



HAL
open science

Crop residues may be a key feedstock to bioeconomy but how reliable are current estimation methods?

Shivesh Kishore Karan, Lorie Hamelin

► To cite this version:

Shivesh Kishore Karan, Lorie Hamelin. Crop residues may be a key feedstock to bioeconomy but how reliable are current estimation methods?. *Resources, Conservation and Recycling*, 2021, 164, 10.1016/j.resconrec.2020.105211 . hal-03102172

HAL Id: hal-03102172

<https://hal.inrae.fr/hal-03102172>

Submitted on 17 Oct 2022

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial 4.0 International License

Crop residues may be a key feedstock to bioeconomy but how reliable are current estimation methods?

Shivesh Kishore Karan^{1,*} and Lorie Hamelin.¹⁺

¹ Toulouse Biotechnology Institute (TBI), INSA, INRAE UMR792, and CNRS UMR5504, Federal University of Toulouse, 135 Avenue de Rangueil, F-31077, Toulouse, France

* Corresponding Author: karan@insa-toulouse.fr, shivesh.karan@gmail.com, Tel.: +33 051-155-9791

⁺hamelin@insa-toulouse.fr

Abstract

Crop residues are acknowledged as a key biomass resource to feed tomorrow's sustainable bioeconomy. Yet, the quantification of these residues at large geographical scales is primarily reliant upon generic statistical estimations based on empirical functions linking the residues production to the primary crop yield. These useful yet unquestioned functions are developed either using direct evidence from experimental results or literature. In the present study, analytical evidence is presented to demonstrate that these methods generate imprecise and likely inaccurate estimates of the actual biophysical crop residue potential. In this endeavor, we applied five of the most used functions to a national case study. France was selected, being the country with the largest agricultural output in Europe. Our spatially-explicit assessment of crop residues production was performed with a spatial resolution corresponding to the level of an administrative department (96 departments in total), also the finest division of the European Union's hierarchical system of nomenclature for territorial units (NUTS), and included 17 different crop residues. The theoretical potential of crop residues for France was found to vary from 987 PJ y⁻¹ to 1369 PJ y⁻¹, using different estimation functions. The difference observed is more than the entire annual electricity consumption of Belgium, Latvia, and Estonia combined. Perturbation analyses revealed that some of the functions are overly sensitive to fluctuation in primary crop yield, while analytical techniques such as the null hypothesis statistical test indicated that the crop residues estimates stemming from all functions were all significantly different from one another.

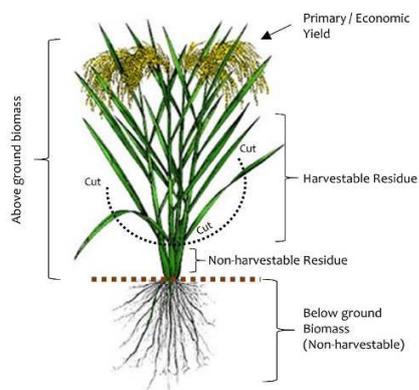
Keywords

Bioeconomy; Residue-to-Product Ratio; Spatial Quantification; Straw; Theoretical Potential.

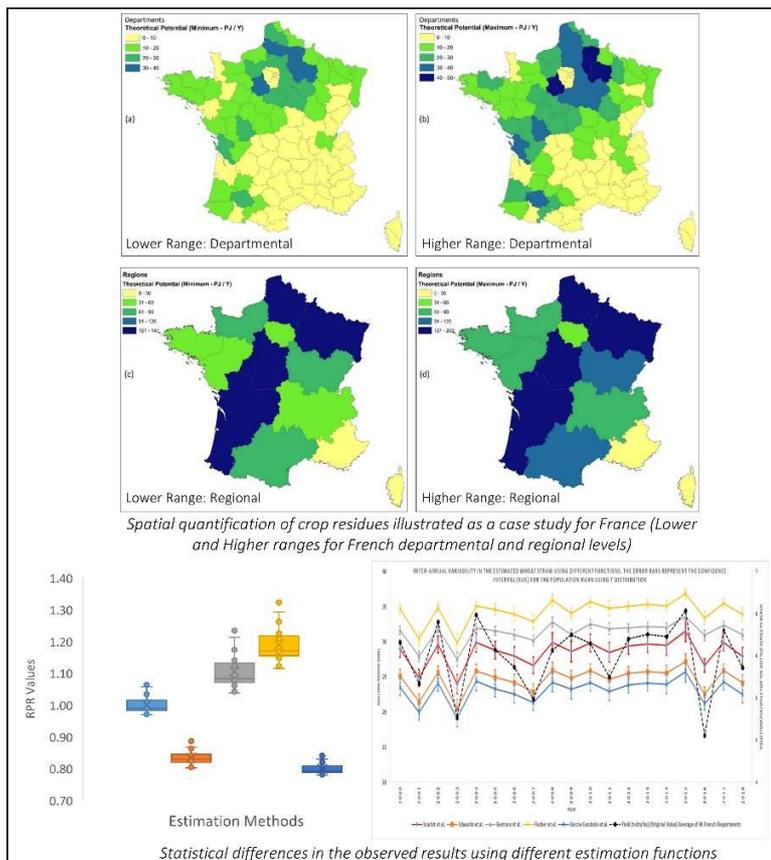
List of abbreviations

CR	Crop Residues
GJ/t	Giga joule per tonne
HI	Harvest Index
km ²	Square kilometers
kt	Kilo tonne
LHV	Lower Heating Value
MJ/kg	Megajoule per kilogram
Mt	Million tonne
NUTS	Nomenclature of Territorial Units for Statistics
PJ Y ⁻¹	Peta joule per year
R ²	Coefficient of determination
RPR	Residue-to-product ratio
SD	Standard Deviation
t/ha	tonne per hectare
THP	Theoretical Potential

43 Graphical abstract



- Method 1
 - Method 2
 - Method 3
- Residue estimation using different methods.
 - Comparison of methods and results.
 - Uncertainty analysis.

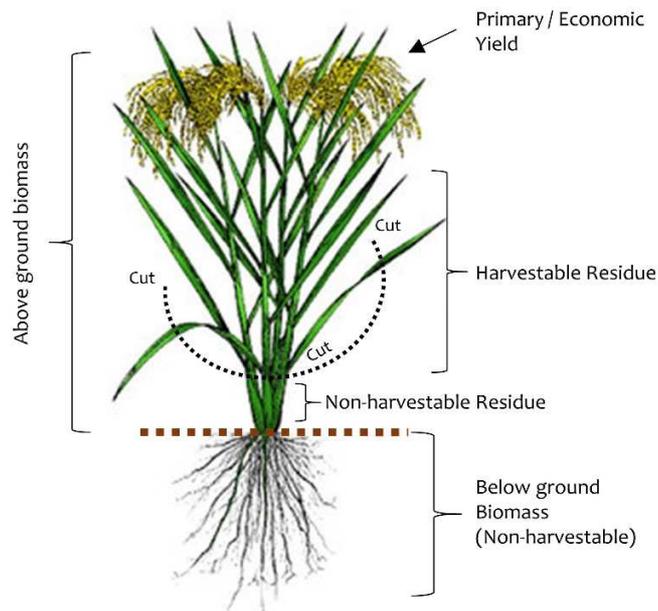


44

45 **1. Introduction**

46 Terrestrial lignocellulosic biomass from crop residues (CR) (e.g., cereal straw) is a significant carbon
 47 feedstock source to feed a well-below 2°C economy with the non-fossil carbon it requires (Bentsen et al.,
 48 2014; Hamelin et al., 2019; Williams et al., 2016). In fact, fossil fuel carbon dioxide (CO₂) emissions are the
 49 leading cause of human-induced climate change, counting with ca. 69% of global greenhouse gas (GHG)
 50 emissions (WRI, 2020). Substituting, to the extent possible, the use of fossil carbon by biogenic carbon
 51 from residual sources like CR furthermore allows supplying a carbon source decoupled from the demand
 52 for additional arable land.

53 CR has been subjected to scientific scrutiny for many years, particularly in the last two decades, and is
 54 typically defined as an agrarian by-product. CR mainly consists of the dry stalks and leaves of cereal and
 55 oilseed crops after the product of interest (i.e., grain, seed, or cobs) is harvested, and of the top stem and
 56 leaves from tuber tops (e.g., potato or beetroot). Figure 1 shows the typical representation of a cereal and
 57 oilseed crop in terms of the above-ground and below-ground repartition of the biomass. The above-
 58 ground biomass is partitioned as primary crop yield (to be harvested, also referred to as economic yield),
 59 harvestable residues (may or may not be collected), and non-harvestable residues (the machinery or
 60 specific farm management does not permit the harvest of these in most cases) (Hamelin et al., 2012). In
 61 the case of tubers, the primary crop yield is below-ground, and harvestable residues above-ground.



62
 63 **Figure 1: Generic repartition of the above- and below-ground biomass for cereal and oilseed crops. Although**
 64 **the whole above-ground biomass could be harvested, a portion of the above-ground often remains**
 65 **unharvested and considered as non-harvestable due to the specific farm management or harvester used.**

66 Although CR and, in particular, cereal straw represent an important non-fossil carbon source in terms of
 67 quantity generated all over the World, it is only the amount generated in surplus of current uses that can
 68 be directly available for the bioeconomy, at least to avoid inducing market reactions caused by a change in
 69 supply. Apart from use in bioenergy production (e.g., straw-firing heat plants), CR already serve several
 70 competitive demands ranging from fodder and bedding in animal husbandry, as a substrate for mushroom
 71 cultivation or as a mulch in farms, among other applications (Haase et al., 2016; Scarlat et al., 2019; Tonini
 72 et al., 2016b). Furthermore, one inherent essential function of CR is its role as a vital source of organic
 73 matter for soils, including a supply of carbon, nitrogen, and other nutrients to soils. CR are also known for
 74 their ecosystemic functions, such as acting as a preventive layer against erosion (Haase et al., 2016) or
 75 enhancing soil water retention (Blanco-Canqui, 2013). Hence, the plethoric removal of these residues
 76 from agricultural fields can decrease the long-term productivity of soils (Blanco-Canqui, 2013; FAO, 2017).
 77 Therefore, the economic and environmental sustainability of removing CR from fields requires attentive
 78 and site-specific evaluation before any massive investment in CR-based bioeconomy solutions takes place.
 79 This challenge was first acknowledged by Scarlat et al. (2010), who presented a comprehensive
 80 assessment of the availability of CR in the European Union. Based on a literature review, the authors
 81 proposed sustainable removal rates varying between 40% and 50% according to the CR type, these rates
 82 allowing to maintain soil organic matter. The sustainable removal rates published by Scarlat et al., (2010)
 83 have been widely used in bioenergy and bioeconomy studies (Daioglou et al., 2016; Monforti et al., 2013;
 84 Searle and Malins, 2015). Apart from the study of Scarlat et al. (2010), several studies at scales varying
 85 from regional to global have proposed a variety of indicators to quantify the sustainable CR removal rates
 86 (Hansen et al., 2020; Muth et al., 2013; Ronzon and Piotrowski, 2017; Scarlat et al., 2019).

87 Yet, when it comes to bioeconomy planning, the starting point is to ascertain the total annual biophysical
 88 quantity of these residues, i.e., prior to applying any restrictions, whether of sustainability or feasibility
 89 nature. This quantity is typically referred to as the theoretical potential (THP) (Bentsen and Felby 2012).
 90 Providing THP estimates, although these do involve their load of uncertainties, has the merit to supply a
 91 transparent quantitative basis for decision-making. Scaler multipliers may subsequently be applied to the
 92 THP estimates, at the convenience of stakeholders in charge of the planning to reflect techno-economic or

93 environmental constraints (Ericsson and Nilsson, 2006; Haberl et al., 2010; Kadam and McMillan, 2003).
94 Thus, in this study, we focus on the methods for estimating the THP of CR.

95 Actual field measurements would probably supply the most accurate method for quantifying CR THP in a
96 given plot. Yet, because CR are a seldom traded market commodity, and because of the related time and
97 cost constraints associated with measurements of unharvested CR, these measurements are rarely
98 available nor performed. To derive THP estimates at global, national, or even at regional levels, statistical
99 and empirical estimation methods have typically been used (Bentsen et al., 2014; García-Condado et al.,
100 2019; Scarlat et al., 2010). Usually, the estimation of CR production has been realized based on
101 assumptions on the mathematical relationship between the crop and the residue yield. This relationship is
102 generally derived as a factor based on the ratio between the primary crop yield and the residue yield,
103 commonly referred to as the residues-to-product ratio (RPR). Some studies also use Harvest Indexes (HI)
104 for estimating CR (e.g., Sommer et al., 2016). HI is defined as the primary crop yield expressed as a
105 fraction of the total above-ground biomass produced.

106 Several studies suggest that RPR is better represented as a function of primary crop yield rather than as a
107 fixed value (Bentsen et al., 2014; Scarlat et al., 2010). As reported in Ronzon and Piotrowski (2017), the
108 functions so-far proposed for estimating the residue yield are somewhat diverse, including linear (Fischer
109 et al., 2007), logarithmic (Scarlat et al., 2010), hyperbolic (Bodirsky et al., 2012), inverse tangential
110 (Edwards et al., 2005) or exponential (Bentsen et al., 2014). In reality, the quantity of CR generated at
111 large geographic regions can encapsulate significant variations due to a plethora of factors such as soil
112 type, prevailing meteorological conditions, harvesting practices, and primary crop yield, among other
113 things. Some studies also reported that drought has an impact on the residue-to-product ratio that may
114 either decrease or increase if drought occurs at earlier or later growth stages, respectively (McCartney et
115 al., 2006). Because of this diversity in the factors affecting the residue yield, there is no clear standard or
116 set of rules for the quantification of crop residues THP at large geographical scales. Yet, it appears that
117 despite the heavy focus on quantifying sustainable removal rates, studies never challenged nor addressed
118 the potential significance of the choice of selecting the initial THP estimation method in the first place,
119 whether based upon HI or RPR functions.

120 Hence, the overall goal of this study is to evaluate the magnitude of eventual differences in CR THP
121 estimates resulting from the use of the most commonly reported functions for CR estimation. This is
122 illustrated with a national case study for Metropolitan France, the European Union country with the
123 largest agricultural output, in economic terms (European Commission, 2020). We further address three
124 specific sub-questions (SQ):

- 125 ○ SQ-1: How variations in primary crop yield affect the estimation of CR yield for the assessed
126 functions;
- 127 ○ SQ-2: How uncertainties in primary crop yield overshadow the differences observed in the
128 estimated CR stemming from the functions assessed herein and;
- 129 ○ SQ-3: Is there any significant differences in the RPRs estimated from the different estimation
130 functions.

131 2. Materials and methods

132 2.1 Scoping

133 The assessment considers all major crops grown in France and reported in the national statistics (Agreste,
134 2020), here grouped into four categories (Cereal crops, Roots and Tubers, Protein Crops, and Oil crops),
135 which comprises 16 crops in total (Table 1). These represent ca. 20% of the overall land cover. The annual
136 data on their production and surface area was obtained from the national agricultural statistics (Agreste,
137 2020) at the French departmental administrative level (corresponding to NUTS-3 division in Eurostat's

138 Nomenclature of Territorial Units for Statistics; Eurostat, 2020a). For each department, average yields
139 were calculated from 19 years of production and surface area data (2000 – 2018), as shown in Eq. 1:

$$140 \quad \text{Primary Yield}(Y_{i,j}) = \frac{\text{Production}_{i,j} \text{ (Tonne)}}{\text{Surface area}_{i,j} \text{ (Hectare)}} \quad (\text{Eq. 1}).$$

141 Where *Primary crop Yield_{i,j}* is the economic (cereal) yield for crop *i* in department *j*, *Production_{i,j}* is the
142 production of crop *i* in department *j*, and *Surface area_{i,j}* is the corresponding agricultural surface for crop *i*
143 in department *j*.

144 As detailed in the Supplementary Material 1 (SM1), the minimum and maximum records of crop
145 production and surface area were identified for each crop and department in order to incorporate the
146 range of annual variability in crop yield.

147 2.2 Estimation of crop residues using empirical functions

148 RPR is mathematically defined as the ratio of the above-ground harvestable biomass residue, here defined
149 as residue yield, *R*, to the primary crop yield, *Y* (García-Condado et al., 2019), as shown in Eq. (2), which
150 also presents the correspondence between RPR and HI:

$$151 \quad RPR = \frac{1-HI}{HI} = \frac{R_{i,j}}{Y_{i,j}} \quad \text{Eq. (2)}$$

152 It should be noted that Eq. (2) was also presented in García-Condado et al. (2019), and is only valid to the
153 extent *R* refers to the overall generated residue (harvestable and non-harvestable; Figure 1).

154 The rationale for selecting different empirical functions for RPR varies for different studies. Still, the
155 essential notion behind most functions is that the residue yield is directly proportional to the primary crop
156 yield (Scarlat et al., 2010). Based on this, Bentsen et al. (2014) as well as Ronzon and Piotrowski (2017),
157 proposed an exponential relation between the crop and the residue yields. Scarlat et al. (2010), on the
158 other hand, derived best-fit logarithmic function curves for RPR by plotting the values for RPR and primary
159 crop yield based on data available in the literature. Edwards et al. (2005) derived RPR functions for wheat
160 and barley, based on grain yields and empirical ranges of harvest indexes taken from de Vries (1999). The
161 study of Fischer et al. (2005) proposed negative linear RPR functions, which do not limit the production of
162 crop residues to a threshold. This, however, mathematically implies that residue yields may decrease at
163 very high levels of primary crop yields, as highlighted by Ronzon and Piotrowski (2017). On the other hand,
164 Bentsen et al. (2014) argue that plant breeding has led to an increase in the HI without changing the
165 overall plant biomass (Hay, 1995), indicating an asymptotic development of residue yield to a theoretical
166 threshold only limited by physiological constraints. Thus they considered piecewise continuous functions
167 to derive RPR estimates. García-Condado et al., (2019) used empirical models to predict crop residues
168 from annual yield statistics. Their models were developed based on experimental data from the scientific
169 literature. The functions mentioned above are summarized in Table 1. It can also be noted from Table 1
170 that although RPR functions typically differ from one crop to the other, there are also cases where exactly
171 the same functions are proposed (e.g. wheat and barley RPR functions of Edwards et al. (2005). **Moreover,**
172 **it is not always the same nor clear which exact fraction of the residues is considered in these studies (e.g.,**
173 **harvestable CR only or the entire aboveground CR), as highlighted in Table 2.**

174 **Table 1: RPR functions, Lower Heating Values and Dry matter for the selected crops^a**

Crop Type	Crop	Lower Heating Values	Dry matter (%)	RPR Function ^b	R ² (if provide d)	Source
Cereal Crops	Wheat	15.2 MJ kg ⁻¹ (Phyllis2, 2020)	90 % (Wirsenius, 2000)	RPR = -0.3629*ln(Y)+1.6057	0.28	(Scarlat et al., 2010)
				$RPR = \frac{0.769-0.129*\arctan((Y)-6.7)/1.5}{Y}$	-	(Edwards et al., 2005)
				RPR = 2.186*exp(-0.127*Y)	0.52	(Bentsen et al., 2014)
				RPR = -0.14Y+1.96	-	(Fischer et al., 2007)
	Barley	16.19 MJ kg ⁻¹ (Phyllis2, 2020)	90% (Wirsenius, 2000)	RPR = 1.822*exp(-0.149*Y)	0.51	(Bentsen et al., 2014)
				$RPR = \frac{0.769-0.129*\arctan((Y)-6.7)/1.5}{Y}$	-	(Edwards et al., 2005)
				RPR = -0.27*Y+2.77	-	(Fischer et al., 2007)
	Maize	17.41 MJ kg ⁻¹ (Phyllis2, 2020)	85% (Wirsenius, 2000)	RPR = -0.1807*ln(Y)+1.3373	0.17	(Scarlat et al., 2010)
				RPR = 2.656*exp(-0.103*Y)	0.49	(Bentsen et al., 2014)
				RPR = -0.13*Y+2.20	-	(Fischer et al., 2007)
	Oats	18.45 MJ kg ⁻¹ (Phyllis2, 2020)	92% (Phyllis2, 2020)	RPR = 1.868*exp(-0.250*Y)	-	(Ronzon and Piotrowski, 2017)
				RPR = -0.1874*ln(Y)+1.3002	0.21	(Scarlat et al., 2010)
				RPR = -0.20*Y+2.70	-	(Fischer et al., 2007)
	Triticale	15.45 MJ kg ⁻¹ (Ruiz et al., 2015)	90% (Average of all cereals)	RPR = 1.880*exp(-0.120*Y)	-	(Ronzon and Piotrowski, 2017)
	Rye	15.24 MJ kg ⁻¹ (Phyllis2, 2020)	89% (CCOF, 2013)	RPR = 1.964*exp(-0.250*Y)	-	(Ronzon and Piotrowski, 2017)
RPR = -0.3007*ln(Y)+1.5142				0.22	(Scarlat et al., 2010)	
RPR = -0.20*Y+2.70				-	(Fischer et al., 2007)	
Sorghum	14.27 MJ kg ⁻¹ (Phyllis2, 2020)	85% (Wirsenius, 2000)	RPR = -0.55*Y+4.55	-	(Fischer et al., 2007)	
			RPR = 2.302*exp(-0.100*Y)	-	(Ronzon and Piotrowski, 2017)	
Rice	16.38 MJ kg ⁻¹ (Phyllis2, 2020)	90% (Wirsenius, 2000)	RPR = -1.2256*ln(Y)+3.845	0.57	(Scarlat et al., 2010)	
			RPR = 2.450*exp(-0.084*Y)	0.22	(Bentsen et al., 2014)	
			RPR = -0.22*Y+2.56	-	(Fischer et al., 2007)	
Roots and Tubers	Beet	16.6 MJ kg ⁻¹ (Koga, 2008)	20% (Wirsenius, 2000)	RPR = 1.328*exp(-0.060*Y)	-	(Ronzon and Piotrowski, 2017)
				RPR = -0.005*Y+0.75	-	(Fischer et al., 2007)
	Potato	13.6 MJ kg ⁻¹ (Koga, 2008)	20% (Wirsenius, 2000)	RPR = 1.916*exp(-0.108*Y)	-	(Ronzon and Piotrowski, 2017)
RPR = -0.01*Y+1.10				-	(Fischer et al., 2007)	
Protein Crops	Beans	16.24 MJ kg ⁻¹ (Phyllis2, 2020)	95% (Wirsenius, 2000)	RPR = 3.232*exp(-0.300*Y)	-	(Ronzon and Piotrowski, 2017)
	Protein Pea	13.57 MJ kg ⁻¹ (Özyuğuran et al., 2018)	95% (Wirsenius, 2000)	RPR = 3.644*exp(-0.300*Y)	-	(Ronzon and Piotrowski, 2017)
	Sweet Lupine	14.90 MJ kg ⁻¹ (Taken as average of above two)	95% (Taken as average of above two)	RPR = 3.232*exp(-0.300*Y)	-	(Ronzon and Piotrowski, 2017)
Oil Crops	Rape	16.33 MJ kg ⁻¹ (Phyllis2, 2020)	87.3% (Karaosmanoğlu et al., 1999)	RPR = 3.028*exp(-0.200*Y)	-	(Ronzon and Piotrowski, 2017)
				RPR = -0.452*ln(Y)+3.2189	0.17	(Scarlat et al., 2010)
	Sunflower	13.9 MJ kg ⁻¹ (Lindley and Smith, 1988)	90% (Wirsenius, 2000)	RPR = 2.580*exp(-0.200*Y)	-	(Ronzon and Piotrowski, 2017)
				RPR = -1.1097*ln(Y)+3.2189	0.26	(Scarlat et al., 2010)
	Soy	14.3 MJ kg ⁻¹ (Teagasc, 2010)	90% (Wirsenius, 2000)	RPR = 3.869*exp(-0.178*Y)	0.45	(Bentsen et al., 2014)
				RPR = -0.80*Y+3.90	-	(Fischer et al., 2007)
Others ^c	14.8 MJ kg ⁻¹ (Taken as average of above three)	89.1% (Taken as average of above three)	RPR = 2.148*exp(-0.200*Y)	-	(Ronzon and Piotrowski, 2017)	

175 ^a For primary crop yields, see SM1. These are not presented herein, as derived at the department level.

176 ^b To maintain consistency with the terms used in the present study, the terminology used in the functions original have been adapted to the one used herein.

177 ^c As per (Agrete, 2020), other oil crops include flax, castor and oeillette.

178

179 **Table 2: Qualitative overview of the residue portion considered in the RPR functions of the studies inventoried.**

Functions	Residue portion considered in the RPR functions (applies to all crops of the study)
(Scarlat et al., 2010)	Unclear if the residue is a fraction of total above-ground residue or the harvestable portion only.
(Bentsen et al., 2014)	
(Fischer et al., 2007)	
(Edwards et al., 2005)	Residue from the entire above-ground portion of the crop.
(García-Condado et al., 2019)	
(Ronzon and Piotrowski, 2017)	Residue from the harvestable portion of the above-ground biomass.

180
 181 The RPR functions presented in Table 1 were used to estimate spatially-explicit residue yields considering,
 182 for each administrative department, the primary crop yield and surface area data for each of the 16 crops
 183 included in this case study (Eq. 3).

184
$$Residue\ Production\ (RP_{i,j}) = RPR_{i,j} \times Primary\ Yield(Y_{i,j}) \times Surface\ area_{i,j} \quad Eq. (3)$$

185 Where *Residue Production* ($RP_{i,j}$) is the amount of residue produced for crop i in department j , and $RPR_{i,j}$ is
 186 the residue-to-product ratio of crop i in department j . **The aggregated spatially-explicit residue production**
 187 **is presented in terms of energy units using the LHV values shown in Table 1. For detailed department-wise**
 188 **crop-specific residue production in terms of mass, see SM1 (Sheet: Crop Residues DM and Energy).**

189 **2.3 Uncertainty assessment**

190 Uncertainty assessment was used to address **the** three specific sub research questions. Three tests were
 191 performed by considering wheat cereal as a case-example, as it represents a significant share of the
 192 generated CR (39% by production volume in France).

193 In the first test, **SQ-1 was addressed. Here** the extent to which the variation (or sensitivity) in primary crop
 194 yield affected the estimated residue yield was evaluated by performing a one-at-a-time (OAT)
 195 perturbation analysis (Bisinella et al., 2016). In the OAT analysis, primary crop yield values were changed
 196 by $\pm 10\%$ and $\pm 50\%$ of the original values, and residue yields were recalculated accordingly, using all the
 197 functions presented in Table 1.

198 In the second test, **SQ-2 was addressed. Here** we evaluated how the actual uncertainty in primary crop
 199 yield overshadows the differences we observe in the estimated residues using the functions listed in Table
 200 1. For performing this test, each of the 96 French departments was considered as an individual sample,
 201 and the mean and standard deviation (SD) of primary crop yield **for the whole of France using the data**
 202 **from the 96 French departments** was calculated on a year per year basis for the period considered here
 203 (2000 – 2018). To incorporate this uncertainty in the estimated annual results, residue yields were
 204 recalculated with the original primary crop yield $\pm SD$ values for all the 19 years of data, and a chart was
 205 plotted to observe the overshadowed differences as confidence interval using the student's t distribution
 206 (Supplementary Material 2: SM2).

207 Finally, in the third test, we evaluated, through a two-tailed t -test, if there are any significant differences
 208 in the RPRs obtained using the different estimation functions presented in Table 1 (**SQ-3**). The RPR values
 209 were calculated using the mean annual primary crop yield value calculated in the second test (detailed in
 210 SM2). In the case of the function from García-Condado et al. (2019), the RPR values were calculated
 211 indirectly using Eq. (2). Each given RPR result was paired to the RPR result of the other functions for the
 212 corresponding year, thus creating a sample size of nineteen. For this test case, the null hypothesis and the
 213 alternate hypothesis were formulated as:

214 H_0 : There are no significant differences in the estimated RPRs of wheat cereal using different
 215 functions.

216 H_1 : There is a significant difference in the estimated RPRs of wheat cereal using different
 217 functions.

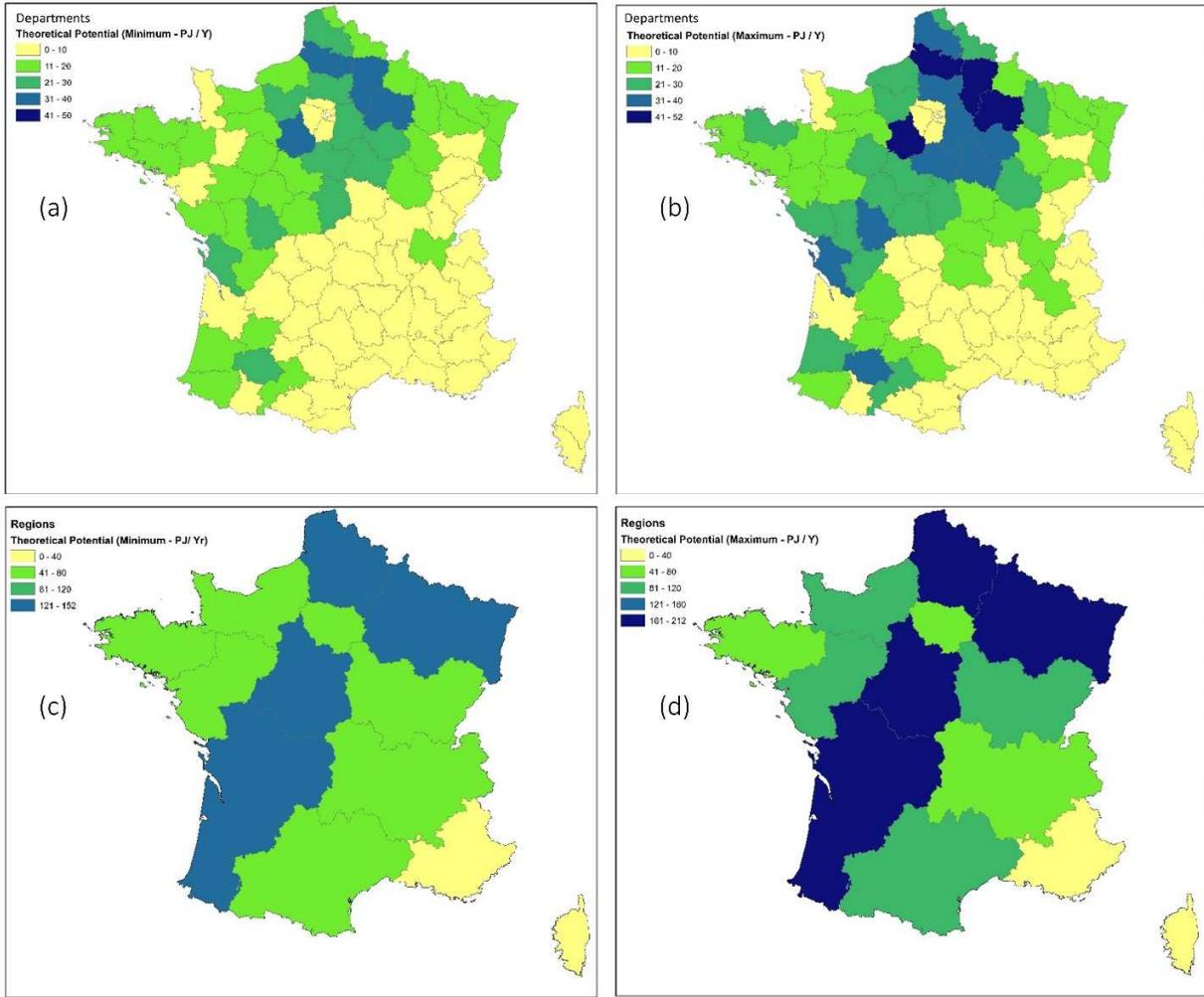
218 The t-test was performed at a significance level of $\alpha = 0.05$ (95% confidence) (SM2).

219 3. Results and Discussion

220 In this study, we examined with a national case for France, the use of different estimation methods for
221 quantifying CR. For each crop, the CR THP of a given spatially-explicit unit was separated into two ranges,
222 i.e., (i) higher range, which includes the maximum CR estimate for the given crop, and (ii) lower range,
223 which includes the lowest CR estimate for the given crop. The aggregated spatially-explicit results (i.e., for
224 all crops) are shown in Figure 2 in terms of energy units, both at the French departmental (NUTS-3) and
225 regional (NUTS-2) level. The THP of CR considering the selected sixteen crops varied from 987 PJ Y⁻¹ to
226 1369 PJ Y⁻¹. These estimates are considerable, equivalent to about 60% - 80% of the annual French
227 electricity consumption (For the year 2017, Eurostat, 2020).

228 The THP, by definition, does not consider any competitive use (animal feed, bedding, etc.). The
229 competitive uses of CR can be substantial; for example, Monforti et al. (2013) estimated that about 16% of
230 the collectible CR is needed as animal bedding. Furthermore, in reality, not all of the estimated residues
231 are collectible, and their removal from fields is not always suitable. Several studies have reported that
232 about 40%-70% of these residues should not be collected, considering a variety of sustainability goals and
233 premises (Einarsson and Persson, 2017; Scarlat et al., 2019, 2010; Hansen et al., 2020). Consequently, it
234 should be kept in mind that the ranges presented in Figure 2 are higher than what can actually be used as
235 a replacement for fossil carbon. However, mobilizing even just 20% of the potentials presented in Figure 2
236 could substitute about 3% - 5% of the 2017 French electricity consumption, considering an electrical
237 conversion efficiency of 27% (Tonini et al., 2016a).

238 From Figure 2, it can be observed that the CR production is mainly concentrated in the Centre-Val de
239 Loire, Hauts-de-France, Grand Est, and the Nouvelle-Aquitaine regions of France, which are also the
240 primary cereal producing regions. The overall THP of CR at the Regional (NUTS-2) level is shown in Table 3,
241 while THPs at the department (NUTS-3) level and crop-specific maps of the estimated THP using different
242 functions are available in SM1.



243

244

245

Figure 2: Theoretical potential of crop residues at the French departmental (a: minimum; b: maximum) and regional level (c: minimum; d: maximum).

246 **Table 3: Crop residues theoretical potential at the regional (NUTS-2) level, all crops^a**

Region Name	Overall potential (Minimum) PJ Y ⁻¹	Overall potential (Maximum) PJ Y ⁻¹	Δ%
Ile-de-France	42.36	57.89	37%
Centre-Val de Loire	128.7	184.7	43%
Bourgogne-Franche-Comte	74.60	109.6	47%
Normandie	64.42	88.05	37%
Hauts-de-France	137.4	179.8	31%
Grand Est	151.6	212.4	40%
Pays de la Loire	60.39	83.62	38%
Brittany	52.51	71.52	36%
Nouvelle-Aquitaine	139.9	191.4	37%
Occitanie	76.66	109.0	42%
Auvergne-Rhone-Alpes	52.33	71.59	37%
Provence-Alpes-Cote d'Azur	6.460	9.467	47%
Corse	0.1457	0.1844	27%
Total	987.5	1369	39%

247 ^a All values are presented with a maximum of four significant digits, but it should not be seen as an indication of precision.

248 The results presented in Table 3 reveal high variability. At the national scale, this corresponds to about
 249 39% difference (987 – 1,369 PJ Y⁻¹). This 382 PJ Y⁻¹ difference is almost equal to about 22% of the overall
 250 annual electricity consumption in France, also equivalent to more than the overall electricity consumption
 251 of Belgium, Latvia, and Estonia combined (year 2017, Eurostat, 2020). At the regional level, the maximum
 252 difference was observed in the region of Grand Est with nearly 61 PJ Y⁻¹, which itself is nearly twice the
 253 entire electricity consumption of a small country like Estonia. These considerable differences are isolating
 254 the “RPR function” effect only, as the primary crop yield considered for a given crop-department
 255 combination remains constant.

256 The estimated THP of CR of our study falls within the range of a recent study by Scarlat et al. (2019),
 257 where an average THP of 1067.5 PJ Y⁻¹ was estimated for France, considering a LHV of 17.5 MJ kg⁻¹ DM.
 258 However, in their study, they only considered eight crops, namely wheat, rye, barley, oats, maize, rice,
 259 rapeseed, and sunflower. When compared to the estimates of Monforti et al. (2013), our estimates
 260 (62,182 kt – 86,178 kt) are 4 – 44% higher than the 59,569 kt Y⁻¹ presented in Monforti et al. (2013).

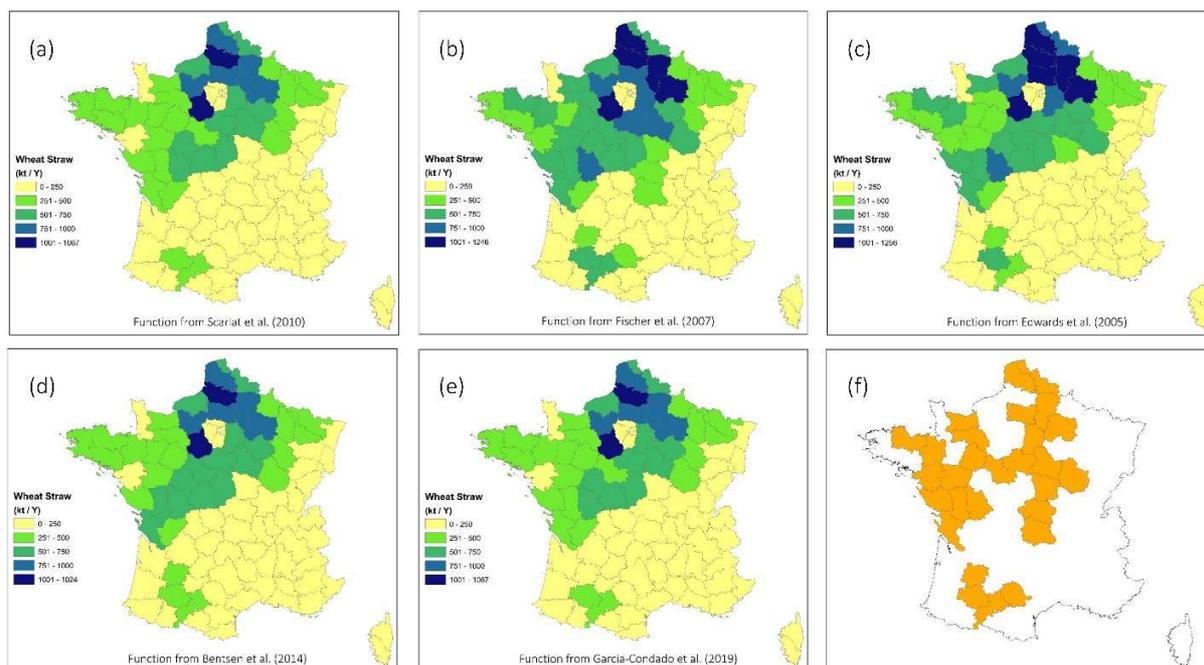
261 The average residue production (Mt) and the residue yield (t/ha) range of the crops selected in this study
 262 are shown in Table 4, based on the RPR function used. In terms of absolute volume, the maximum
 263 difference in the residue production was observed for wheat straw between the functions proposed by
 264 Fischer et al. (2007) and García-Condado et al. (2019), with a difference of 9.3 Mt Y⁻¹ of wheat straw.

265 **Table 4: Average residue yield and residue production for the selected crops, using the different RPR**
 266 **functions assessed in this study.**

Crops	Average (2000 – 2018) residue production in M tonne DM per year, national level ^a						Residue Yield (Min-Max) (Tonne / ha)
	(Scarlat et al., 2010)	(Edwards et al., 2005)	(Bentsen et al., 2014)	(Fischer et al., 2007)	(García-Condado et al., 2019)	(Ronzon and Piotrowski, 2017)	
Wheat	30.73	25.70	31.89	34.59	25.34		4.4 – 6.3
Barley	8.73	8.01	7.48	11.60	10.27		4.2 – 6.7
Maize	12.24		15.26	15.14			7.1 – 9.0
Oats	0.43			0.77		0.2707	2.6 – 7.2
Rice	0.16		0.14	0.13	0.16		7.1 – 7.8
Rye	0.15			0.21		0.07846	2.8 – 7.4
Sorghum				0.48		0.3738	6.6 – 8.6
Triticale						2.048	4.7
Rape	10.88					6.714	4.6 – 7.3
Soy			0.45	0.35			4.6 – 5.8
Sunflower	3.26			3.21		2.302	3.5 – 5.0
Other Oil Crops				0.06			2.50
Lupine						0.006225	0.83
Pea						0.3602	1.30
Beans						0.09343	0.97
Potato				1.41		1.047	5.8 – 6.3
Beetroot				4.51		3.284	7.7 – 9.9

267 ^a Empty cells mean that a given study did not supply RPR functions for the crop under consideration. All values are presented with
 268 a maximum of four significant digits, but it should not be seen as an indication of precision.

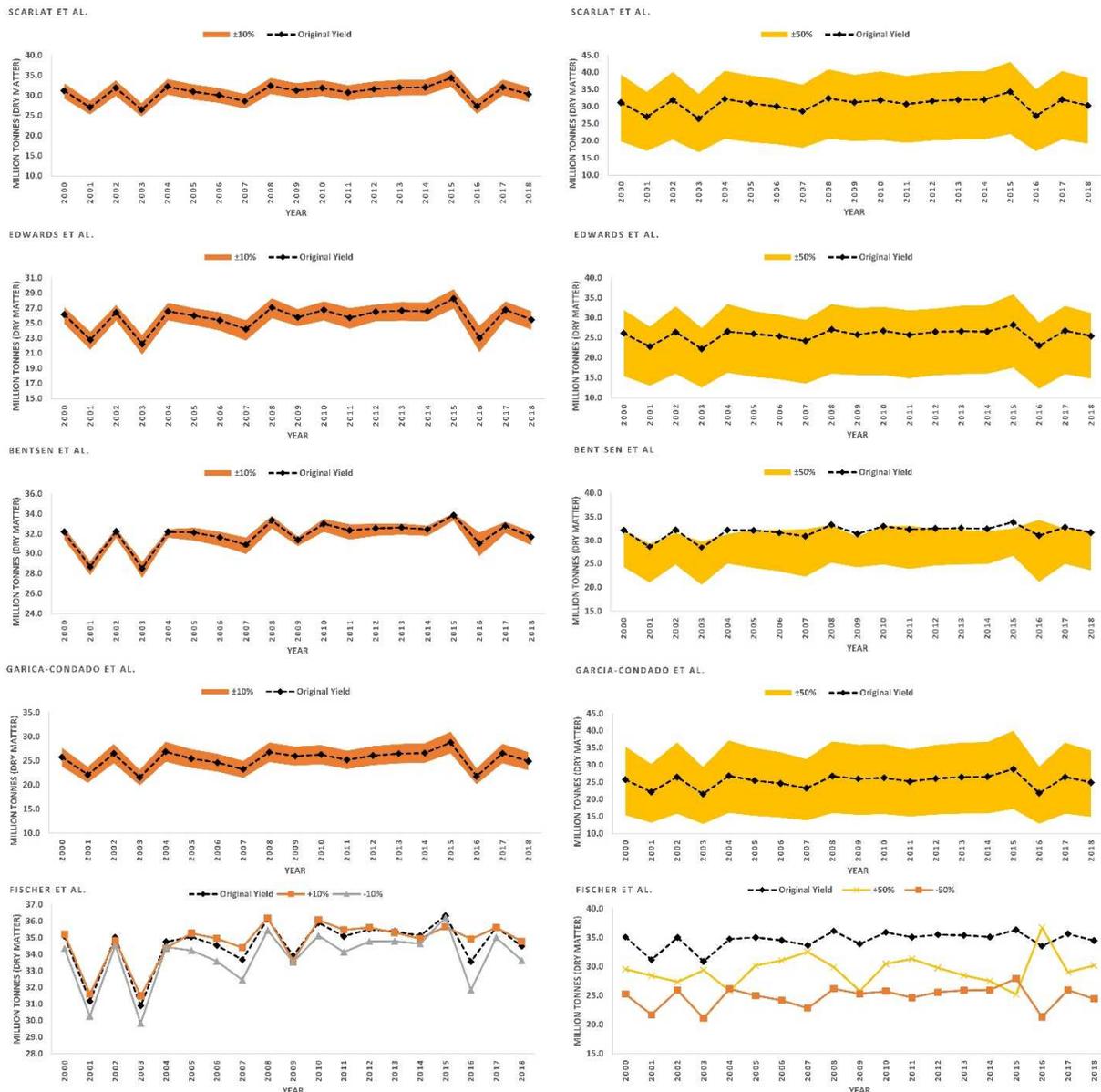
269 Figure 3 (a-e) shows the spatial distribution of wheat straw estimated using different empirical functions.
 270 Wheat straw is used here as a representative example since it contributes with ca. 40% of the THP-energy
 271 (385.1 PJ Y⁻¹ – 525.8 PJ Y⁻¹), but the details for all other CR can be found in SM1- CR (DM and Energy).
 272 Figure 3 (f) highlights the departments which are associated with two or more ranges of wheat straw
 273 potential, according to the RPR function used for the estimation. In total, 29 out of the 96 French
 274 departments have different ranges of wheat straw potential associated with them.



275
 276 **Figure 3: Department (NUTS-3) spatial distribution of wheat straw THP using different empirical functions (a**
 277 **– e), and (f) Departments associated with two or more ranges of wheat straw potential.**

278 3.1 SQ-1 Effect of variations in primary crop yield on estimated CR

279 In order to evaluate the sensitivity of the empirical functions to the fluctuations in primary crop yield, OAT
 280 perturbation analysis was performed by changing the primary crop yield value by $\pm 10\%$ and $\pm 50\%$ of the
 281 original. For three out of the five functions (Edwards et al., 2005; García-Condado et al., 2019; Scarlat et
 282 al., 2010), a proportional increasing or decreasing trend was observed, i.e., with the increase in primary
 283 crop yield, the estimated residues also increased and vice versa (Figure 4). For the function by Bentsen et
 284 al. (2014), when the primary crop yield values were changed by $\pm 10\%$, the estimated results were
 285 observed to be tightly bound to the results estimated using the original primary crop yield values.
 286 However, when the primary crop yield values were changed by $\pm 50\%$, disproportionate changes were
 287 observed in the estimated straw, reflecting the very nature of the piecewise functions proposed by the
 288 authors, which limits the CR production (and indirectly possible yield increases) to a certain threshold.
 289 Similarly, yield variations generated rather erratic results when using the RPR function of Fischer et al.
 290 (2007), especially with a $\pm 50\%$ yield variation. Mathematically, the linear function proposed by Fischer et
 291 al. (2007) has a general structure of $RPR = -0.14 * yield + 1.96$ (Table 1); hence if the primary crop yield
 292 values are increased, the estimated residues are bound to decrease.

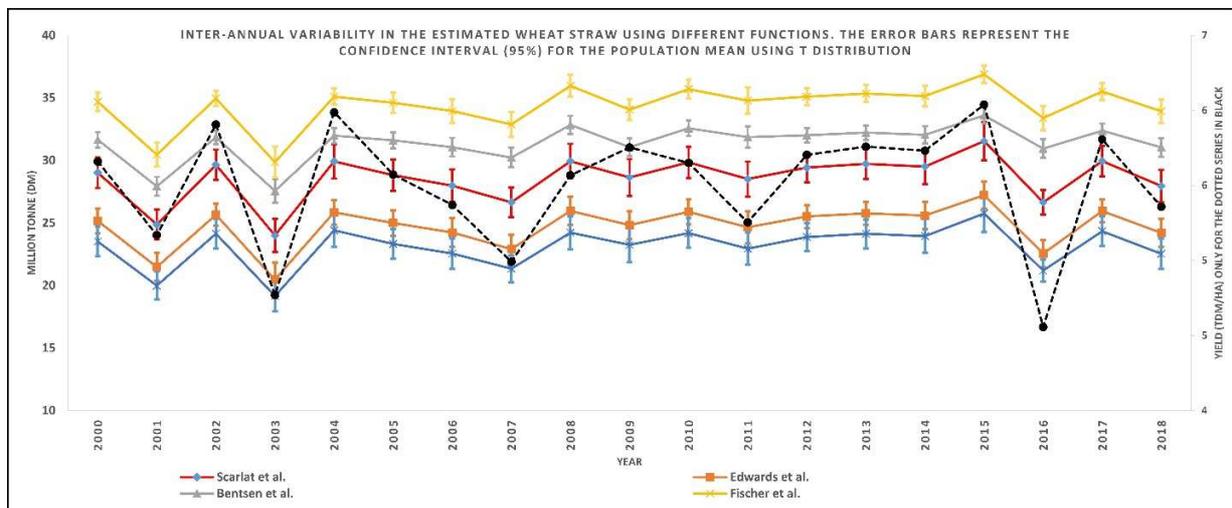


293
 294 **Figure 4: Observed variation in the estimated wheat residue (straw) over a period of 19 years by changing**
 295 **the primary crop yield by $\pm 10\%$ and $\pm 50\%$ of the original values (OAT analysis), using different**
 296 **estimation functions.**

297 **3.2 SQ-2 Uncertainties in primary crop yield overshadow the differences observed in the estimated CR**

298 The chart shown in Figure 5 highlights the inter-annual variability of residues estimated using different
 299 functions along with the 95% confidence interval shown as error bars. From the figure, it can be observed
 300 that the results obtained using the functions from Edwards et al. (2005) and García-Condado et al. (2019)
 301 are mostly overlapping in the confidence intervals. This might be because both functions use HI directly or
 302 indirectly to estimate the residues. In terms of inter-annual variation of estimated residues, sharp
 303 decreases were observed for the years 2001, 2003, and 2016. These decreases followed the sharp
 304 decreasing trend observed in the primary crop yield values (highlighted in the black dotted series).

305 However, this trend is not general; for example, the primary crop yield value increased in the year 2009,
 306 but the estimated residues for that year shows a decreasing trend using all the functions (SM2: Effect).



307
 308 **Figure 5: Average inter-annual variation of wheat straw using different empirical functions. The error bars**
 309 **represent the confidence interval at $\alpha = 0.05$**

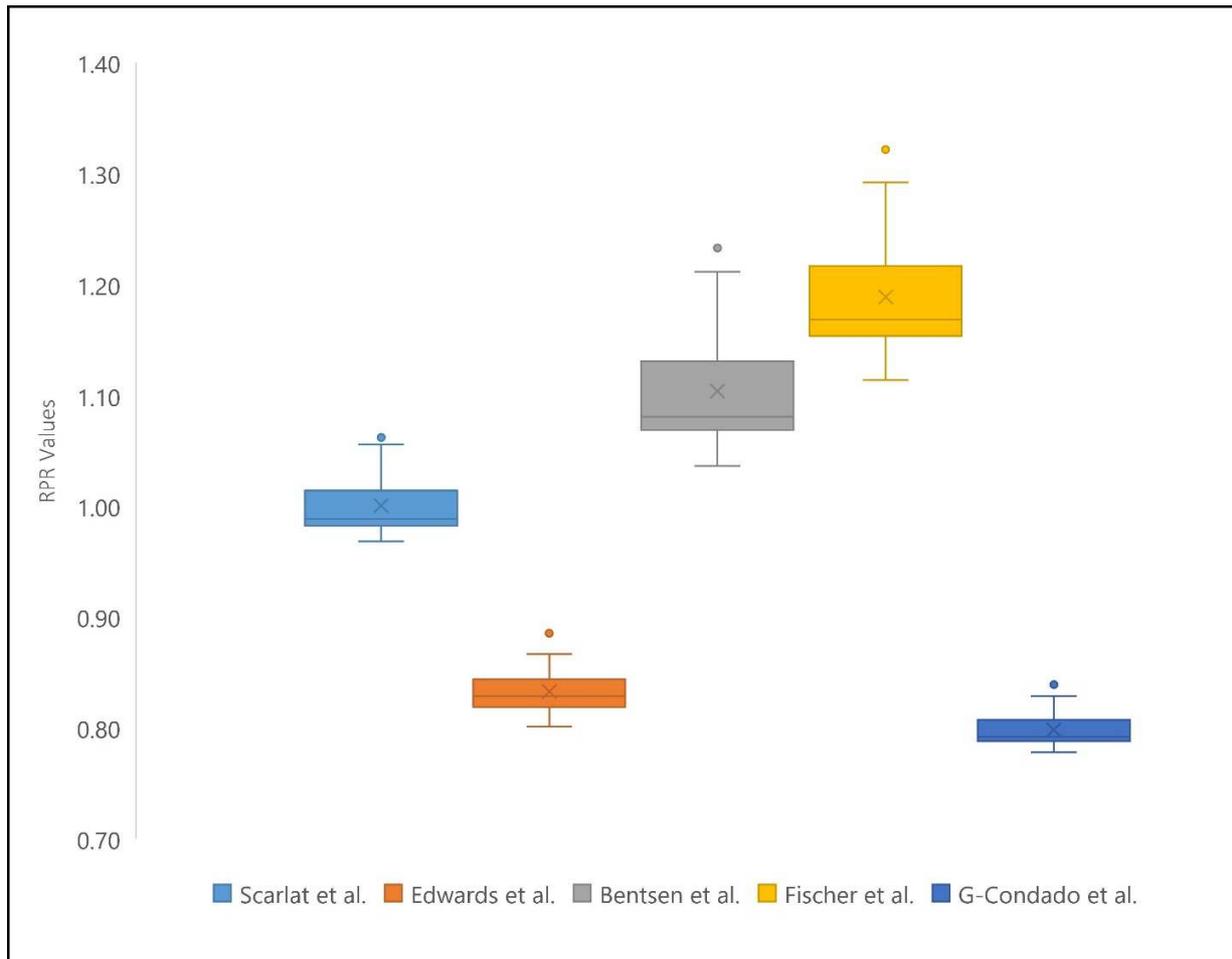
310 **3.3 SQ-3 Differences in the RPRs estimated using different estimation functions**

311 The results of the null-hypothesis test are shown as pairwise comparisons in Table 5 (SM2: T-test RPR).
 312 The results of the t-test revealed that for each pair compared, the CR estimates were significantly
 313 different, with $P(T < t) < 1.96$. Thus the null hypothesis (H_0 = There is no significant difference in the
 314 estimated RPR using different functions) was rejected, and the alternate hypothesis (H_1) was accepted. In
 315 other words, none of the results obtained with each RPR function presented in Table 1 can be considered
 316 equivalent, meaning that the function selected for estimating CR is not a simple choice without
 317 consequences.

318 This is further clarified in Figure 6, which illustrates, **as boxplots, the variability of RPR values over the**
 319 **period 2000 – 2018 using different functions (SM2). It can be noticed, among others, that no two boxes**
 320 **overlap with each other.** Figure 6 also illustrates that results from the functions of Bentsen et al. (2014)
 321 and Fischer et al. (2007) have broader ranges indicating a wider distribution and more scattered output
 322 results. Conversely, the short boxes in the functions of Edwards et al. (2005), García-Condado et al. (2019)
 323 and Scarlat et al. (2010) indicate that the RPR results range consistently hover around the center values.

324 **Table 5: Pairwise comparison of different functions used for estimating the RPR for wheat cereal. Values**
 325 **represent the $P(T \leq t)$ two-tail results, which are all significantly inferior to $P(T < t) < 1.96$.**

P ($\alpha=0.05$)	Scarlat et al. (2010)	Edwards et al. (2005)	Bentsen et al. (2014)	Fischer et al. (2007)	G-Condado et al. (2019)
Scarlat et al. (2010)		1.2×10^{-25}	3.9×10^{-12}	2.7×10^{-16}	3.2×10^{-24}
Edwards et al. (2005)			6.8×10^{-18}	7.8×10^{-20}	3.3×10^{-15}
Bentsen et al. (2014)				1.1×10^{-26}	8.8×10^{-18}
Fischer et al. (2007)					1.8×10^{-19}
G-Condado et al. (2019)					



327
 328 **Figure 6: RPR of the different functions for wheat residues represented as Box-plots. Crosses represent**
 329 **averages.**

330 The RPR functions developed by Bentsen et al. (2014) and Scarlat et al. (2010) are also accompanied by
 331 their coefficients of determination (R^2) values (Table 1), which at best reaches 0.52. This implies that
 332 approximately half of the observed variation in the estimated residues can be explained by the function's
 333 variable, here the yield. This makes the estimation functions highly uncertain. Furthermore, it is not
 334 always clear with these functions, whether they capture the entire generated residual biomass, or just the
 335 portion that is harvestable, as shown in Table 2. According to Kristensen Fløjgård (2012), this non-
 336 harvestable portion (or loss) can represent 10-15% of the overall CR in the case of cereals.

337 While carrying out such resource assessment studies at large geographic scales (country, continental,
 338 global), empirical or statistical functions as those used here remain the most convenient tool for CR
 339 estimation. However, as shown in this study, the functions available at present appear little reliable, and
 340 additional experimental research to improve these would be rather beneficial in the perspective of
 341 bioeconomy action plans.

342 **3.4 Perspectives**

343 Our results have shown that resource assessments for crop residues can be uncertain due to the diversity
 344 of available methods and the lack of empirical validation. One alternative could be to standardize the
 345 assessments by specifying the method to use for all types of residues. Having a harmonized method for
 346 large geographical scale resource assessment will indeed carry an uncertainty of its own due to several

347 factors such as heterogeneity in local farm management, differences in soil properties, crop genetics,
348 diverse climatic regime, among other things. It is expected that this uncertainty is going to be present
349 regardless of the choice of method. Thus, uncertainty accounting is recommended in such assessments to
350 provide the bioeconomy planners with a range or confidence intervals of these estimates. This can be
351 done using the standard uncertainty propagation methods (JCGM, 2008), as exemplified in the case of
352 assessment of primary forestry residues in France (Karan and Hamelin, 2020).

353 Furthermore, in large-scale assessments, the choice of resource estimation method should not be
354 random. When no clear evidence is available to support the obtained estimates, the use of the
355 precautionary principle is recommended to report estimates that are more conservative, according to the
356 intended use. In addition, an alternative ranking of methods in the perspective of bioeconomy can be
357 proposed to select the most appropriate estimation method. For example, Sanderson et al. (2015)
358 provided a ranking scheme for earth system models based on a stepwise model elimination procedure. A
359 similar approach could be adapted for selecting the most relevant method for estimating the crop residue
360 potential, building upon our wheat straw demonstration, but considering all types of crop residues.

361 **4. Conclusions**

362 A comprehensive assessment of crop residues theoretical potential was performed for metropolitan
363 France, considering 16 major crops. The spatially-explicit estimation of crop residues was performed at
364 the French departmental (NUTS-3) + regional level (NUTS-2). Empirical functions commonly used in the
365 literature were used to estimate the CR by considering a ratio (RPR), which partitions the total above-
366 ground biomass into primary crop yield (the main cereal component of the crop) and CR. The results and
367 uncertainties obtained with the different empirical functions were thoroughly analyzed.

368 The key conclusion of this study is that existing RPR functions, albeit rather unquestioned, are highly
369 unreliable and would greatly benefit from additional experimental research. In fact, we showed, with a
370 case study on wheat produced in France in the period 2000 – 2018, that none of the assessed functions
371 produced a CR estimate that can be considered as statistically comparable with one another.

372 **CRedit Author contributions**

373 *Shivesh Kishore Karan:* Conceptualization; Data curation; Formal analysis; Investigation; Methodology;
374 Software; Writing – original draft

375 *Lorie Hamelin:* Conceptualization; Funding acquisition; Investigation; Resources; Supervision; Writing –
376 review & editing

377 **Declaration of interest**

378 The authors declare no conflict of interest.

379 **Acknowledgments**

380 This work was carried out within the framework of the research project Cambioscop
381 (<https://cambioscop.cnrs.fr>) and was partly financed by the French National Research Agency, Programme
382 Investissement d'Avenir (ANR-17-MGPA-0006) and Region Occitanie (18015981). The authors
383 acknowledge the support provided by Sandrine Laguerre for the statistical part of the study.

384 **References**

385 Agreste, 2020. Cultures développées [WWW Document]. URL [https://agreste.agriculture.gouv.fr/agreste-](https://agreste.agriculture.gouv.fr/agreste-saiku/?plugin=true&query=query/open/SAANR_DEVELOPPE_2#query/open/SAANR_DEVELOPPE_2)
386 [saiku/?plugin=true&query=query/open/SAANR_DEVELOPPE_2#query/open/SAANR_DEVELOPPE_2](https://agreste.agriculture.gouv.fr/agreste-saiku/?plugin=true&query=query/open/SAANR_DEVELOPPE_2#query/open/SAANR_DEVELOPPE_2)
387 (accessed 8.26.20).

388 Bentsen, N.S., Felby, C., 2012. Biomass for energy in the European Union - A review of bioenergy resource

389 assessments. *Biotechnol. Biofuels* 5, 1–10. <https://doi.org/10.1186/1754-6834-5-25>

390 Bentsen, N.S., Felby, C., Thorsen, B.J., 2014. Agricultural residue production and potentials for energy and
391 materials services. *Prog. Energy Combust. Sci.* 40, 59–73. <https://doi.org/10.1016/j.pecs.2013.09.003>

392 Bisinella, V., Conradsen, K., Christensen, T.H., Astrup, T.F., 2016. A global approach for sparse
393 representation of uncertainty in Life Cycle Assessments of waste management systems. *Int. J. Life
394 Cycle Assess.* 21, 378–394. <https://doi.org/10.1007/s11367-015-1014-4>

395 Blanco-Canqui, H., 2013. Crop Residue Removal for Bioenergy Reduces Soil Carbon Pools: How Can We
396 Offset Carbon Losses? *BioEnergy Res.* 6, 358–371. <https://doi.org/10.1007/s12155-012-9221-3>

397 Bodirsky, B.L., Popp, A., Weindl, I., Dietrich, J.P., Rolinski, S., Scheffele, L., Schmitz, C., Lotze-Campen, H.,
398 2012. N₂O emissions from the global agricultural nitrogen cycle – current state and future
399 scenarios. *Biogeosciences* 9, 4169–4197. <https://doi.org/10.5194/bg-9-4169-2012>

400 CCOF, 2013. (California Certified Organic Farmers), Average dry matter percentages for different livestock
401 feeds.

402 Daioglou, V., Stehfest, E., Wicke, B., Faaij, A., van Vuuren, D.P., 2016. Projections of the availability and
403 cost of residues from agriculture and forestry. *GCB Bioenergy* 8, 456–470.
404 <https://doi.org/10.1111/gcbb.12285>

405 de Vries, S.S., 1999. Kansen voor bioenergie uit biomassa. Delft University of Technology.

406 Edwards, R.A.H., Šúri, M., Huld, T.A., Dallemand, J.F., 2005. GIS-based assessment of cereal straw energy
407 resource in the European Union. *Proc. 14th Eur. Biomass Conf. Exhib. Biomass Energy, Ind. Clim.
408 Prot.*

409 Einarsson, R., Persson, U.M., 2017. Analyzing key constraints to biogas production from crop residues and
410 manure in the EU - A spatially explicit model. *PLoS One* 12, 1–23.
411 <https://doi.org/10.1371/journal.pone.0171001>

412 Ericsson, K., Nilsson, L.J., 2006. Assessment of the potential biomass supply in Europe using a resource-
413 focused approach. *Biomass and Bioenergy*. <https://doi.org/10.1016/j.biombioe.2005.09.001>

414 European Commission, 2020. Statistical Factsheet. [https://ec.europa.eu/info/sites/info/files/food-farming-
415 fisheries/farming/documents/agri-statistical-factsheet-eu_en.pdf](https://ec.europa.eu/info/sites/info/files/food-farming-fisheries/farming/documents/agri-statistical-factsheet-eu_en.pdf)

416 Eurostat, 2020a. Statistical regions in the European Union and partner countries. NUTS and statistical
417 regions 2021. <https://doi.org/10.2785/850262>

418 Eurostat, 2020b. Energy Statistics [WWW Document]. URL
419 <https://ec.europa.eu/eurostat/web/energy/data/database> (accessed 6.11.20).

420 FAO, 2017. Soil Organic Carbon: the hidden potential. Food and Agricultural Organization of the United
421 Nations. Rome Italy.

422 Fischer, G., Hizsnyik, E., Prieler, S., van Velthuisen, H., 2007. Assessment of biomass potentials for bio- fuel
423 feedstock production in Europe : Methodology and results, Refuel.

424 García-Condado, S., López-Lozano, R., Panarello, L., Cerrani, I., Nisini, L., Zucchini, A., Van der Velde, M.,
425 Baruth, B., 2019. Assessing lignocellulosic biomass production from crop residues in the European
426 Union: Modelling, analysis of the current scenario and drivers of interannual variability. *GCB
427 Bioenergy* 11, 809–831. <https://doi.org/10.1111/gcbb.12604>

428 Haase, M., Rösch, C., Ketzer, D., 2016. GIS-based assessment of sustainable crop residue potentials in
429 European regions. *Biomass and Bioenergy* 86, 156–171.
430 <https://doi.org/10.1016/j.biombioe.2016.01.020>

431 Haberl, H., Beringer, T., Bhattacharya, S.C., Erb, K.H., Hoogwijk, M., 2010. The global technical potential of
432 bio-energy in 2050 considering sustainability constraints. *Curr. Opin. Environ. Sustain.* 2, 394–403.

433 <https://doi.org/10.1016/j.cosust.2010.10.007>

434 Hamelin, L., Borzęcka, M., Kozak, M., Pudełko, R., 2019. A spatial approach to bioeconomy: Quantifying
435 the residual biomass potential in the EU-27. *Renew. Sustain. Energy Rev.* 100, 127–142.
436 <https://doi.org/10.1016/j.rser.2018.10.017>

437 Hamelin, L., Jørgensen, U., Petersen, B.M., Olesen, J.E., Wenzel, H., 2012. Modelling the carbon and
438 nitrogen balances of direct land use changes from energy crops in Denmark: A consequential life
439 cycle inventory. *GCB Bioenergy* 4, 889–907. <https://doi.org/10.1111/j.1757-1707.2012.01174.x>

440 Hansen, J.H., Hamelin, L., Taghizadeh-Toosi, A., Olesen, J.E., Wenzel, H., 2020. Agricultural residues
441 bioenergy potential that sustain soil carbon depends on energy conversion pathways. *GCB Bioenergy*
442 *gcb.12733*. <https://doi.org/10.1111/gcbb.12733>

443 Hay, R.K.M., 1995. Harvest index: a review of its use in plant breeding and crop physiology. *Ann. Appl. Biol.*
444 126, 197–216. <https://doi.org/10.1111/j.1744-7348.1995.tb05015.x>

445 JCGM, 2008. Evaluation of measurement data — Guide to the expression of uncertainty in measurement.
446 *Int. Organ. Stand. Geneva ISBN 50, 134*. <https://doi.org/10.1373/clinchem.2003.030528>

447 Kadam, K.L., McMillan, J.D., 2003. Availability of corn stover as a sustainable feedstock for bioethanol
448 production. *Bioresour. Technol.* 88, 17–25. [https://doi.org/10.1016/S0960-8524\(02\)00269-9](https://doi.org/10.1016/S0960-8524(02)00269-9)

449 Karan, S.K., Hamelin, L., 2020. Towards local bioeconomy: A stepwise framework for high-resolution
450 spatial quantification of forestry residues. *Renew. Sustain. Energy Rev.* 134, 110350.
451 <https://doi.org/10.1016/j.rser.2020.110350>

452 Karaosmanoğlu, F., Tetik, E., Gürboy, B., Şanlı, İ., 1999. Characterization of the straw stalk of the rapeseed
453 plant as a biomass energy source. *Energy Sources* 21, 801–810.
454 <https://doi.org/10.1080/00908319950014353>

455 Koga, N., 2008. An energy balance under a conventional crop rotation system in northern Japan:
456 Perspectives on fuel ethanol production from sugar beet. *Agric. Ecosyst. Environ.* 125, 101–110.
457 <https://doi.org/10.1016/j.agee.2007.12.002>

458 Kristensen Fløjgård, E., 2012. Tekniske muligheder for at bjerge en større del af den producerede
459 halmmængde.

460 Lindley, J.A., Smith, G.M., 1988. Heat Energy From Sunflower Residue. *Trans. Am. Soc. Agric. Eng.* 31,
461 1197–1202. <https://doi.org/10.13031/2013.30844>

462 McCartney, D.H., Block, H.C., Dubeski, P.L., Ohama, A.J., 2006. Review: The composition and availability of
463 straw and chaff from small grain cereals for beef cattle in western Canada. *Can. J. Anim. Sci.* 86, 443–
464 455. <https://doi.org/10.4141/A05-092>

465 Monforti, F., Bódis, K., Scarlat, N., Dallemand, J.F., 2013. The possible contribution of agricultural crop
466 residues to renewable energy targets in Europe: A spatially explicit study. *Renew. Sustain. Energy*
467 *Rev.* 19, 666–677. <https://doi.org/10.1016/j.rser.2012.11.060>

468 Muth, D.J., Bryden, K.M., Nelson, R.G., 2013. Sustainable agricultural residue removal for bioenergy: A
469 spatially comprehensive US national assessment. *Appl. Energy* 102, 403–417.
470 <https://doi.org/10.1016/j.apenergy.2012.07.028>

471 Özyüğüran, A., Yaman, S., Küçükbayrak, S., 2018. Prediction of calorific value of biomass based on
472 elemental analysis. *Int. Adv. Res. Eng. J.* 02, 2.

473 Phyllis2, 2020. Phyllis2 - Database for (treated) biomass, algae, feedstocks for biogas production and
474 biochar, <https://phyllis.nl/> [WWW Document]. URL <https://phyllis.nl/> (accessed 5.13.20).

475 Ronzon, T., Piotrowski, S., 2017. Are Primary Agricultural Residues Promising Feedstock for the European
476 Bioeconomy? *Ind. Biotechnol.* 13, 113–127. <https://doi.org/10.1089/ind.2017.29078.tro>

477 Ruiz, P., Sgobbi, A., Nijs, W., Thiel, C., Dalla Longa, F., Kober, T., Elbersen, B., Hengeveld, G., Alterra, 2015.
478 The JRC-EU-TIMES model. Bioenergy potentials for EU and neighbouring countries. European
479 Commission Joint Research Centre. <https://doi.org/10.2790/39014>

480 Sanderson, B.M., Knutti, R., Caldwell, P., 2015. A Representative Democracy to Reduce Interdependency in
481 a Multimodel Ensemble. *J. Clim.* 28, 5171–5194. <https://doi.org/10.1175/JCLI-D-14-00362.1>

482 Scarlat, N., Fahl, F., Lugato, E., Monforti-Ferrario, F., Dallemand, J.F., 2019. Integrated and spatially explicit
483 assessment of sustainable crop residues potential in Europe. *Biomass and Bioenergy* 122, 257–269.
484 <https://doi.org/10.1016/j.biombioe.2019.01.021>

485 Scarlat, N., Martinov, M., Dallemand, J.F., 2010. Assessment of the availability of agricultural crop residues
486 in the European Union: Potential and limitations for bioenergy use. *Waste Manag.* 30, 1889–1897.
487 <https://doi.org/10.1016/j.wasman.2010.04.016>

488 Searle, S., Malins, C., 2015. A reassessment of global bioenergy potential in 2050. *GCB Bioenergy* 7, 328–
489 336. <https://doi.org/10.1111/gcbb.12141>

490 Sommer, S.G., Hamelin, L., Olesen, J.E., Montes, F., Jia, W., Chen, Q., Triolo, J.M., 2016. Agricultural Waste
491 Biomass, in: *Supply Chain Management for Sustainable Food Networks*. John Wiley & Sons, Ltd,
492 Chichester, UK, pp. 67–106. <https://doi.org/10.1002/9781118937495.ch3>

493 Teagasc, 2010. *Straw for Energy*.

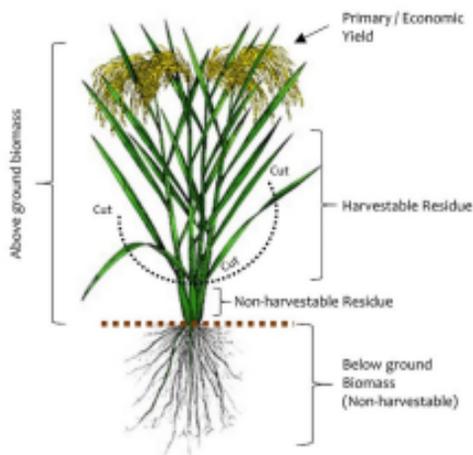
494 Tonini, D., Hamelin, L., Alvarado-Morales, M., Astrup, T.F., 2016a. GHG emission factors for bioelectricity,
495 biomethane, and bioethanol quantified for 24 biomass substrates with consequential life-cycle
496 assessment. *Bioresour. Technol.* 208, 123–133. <https://doi.org/10.1016/j.biortech.2016.02.052>

497 Tonini, D., Hamelin, L., Astrup, T.F., 2016b. Environmental implications of the use of agro-industrial
498 residues for biorefineries: application of a deterministic model for indirect land-use changes. *GCB*
499 *Bioenergy* 8, 690–706. <https://doi.org/10.1111/gcbb.12290>

500 Williams, C.L., Westover, T.L., Emerson, R.M., Tumuluru, J.S., Li, C., 2016. Sources of Biomass Feedstock
501 Variability and the Potential Impact on Biofuels Production. *Bioenergy Res.* 9, 1–14.
502 <https://doi.org/10.1007/s12155-015-9694-y>

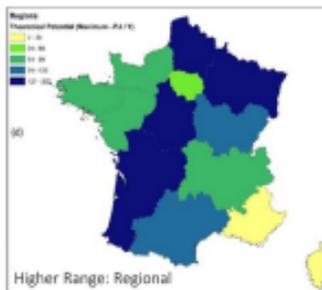
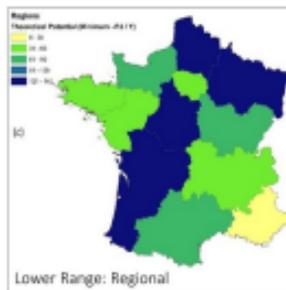
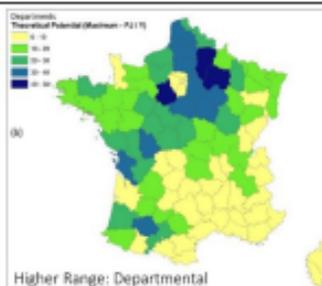
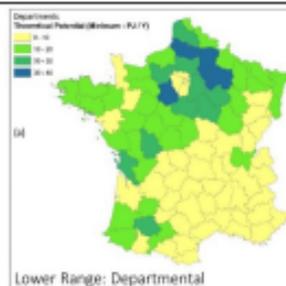
503 Wirsenius, S., 2000. Human use of land and organic materials: modeling the turnover of biomass in the
504 global food system. *Doktorsavhandlingar vid Chalmers Tek. Hogsks.* Chalmers University of
505 Technology and Goteborg University.

506 WRI, 2020. *World Greenhouse Gas Emissions: 2016* | World Resources Institute [WWW Document]. URL
507 <https://www.wri.org/resources/data-visualizations/world-greenhouse-gas-emissions-2016> (accessed
508 7.14.20).



- Method 1
- Method 2
- Method 3

- Residue estimation using different methods.
- Comparison of methods and results.
- Uncertainty analysis.



Spatial quantification of crop residues illustrated as a case study for France (Lower and Higher ranges for French departmental and regional levels)

