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Research



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# Non-stomatal processes reduce gross primary productivity in temperate forest ecosystems during severe edaphic drought

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Severe drought events are known to cause important reductions of gross primary productivity (*GPP*) in forest ecosystems. However, it is still unclear whether this reduction originates from stomatal closure (Stomatal Origin Limitation) and/or non-stomatal limitations (Non-SOL). In this study, we investigated the impact of edaphic drought in 2018 on *GPP* and its origin (SOL, NSOL) using a dataset of 10 European forest ecosystem flux towers. In all stations where *GPP* reductions were observed during the drought, these were largely explained by declines in the maximum apparent canopy scale carboxylation rate  $V_{\text{CMAX,APP}}$  (NSOL) when the soil relative extractable water content dropped below around 0.4. Concurrently, we found that the stomatal slope parameter ( $G_1$ , related to SOL) of the Medlyn *et al.* unified optimization model linking vegetation conductance and *GPP* remained relatively constant. These results strengthen the increasing evidence that NSOL should be included in stomatal conductance/photosynthesis models to faithfully simulate both *GPP* and water fluxes in forest ecosystems during severe drought.

This article is part of the theme issue ‘Impacts of the 2018 severe drought and heatwave in Europe: from site to continental scale’.

## 1. Introduction

With global climate change, droughts are likely to be more intense [1,2]. In 2018, a severe drought event occurred in Northern and central Europe causing forest fires and crop yield losses [3]. Europe experienced a major reduction of gross primary productivity (*GPP*) and transpiration (*E*) similarly to previous extreme events such as the 2003 Europe drought-heatwave [4] mostly because of soil water limitation [5,6]. Continuous measurements of ecosystem  $\text{CO}_2$  and water fluxes captured throughout Europe at eddy covariance (EC) flux tower stations thus provide a great large-scale ‘natural experiment’ to study the impact of drought on *GPP* and *E* [7].

There is increasing evidence that *GPP* reductions due to droughts could originate from both changes in stomatal behaviour (stomatal origin limitation, SOL) and non-stomatal traits (non-stomatal origin limitation, NSOL) [8–12]. Proposed NSOL mechanisms are reduced Rubisco activity (carboxylation rate) and/or electron transport activity [13], reduced active leaf area index (LAI) [13], reduced mesophyll conductance (including the intercellular airspace, cell walls, plasma membranes, cytoplasm and the chloroplast envelopes [14],  $g_m$ ) [15] or a combination of those [11,16]. The cause of *GPP* reduction is still subject to debate [17,18] and the modelling of SOL and NSOL and their (de)coupling is still poorly constrained by data. As a result, there is a strong need to examine whether different mechanisms are relevant and if models could be improved by developing more evidence-based functions for the impact of drought stress [19,20].

In leaf/canopy photosynthesis models, gross primary assimilation ( $A$ ) is very often modelled using the Farquhar *et al.* [21] photosynthesis model for C3 species [9,10,13]. In this model, Rubisco-limited photosynthesis (usually close to light saturation) is a function of the maximum carboxylation rate ( $V_{\text{cmax}}$ ) and the internal  $\text{CO}_2$  leaf concentration ( $C_i$ ) which implicitly considers that  $C_i$  is equal to the  $\text{CO}_2$  concentration in the chloroplasts ( $C_c$ ). As  $C_i$  cannot be measured directly, it is usually approximated by employing Fick's diffusion law through the stomata using a stomatal conductance ( $g_s$ ). This representation requires the determination of stomatal conductance by modelling. In this study, following Zhou *et al.* [10], we use the concept of apparent  $V_{\text{cmax}}$  ( $V_{\text{cmax,app}}$ ) recognizing that variations in  $V_{\text{cmax,app}}$  can result either from changes in the actual maximum rate of carboxylation or from changes in  $g_m$  which are not explicitly represented in this diffusion model. Consequently, when drought occurs, it impacts directly stomatal behaviour (closure) and then photosynthesis by limiting the diffusion of  $\text{CO}_2$  into the leaf which results in reduced  $C_i$  (SOL) or/and it impacts non-stomatal mechanisms (NSOL) which result in decreases of  $V_{\text{cmax,app}}$  [10,20].

A long standing stomatal conductance model from Cowan & Farquhar [22] states that stomata should act to maximize carbon gains while minimizing water losses (transpiration,  $E$ ), that is to maximize the integrated sum of  $A-E$  where  $\lambda$  ( $\text{mol C mol}^{-1} \text{H}_2\text{O}$ ) is the carbon cost of water gain  $\partial A/\partial E$  or marginal water-use efficiency [23] (note that we inverted the original expression). Medlyn *et al.* [24] proposed a reconciliation of the optimal stomatal behaviour theory [22] with empirical stomatal models linking  $g_s$  and  $A$ . Their work resulted in a unified stomatal optimization model (USO) with a form similar to former empirical expressions [25,26] (see equation (2.3)) where the slope between  $g_s$  and  $A^f$  ( $g_1$ ) is a key parameter (called the stomatal slope parameter).  $g_1$  is directly interpretable as inversely related to  $\lambda$  and to intrinsic water-use efficiency ( $i\text{WUE}$ ,  $A/g_s$ ) normalized by vapour pressure deficit ( $VPD$ ) and  $\text{CO}_2$  air concentration ( $C_a$ ) [27].

The USO model has been used both at the leaf level using leaf gas exchange data [28] and at the ecosystem level using EC flux observations [29] during non-water limited periods. During water limited periods, various responses of  $g_1$  (leaf level, SOL) to soil moisture were found [10] for a large range of species while a more consistent pattern of decreasing  $V_{\text{cmax,app}}$  was found. In a recent work, a good correlation between leaf scale and ecosystem scale  $g_1$  (or  $G_1$ ) response to soil moisture was found in a woodland dominated by *Acacia*

trees thereby demonstrating the ability of both leaf and ecosystem scale approaches to quantify drought effect [30].

In this study, we used the USO model combined with the Farquhar C<sub>3</sub> model (considering that  $C_i = C_c$ ) to study the origin of edaphic drought impacts on *GPP* (SOL and/or NSOL) in forest ecosystems using EC flux measurements by replacing leaf-level variables by their ecosystems analogues using a big leaf framework [27,31]. The surface conductance ( $G_s$  analogous to  $g_s$ ) was estimated by inverting the Penman–Monteith equation [32]. We then inferred the bulk stomatal slope parameter ( $G_1$  analogous to  $g_1$ ) and the maximum apparent carboxylation rate of the ecosystem ( $V_{\text{cmax,APP}}$ ) [33] at a daily time step for each ecosystem. The study was restricted to the growing period excluding any autumn senescence or spring leaf emergence influence on the variation of  $V_{\text{cmax,APP}}$ .

In addition, drought intensity was quantified using the relative extractable water (*REW*) as proposed by Granier *et al.* [5], which is a normalized index of soil water deficit varying from 0 to 1 that allows for edaphic status inter-site comparisons. This index was used in previous studies [5,34] and, based on their results, we hypothesize that both  $E$  and *GPP* reductions will occur when *REW* falls below  $\approx 0.4$ .

The objective of this work is to examine the response of  $G_1$ , as a measure for SOL, and  $V_{\text{cmax,APP}}$  as a measure for NSOL, to soil water deficit using EC data collected in forests during the 2018 European drought. More specifically, we intend to answer the following questions: (1) how was *REW* impacted by the drought in forest sites in 2018? (2) Can we confirm the *REW* threshold of  $\approx 0.4$  for *GPP* reductions found in previous studies [5,34]? (3) To what degree did SOL and NSOL impact *GPP* during the drought? (4) What were  $G_1$  and  $V_{\text{cmax,APP}}$  responses to *REW* function shapes and how did these responses vary across sites?

## 2. Material and methods

### (a) Site and data description

Data have been processed by the Ecosystem Thematic Centre of the Integrated Carbon Observation System (ICOS) and form the 2018 drought ICOS/Fluxnet dataset [35], which is a compilation of EC fluxes, meteorological and edaphic data during the 2018 European drought at half-hourly resolution. Only sites with a sufficiently resolved vertical profile of soil water content sensors were selected. The main site characteristics are summarized in table 1. Flux data followed the standard FLUXNET processing [46], including friction velocity ( $u_*$ ) filtering [47] and *GPP* determination by night-time flux partitioning [48]. Only data marked with highest quality flags were used for this study. Latent heat fluxes were not corrected for energy balance closure.

### (b) Quantification of drought

The intensity of edaphic drought was quantified by computing the *REW* content [5] at each time step and for the entire root depth using:

$$REW = \frac{\sum_i ((SWC_i - SWC_{WP,i}) / (SWC_{FC,i} - SWC_{WP,i})) \Delta h_i}{h_{\text{max}}}, \quad (2.1)$$

where  $i$  is the index of each soil layer over the rooting depth,  $SWC_i$  is the actual soil water content,  $SWC_{WP}$  is the soil water content at the wilting point,  $SWC_{FC}$  is the soil water content at field capacity,  $\Delta h_i$  is the thickness of each layer and  $h_{\text{max}}$  is the maximum rooting depth. Each soil horizon was divided into

**Table 1.** Main characteristics of the flux tower sites included in this study. The LAI corresponds to the maximum LAI typically observed at the sites.

site ID	country	latitude	longitude	dominating species	LAI m <sup>2</sup> m <sup>-2</sup>	soil texture	rooting depth m	ref
BE-Bra	Belgium	51.308	4.52	<i>Pinus sylvestris</i>	3	sand	1	[36]
BE-Vie	Belgium	50.305	5.998	<i>Fagus sylvatica</i> / <i>Pseudotsuga menziesii</i>	5	silty clay loam	1.4	[37]
CZ-Lnz	Czechia	48.682	16.946	<i>Quercus robur</i> / <i>Fraxinus angustifolia</i> / <i>Carpinus betulus</i> / <i>Tilia cordata</i>	6.5	sandy loam	1.2	[38]
CZ-Raj	Czechia	49.444	16.697	<i>Picea abies</i>	5	sandy loam	0.7	[39]
CZ-Stn	Czechia	49.036	17.97	<i>Fagus sylvatica</i>	5.5	sandy loam	0.7	[40]
DE-Hai	Germany	51.0792	10.453	<i>Fagus sylvatica</i>	6	clay loam	0.7	[41,42]
DK-Sor	Denmark	55.486	11.645	<i>Fagus sylvatica</i>	5	sandy clay loam	1	[43]
FR-Bil	France	44.494	-0.956	<i>Pinus pinaster</i>	2.5	sand	1.1	[44]
FR-Hes	France	48.674	7.065	<i>Fagus sylvatica</i>	6.5	silty clay loam	1.6	[45]
IT-Sr2	Italy	43.732	10.291	<i>Pinus pinea</i>	2.5	sand	1.2	

soil layers corresponding to the number of sensors installed in the horizon. The layer boundaries were the horizon limits or the midway point between two sensors. Soils related data are summarized in the supplements (electronic supplementary material, table S1). For each layer,  $SWC_{WP}$  and  $SWC_{FC}$  were estimated using soil retention curves based on either measurements (by research teams) or modelling (based on soil textures) and checked for consistency with  $SWC_{48h}$  after a rain event for  $SWC_{FC}$  and with minimum  $SWC$  values observed at the site for  $SWC_{WP}$  to avoid negative  $REW$  values. When not available, the maximum depth was defined as the bedrock depth [5]. When data were available (BE-Vie and BE-Bra),  $REW$  was corrected for the coarse fraction by applying a correction factor for each layer. According to Granier *et al.* [5,34] it is expected that both  $GPP$  and  $G_s$  start to decrease when  $REW$  drops below  $\approx 0.4$ . The evolution of  $REW$  at each site in 2018 is presented in the electronic supplementary material, figure S1.

### (c) Canopy surface variables

Detailed computation procedures for canopy surface variables are fully described in Knauer *et al.* [31]. First the aerodynamic conductance to water transfer ( $G_{aw}$ ) was computed as a combination of an aerodynamic conductance to momentum (first term) and a boundary layer conductance (second term) as follows:

$$G_{aw} = \left( \frac{u_*^2}{u(z)} + (6.2u_*^{-0.67})^{-1} \right)^{-1}, \quad (2.2)$$

where  $u_*$  is the friction velocity ( $m\ s^{-1}$ ), and  $u(z)$  the wind speed at measurement height ( $z$ ). Canopy surface conductance for water ( $G_s$ ,  $m\ s^{-1}$ ) was computed by inverting the Penman–Monteith equation [32]:

$$G_s = \frac{LE G_{aw} \gamma}{s(R_n - G - S) + \rho c_p G_{aw} VPD_a - \lambda E(s + \gamma)}, \quad (2.3)$$

where  $LE$  is the latent heat flux ( $W\ m^{-2}$ ),  $\gamma$  is the psychrometric constant ( $Pa\ K^{-1}$ ),  $s$  is the slope of the saturation vapour pressure curve at air temperature ( $Pa\ K^{-1}$ ),  $R_n$  is the net radiation ( $W\ m^{-2}$ ),  $G$  is the ground heat flux ( $W\ m^{-2}$ ),  $S$  is the sum of all storage terms ( $W\ m^{-2}$ ),  $c_p$  is the heat capacity of dry air ( $1005\ J\ kg^{-1}\ K^{-1}$ ) and  $VPD_a$  is the VPD of ambient air ( $P_a$ ).  $G$  was considered negligible when not available while  $S$  was not available and was set to 0 at all sites.

The  $CO_2$  concentration at the canopy surface ( $C_s$ ), needed in the USO and diffusion equations, was computed as:

$$C_s = C_a + \frac{NEE}{(G_{aw}/1.32)}, \quad (2.4)$$

where  $C_a$  ( $\mu mol\ CO_2\ mol^{-1}$ ) is the  $CO_2$  air concentration at the measurement height,  $NEE$  ( $\mu mol\ CO_2\ m^{-2}$ ) is the net  $CO_2$  ecosystem exchange and the factor 1.32 is the ratio of diffusivities of  $CO_2$  and water vapour in the boundary layer. The VPD at the canopy surface ( $VPD_s$ , Pa) was also computed (see [31] for more details).  $G_s$  is a good predictor of bulk stomatal conductance only when evaporation is small compared to transpiration; data collected during a period of 48 h following a rain event were discarded. Secondly, the analysis was restricted to the growing season, avoiding senescence and leaf emergence periods. We defined this period as the days when the daily  $GPP$  (averaged over all the available years) smoothed with a 15 days moving average window was higher than 70% of the 95th percentile of the daily  $GPP$  distribution.  $G_s$  data were also filtered excluding half hour with  $LE < 0$  or  $R_n < 0$ . Negative  $G_s$  values were filtered and  $G_s$  outliers were also discarded by removing data when  $G_s$  were higher than the 98th percentile of the  $G_s$  distribution.

### (d) Stomatal origin limitations

Similarly to previous work [10,49], reductions of  $GPP$  originating from SOL were assessed by analysing dependence on  $REW$  of the  $G_T$  parameter used in the USO model developed by Medlyn *et al.* [24] but adapted to the ecosystem scale using bulk ecosystem parameters [27]:

$$G_s = G_0 + 1.6 \left( 1 + \frac{G_T}{\sqrt{VPD_s}} \right) \frac{GPP_{high}}{C_s}, \quad (2.5)$$

where,  $GPP_{high}$  is the  $GPP$  at high radiation ( $R_g > 500\ W\ m^{-2}$ ) and replaces net assimilation in the original leaf scale expression of the model,  $G_s$  replaces stomatal conductance and leaf surface variables were replaced by their corresponding canopy surface values [50] ( $C_s$  the air  $CO_2$ ,  $VPD_s$ ). In this expression, the nocturnal stomatal conductance  $G_0$  was set to 0 as its magnitude can be considered negligible when compared to the other terms at saturating daylight conditions [27].  $G_T$ , the slope parameter, is a physiologically meaningful parameter as it was shown to be inversely related to  $\lambda$  [24] and to  $iWUE$  [27].

$G_1$  was obtained by inverting equation (2.5) using half-hourly measurements. Because leaf respiration was neglected in equation (2.5), the equation was inverted only for high radiation data so that  $GPP_{\text{high}}$  should be much higher than leaf respiration. Negative  $G_1$  values were filtered and outliers were also discarded by removing data when absolute  $G_1$  was two times higher than the average absolute deviation from the median. Finally, daily  $G_1$  averages were then computed for days with at least five valid half-hourly values.

The response of  $G_1$  to  $REW$  was fitted with a segmented linear response curve (two segments) in order to test the presence of an  $REW$  threshold (break point) above which  $G_1$  is constant (no effect range) and under which stomatal regulation occurs ( $G_1$  decreases or increases) [51]. After a first fit, outliers were removed by exclusion of the  $G_1$  value having absolute residuals more than 2.2 times the standard deviation of the residuals distribution. A second fit was then done. Parameters obtained from a second fit were  $G_1^*$ , i.e. the average  $G_1$  value within the ‘no effect range’, the break point  $REW$  value ( $REW_{B,G_1}$ ) and the slope/intercept of the  $G_1$  decrease/increase. The presence of the break point was further tested by comparing the residuals of the model to those of a simple linear regression model using an  $F$ -test.

### (e) Non-stomatal origin limitations

Reductions of  $GPP$  originating from NSOL were studied by assessing the effect of water stress on apparent bulk  $V_{\text{Cmax}}$  hereafter called  $V_{\text{Cmax,APP}}$ . It was obtained by inverting the expression of Rubisco-limited photosynthesis during high radiation conditions [21]:

$$V_{\text{Cmax,APP}} = \frac{GPP_{\text{high}}(C_i + K_m)}{(C_i - \Gamma^*)}, \quad (2.6)$$

where  $V_{\text{Cmax,APP}}$  is expressed per  $\text{m}^2$  of soil and not of leaf as usual.  $K_m$  is the effective Michaelis Menten coefficient kinetics and  $\Gamma^*$  is the  $\text{CO}_2$  compensation point, which were both computed using temperature responses following Bernacchi *et al.* [52],  $GPP_{\text{high}}$  is the  $GPP$  at  $R_g > 500 \text{ W m}^{-2}$  while  $C_i$  was computed using Fick’s diffusion law [53]:

$$C_i = C_s - \frac{GPP_{\text{high}}}{(G_s/1.6)}. \quad (2.7)$$

Note that in equations (2.5)–(2.7), leaf respiration was neglected which should have a small effect on the results as, at high radiation, leaf respiration should be much smaller than  $GPP$ . Half-hourly values of  $V_{\text{Cmax,APP}}$  were then normalized for temperature to  $25^\circ\text{C}$  using an Arrhenius equation [31] fitted for each decile of  $REW$  as  $V_{\text{Cmax,APP}}$  response to temperature was found to decrease under drought conditions. Finally,  $V_{\text{Cmax,APP}}$  was averaged on a daily basis and days with less than 5 half-hourly values were discarded. Considering the way we estimated  $V_{\text{Cmax,APP}}$  a decrease in  $V_{\text{Cmax,APP}}$  indicates NSOL of  $GPP$  including either changes in mesophyll conductance or in actual  $V_{\text{Cmax}}$  or other processes limiting  $GPP$  (apart from stomatal closure). The response of  $V_{\text{Cmax,APP}}$  to  $REW$  was assessed using the same segmented linear regression model as explained in the previous section.  $V_{\text{Cmax,APP}}^*$  and  $REW_{B,V_{\text{Cmax}}}$  were defined as the average of  $V_{\text{Cmax,APP}}$  values (normalized at  $25^\circ\text{C}$ ) within the no effect range and the break point  $REW$  value, respectively.

### (f) Degree of stomatal and non-stomatal origin limitation

To illustrate the degree of SOL and NSOL due to edaphic drought, following Zhou *et al.* [10], a sensitivity analysis was performed to explore the impact of drought-induced changes in  $G_1$  and  $V_{\text{Cmax,APP}}$  on  $GPP$ . The impact of NSOL was assessed by comparing measured  $GPP$  to a theoretical non-affected value corresponding to a ‘modelled’  $GPP$  computed using inverted equation (2.5) with  $V_{\text{Cmax,APP}}^*$  instead of  $V_{\text{Cmax,APP}}$  and a  $C_i$  obtained from equation (2.6) using observed  $G_s$  values (equation (2.2)) and measured  $GPP$

values. Similarly, the impact of SOL was assessed by comparing measured  $GPP$  to ‘modelled’ unaffected  $GPP$  computed using constant  $G_1^*$  values and observed  $V_{\text{Cmax,APP}}$  values. The degree of limitation (DoL) was computed as the ratio of modelled ‘unaffected’  $GPP$  against measured  $GPP$  and represents the factor by which  $GPP$  was divided because of SOL or NSOL.

## 3. Results

### (a) $G_s$ and $GPP_{\text{high}}$

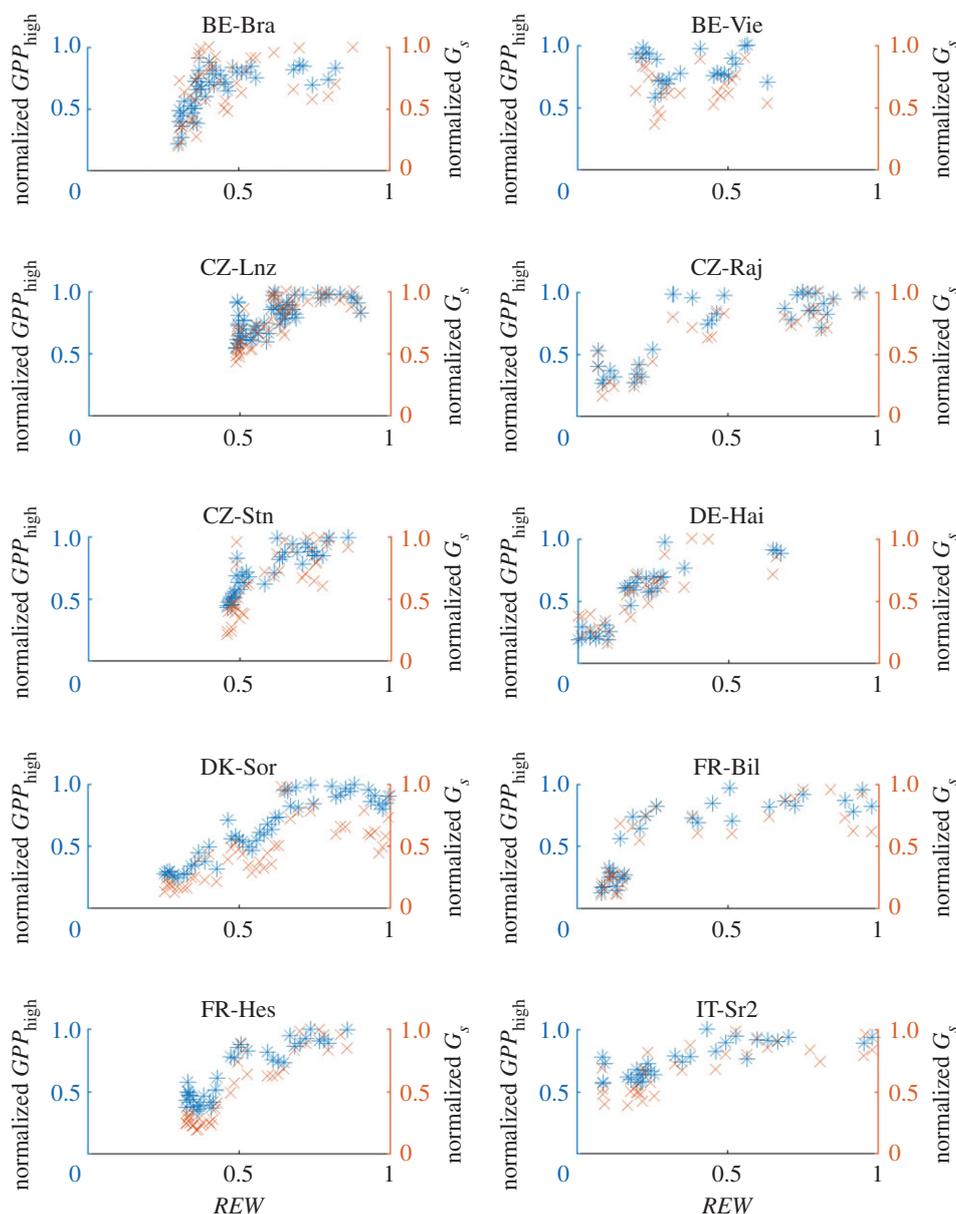
$GPP_{\text{high}}$  and  $G_s$  (for  $R_g > 500$ ) normalized by their respective maximum values in relation with  $REW$  are presented in figure 1. At all sites, we can observe that both  $GPP_{\text{high}}$  and  $G_s$  behave similarly. High values of both  $GPP_{\text{high}}$  and  $G_s$  were observed for high  $REW$  while both variables decreased simultaneously with  $REW$ . There was an exception at BE-Vie where such a pattern was not observed although both variables still behaved similarly. The lowest  $GPP_{\text{high}}$  and  $G_s$  values were observed at sites such as CZ-Raj, FR-Bil and DE-Hai where very low  $REW$  values (lower than 0.15) were reached. At IT-Sr2,  $G_s$  and  $GPP_{\text{high}}$  were still quite high (around half of maximum values) even for very low  $REW$  values most probably because, in this sandy soil, rooting depth was probably deeper than the deepest available SWC sensor (1.2 m, see electronic supplementary material, table S1) which caused an underestimation of  $REW$ .

### (b) Response of $G_1$ to edaphic drought

$G_1$  was found to be constant at all sites apart from DE-Hai even for sites where  $REW$  values lower than 0.4 were observed (figure 2). In BE-Bra  $G_1$  seems to be enhanced at low  $REW$  but the segmented model did not perform significantly better than the linear one. In DE-Hai,  $G_1$  was found to increase when  $REW$  dropped below a very low value of 0.2, which is quite close to the wilting point. Such low  $REW$  were also observed during the growing season at CZ-Raj and IT-Sr2 but no similar behaviour was observed. The lowest  $G_1^*$  ( $1.5 \text{ kPa}^{0.5}$  at CZ-Raj) was three times lower than the highest value ( $4.5 \text{ kPa}^{0.5}$  at FR-Bil, table 2).

### (c) Response of $V_{\text{Cmax,APP}}$ to edaphic drought

The effect of NSOL caused by drought was studied by analysing the dependence of the temperature normalized  $V_{\text{Cmax,APP}}$  values on  $REW$  (figure 3). At all sites that experienced low  $REW$  conditions (below  $\approx 0.4$ ) apart from BE-Vie and IT-Sr2, constant  $V_{\text{Cmax,APP}}$  were observed for large  $REW$  values followed by a decrease when  $REW$  declined below an  $REW_{B,V_{\text{Cmax}}}$  threshold. The  $REW_{B,V_{\text{Cmax}}}$  were not significantly different (according to the confidence intervals) than the value of 0.4 which was found in previous studies, with an exception at DK-Sor where  $REW_{B,V_{\text{Cmax}}}$  was higher ( $REW_{B,V_{\text{Cmax}}} = 0.8 \pm 0.09$ , table 2). The high  $REW_{B,V_{\text{Cmax}}}$  observed at DK-Sor might result from an overestimation of  $REW$  as the shallowest available SWC probe was at 15 cm depth (see electronic supplementary material, table S1) and was not able to catch the beginning of the progressive drying of the upper layers that contain a large amount of roots. The most impacted site was DE-Hai where  $REW$  almost reached the wilting point ( $REW \approx 0$ ) with very low  $V_{\text{Cmax,APP}}$  values ( $\approx 15 \mu\text{mol m}^{-2} \text{ s}^{-1}$ ) probably because of shallow soil and rooting depth (0.6 m).



**Figure 1.** Dependence of  $GPP_{high}$  (blue, left axis) and  $G_s$  at high radiation ( $R_g > 500$ , red, right axis) normalized by their maximum value on  $REW$  for each site. (Online version in colour.)

#### (d) Degree of stomatal and non-stomatal limitation

DoL reached values of 5 for NSOL at DE-Hai while it remained close to 1 for SOL at all sites (figure 4). This analysis therefore confirms that, at all sites, NSOL as the dominant mechanism. As a result, reducing  $V_{CMAX,APP}$  while maintaining  $G_1$  constant could capture the variations of both  $GPP_{high}$  and  $G_s$  with drought (identical conclusions are obtained if we focus the analysis on  $G_s$  instead of  $GPP$ , data not shown). It is also worthwhile noticing that the increasing  $G_1$  observed at DE-Hai (and at BE-Bra to a lesser extent) did not lead to important changes in  $GPP$ .

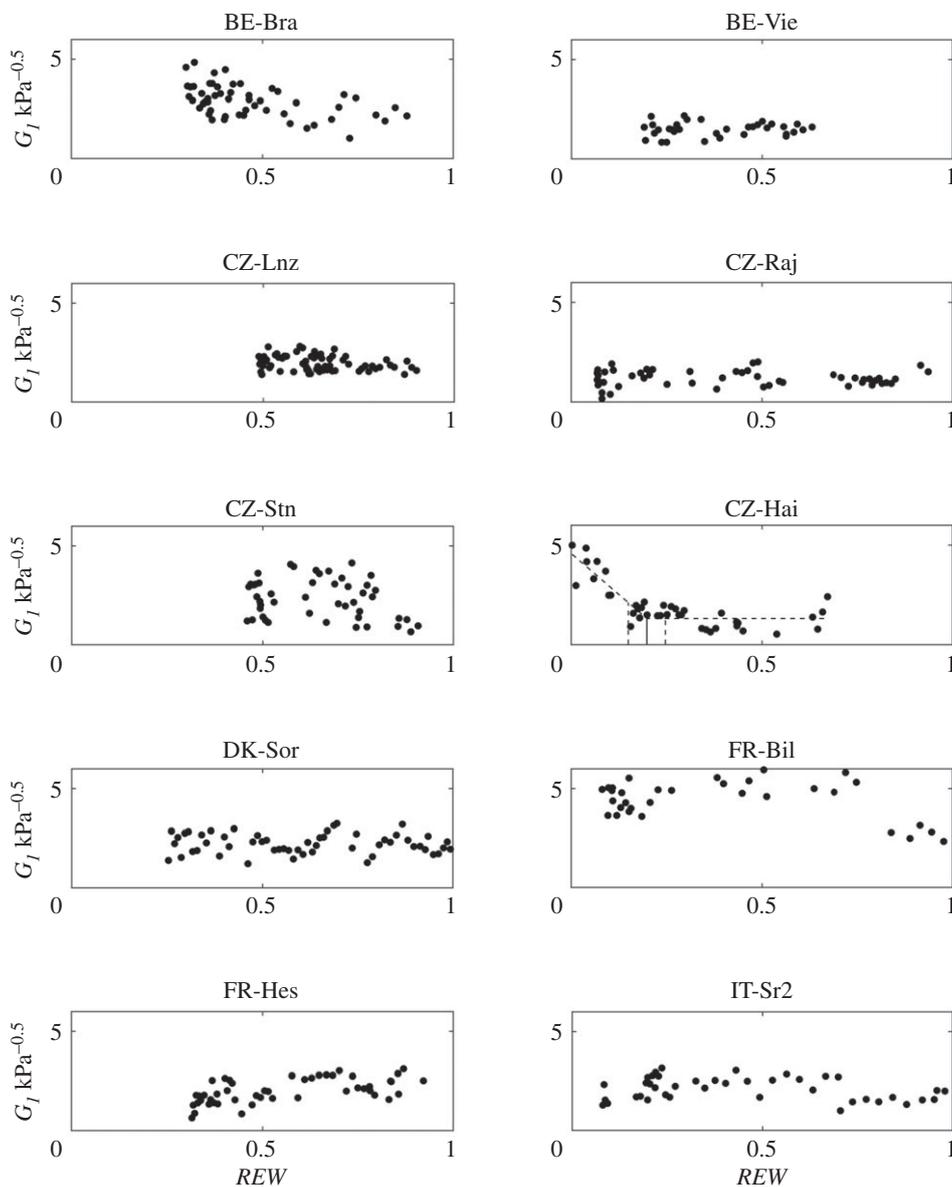
## 4. Discussion

### (a) Methodological limitations

Although the responses of  $V_{CMAX,APP}$  and  $G_1$  to  $REW$  were relatively consistent, some sites showed unexpected behaviours. For example,  $REW_{B,VC_{MAX}}$  at DK-Sor was much

higher than expected (0.85) because some SWC sensors experienced failures during the drought. At IT-Sr2, no limitation of  $GPP$  was found although very low  $REW$  values were estimated, probably because the SWC sensor profile was not deep enough to capture the whole rooting depth. Multiple and deeper sensor profiles (with matching wilting points and field capacities) would certainly help to reduce these uncertainties. Complementary measurements such as predawn leaf water potential and soil matric potential [17,54], when  $REW$  approaches values close to 0.4 at the site, would also be useful.

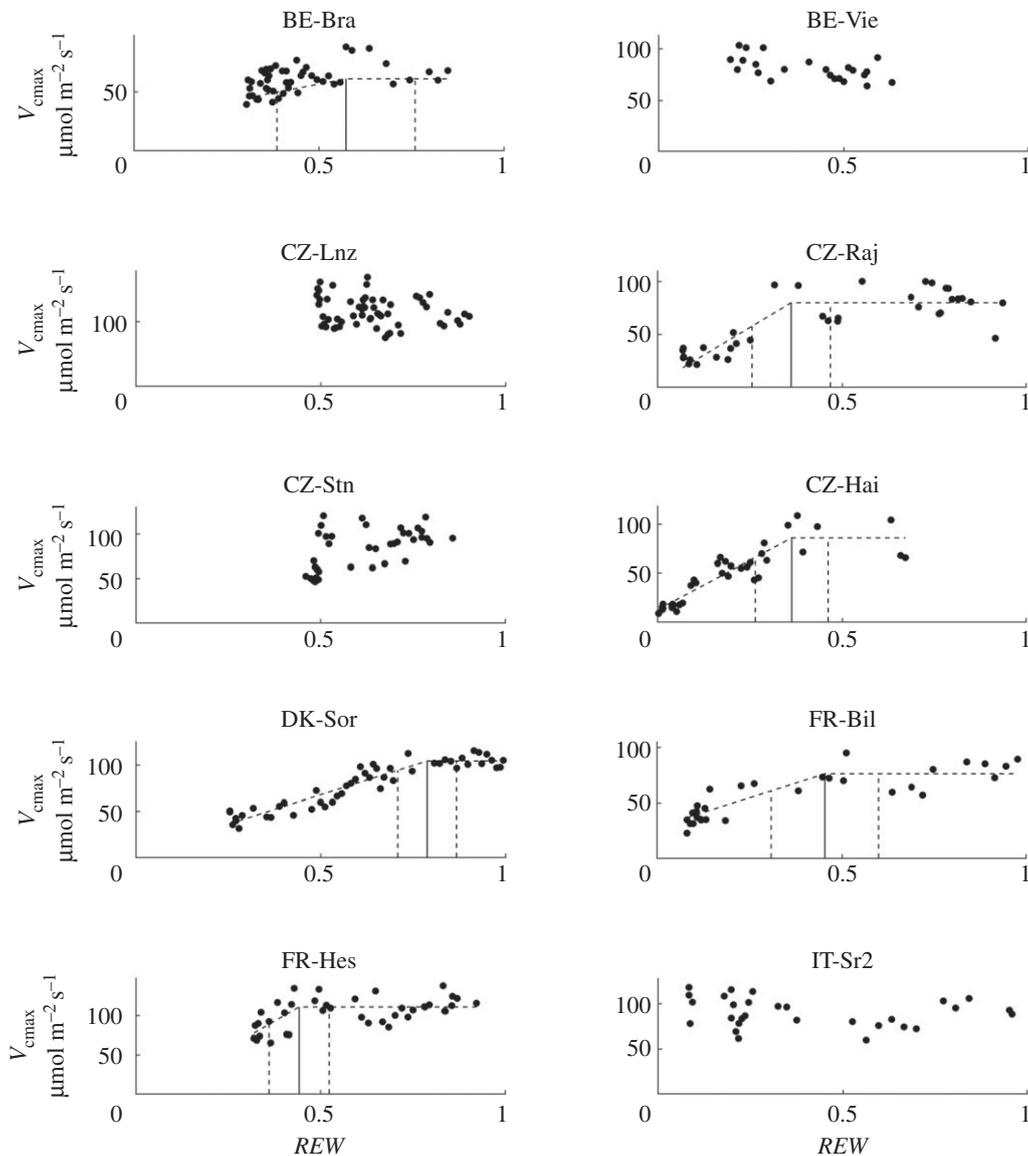
The big leaf approach used in this study also has several limitations [31] which could be critical when comparing leaf scale-derived parameters to big leaf canopy scale estimates [27,29,31] or when attributing a behaviour to a specific species. First, the approach is only able to derive bulk parameters and is unable to distinguish the vertical and horizontal distribution of the properties. Horizontal heterogeneity is especially crucial at mixed forest sites where different species could show different responses to drought (and different root depth and therefore  $REW$ ) which would blur their



**Figure 2.** Dependence of  $G_1$  on  $REW$  for each site.  $REW_{b,G_1}$  are marked by a vertical solid line with 95% confidence intervals (dashed vertical lines). Regression lines are only shown when  $F$ -test  $p$ -values (comparison with linear regression) were smaller than 0.05.

**Table 2.** Maximum extractable water (EW), minimum observed  $REW$  in 2018 during the growing season,  $REW_{b,VC_{MAX}}$  and  $REW_{b,G_1}$  ( $REW$  break points for  $V_{C_{MAX,APP}}$  and  $G_1$ , respectively) given with 95% confidence intervals.  $V_{C_{MAX,APP}}^*$  and  $G_1^*$  ( $V_{C_{MAX,APP}}$  and  $G_1$  values in unstressed conditions) given with 95% confidence intervals.  $P$ -values are given for the  $F$ -test comparing the segmented mode (three parameters) to the linear model (two parameters).

site ID	EW mm	min $REW$ —	$V_{C_{MAX,APP}}^*$ $\mu\text{mol m}^{-2} \text{s}^{-1}$	$REW_{b,VC_{MAX}}$ —	$p$ -value —	$G_1^*$ $\text{kPa}^{-0.5}$	$REW_{b,G_1}$ —	$p$ -value —
BE-Bra	133	0.30	$58 \pm 3$	$0.57 \pm 0.19$	<0.05	$2.9 \pm 0.2$	—	0.28
BE-Vie	215	0.19	$81 \pm 5$	—	1	$1.5 \pm 0.1$	—	1
CZ-Lnz	241	0.49	$116 \pm 10$	—	1	$2.0 \pm 0.2$	—	1
CZ-Raj	92	0.07	$80 \pm 7$	$0.36 \pm 0.11$	<0.001	$1.2 \pm 0.8$	—	1
CZ-Stn	236	0.46	$120 \pm 7$	—	1	$2.3 \pm 0.3$	—	1
DE-Hai	143	0.00	$86 \pm 17$	$0.36 \pm 0.07$	<0.001	$1.3 \pm 0.2$	$0.2 \pm 0.05$	<0.001
DK-Sor	176	0.25	$104 \pm 3$	$0.80 \pm 0.09$	<0.001	$2.2 \pm 0.2$	—	1
FR-Bil	159	0.08	$76 \pm 7$	$0.45 \pm 0.15$	<0.05	$4.5 \pm 0.4$	—	1
FR-Hes	338	0.33	$121 \pm 13$	$0.44 \pm 0.08$	<0.05	$2.0 \pm 0.5$	—	1
IT-Sr2	108	0.08	$89 \pm 6$	—	1	$2.1 \pm 0.2$	—	1



**Figure 3.** Dependence of  $V_{\text{CMAX,APP}}$  normalized at 25°C on  $REW$  for each site.  $REW_{B,VC_{\text{MAX}}}$  are marked by a vertical solid line with 95% confidence intervals (dashed vertical lines). Regression lines are only shown when  $F$ -test  $p$ -values (comparison with linear regression) were smaller than 0.05.

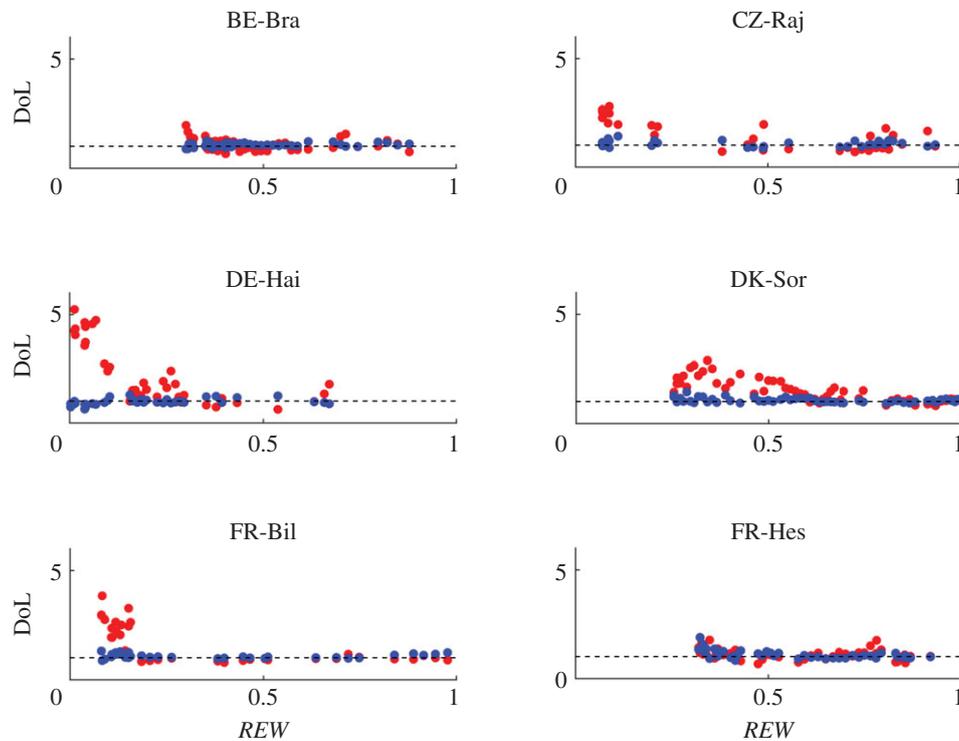
respective responses in the measured signal. This is especially critical at BE-Vie where the two most frequent wind directions (southwest and northeast) correspond to different stands (coniferous and beech stands) with possibly different root depth,  $REW$  and weather conditions [37]. Separating the data between each sector did not however improve the relation because of the lack of data (data not shown).

At sites with dense canopy and high LAI, vertical gradients of the parameters ( $G_s$ ,  $G_1$  and  $V_{\text{CMAX,APP}}$ ) could result from vertical gradients within plants of the same species or from physiological differences across species [27]. Sun leaves, developed under high irradiance, usually exhibit higher WUE (lower  $G_1$ ) than those developed under shady conditions [55], primarily because of higher photosynthetic capacities. However, more critical for this study, little is known about to what degree these vertical gradients (within and across species) could affect the response of  $G_1$  and  $V_{\text{CMAX,APP}}$  to drought. To our knowledge, in most Earth system models, the same reduction functions of photosynthesis during edaphic drought (either NSOL or SOL) are used for sun and shade leaves [17]. More complex multiple layers and/or sun-shade models as well as additional data gathered at multiple canopy layers would be needed to assess this question more closely.

Moreover, soil and vegetation components cannot be distinguished so that critical variables such as  $G_s$  (and variables depending on it, such as  $G_1$  and  $C_i$ ) will inevitably contain some signal from the soil. This signal can be reduced by filtering the data after rain events [27] and should be small for dense canopies with LAI higher than 2–3 [56] (which is the case in all sites) and even smaller when the upper soil layer dries.

Finally, systematic errors (energy balance non-closure [57]) in EC fluxes are also major sources of uncertainties that affect  $G_1$  and  $V_{\text{CMAX,APP}}$  magnitudes. However, it was found that, at multiple flux tower sites, the surface energy balance was not modified during the 2018 drought [58]. This source of error is therefore unlikely to affect  $G_1$  and  $V_{\text{CMAX,APP}}$  responses to  $REW$ .

Nevertheless, despite all the limitations of the big leaf approach detailed above, this framework was very suitable for this multi-site study as it relied on very few ancillary data [31]. If the comparison of  $G_1$  and  $V_{\text{CMAX,APP}}$  (which inherently has a different meaning than leaf-level  $V_{\text{cmmax,app}}$ ) is not straightforward, analysing the dynamics of these parameters inferred from *in situ* EC data during drought provides very useful information about how forest ecosystems reacted to these events.



**Figure 4.** Degree of limitation (DoL) by both SOL and NSOL. The degree of limitation was computed as the ratio of modelled *GPP* on measured *GPP* at high radiation. Modelled *GPP* was computed by using either fixed  $V_{\text{CMAX,APP}}^*$  (red points, NSOL) and varying  $G_1$  or fixed  $G_1^*$  and observed  $V_{\text{CMAX,APP}}$  (blue points, SOL). (Online version in colour.)

### (b) Implications of non-stomatal origin limitation for the modelling of gross primary productivity and transpiration

In this study, similarly to Granier *et al.* [5], we found that *GPP* and  $G_s$  reductions can be expected when *REW* drops below  $\approx 0.4$ . To account for these reductions, using empirical reduction factors (ranging from 1 to 0) when soil water content falls below a given threshold is a widely used approach [17]. However, it is questionable whether the reduction factors should be applied to SOL and/or to NSOL. This was previously investigated in Mediterranean ecosystems [8,13,59] and it was found that, calibrating the model on either *GPP* or transpiration (*E*) (not both) by considering only SOL during edaphic drought conditions systematically led to overestimates of WUE which did not allow the correct simulation of both fluxes. Surprisingly, it was found that applying NSOL only was sufficient to correctly simulate both *GPP* and *E*.

In this study, we found that reducing  $V_{\text{CMAX,APP}}$  when *REW* dropped below  $\approx 0.4$  and using a constant  $G_1$  parameter (from the USO model [24]) allowed the capture of both *GPP* and  $G_s$  reductions at European forest sites. Similar conclusions were also found by Chen *et al.* [60] in four different ecosystems (temperate grassland, tropical savannah, boreal and one temperate forest). More specifically, relatively consistent behaviour was observed at the three beech (*Fagus sylvatica*) forest sites (FR-Hes, DK-Sor and DE-Hai) where NSOLs were the main source of photosynthesis reductions with relatively constant  $G_1$ . Similarly, in a study carried out on adult beech using leaf-level measurements, the Ball–Berry slope was found to be almost insensitive to soil water potential [61]. Our results are in agreement with Granier *et al.* [5] who observed constant WUE even for very low *REW* during the 2003 drought and with Hentschel *et al.* [62] who also found unchanged annual iWUE derived from tree ring carbon isotopic composition.

Studies were also performed at the leaf scale to study the impact of drought on NSOL and SOL. In their meta-analysis, Zhou *et al.* [10] found highly variable responses of  $g_1$  (leaf level) for woody species ranging from rather constant to severely decreasing  $g_1$  with drought. Decreasing  $V_{\text{cmax,app}}$  (NSOL) were however found for all species. It was also highlighted that NSOL was the main factor limiting photosynthesis under severe stress in 10 Mediterranean herbs and shrubs species [11,63] and, more importantly for this study, for four tree species [20]. Unfortunately, a direct comparison with our results cannot be carried out as the studied species were different. Such direct comparisons have been carried out in a woodland dominated by *Acacia* trees by Tarin *et al.* [30] who found a close agreement between  $G_1$  and  $g_1$  estimated from ecosystem (EC big leaf) and leaf-level approaches, respectively.

According to Keenan *et al.* [64], NSOL could be caused by the variations of a finite mesophyll conductance with soil water availability [65] which, if not taken into account, leads to wrong estimates of actual  $V_{\text{cmax}}$  [66] (which was implicitly taken into account by using  $V_{\text{CMAX,APP}}$ ). In addition, the hypothesis that, under severe droughts, *GPP* can be directly impacted by biochemical limitations which cause the reduction of actual  $V_{\text{CMAX}}$  should not be discarded [67]. Separating NSOL between these two mechanisms (mesophyll conductance and actual  $V_{\text{CMAX}}$ ) was not done in this study. Currently, without additional leaf-level data to better understand the mechanisms underlying mesophyll conductance changes during droughts, we use an apparent  $V_{\text{CMAX,APP}}$  [15,17,20].

### (c) Optimal stomatal behaviour during drought and intrinsic water-use efficiency

We did not find a general pattern of systematically decreasing  $G_1$  during drought or, in other words, an increasing iWUE for stomatal closure (increasing  $\lambda$ ) across ecosystems as

theoretically predicted [23,68]. To the contrary, we found constant  $G_T$  (and therefore  $\lambda$ ) values at most sites, and even increasing values at DE-Hai (and BE-Bra to a lesser extent). This result (a constant  $G_T$ ) is rather surprising as it would suggest that changes in stomatal conductance responses were not needed to model  $G_s$  under long-term water stress events [59] and that  $\lambda$  does not increase with drought. We argue that this is caused by the fact that NSOLs were not considered by Manzioni *et al.* [23] and Mäkelä *et al.* [68] in their analyses as stomatal closure (reduced  $G_s$ ) is known to regulate leaf water flows in response to soil water availability [69]; without such mechanisms, leaves would be quickly dehydrated. However, one should consider that stomatal closure in response to drought does not necessarily lead to a decrease in  $G_T$  as, in USO, any reduction of  $V_{\text{CMAX,APP}}$  lead to a reduction in stomatal conductance [10]. At very low  $REW$  values, previous studies showed that  $C_i$  could even increase because of NSOL [70], which would explain the increase of  $G_T$  we observed at DE-Hai.

Another more complex approach to stomatal conductance modelling is to model stomatal conductance in the function of leaf water potential which is expected to regulate  $G_s$  [71]. This approach requires a complete model of water flow from the soil through the plant to the atmosphere [72]. This kind of model was tested by reducing the stomatal slope of the Ball–Berry–Leuning model [26] with leaf water potential, but the model did not account for NSOL [54]. Recently, Dewar *et al.* [73] proposed a new optimization model in which stomatal behaviour maximizes photosynthesis and where the costs of stomatal closure arise from NSOL (mesophyll conductance and/or carboxylation rate) and/or loss of hydraulic conductance [74]. This results in a parameter, equivalent to  $G_T$ , which is expressed as a function of measurable variables such as hydraulic conductivity, leaf water potential and  $V_{\text{cmax}}$ . This model has been successfully tested on saplings for different plant functional types [75] and fitted well sub-daily leaf scale observations; however, this still needs to be tested for longer term *in situ* ecosystem droughts. This could not be done in this study as leaf and soil water potentials were lacking. It does, however, highlight a promising research path for the future.

## 5. Conclusion

In this study, we used a big leaf framework to investigate the origin of edaphic drought impacts on  $GPP$  (stomatal origin limitation and non-stomatal origin limitation) in European forest ecosystems during the 2018 drought. In agreement with Granier *et al.* [5], we found that  $GPP$  and  $G_s$  were both greatly affected by soil moisture depletion at many sites. We

went a step further by showing that these reductions could be faithfully modelled by decreasing  $V_{\text{CMAX,APP}}$  (NSOL) when the  $REW$  dropped below around 0.4 while keeping the  $G_T$  (SOL) parameter from the USO model [24] constant. These results were rather unexpected as it would suggest that stomatal closure was not responsible for  $GPP$  reductions with drought. We argue that this was caused by the fact that  $G_T$  was not representative of stomatal behaviour during drought because  $GPP$  was not only regulated by stomatal closure but also by NSOL. Nevertheless, these results strengthen the increasing evidence that NSOL should be included in stomatal conductance/photosynthesis models to faithfully simulate both  $GPP$  and water fluxes in forest ecosystems.

**Data accessibility.** Most of the flux data used in this study are available on the ICOS carbon portal at: <https://doi.org/10.18160/PZDK-EF78>. Ancillary information about the soil is provided in the electronic supplementary material.

**Authors' contributions.** L.G.dIM. was the leader of the data analysis and writing of the paper. Q.B. and B.H. participated in the design of the study, gave intellectual inputs and feedbacks to the paper. B.L. designed the study and followed the data analyses and writing process closely. All the other co-authors were involved in the data acquisition as principal investigators for each site and actively contributed to the paper by commenting on intermediate versions.

**Competing interests.** We declare we have no competing interests

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