: High PEOM organic N content samples: Agri_1, FFMSW_1, Centr_2, FFMSW_2 and Sludge_2. PEOM extraction targets holocellulose-like compounds found in fibrous digestates as Agri_1, FFMSW_1 and 2. Phase separation concentrates OM and fibers in the solid phase (Guilayn et al., 2019) as Centr_2. Finally, the compost Sludge_2 was also clustered in this group because of its co-composting with green wastes.

P was extracted in each fraction as reported in Figure 2 b. Phosphorus fractionation showed that inorganic P content was higher than organic P in all the fractions, as reported in the studies of Grigatti et al. (2015, 2017 and 2019) on digestates and composts, and as observed by He et al. (2010) on poultry litter and dried wastewater sludge. Mazzini et al. (2020) showed that 78 to 93% of TP was inorganic following the SCE from six types of animal slurries digestates. The authors showed relevant NaOH (5-43% of organic P) and HCl (2-25% of organic P) extractable organic P, relating this to inorganic P microbial immobilization during anaerobic digestion for microorganism growth. Indeed, soluble inorganic P would be transformed into organic P compounds such as phosphates monoesters or DNA. This biological organic-P may be rapidly mineralized once in soil (He et al., 2010). In this context, NaOH and H_2O were the most organic-P rich fractions from the samples investigated in this work. Furthermore, Agri_1, 2 and 3, and FFMSW_1 mainly showed NaOH extractable organic P. Agri_3 contained the highest organic P content ($\approx$40%), and 1.5 to 4- folds higer than other samples. In water Agri_2 and 3 showed 60% of organic P, higher than BD (49%).

The results obtained by He et al. (2010) on poultry litter showed that a large part of organic P was in the NaOH fraction and was mainly related to phytate-like. Mazzini et al. (2020) showed similar results on crop residues digestates. The authors showed also that organic P were mainly extracted in NaOH fraction of 6 agro-wastes digestates as shown by Agri_1 and 3. Organic P was observed in HCl fraction from Sludge_1 and 2 in this study. He et al. (2010), observed also organic P in HCl fractions related to non-hydrolysable fractions in the dried wastewater sludge. The $NaHCO_3$ and the HCl fractions showed to be the most inorganic P rich factions from the samples tested in this work thus fitting to available P and Ca-bounded P. The labile-P fraction (H_2O + $NaHCO_3$) was found at the highest level in the agricultural residue digestates (Agri_1, 2 and 3). Similar observations on bovine manure and energy crops digestate (BD) were made by Grigatti et al. (2019). Then the biowaste and agro-food industries digestates (BW_1, Centr_1 and 2) had also high labile-P. The FFMSW samples showed intermediate P accessibility. The sludge samples were characterized by a high NaOH extractable P, related to metal-bounded P. Indeed, this result was consistent with both Al and Fe content (≈ 7 and ≈ 37 g kg^{-1}). Li et al. (2018), showed mainly NaOH extractable Al, while Fe was NaOH soluble being also extracted by HCl in sewage sludge, thus showing Fe-bounded P was partially occluded. Amongst the others samples, compost Sludge_2 showed important poorly available HCl-P, being very similar to FFMSW_2, these results were consistent with the P fractionation of composts showed by Grigatti et al. (2015, 2017, and 2019).

Finally, HCA showed six groups (clustering not shown). Groups were ordered according to the chemical availability level (i.e. Labile-P> NaOH-P> HCl-P) as follow: (i) Very high P accessibility related to high H_2O-P fractions: digestates from liquid feedstock (no manure), biowastes and winery (BW_1, D1 and D2). These samples were characterized by amorphous P and crystalline P (quartz and struvite); (ii) High P accessibility associated with organic H_2O-P and $NaHCO_3$-P: digestates from liquid manure (BD and Agri_2). These samples were characterized by crystalline P (struvite); (iii) Moderate P accessibility (intermediary P speciation): digestates from organic fraction from municipal wastes digestates and centralised digestates (FFMSW_1, FFMSW_2, Centr_1 and Centr_2). These samples were characterized by an amorphous phase and mainly quartz and calcite as crystalline phases; (iv) Poor-P accessibility associated with organic $NaHCO_3$-P and NaOH-P, associated with digestates from fibrous substrates as cow manure with crops residues and wheat straw (Agri_1 and Agri_3). These samples were

392 characterized by a high amorphous phase; (v) Very poor P accessibility associated with inorganic NaOH-

393 P and low fraction of available P: digestate from solid phase of sludge (Sludge_1). This sample was

394 characterized by a crystalline P phase (mainly quartz and calcite); (vi) The poorest accessible P

395 associated with inorganic HCl-P was compost of sludge digestate (Sludge_2). This sample was

396 characterized by less crystalline P than Sludge_1 with calcite and quartz.

397 For some groups, the feedstock type seemed to have an impact on the P accessibility of digestates (solid

398 phase of sludge, liquid manure, liquid feedstocks, fibrous feedstock). Another observation was that the

399 amorphous and crystalline characteristics also seemed to be associated with some groups: fibrous

400 digestates were mainly composed of amorphous phase and contained organic P in $NaHCO_3$ and NaOH.

401 The labile P fractions had in common amorphous P and struvite as crystalline P. The most crystalline P-

402 rich fractions were found in the solid phase of sludge digestates before and after composting, containing

403 low available P and high sparingly soluble HCl-P.

404 The eight tested digestates represented a wide range of inherent characteristics and accessibility patterns

405 for both N and P. In this framework, the soil and plant test results can give a deeper insight to these

406 issues as discussed in the following section.

3.2. Soil incubation and plant pot trials

408 Table 4 shows the cumulative amount of biomass harvested in ryegrass plant pot tests. Any significant

409 differences between treatments for tissue and root (Kruskal-Wallis test p=0.89 and 0.23 respectively).

410 Nevertheless, it appears that poor biomass was obtained in Agri_2 treatment (pig manure digestate),

411 and in FFMSW_2 treatment (FFMSW digestate compost), close to the negative control. The best

412 results were obtained in Centr_1 and BW_1 (solid phase of a centralised digestate and biowaste

413 digestate). The total plant biomass (shoot+ root) was the highest (g pot^{-1}) in Centr_1 (4.11) ≥ Sludge_2

414 (3.75) ≥ BW_1 (3.72) ≥ FFMSW_1 (3.70) ≥ Ctrl$^+$ (3.65) ≥ Sludge_1 (3.48) ≥ Agri_3 (3.38) ≥ Centr_2

415 (3.23) ≥ Agri__1 (3.18) ≥ Agri_2 (3.02) ≥ FFMSW_2 (2.93) ≥ Ctrl$^-$ (2.90).

3.2.1. Nitrogen fate

417   The N mineralization data from soil incubation (Figure 3) showed the net cumulated mineral N (i.e.

418    $NH_4 + NO_3$ mass) during soil incubation (Figure 3a) and the net cumulated mineralization rate of

419   organic N (Figure 3b).

420   Concerning the available mineral N (N-min), BW_1, Centr_2, Sludge_1, Sludge_2, Agri_2 and

421   FFMSW_1 achieved the highest cumulated values after 84 days of incubation. On the contrary,

422   Centr_1, FFMSW_2, Agri_3 and Agri_1 performed poorly cumulated N-min. These latter were more

423   fibrous samples, with high C: N ratios.

424   The cumulated organic N mineralization rate was calculated (Figure 3b). Two groups of treatments

425   appeared and were classified in the same order than the cumulated N-min: (i) positive mineralization

426   rate associated with the Sludge_1, Centr_2, BW_1, Agri_2, Sludge_2 and FFMSW_1 treatments and

427   (ii) negative mineralization rate associated with Centr_1, FFMSW_2, Agri_1 and Agri_3. Sludge_1

428   showed the highest N mineralization rate (30%) suggesting that this digestate was rich in hydrolysable

429   proteins. Organic N from the liquid digestates Centr_2 and Agri_2 showed respectively 21% and 18%

430   of mineralized organic N, following Sludge_1 (30%). This result is consistent with Guilayn et al.

431   (2019) who reported that the highest organic N came from sludge digestate and pig-slurry digestate

432   group. Sludge_2 and FFMSW_1 performed lower N mineralization rates (11%). Similar results were

433   observed in liquid or solid digestates by Rigsby et al. (2013). They showed that the solid phase from

434   municipal solid wastes digestate (C: N ratio of 11) had negative N mineralization rate whatever the

435   soil composition used (sandy, silty or clay). On the contrary, Rigsby et al. (2013) showed that the

436   liquid digestates from slurry (C: N ratio of 4), had 16% to 40% N mineralization while food waste

437   digestate (C: N ratio of 14) reached 30% N mineralization.

438   The negative mineralization rates group was associated to an immobilization of mineral nitrogen by

439   the soil microorganisms during their growth (De la Fuente et al., 2013). This was related to the

440   increased microbial activity following the OW addition to the soil. Indeed, nitrogen is the main

441   limiting nutrient for plant growth, especially for a fast growing, highly demanding species such as

442   Italian ryegrass (Grigatti et al., 2011). These results are in agreement with Cavalli et al. (2017) which

443   reported immobilization for high C: N ratio anaerobic digestates. The authors showed the low C: N

444   ratio (5 to 7) associated to the raw digestate and its liquid phase (cattle slurry/mais silage) induced net

445   N mineralization ($\approx$ 30%). On the contrary, the cellulose- and volatile fatty acids-rich solid phase (C:

446   N ratio, 20) induced lower net N mineralization (9–16%). Indeed, negative correlations were found

447   between the N mineralization rate and C: N ratio (r= - 0.73, p = 0.029) as observed by Morvan et al.

448   (2006) on animal manure wastes.

449   Considering the impact of process treatment of digestates, composting appeared to negatively affect N

450   mineralization. N mineralization rate from Sludge_2 (11%) was lower than Sludge_1 (30%) and

451   FFMSW_1 (11%) had highest mineralized values than FFMSW_2 (-2%). Phase separation impacted

452   also organic N mineralization as shown by Cavalli et al. (2017). Solid phase (Centr_1) reached

453   negative values of minerlization whereas Centr_2 reached 21% of mineralized N.

454

455   Plant growth experiments were affected by fertilizer treatment for N-ARF (Kruskal-Wallis, p = 0.026)

456   only for shoot plant tissues. Cumulated N recovery by plant tissues was plotted in the Figure 4. The

457   chemical treatment showed the greatest shoot N recovery (70%) with a total of 76% of N-ARF. N

458   uptake recovery was mainly observed in shoots (64% to 87% of the total recovery). Root N-ARF was

459   measured between 7% and 36% of the total N-ARF, except for Agri_1 and Agri_3 treatments where

460   negative to zero values were obtained.

461   Sludge_1 and Sludge_2 achieved the best N-ARF amongst the digestates treatments (16.5 and 15.1%).

462   These results were consistent with the Sludge_1 treatment yielding the highest N-min available for

463   plant growth. Centr_2 (13.7%), BW_1 (13.2%) and Centr_1 (11.6%) treatments showed intermediate

464   total N recovery. The FFMSW_1, FFMSW_2 treatments followed with 8.2% and 6.4% respectively.

465   Finally, a last group formed by Agri_1 and Agri_3 was observed with low N-ARF of 1.8% and 2.6%

466   respectively. In soil incubation tests, Agri_1 and 3 showed N immobilization. Accordingly to these

467   outcomes, the plant N-ARF was negative or close to 0 and close to the unfertilized soil treatment.

468   Considering the digestate treatments impact, composting (i.e. FFMSW_2 and Sludge_ 2) lowered the

469   fertilizing N value of its associated digestate (i.e. FFMSW_1 and Sludge_1 respectively). Phase

470   separation impacted also the N fertilizing potential as shown by Tambone et al. (2017). Centr_2

471   treatment induced higher N mineralization rates and N-ARF than Centr_1 treatment. According to N

472   speciation and characterization, the solid phase Centr_1 contained more fibers than the liquid phase

473   Centr_2.

474   PCA and HCA were performed to find correlations between the N availability data, digestate

475   characteristics, mineralized N percentages and N-ARF from shoots and roots as reported in Figure A.2

476   a (supplementary material). No significant correlation was observed between the total N content, the

477   mineralization rate and the N-ARF. However, a strong correlation was obtained between the C: N ratio

478   and N-ARF from shoots-N (r=-0.82, p=0.004) as in the mineralization N tests without plants. This was

479   described by a linear equation (Equation 8) based on the data of this study and validated by data from

480   digestates and composts studied in Grigatti et al. (2015). Similarly, Decoopman et al. (2017) found a

481   correlation between C: N ratio of agricultural digestates and plant N-ARF measured in field

482   experiments on cereals.

483   $\text{Shoots}_N = -0.0085 \times \frac{C}{N} + 0.1782, \ r^2 = 0.6718$ Equation 8

484   The fiber component of digestates had a negative impact on mineralised N and N recovery by plant

485   tissues. Similarly, the fiber fraction PEOM of digestates was negatively correlated with cumulated

486   mineralized N (-0.58, p=0.08) and shoots-N although to a lesser extent (r=-0.58, p = 0.44). Moreover,

487   cumulative mineralized N and shoots-N ARF were positively correlated (r=0.72, p=0.02) which was

488   consistent with the availability of N for plant growth. Concerning roots, there was a positive

489   correlation between roots N-ARF and nitrates content (r=0.72, p=0.017), SPOM_NH$_4$ (r= 0.61,

490   p=0.06) whereas SPOM_org and roots_N ARF were negatively correlated (r=-0.59, p=0.07). A linear

491   equation (Equation 9) was found between Roots N-ARF and water extracted mineral N (i.e.

492   $N_{soluble\_min} = \text{NO}_3^- + \text{SPOM\_NH}_4$).

493   $\text{Roots}_N = 5E^{-5} \times N_{soluble min} + 0.0138, \ r^2 = 0.5313$ Equation 9

494   According to Gunnarsson et al. (2010), plants respond to N availability with a different root/shoot

495   nutrient allocations and plant growth rate could be influenced by the ammonia/nitrates ratio in soil

496   solution due to the different N sources. Authors observed an increase in root biomass as a result of the

497   availability of a high amount of ammonia.

498   HCA revealed four clusters (Supplementary material Figure A.2.a) as follows: (i) Group 1: Agri 3

499   associated with fibrous digestate with high C: N (>25), high SEOM_N and PEOM_N and low shoots-

500   N ARF. This type of digestate is not suitable as N fertilizer; (ii) Group 2: FFMSW_2 and Agri_1

501   15<C/N<13, intermediary protein-like and fibrous substrate digestates with high organic SPOM_N

502   and PEOM_N for Agri_1 as compost FFMSW_2. They recovered very low roots N-ARF and

503   intermediate level of shoots N-ARF; (iii) Group 3: Centr_1, FFMSW_1 and BW_1 municipal solids

504   and biowastes digestates, with C: N ratios of 15 in average, high ammonium and nitrate levels in the

505   labile fraction and high levels of total N recovery; (iv) Group 4: The protein-like digestates (Sludge_1,

506   Centr_2, Agri_2) which have a poor C: N (<10), poor PEOM_N, high SPOM_Norg content and the

507   highest total N recovery. Sludge_2 was classified in the Group 4 because of its lower C: N ratio (9.7)

508   and its high N mineralization level.

509   C: N ratio is associated to the fibrous level of an OW and seemed to be discriminant enough to predict

510   N-ARF and organic N mineralization rate. Digestates with C: N ratios $\geq$15 were not suitable as N

511   fertilisers whereas digestates with lower C: N ratios had a high potential of N fertilizer.

512   Accessibility fractionation of N allowed a consistent digestate classification. Furthermore, interesting

513   correlation between roots-ARF N and water extractable nitrates and ammonia were found.

514   **3.2.2.Phosphorous fate**

515   The Olsen-P course during the soil incubation is shown by Figure 5a. The Olsen-P evolution in soil

516   depends on several phenomena: (i) the physico-chemical conditions of the soil can trap some weakly

517   bound P, (ii) the organic part of P can be mineralized and (iii) P can be taken up by soil

518   microorganisms for their growth.

519   The control soil had the lowest Olsen-P content throughout the incubation (12-14 mg kg soil^{-1}),

520   whereas the compared treatments exhibited different Olsen-P courses in time. All the tested samples

521   showed an available P depletion during the first two weeks. The chemical control (Ctrl$^+$) performed

522   the fastest depletion. Indeed, the chemical P treatment (phosphate salt) can be rapidly fixed with Ca

523   components in calcareous soil (Alburquerque et al., 2012). In this context, the presence of organic

524 matter and the different P forms from digestates can have positive interactions with soil, performing

525 lower fixation and higher efficiency in comparison to chemical P sources (Alburquerque et al., 2012).

526 In some treatments, an Olsen-P increase appeared between 14 and 28 days (Agri_2, Centr_1, BW_1

527 and FFMSW_1) probably due to organic P mineralization before another decrease until 56 days. This

528 latter decrease led to a lower quantity of available P for plant growth. The RPE obtained was between

529 50 and 84%. The RPE of the Centr_2 treatment was the highest with 84% vs. the chemical P

530 treatment, followed by Sludge_2 (74%). FFMSW_1, Agri_2 and Agri_3 were the samples where the P

531 fixation was the highest (RPE of 60% in average). The other digestates were intermediate between the

532 two groups. FFMSW_1, Agri_2 and Agri_3 contained high fractions of water P.

533 The pot tests showed the treatments affected the plant P tissue (p-value = 0.021). In this context the

534 control soil performed the worst at the first cut (1.93 mg.pot^{-1}), showing very poor performance during

535 the plant pot test reaching 4.55 mg pot^{-1} on average (Figure 5b and 5c). The shoot-P ARF is shown in

536 Figure 4a and b. P-ARF from Agri_2 and Agri_3 was negative from the beginning. This result was

537 consistent with the soil incubation observations. This can be linked with high fixation of P in soil as

538 suggested by Ahmad et al. (2018). In the second cut (day 56), the cumulative shoots P-ARF decreased

539 in all the treatments. Only FFMSW_1 and Centr_2 showed opposite outcomes. Then, between the

540 second and the third cut (day 56; 84), shoot-P ARF further increased. In the end the whole P-ARF (%)

541 was: Centr_1 (5.78)≈ FFMSW(5.31)≥BW_1 (4.35)>Centr_2(3.78)>Sludge_1 (2.72)≥Agri_1

542 (2.44)≥Sludge_2 (2)>FFMSW_2 (1.6)>Agri_2 (0.6)≈Agri_3 (0.5%).

543 FFMSW_2, Agri_2 and 3 showed lower shoots P-ARF than the chemical treatment (1.81%). That

544 means they were not suitable for P fertilization on calcareous soil.

545 The root-P content was unaffected by the treatment. However, the root P-ARF (%) was: Agri_1

546 (2.46)> Agri_3 (2.04)> FFMSW_1 (1.18) = Sludge_1 (1.17) = Sludge_2 (1.16)> Centr_1 (1.04)>

547 BW_1 (0.3)> Centr_2 (0.00)> FFMSW_2 (-0.08)>Agri_2 (-0.21). Shoots-P and roots-P recovery

548 potential were different depending on the treatment, except for Agri_2 and FFMSW_2 which had the

549 lowest ARF P recovery for both.

550   Briat et al. (2020) reported that availability of soil P (and K) for plants highly depends on the N

551   availability. They reported that P concentrations on shoots were correlated with N concentrations in

552   shoots. Indeed, a significant correlation was observed in this study between total N-ARF and total P-

553   ARF (r = 0.8152, p = 0.0022) with a linear relationship (ARF_N/ARF_P = 7.7535, r^2 = 0.6626) thus

554   proving N and P recovery by plants were linked.

555   PCA and HCA analysis were performed on both shoots and roots P uptake and P speciation for the 8

556   digestates studied beside the three digestates from Grigatti et al. (2019), as shown by Figure A.2 b in

557   Supplementary Material. Six groups were found according HCA analysis. The results showed that the

558   P-ARF was classified not only according to the digestate's nature but also according to their P

559   accessibility. The groups found were: (i) group 1 : Sludge_2, compost of solid phase of sludge

560   digestate, HCl rich with an intermediate level of P recovered in shoots and roots; (ii) group 2 :

561   Sludge_1, solid phase of sludge digestate (inorganic $NaHCO_3$ and NaOH fractions rich), with an

562   intermediate level of P recovered in shoots and roots; (iii) groups 3-4 : Centr_1, FFMSW_1,

563   FFMSW_2, Centr_2, Agri_2 (intermediate to low total P-ARF); (iv) groups 5-7 : BW_1, D1, D2 and

564   BD (high total P-ARF and highest inorganic H_2O-P); (v) group 6 : Agri_1, Agri_3, fibrous digestates.

565   Agri_1 has high P total recovery in both shoots and roots as BD, but Agri_3 shows a high recovery

566   only in roots. It seems that the fibrous characteristic enhances P recovery above all in in plants roots

567   through the organic fraction of NaOH-P.

568   The groups obtained were similar to digestate typology groups, except for Agri_2 and BD which were

569   not in a same group. When doing statistical treatments on the shoots results and P speciation for the 8

570   digestates used in addition to the 3 digestates used by Grigatti et al. (2019), some correlations were

571   found. Water-P fraction was positively correlated with shoots (r=0.49, p=0.09). This result was also

572   observed by Grigatti et al. (2017) for composts. As reported by Ahmad et al., 2018, available P consists

573   in the exchangeable and soluble P corresponding to water and Na-bounded P fractions (H_2O+ $NaHCO_3$)

574   whereas occluded P consists into P bound with metals (Fe, Al) and Ca. However, a simple linear model

575   did not fit between inorganic H_2O-P fraction (H_2O-P_i) and shoots P-ARF. PLS model was used showing

576   that two main significant variables were needed to predict shoots P-ARF (Equation 10). H_2O-P_i fraction

577   impacted positively shoots P-ARF, as expected, whereas organic H_2O-P fraction (H_2O-P_o) had a negative

578   impact. Therefore, available form of P for shoots growth was associated to H_2O-P_i fraction, as expected.

579   H_2O-P_o fraction is accessible but has to be mineralized to be available.

580     $Shoots_P = 0.1153 \times H_2O\,P_i - 0.2777 \times H_2O\,P_o + 0.0249$, r²=0.71   Equation 10

581   No correlations were found between ARF results and TP or phosphates concentrations (r=-0.12,

582   p=0.74, r=-0.26, p=0.46 respectively) meaning that the TP concentration is not enough to predict P

583   uptake by plants.

584   Organic $NaHCO_3$-P fraction ($NaHCO_3$-P_o) was positively correlated with roots (r=0.62, p=0.025 with

585   D1, D2 and BD). Organic NaOH-P fraction (NaOH-P_o) was also positively correlated with roots

586   (r=0.79, p=0.006). A PLS regression model was found between roots P-ARF and inorganic and

587   organic P content in $NaHCO_3$-P and NaOH-P fractions (Equation 11).

588     $Roots_P = 0.0955 \times NaHCO_3P_o + 0.0955 \times NaOH\,P_o - 0.0584 \times NaHCO_3P_i + 0.0128$, r²=0.8641

589   Equation 11

590   Negative impact of inorganic labile $NaHCO_3$ was found whereas organic P forms impacted positively

591   the roots P-ARF. As previously mentioned, phytate and others phosphates esters can be extracted in

592   the NaOH-P fraction (He et al., 2010). Roots can exude phosphatases able to hydrolyse phytate and

593   organic P forms able to be absorbed by roots (Lambers et al., 2006; Gerke et al., 2015).

594   Interestingly, roots and shoots P recoveries were correlated with different P fractions. Finally,

595   chemical nutrient equivalents were calculated for all the treatments. The obtained values were above

596   100% except for Agri_2 (35%) and Agri_1 if only shoots were considered. That means that all the

597   studied digestates (except from pig slurry) can substitute chemical P needs.

598   Similarly to N results, phase separation impacted the P fertilizing potential. The solid phase of

599   centralised digestate Centr_1 was more suitable as P fertilizer than Centr_2. Composting impacted

600   also the fertilizing value of the municipal waste digestates by lowering the P fertilizer potential relative

601   to its respective digestate. This trend was also showed by Sludge_1 (digestate) versus Sludge_2

602   (composted digestate). Knowledge of the availability of P and its effects makes possible to anticipate

603   fertilisation according to the soil composition (Alburquerque et al., 2012) and not only based on overall

604   P analysis. Indeed, this was the case of sludge digestates Sludge_1 and 2 which had the highest P

605   contents (same level of Centr_1) but their treatments obtained moderate P-ARF. From P results, first

606   guidelines can be given according to the digestate feedstock typology for the studied calcareous soil

607   and ryegrass. Pig manure digestate (Agri_2) seemed not suitable for P fertilization. Fibrous digestates

608   rich on wheat straw Agri_1 and 3 had very low P fertilizing potential whereas FFMSW_1 presented

609   higher potential as the solid phase Centr_1. The compost FFMSW_2 had also low P fertilizing

610   potential and the liquid phase Centr_2 was intermediate. Sludge 1 and 2 had intermediary results and

611   biowaste digestate BW_1 showed high N and P fertilizing potential.

612   **4. Conclusions**

613   This study focused on different typologies of digestates classified according to their process and to

614   their feedstock. In this context, both P and N speciations showed a wide accessibility range according

615   to feedstocks charateristics. The chemical accessibility indicators described the nutrient availability for

616   plants and allowed the digestates classification on N and P fertilizing potential basis. The N and P

617   speciation impacted the results from incubations with bare soil as well as the apparent coefficients of

618   the use of N and P by the plant for its growth. Depending on the tissue collected (shoot or root), the

619   speciation variables having a significant impact were different for P and N, for the type of calcacerous

620   soil used. First models were found to predict P recovery in shoots and roots using P speciation.

621   Furthermore, C: N ratio value was significant and could be used for shoots N-ARF prediction wheras

622   mineral water extracted N could be used for roots N-ARF prediction. Thus, a more detailed knowledge

623   of the digestates would allow more adequate control of fertilization. Moreover, composting and phase

624   separation impacted the nutrient recovery and can be used as an actuator to propose different organic

625   fertilizers type. In terms of perspectives, field trials on contrasted soils qualities for crops with

626   contrasting nutrient needs should be carried out in order to offer a guide to fertilization by type of

627   digestate. Finally, N and P speciation studied could be used in dynamic models to improve soil and

628   plant model predictions for digestate use in agriculture.

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734   Table 1: Definition and origins of digestate samples

Sample Name	Type	Scale	Anaerobic Digestion conditions: type (dry or liquid), feed mode (batch or continuous) and temperature	Digester Hydraulic retention time	Post-treatment	Substrate	Reference
Agri_1	digestate	Operating scale (farmer)	Dry batch mesophilic	60 days	-	Cow manure	
Agri_2	digestate	Operating scale (farmer)	Liquid continuous mesophilic	60 days	-	Pig manure co-digested with energetic crop and vegetable residues	
Agri_3	digestate	Lab scale	Dry batch mesophilic	100 days	-	wheat straw silage	
Sludge_1	Solid phase from digestate	Operating scale (private agency)	Liquid continuous mesophilic	20 days	Press filter	wastewater treatment sludge	
Sludge_2	compost of solid digestate	Operating scale (private agency)	Liquid continuous mesophilic	20 days	Press filter and composting	digestate of wastewater treatment sludge (1/3) and green wastes (2/3)	This study
FFMSW_1	digestate	Operating scale (private agency)	Dry continous thermophilic	20 days	-	Fermentable fraction from municipal solid waste (FFMW)	
FFMSW_2	compost of digestate	Operating scale (private agency)	Dry continuous thermophilic 20	20 days	Composting	96%FFMSW, 50% green waste, 14% agro-industrial wastes	
BW_1	digestate	Operating scale (governmental)	Liquid continous mesophilic	60 days	-	Biowastes (60%) from supermarkets co-digested with agro-industrial wastes (28%) and crop residues (12%)	
Centr_1	Solid phase of digestate	Operating scale (governmental)	Liquid continous mesophilic	45 days	screw press	Oil (20%), crop residues (10%), agro_industrial wastes (55%), sewage sludge (10%), biowaste (5%)	
Centr_2	Liquid phase of digestate	Operating scale (governmental)	Liquid continous mesophilic	45 days	screw press		
D1	Digestate	Operating scale	Liquid continous thermophilic	na	-	Wastewater treatment sludge	Grigatti et al. (2019)
D2	Digestate	Operating scale	Liquid continous mesophilic	na	-	Wine sludge	
BD	digestate	Operating scale	Liquid continous thermophilic	na	-	Bovine slurry and energy crops	

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737 Table 2: Physico-chemical characteristics of the investigated digestates

Parameters	Units	Agri_1	Agri_2	Agri_3	Sludge_1	Sludge_2	FFMSW_1	FFMSW_2	BW_1	Centr_1	Centr_2	D1	D2	BD
TS	%	17.1%	4.4%	14%	22.4%	59.1%	19.7%	53.3%	24.9%	26.4%	6%	3.94%	3.09%	5.05%
VS	%TS	70.2%	70.1%	87.0%	51.7%	48.5%	48.7%	41.8%	79.0%	82.2%	60.0%	58.2%	59.4%	68.3%
TOC	$g\ kg^{-1}$	382.90	404.71	454.83	283.35	258.01	279.30	232.55	438.64	438.17	320.53	344.2	397.8	515.0
COD	$g.kg^{-1}$	1104.00	1400.56	1235.50	679.00	726.00	712.09	536.00	1291.00	1245.00	1497.58	nd*	nd	Nd
TKN**	$g.kg^{-1}$	25.85	115.00	nd	40.36	26.40	22.20	13.61	29.79	23.33	75.67	nd	nd	nd
TKN***	$g.kg^{-1}$	27.98	**44.53**	17.72	43.06	26.61	17.78	15.39	30.89	28.62	**45.47**	47.8	44.9	43.0
C/N		13.68	3.52	25.67	6.58	9.70	15.71	15.11	14.20	15.31	7.05	7.2	8.9	12.0
NH_4^+-N**	$g.kg^{-1}$	6.28	87.85	2.99	10.80	0.92	10.5	0.04	27.68	6.81	34.74	nd	nd	nd
P	$g.kg^{-1}$	5.36	10.37	4.00	20.44	16.15	4.21	4.23	10.79	6.93	17.98	7.1	18.4	6.2
K	$g.kg^{-1}$	21.22	27.41	12.98	2.15	4.90	11.49	8.22	7.25	4.86	16.10			
S	$g.kg^{-1}$	3.59	6.23	1.63	6.23	5.33	2.22	5.00	2.16	6.27	7.58		nd*	
Al	$g.kg^{-1}$	2.31	3.23	0.65	6.66	12.56	8.68	2.73	8.46	2.68	9.33			
Ca	$g.kg^{-1}$	18.53	23.08	13.44	69.08	41.25	37.05	29.72	42.31	17.56	27.08	10	37	11
Fe	$g.kg^{-1}$	1.58	2.81	0.94	42.60	37.48	5.79	5.06	7.89	11.54	12.07	1.9	8.6	1.0
Ca/P		0.29	0.27	0.24	2.08	2.32	1.38	1.19	0.73	1.67	0.67	1.4	2.0	1.7
Mg	$g.kg^{-1}$	5.49	6.28	1.56	3.52	3.72	3.24	4.29	3.47	1.98	5.13	2.8	11.6	5.2
References						This study							Grigatti et al. (2019)	

738   * nd : Not determined

739   ** Measured on raw samples

740   *** Measured on from freeze-dried samples

741 Table 3: Dry biomass (DW), total P uptake and N uptake in ryegrass shoots during three successive

742 harvests (day 28-84) and roots at the final harvest (day 84)

Treatment	Days after sowing					Roots	Total
	Shoots						
	28	56	84	*Mean*	0-84	84	84
DW (g pot^{-1})							
Ctrl-	0.87	0.53	0.25	*0.55*	1.65	1.25	2.90
Ctrl+	1.00	0.70	0.39	*0.70*	2.09	1.56	3.65
Agri_1	0.88	0.50	0.29	*0.56*	1.67	1.51	3.18
Agri_2	0.86	0.58	0.35	*0.60*	1.79	1.23	3.02
Agri_3	0.84	0.44	0.27	*0.52*	1.56	1.82	3.38
Sludge_1	0.94	0.58	0.42	*0.65*	1.94	1.55	3.49
Sludge_2	0.94	0.57	0.37	*0.63*	1.88	1.87	3.75
FFMSW_1	0.85	0.58	0.36	*0.60*	1.79	1.91	3.70
FFMSW_2	0.88	0.47	0.26	*0.54*	1.61	1.32	2.93
BW_1	0.99	0.53	0.41	*0.64*	1.92	1.79	3.72
Centr_1	0.89	0.54	0.37	*0.60*	1.81	2.30	4.11
Centr_2	0.96	0.58	0.41	*0.65*	1.95	1.28	3.23
P uptake (mg pot^{-1})							
Ctrl-	1.93	1.60	1.02	*1.52*	4.55	2.25	6.80
Ctrl+	2.17	1.48	1.43	*1.70*	5.09	2.24	7.33
Agri_1	2.41	1.14	1.61	*1.72*	5.16	2.87	8.03
Agri_2	1.81	1.15	1.76	*1.58*	4.73	2.19	6.92
Agri_3	1.87	1.07	1.76	*1.57*	4.70	2.86	7.56
Sludge_1	2.28	1.23	2.70	*2.07*	6.21	2.96	9.17
Sludge_2	2.29	1.53	2.32	*2.05*	6.14	3.18	9.32
FFMSW_1	2.05	1.76	2.41	*2.07*	6.21	2.62	8.84
FFMSW_2	2.65	1.13	1.35	*1.71*	5.13	2.22	7.35
BW_1	2.76	1.40	2.38	*2.18*	6.54	2.39	8.93
Centr_1	2.77	1.36	2.26	*2.13*	6.39	2.58	8.97
Centr_2	2.67	1.53	2.31	*2.17*	6.51	2.25	8.76
N (mg pot^{-1})							
Ctrl-	25.01	9.32	6.52	*13.62*	40.85	12.44	53.28
Ctrl+	45.70	16.69	9.23	*23.87*	71.62	14.89	86.51
Agri_1	25.29	7.35	8.21	*13.62*	40.85	14.76	55.61
Agri_2	29.41	9.03	8.98	*15.81*	47.42	13.31	60.73
Agri_3	21.84	6.48	6.80	*11.70*	35.11	14.71	49.82
Sludge_1	36.91	10.51	12.39	*19.94*	59.81	15.13	74.93
Sludge_2	35.32	9.72	10.52	*18.52*	55.55	17.49	73.05
FFMSW_1	28.77	10.34	9.67	*16.26*	48.78	15.35	64.13
FFMSW_2	29.80	8.97	7.41	*15.39*	46.17	15.45	61.62
BW_1	32.03	9.97	11.88	*17.96*	53.88	16.51	70.39
Centr_1	30.74	9.28	10.00	*16.67*	50.02	18.45	68.47
Centr_2	34.63	10.39	12.16	*19.06*	57.18	14.04	71.22

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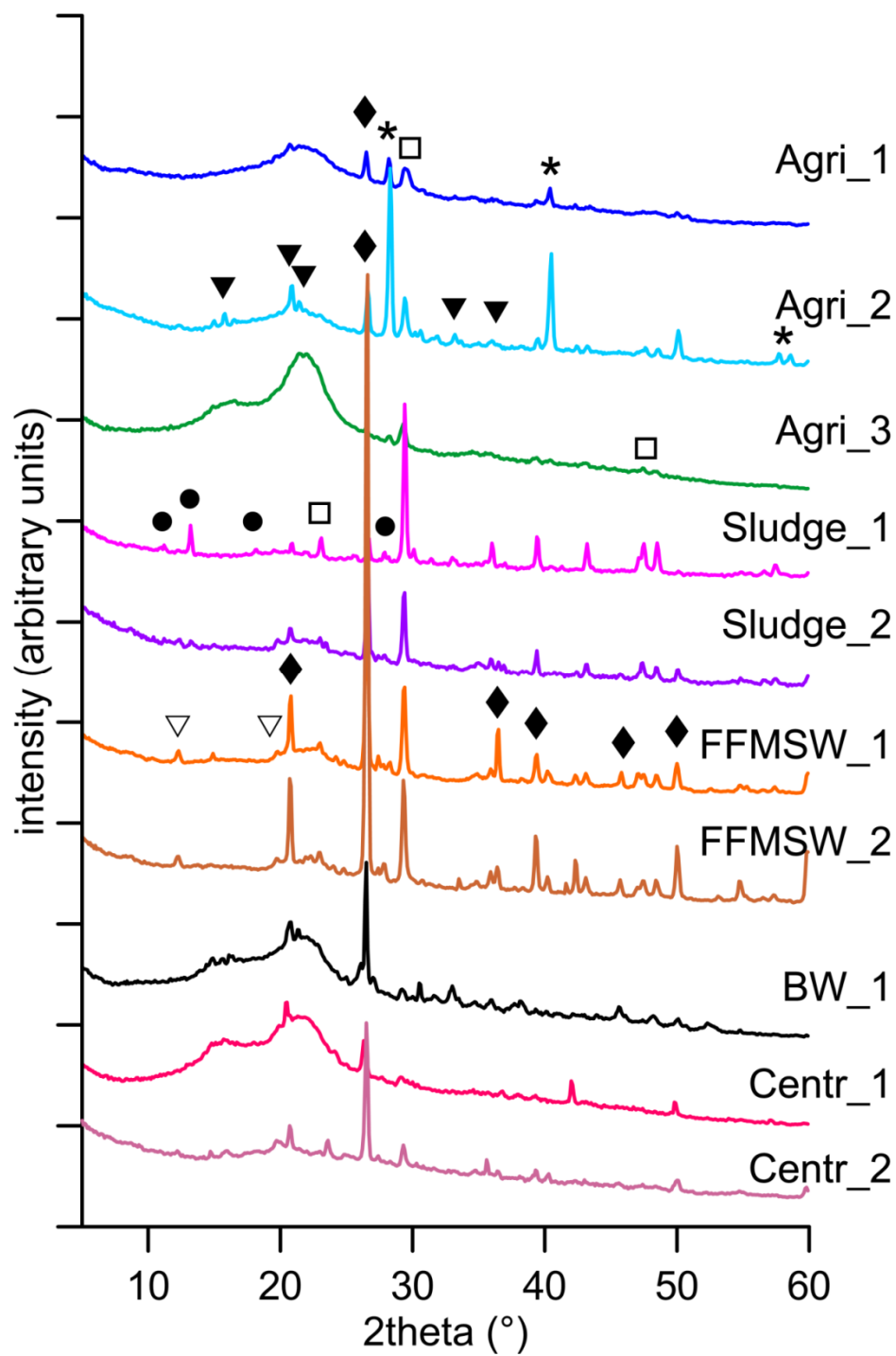
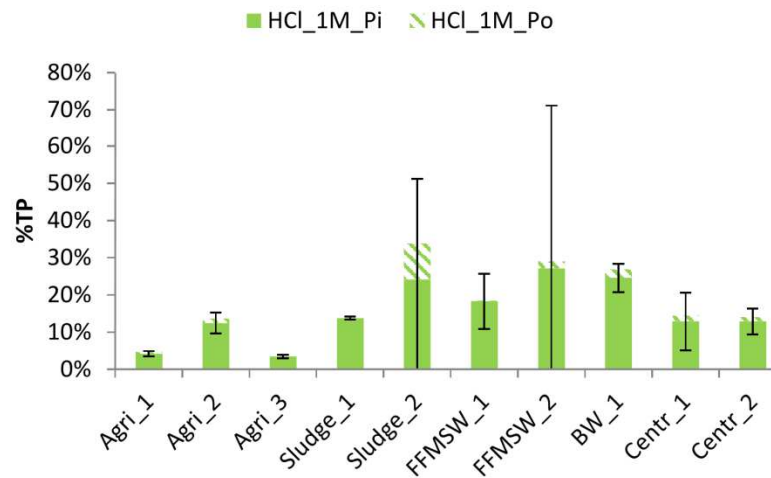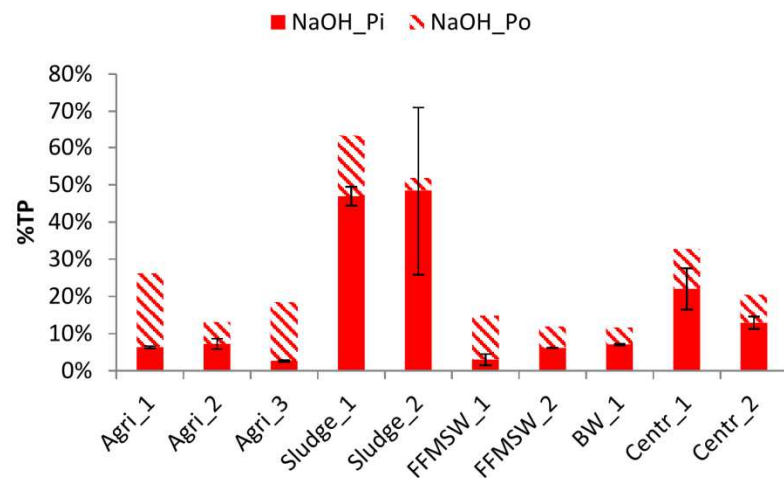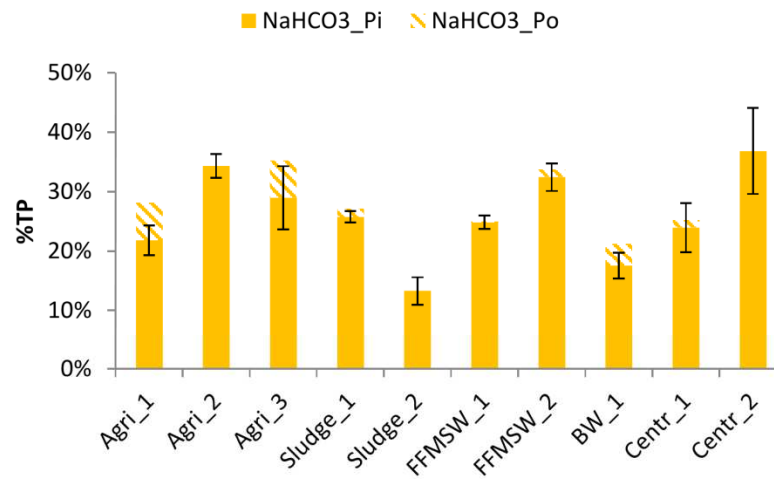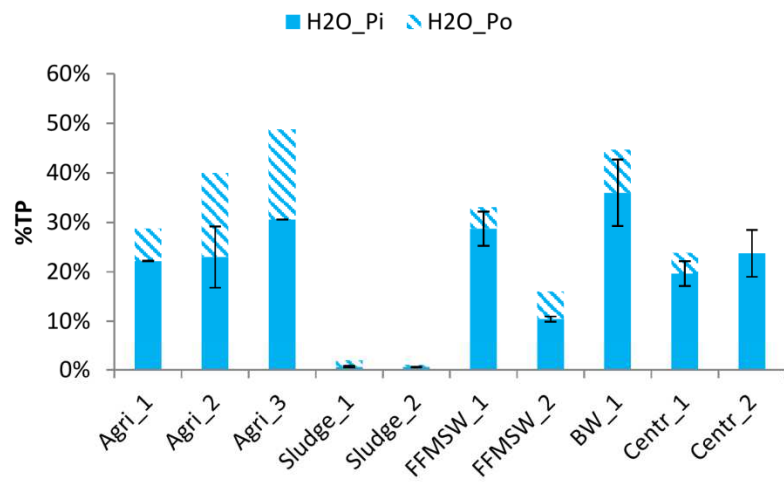Figure 1: X-ray diffraction of the digestates. The main peaks of different crystalline phases are identified by symbols: ▼ struvite; ♦ quartz; □ calcite; ● vivanite; * sylvine; ▽ kaolinite.
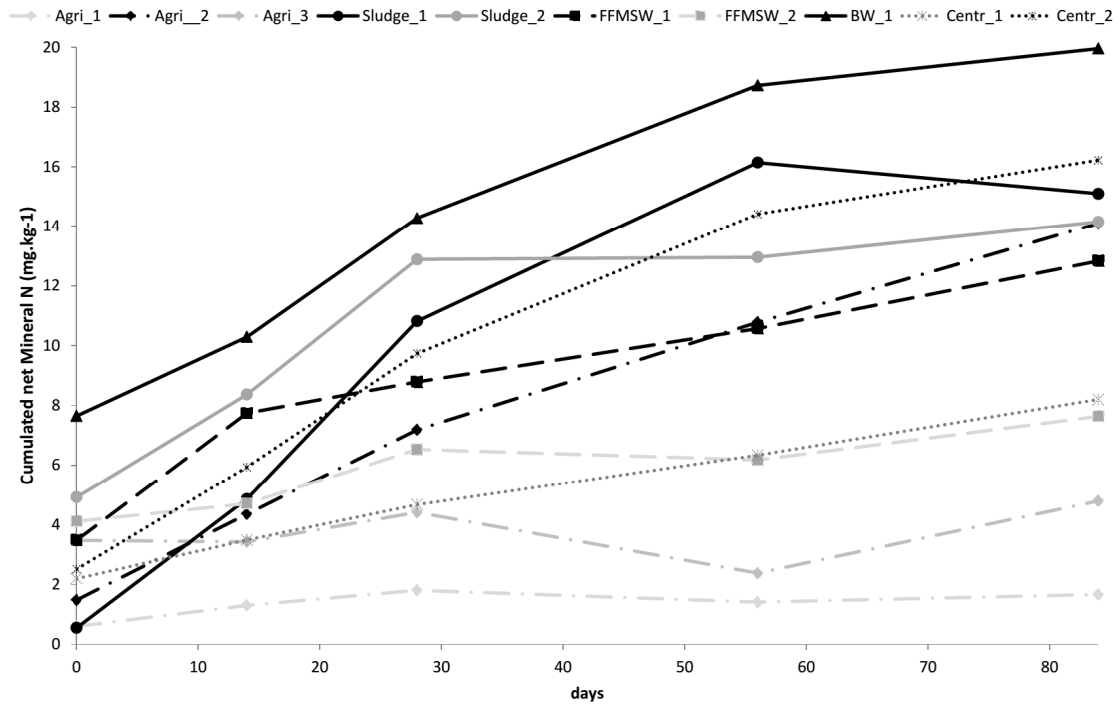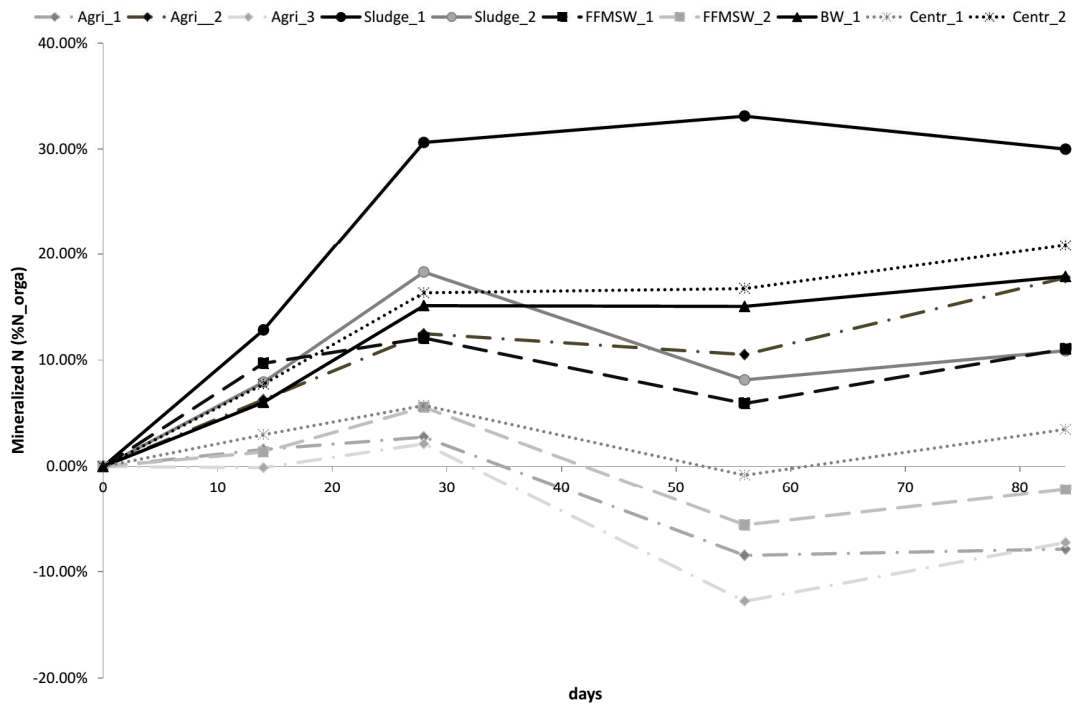
746

747   (a)

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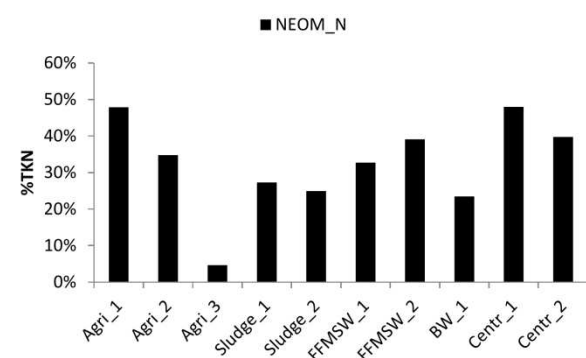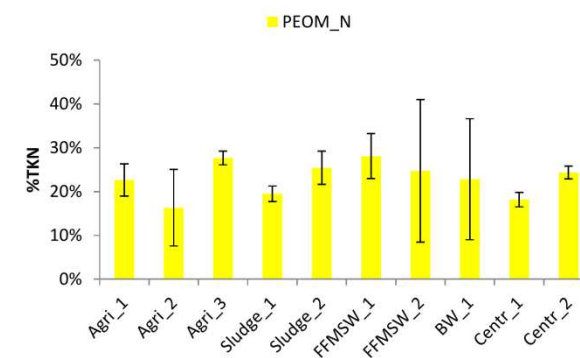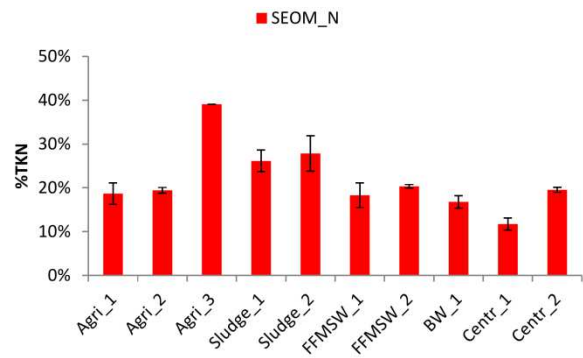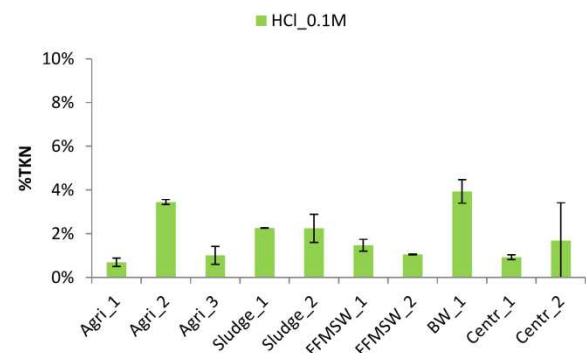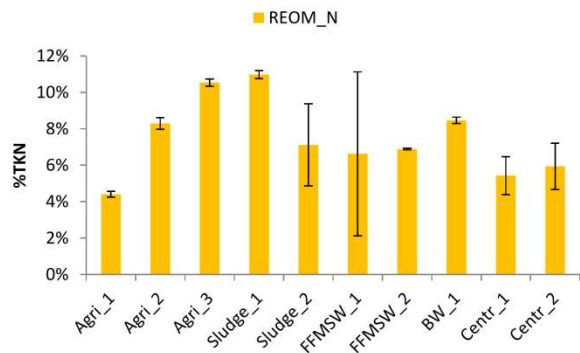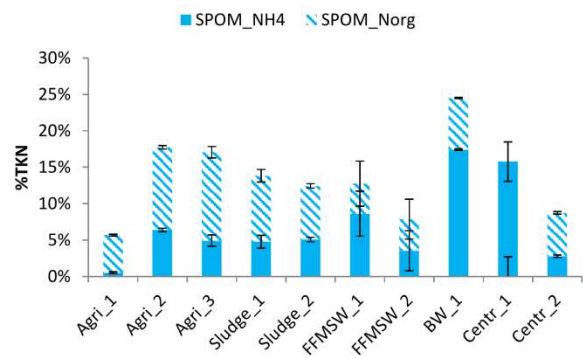
Figure 2: SCE fractions of TKN (a) and TP (b) obtained for all the digestates, according chemical accessibility. Error bars, SE (n=2)

751

752   (a)

753
754   (b)

Figure 3: Cumulated mineral nitrogen concentration evolution (a) and cumulated mineralized organic
nitrogen percentage (b) during soil incubation. Net values. *Black: high mineralized N group; Grey:
low mineralized N group (circle+full line: sludge digestates, square+large dashes: FFMSW
digestates, diamond+ dashes and dots line: Agrowastes digestates, triangle: biowaste digestate,
cross+dotted line: centralised digestates)*

760   (a)
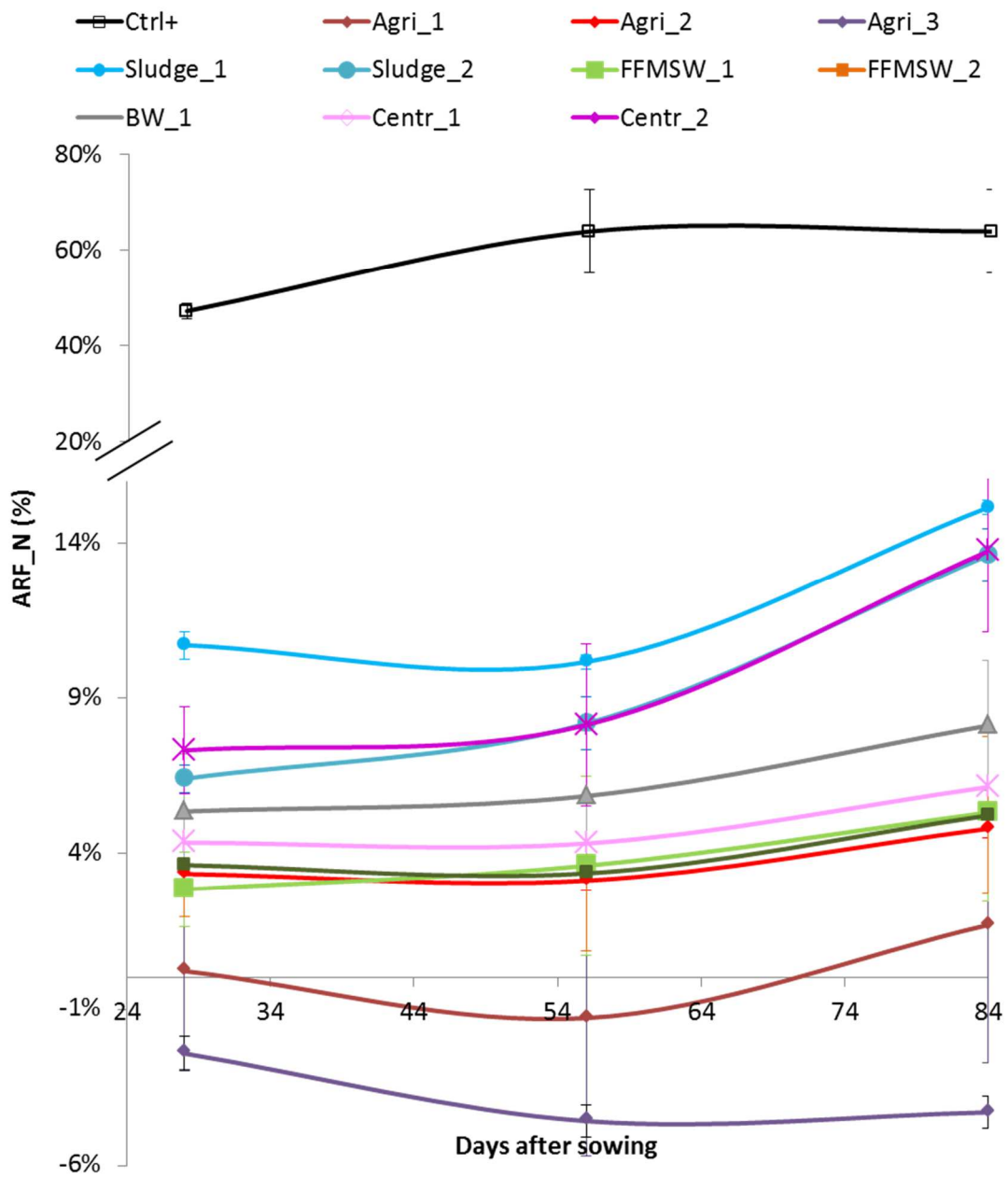
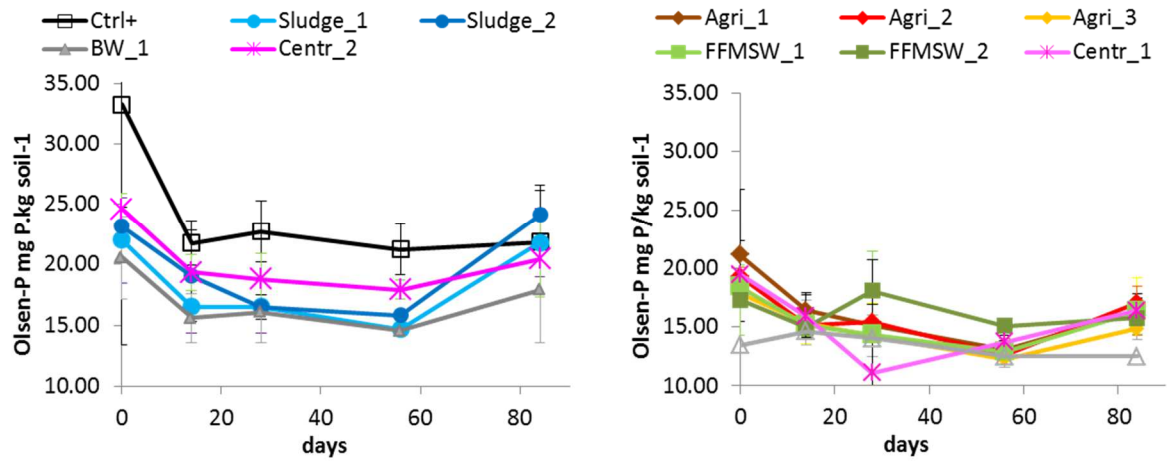761   (a)
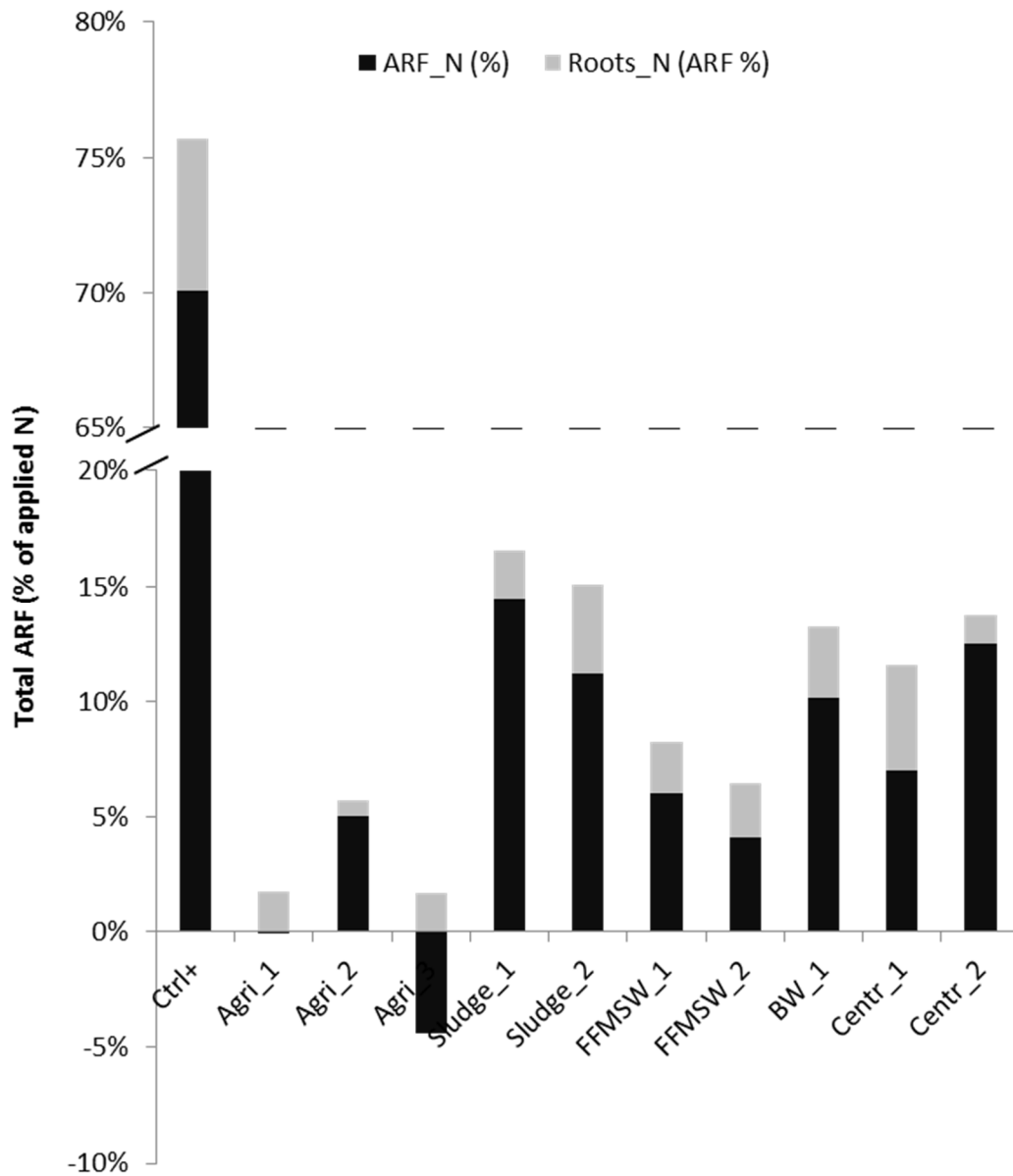
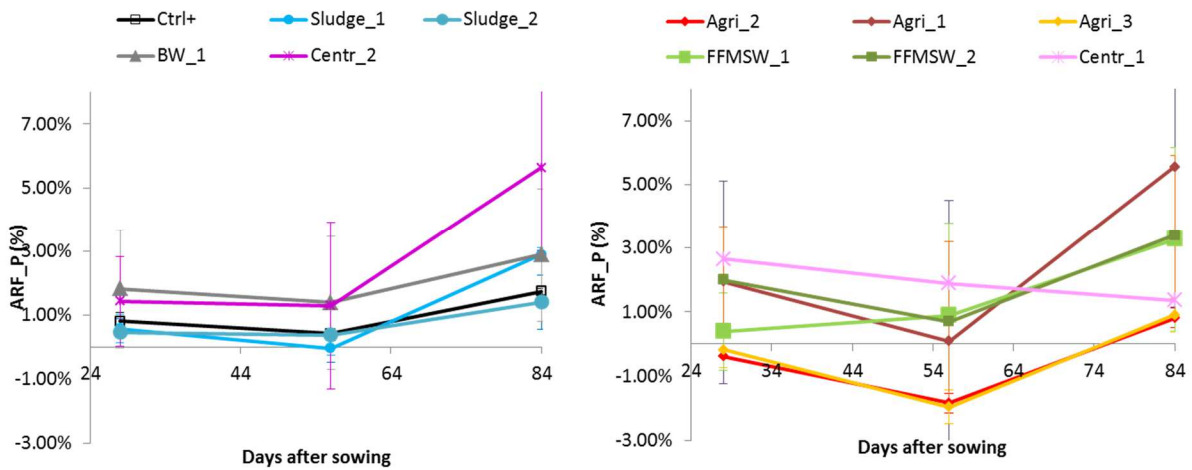762   (b)

763   Figure 4: Nitrogen apparent recovery fraction (ARF) of ryegrass shoots (a) and in both shoots and

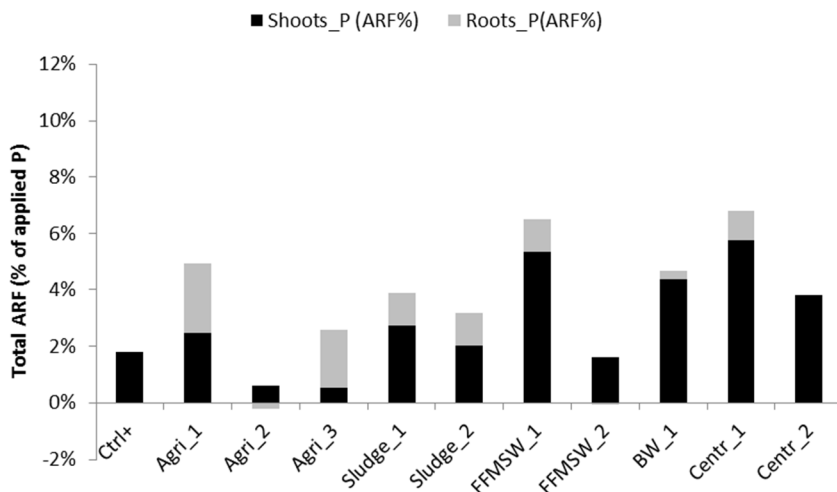764   roots (b) in three harvests during pot trial. Error bars, SE (n=3)

(a)

(b)

(c)

Figure 5: Olsen-P evolutions during the soil incubation of the tested digestates (a), cumulative phosphorous apparent recovery fraction (ARF) of ryegrass shoots (b) and in both shoots and roots (c) in three harvests during pot trial. Error bars, SE (n=3)