



**HAL**  
open science

## Search for top-down and bottom-up drivers of latitudinal trends in insect herbivory in oak trees in Europe

Elena Valdés-correcher, Xoaquín Moreira, Laurent Augusto, Luc L. Barbaro, Christophe Bouget, Olivier Bouriaud, Manuela Branco, Giada Centenaro, Gyorgy Csoka, Thomas Damestoy, et al.

### ► To cite this version:

Elena Valdés-correcher, Xoaquín Moreira, Laurent Augusto, Luc L. Barbaro, Christophe Bouget, et al.. Search for top-down and bottom-up drivers of latitudinal trends in insect herbivory in oak trees in Europe. *Global Ecology and Biogeography*, 2021, 30 (3), pp.651-665. 10.1111/geb.13244 . hal-03103928

**HAL Id: hal-03103928**

**<https://hal.inrae.fr/hal-03103928>**

Submitted on 25 May 2022

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

## 1 **Search for top-down and bottom-up drivers of latitudinal trends in insect** 2 **herbivory in oak trees in Europe**

3 Elena Valdés-Correcher<sup>1</sup>, Xoaquín Moreira<sup>2</sup>, Laurent Augusto<sup>3</sup>, Luc Barbaro<sup>4,5</sup>, Christophe Bouget<sup>6</sup>, Olivier  
4 Bouriaud<sup>7</sup>, Manuela Branco<sup>8</sup>, Giada Centenaro<sup>9</sup>, György Csóka<sup>10</sup>, Thomas Damestoy<sup>1</sup>, Jovan  
5 Dobrosavljević<sup>11</sup>, Mihai-Leonard Duduman<sup>7</sup>, Anne-Maïmiti Dulaurent<sup>12</sup>, Csaba B. Eötvös<sup>10</sup>, Maria Faticov<sup>13</sup>,  
6 Marco Ferrante<sup>14,15</sup>, Ágnes Fürjes-Mikó<sup>10</sup>, Andrea Galmán<sup>2</sup>, Martin M. Gossner<sup>16</sup>, Arndt Hampe<sup>1</sup>, Deborah  
7 Harvey<sup>17</sup>, Andrew Gordon Howe<sup>18</sup>, Yasmine Kadiri<sup>1</sup>, Michèle Kaennel Dobbertin<sup>16</sup>, Julia Koricheva<sup>17</sup>,  
8 Alexander Kozel<sup>19</sup>, Mikhail V. Kozlov<sup>20</sup>, Gábor L. Lövei<sup>14</sup>, Daniela Lupaştean<sup>7</sup>, Slobodan Milanović<sup>21,22</sup>, Anna  
9 Mrazova<sup>23,24</sup>, Lars Opgennoorth<sup>25,26</sup>, Juha-Matti Pitkänen<sup>27</sup>, Anna Popova<sup>20</sup>, Marija Popović<sup>21</sup>, Andreas  
10 Prinzing<sup>28</sup>, Valentin Queloz<sup>16</sup>, Tomas Roslin<sup>29</sup>, Aurélien Sallé<sup>30</sup>, Katerina Sam<sup>23,24</sup>, Michael Scherer-  
11 Lorenzen<sup>31</sup>, Andreas Schuldt<sup>32</sup>, Andrey Selikhovkin<sup>33,34</sup>, Lassi Suominen<sup>35</sup>, Ayco J. M. Tack<sup>13</sup>, Marketa  
12 Tahadlova<sup>23,24</sup>, Rebecca Thomas<sup>17</sup> and Bastien Castagneyrol<sup>1\*</sup>.

- 13  
14 <sup>1</sup> BIOGECO, INRAE, Univ. Bordeaux, 33610 Cestas, France  
15 <sup>2</sup> Misión Biológica de Galicia (MBG-CSIC), Pontevedra, Galicia, Spain  
16 <sup>3</sup> INRAE, Bordeaux Sciences Agro, UMR 1391 ISPA, 33882 Villenave-d'Ornon Cedex, France  
17 <sup>4</sup> DYNAFOR, University of Toulouse, INRAE, Castanet-Tolosan, France  
18 <sup>5</sup> CESCO, Sorbonne University, MNHN, CNRS, Paris, France  
19 <sup>6</sup> INRAE, 'Forest Ecosystems' Research Unit – Biodiversity team Domaine des Barres F-45290 Nogent-sur-  
20 Vernisson FRANCE  
21 <sup>7</sup> "Ştefan cel Mare" University of Suceava, Forestry Faculty, Applied Ecology Laboratory, Universităţii Street  
22 13, Suceava, Romania  
23 <sup>8</sup> Centro de Estudos Florestais, Instituto Superior de Agronomia, Universidade de Lisboa, Portugal  
24 <sup>9</sup> Department of Land, Environment, Agriculture and Forestry, University of Padova, Viale dell' Università  
25 16, I-35020 Legnaro (PD), Italy  
26 <sup>10</sup> NARIC Forest Research Institute, Department of Forest Protection, Hegyalja str. 18, 3232 Mátrafüred,  
27 Hungary  
28 <sup>11</sup> University of Belgrade, Faculty of Forestry, Kneza Višeslava 1, 11000 Belgrade, Serbia  
29 <sup>12</sup> UniLaSalle, AGHYLE, UP.2018.C101, SFR Condorcet FR CNRS 3417, FR-60026 Beauvais, France  
30 <sup>13</sup> Department of Ecology, Environment and Plant Sciences, Stockholm University, SE-106 91 Stockholm,  
31 Sweden  
32 <sup>14</sup> Department of Agroecology, Aarhus University, Flakkebjerg Research Centre, DK-4200 Slagelse,  
33 Denmark  
34 <sup>15</sup> cE3c – Centre for Ecology, Evolution and Environmental Changes/Azorean Biodiversity Group and  
35 University of the Azores, Azores, Portugal  
36 <sup>16</sup> Swiss Federal Research Institute WSL, 8903 Birmensdorf  
37 <sup>17</sup> Department of Biological Sciences, Royal Holloway University of London, Egham, UK TW20 0EX  
38 <sup>18</sup> Department of Geosciences and Natural Resource Management, University of Copenhagen,  
39 Frederiksberg, Denmark

- 40 <sup>19</sup> Belarusian State Technological University, Department of Forest Protection and Wood Science,  
41 Sverdlova str. 13a, 220006 Minsk, Belarus
- 42 <sup>20</sup> Department of Biology, University of Turku, 20014 Turku, Finland
- 43 <sup>21</sup> University of Belgrade, Faculty of Forestry, Kneza Višeslava 1, 11000 Belgrade, Serbia
- 44 <sup>22</sup> Mendel University, Faculty of Forestry and Wood Technology, Zemedelska 3, 61 300 Brno, Czech  
45 Republic
- 46 <sup>23</sup> Biology Centre of Czech Academy of Sciences, Entomology Institute, Ceske Budejovice, 37005, Czech  
47 Republic
- 48 <sup>24</sup> University of South Bohemia, Faculty of Science, Ceske Budejovice, 37005, Czech Republic
- 49 <sup>25</sup> Swiss Federal Research Institute WSL, Biodiversity and Conservation Biology, Ecological Genetics,  
50 Zürcherstrasse 111, 8903 Birmensdorf, Switzerland
- 51 <sup>26</sup> Department of Ecology, Philipps-Universität Marburg, Karl-von-Frisch Strasse 8, 35043 Marburg
- 52 <sup>27</sup> Department of Ecology, Spatial Foodweb Ecology Group, Department of Agricultural Sciences, PO Box  
53 27 (Latokartanonkaari 5), FI-00014 University of Helsinki, Finland
- 54 <sup>28</sup> Research Unit “Ecosystems, Biodiversity, Evolution”; Université de Rennes 1/ Centre National de la  
55 Recherche Scientifique; Campus de Beaulieu, Bâtiment 14A, 35042 Rennes, France
- 56 <sup>29</sup> Department of Ecology, P.O. Box 7044, Swedish University of Agricultural Sciences, SE-750 07 Uppsala,  
57 Sweden
- 58 <sup>30</sup> Laboratoire de Biologie des Ligneux et des Grandes Cultures, INRAE, Université d'Orléans, 45067,  
59 Orléans, France
- 60 <sup>31</sup> Geobotany, Faculty of Biology, University of Freiburg, Schaezlestr. 1, 79104 Freiburg, Germany
- 61 <sup>32</sup> Georg-August-University Göttingen, Forest Nature Conservation, Büsgenweg 3, 37077 Göttingen,  
62 Germany
- 63 <sup>33</sup> St. Petersburg Forest Technical University 5 Institutskiy lane, St. Petersburg 194021, Russia
- 64 <sup>34</sup> St. Petersburg State University 13B Universitetskaya Emb., St. Petersburg 199034, Russia
- 65 <sup>35</sup> Salo upper secondary school, Kaherinkatu 2, 24130 Salo, Finland

66  
67 \* Author for correspondence: [bastien.castagneyrol@inrae.fr](mailto:bastien.castagneyrol@inrae.fr)

## 68 69 **Acknowledgements**

70 This study has been carried out with financial support from the French National Research Agency  
71 (ANR) in the frame of the Investments for the future Programme, within the Cluster of Excellence  
72 COTE (ANR-10-LABX-45). E.V.C was founded by the BiodivERsA project SPONFOREST  
73 (BiodivERsA3-2015-58). The authors warmly thank all young European citizens and their teachers  
74 who have made this study possible. They also thank professional scientists who have kindly  
75 accepted to participate in this study: Stefan K. Müller (Freie evangelische Schule Lörrach), Olga  
76 Mijón Pedreira (teacher IES Rosais 2, Vigo-Spain) and Mickael Pihain (Research Unit  
77 “Ecosystèmes, Biodiversité, Evolution”, University of Rennes 1 / CNRS, 35042 Rennes, France),  
78 and Chloe Mendiondo and Claire Coliaux (Department of Agroecology, Aarhus University,  
79 Flakkebjerg Research Centre, DK-4200 Slagelse, Denmark). The authors declare no competing  
80 financial interests.

81

82 **Biosketch**

83 *Elena Valdés-Correcher*: Interested on plant-herbivore interactions. More specifically, she  
84 investigate the effect of different drivers of plant herbivore interactions including the effect of  
85 landscape composition, climate and tree genotype on plant-herbivore relationships.

86

87 *Bastien Castagneyrol*: Interested on citizen science and on the ecology of plant-herbivore  
88 interactions including predation, insect herbivory and leaf traits and how these relationships are  
89 influenced by factors that act at different scales.

90

91 *Xoaquín Moreira*: Interested on plant-herbivore interactions. More specifically, interested on the  
92 effect of latitude and elevation on biotic relationships.

1 **Title:** Search for top-down and bottom-up drivers of latitudinal trends in insect herbivory in oak  
2 trees in Europe

3 **Running title:** Plant-herbivore interactions in oak in Europe

#### 4 **Abstract**

##### 5 **Aim**

6 The strength of species interactions is traditionally expected to increase toward the Equator.  
7 However, recent studies have reported opposite or inconsistent latitudinal trends in the bottom-  
8 up (plant quality) and top-down (natural enemies) forces driving herbivory. In addition, these  
9 forces have rarely been studied together thus limiting previous attempts to understand the effect  
10 of large-scale climatic gradients on herbivory.

##### 11 **Location**

12 Europe

##### 13 **Time period**

14 2018-2019

##### 15 **Major taxa studied**

16 *Quercus robur*

##### 17 **Methods**

18 We simultaneously tested for latitudinal variation in plant-herbivore-natural enemy interactions.  
19 We further investigated the underlying climatic factors associated with variation in herbivory, leaf  
20 chemistry and attack rates in *Quercus robur* across its complete latitudinal range in Europe. We  
21 quantified insect leaf damage and the incidence of specialist herbivores as well as leaf chemistry  
22 and bird attack rates on dummy caterpillars on 261 oak trees.

##### 23 **Results**

24 Climatic factors rather than latitude *per se* were the best predictors of the large-scale  
25 (geographical) variation in the incidence of gall-inducers and leaf-miners as well as in leaf  
26 nutritional content. However, leaf damage, plant chemical defences (leaf phenolics) and bird  
27 attack rates were not influenced by climatic factors or latitude. The incidence of leaf-miners  
28 increased with increasing concentrations of hydrolysable tannins, whereas the incidence of gall-

29 inducers increased with increasing leaf soluble sugar concentration and decreased with increasing  
30 leaf C:N ratios and lignins. However, leaf traits and bird attack rates did not vary with leaf damage.

### 31 **Main conclusions**

32 These findings help to refine our understanding of the bottom-up and top-down mechanisms  
33 driving geographical variation in plant-herbivore interactions, and urge for further examination  
34 of the drivers of herbivory on trees.

35 **Key words:** leaf chemistry, plant defences, avian insectivory, climate, artificial prey

### 36 **Introduction**

37 Ecological theory predicts that the strength of species interactions increases toward the Equator  
38 due to warmer temperatures, longer growing seasons, and higher species abundance and  
39 diversity at lower latitudes (Janzen, 1970; Schemske *et al.*, 2009). Plant species at lower latitudes  
40 commonly experience higher rates of herbivory than plants growing further from the equator  
41 (Coley & Barone, 1996; Schemske *et al.*, 2009; Lim *et al.*, 2015; Moreira *et al.*, 2018) and thus  
42 tropical plant species may evolve higher levels of anti-herbivore defences (Johnson & Rasmann,  
43 2011; Pearse & Hipp, 2012; Abdala-Roberts *et al.*, 2016; Hahn *et al.*, 2019). While early reviews  
44 reported patterns supporting these predictions (Coley & Aide, 1991; Coley & Barone, 1996; Dyer  
45 & Coley, 2002), several studies in recent decades have found either no evidence for a latitudinal  
46 gradient in herbivory and plant defences (Moles & Westoby, 2003; Gaston *et al.*, 2004; Moles *et al.*,  
47 2011) or increase in herbivory and defences with latitude (Gaston *et al.*, 2004; Stark *et al.*,  
48 2008; Adams *et al.*, 2009; Martz *et al.*, 2009; Pennings *et al.*, 2009; Del-Val & Armesto, 2010;  
49 Woods *et al.*, 2012; Moreira *et al.*, 2018, 2020). Given these inconsistencies, it is of great  
50 importance to identify the mechanisms underlying the substantial variation in herbivory and plant  
51 defences across latitudes, as herbivory is an important ecological process that modulates primary  
52 productivity by altering the recruitment, mortality and growth of plants.

53 Latitudinal gradients can be used as 'natural laboratories' to study the relationship between  
54 climate and plant-herbivore interactions (De Frenne *et al.*, 2013; Kozlov *et al.*, 2015; Lim *et al.*,  
55 2015; Moreira *et al.*, 2018). In the northern extratropical hemisphere, mean annual temperature  
56 drops by 0.73 °C and mean annual precipitation by 4.04 mm per degree of latitude northward (De  
57 Frenne *et al.*, 2013). Latitudinal variation in plant-herbivore interactions is therefore generally  
58 associated with large-scale variability in climatic conditions (Moreira *et al.*, 2018) and numerous  
59 studies demonstrate an effect of temperature and precipitation on plant traits (e.g. leaf N,  
60 phenolic compounds) (Chen *et al.*, 2013; Holopainen *et al.*, 2018; Gely *et al.*, 2019) and herbivory  
61 (Jamieson *et al.*, 2015; Gely *et al.*, 2019). However, many regions deviate from the global trend in  
62 temperature and precipitation toward higher latitudes due to their proximity to oceans or the

63 presence of mountains (De Frenne *et al.*, 2013), which can markedly change the relationship  
64 between latitude and plant-herbivore-predator interactions (Roslin *et al.*, 2017; Loughnan &  
65 Williams, 2019; Moreira *et al.*, 2019).

66 Recent work identified several potential sources of variation in the reported directions and  
67 strengths of latitudinal gradients in herbivory and plant defences (Johnson & Rasmann, 2011;  
68 Anstett *et al.*, 2016). First, theory on latitudinal gradients in herbivory and plant defences assumes  
69 a plant-centred equilibrium in which plants at low latitudes have adapted to higher herbivory  
70 levels by evolving stronger defences. However, most studies have measured either herbivory  
71 patterns or plant defences, but not both (but see Anstett *et al.*, 2015; Moreira *et al.*, 2018),  
72 leading to an incomplete understanding of the relationship between latitudinal clines and plant-  
73 herbivore interactions. Second, little attention has been paid to latitudinal variation in tritrophic  
74 dynamics (Roslin *et al.*, 2017). Herbivore natural enemies, however, can drastically modify  
75 tritrophic interactions by suppressing herbivore populations or reducing herbivore feeding  
76 (Rosenheim, 1998; Maguire *et al.*, 2015). In the few published studies exploring latitudinal  
77 patterns in natural enemy activity, authors have found no variation in parasitism (Dyer & Coley,  
78 2002; Moreira *et al.*, 2015), lower attack rates on artificial prey by ants (Roslin *et al.*, 2017), and  
79 higher (Zvereva *et al.*, 2019) or no variation (Roslin *et al.*, 2017) in attack rates on artificial prey  
80 by birds with increasing latitude. Third, while external feeders are directly exposed to enemies  
81 and adverse abiotic conditions, internal feeders (e.g., leaf-mining and gall-inducing insect  
82 herbivores) benefit from a buffered microhabitat and relative protection against enemies. It is  
83 therefore likely that latitudinal trends in tritrophic interactions would vary across herbivore  
84 feeding guilds. Thus, considering bottom-up and top-down forces simultaneously could be crucial  
85 for a comprehensive understanding of latitudinal clines in tritrophic interactions.

86 We aimed to test for latitudinal variation in plant-herbivore-natural enemy (*i.e.*, tritrophic)  
87 interactions, as well as the underlying climatic factors associated with variation in herbivory,  
88 bottom-up and top-down forces in the pedunculate oak (*Quercus robur*), a long-lived, common  
89 European tree. In particular, we asked the following questions: (1) Are there latitudinal clines in  
90 herbivory? (2) Is latitudinal variation in leaf chemical traits (bottom-up effects) and/or bird attack  
91 rates (top-down effects) on herbivorous insects associated with latitudinal variation in herbivory?  
92 (3) Are climatic correlates of latitude associated with clines in herbivory, leaf chemical traits and  
93 attack rates? We used data collected by professional scientists and schoolchildren across major  
94 parts of the geographical distribution range of *Q. robur*. We quantified insect leaf herbivory, leaf  
95 chemical traits (soluble sugars, nutrients and phenolics) and attack rates on dummy caterpillars  
96 placed on mature oak trees. Overall, our study attempted to refine our understanding of bottom-  
97 up and top-down mechanisms that may drive geographical variation in plant-herbivore  
98 interactions.

99 **Material and methods**

100 The present study involved 30 professional scientists from 14 countries and 82 school teachers  
101 (with their pupils) from 10 countries, giving a total of 112 partners from 17 European countries  
102 and covering most of the native geographic range of the pedunculate oak (**Figure 1**). Every partner  
103 received detailed instructions at the beginning of the project (Castagneyrol *et al.*, 2019). Here, we  
104 only provide a summary of these instructions. Only project partners who provided data that could  
105 be used in the present article were included.

106 **Target species**

107 The pedunculate oak is one of the dominant deciduous tree species in European forests and is of  
108 high ecological, economic and symbolic importance (Eaton *et al.*, 2016). Its distribution ranges  
109 from Central Spain (39°N) to southern Fennoscandia (62°N), thus this species experiences variable  
110 climatic conditions (Petit *et al.*, 2002). Pedunculate oak supports a large community of specialist  
111 and generalist herbivorous insects; especially suckers, chewers, skeletonizers, gall-inducers and  
112 leaf-miners (Southwood *et al.*, 2005; Moreira *et al.*, 2018), as well as xylophagous species  
113 (Marković & Stojanović, 2011). The wide distribution of pedunculate oak and the high diversity of  
114 associated herbivorous insects make it a suitable model species for research on the effect of  
115 climate on biotic interactions.

116 In total, the study included 261 mature oak trees surveyed by professional scientists (n = 115) and  
117 schoolchildren (n = 146) in 2018 (n = 148) and 2019 (n = 113) (**Figure 1**). However, not every  
118 partner measured or provided material allowing measurement of herbivory, bird attack rates and  
119 leaf chemistry simultaneously on every tree (Figure S1.1a, b and c, supplementary material).

120 **Attack rates on dummy caterpillars**

121 To control for latitudinal variation in environmental conditions, we matched the start of the  
122 experiment in each locality to the phenology of local oak trees. Six weeks after oak budburst,  
123 partners installed 20 dummy caterpillars per tree, *i.e.*, five caterpillars on each of four branches  
124 (facing north, south, east and west) with a minimum distance of 15 cm between caterpillars.

125 The project coordinators provided the same green plasticine (Staedler, Noris Club 8421, green[5])  
126 to all partners to make the caterpillars. In order to standardize caterpillar size among partners,  
127 we made caterpillars from a 1 cm diameter ball of plasticine, and gently pressed/rolled this along  
128 a 12 cm long metallic wire until a 3 cm long caterpillar was obtained, with the wire in its center.  
129 Partners attached the caterpillars to branches by twisting the wire and left the caterpillars on  
130 trees for 15 days before recording predation marks. A second survey using the same procedure  
131 immediately followed the first one. In 2018, schoolchildren photographed every caterpillar with  
132 the suspected attack marks from any potential predatory taxon. In 2019, both schoolchildren and  
133 professional scientists sent caterpillars back to the project coordinators.



134 In order to be consistent and reduce bias due to multiple observers, photos and dummy  
135 caterpillars were screened by a single trained observer (first author, EVC). For each oak tree and  
136 survey period, we assessed attack rate as the proportion of dummy caterpillars with at least one  
137 attack mark. Although we asked partners to record attack rate marks left by different types of  
138 predators (in particular birds and arthropods), attacks by arthropod predators could not be  
139 verified on photos because of their low resolution. In addition, the relevance of marks left by  
140 arthropods on plasticine model prey has recently been questioned, in particular after mandibular  
141 marks were observed on lizards or frog models (Rößler *et al.*, 2018). For these reasons, we  
142 decided to discard arthropod attack rate from the study and focused on marks that were  
143 unambiguously attributed to birds, *i.e.*, conic holes or V-shaped beak marks. Attack marks left by  
144 reptiles or rodents were also disregarded, because only a few caterpillars were attacked by these  
145 potential predators. Most bird marks were directed towards the head or the body centre of the  
146 dummy caterpillars, which is typical to bird attacks and indicates prey recognition (Rößler *et al.*,  
147 2018). We therefore refer to the proportion of dummy caterpillars with such marks as bird attack  
148 rate.

149 Between 2018 and 2019, 137 partners installed 12,760 dummy caterpillars on 319 oak trees.  
150 Despite clear instructions regarding caterpillar installation, removal and conditioning prior to  
151 shipping, the material sent by 22 school partners was of poor quality (with no particular  
152 geographic bias) such that only caterpillars returned by 115 partners (*i.e.*, 78.4%, collected on 254  
153 oak trees) were screened for attack marks and included in subsequent analyses (**Table S1.1;**  
154 **Figure 1**).

#### 155 Leaf herbivory

156 Professional scientists and schoolchildren were instructed to collect oak leaves after the second  
157 bird attack rate survey, *i.e.*, roughly 10 weeks after oak budburst, on the same branches where  
158 dummy caterpillars were installed. They haphazardly collected 30 leaves per branch, totalling 120  
159 leaves from which they blindly drew 60 leaves. Professional scientists oven-dried leaves for a  
160 minimum of 48 h at 45°C immediately after collection, and leaves collected by schoolchildren  
161 were oven dried upon receipt by the project coordinators, to ensure optimal conservation prior  
162 to herbivory assessment.

163 We used three response variables to characterise leaf herbivory: leaf damage (the percentage of  
164 leaf area that was consumed or mined by insect herbivores), incidence of leaf-miners (the  
165 proportion of leaves with leaf-mines) and incidence of gall-inducers (the proportion of leaves with  
166 galls). For each leaf, we visually assessed leaf damage (attributed to ectophagous chewing and  
167 leaf-mining organisms) following eight levels of defoliation (0%, >0-5%, >5-10%, >10-15%, >15-  
168 25%, >25-50%, >50-75%, and >75%). We then averaged leaf damage at the tree level using the  
169 midpoint of each percentage class to obtain a mean value per tree. While this measurement also  
170 included the surface covered by leaf mines (*i.e.*, internally chewed by mining larva), it excluded

171 both galls and punctures made by sap feeders. Leaf assessment was made by two trained  
172 observers who were blind to leaf origin to reduce unconscious bias. We expect that most of the  
173 leaf damage will be attributable to insects, as in our experience, mollusc herbivory (e.g. snails and  
174 slugs), although possible, is rare in adult oak trees. As of mites, they mostly cause discolouration  
175 that can easily be differentiated from insect herbivory. There are also few mammals consuming  
176 oak leaves, mostly rodents and ungulates, but usually they mostly consume seedlings and  
177 saplings.

#### 178 Leaf chemical traits

179 We used leaves collected in 2018 to quantify several leaf chemical traits typically recognized as  
180 important determinants of plant quality for insect herbivores associated with oaks. Details of  
181 procedures used to analyse chemical leaf traits are reported in online Appendix S1.1.

182 We quantified leaf phenolics as oak defensive metabolites (Moreira *et al.*, 2018). We used only  
183 leaves collected by professional scientists in 2018. Unfortunately, we were not able to quantify  
184 other leaf defences that need to be estimated on fresh leaves (e.g. leaf toughness) because leaves  
185 were oven dried after collection to ensure optimal conservation. From each tree, we selected 10  
186 mature, dried leaves with no evidence of insect damage and ground them to fine powder. We  
187 identified four groups of phenolic compounds: flavonoids, ellagitannins and gallic acid derivatives  
188 (“hydrolysable tannins” hereafter), proanthocyanidins (“condensed tannins” hereafter) and  
189 hydroxycinnamic acid precursors to lignins (“lignins” hereafter) (see Appendix S1.1 for further  
190 details).

191 We quantified C:N ratio, N:P ratio, cellulose and soluble sugars as proxies for leaf nutritional  
192 content to herbivores (Moreira *et al.*, 2019) as in many plant taxa these variables are correlated  
193 with herbivory (Mattson, 1980; Schoonhoven *et al.*, 2005; Smilanich *et al.*, 2016). We measured  
194 these traits on leaves collected by both professional scientists and schoolchildren. We ground the  
195 50 oven dried leaves on which we scored herbivory to fine powder such that leaf nutritional traits  
196 reflected the content of leaves with different amounts of herbivore damage (see Appendix S1.1  
197 for further details).

#### 198 Statistical analysis

199 We were primarily interested in testing the effect of latitude on herbivory and in identifying the  
200 underlying ecological forces. We aimed to test whether the effect of latitude was driven by  
201 latitudinal gradients in climatic conditions, and by their direct and indirect consequences on leaf  
202 chemistry, herbivory and bird attack rate. We obtained temperature and precipitation data from  
203 the WorldClim database (Hijmans *et al.*, 2005) based on oak coordinates as retrieved on Google  
204 maps by project partners. Specifically, we extracted the mean temperature and precipitation  
205 from April to June, which roughly corresponded to the period when caterpillars were present on  
206 trees, irrespective of latitudinal cline in moth phenology. Yet, latitude was correlated with both

207 temperature (Pearson's  $r = -0.85$ ) and precipitation ( $r = -0.72$ ). To avoid spurious estimates of  
208 model coefficients caused by collinearity among predictors, we tested the effects of latitude and  
209 climate separately.

210 Specifically, we built three types of Linear Mixed-effects Models (LMM): (i) geographic models  
211 analysing the effect of latitude on herbivory, leaf chemistry and bird attack rate, (ii) climatic  
212 models in which we substituted latitude with climatic data (temperature and precipitation) and  
213 (iii) abiotic and biotic models analysing the effects of leaf chemistry and bird attack rate together  
214 with temperature and precipitation or latitude (both linear and quadratic) on herbivory. As  
215 latitudinal gradients in plant-herbivore interactions can be non-linear (following Rodríguez-  
216 Castañeda 2013 and Kozlov *et al.*, 2015), we complemented our linear analyses with quadratic  
217 models. As leaf chemistry was only measured on a subset of trees ( $n = 64$ ), we used a subsample  
218 of the dataset to quantify relationships between herbivory with its top-down and bottom-up  
219 drivers (Table S1.1).

220 In every LMM, we used Partner ID as a random factor to account for the fact that some partners  
221 surveyed multiple trees. For instance, the geographic models were of the form:

$$222 \quad Y = \beta_0 + \beta_1 \times \text{Year} + \beta_2 \times \text{Partner} + \beta_3 \times \text{Latitude} + \beta_4 \times \text{Latitude}^2 + \gamma + \epsilon + \sigma^2_{\text{Partner ID}}$$

223 where  $Y$  was the response variable,  $\beta_i$  model coefficient parameter estimates, Partner was the  
224 effect of partner type (the estimate for schoolchildren being compared with the estimate for  
225 professional scientists that was included in the intercept), Year was the effect of each year (2019  
226 contrasted with 2018), Latitude (and their quadratic terms) the geographic conditions around  
227 sampled oak trees,  $\sigma^2_{\text{Partner ID}}$  the random effect of Partner ID (assuming that  $\gamma \in N(0, \sigma^2_{\text{Partner ID}})$   
228 and  $\epsilon$  the residuals (assuming  $\epsilon \in N(0, \sigma^2_{\epsilon})$ ). When  $Y$  was bird attack rate, we added the survey  
229 (first vs. second) as a fixed effect and Tree ID as a random effect nested within Partner ID to  
230 account for repeated measurements on the same trees. When needed, we used arcsine square-  
231 root (bird attack rate) or  $\ln(x + 1)$  transformation (leaf damage, soluble sugars, N:P ratio and leaf  
232 defences) of the response variable to satisfy model assumptions.

233 We ran geographic and climatic models on the complete dataset including 2018 and 2019 data  
234 collected by both professional scientists and schoolchildren. Note that because not every partner  
235 provided reliable data on both bird attack rates and herbivory, the sample sizes differed between  
236 models using bird attack rate or herbivory as response variables (**Figure 1, Figure S1.1a and b**).  
237 We ran the geographic and climatic models on leaf phenolics as well as the biotic model on the  
238 2018 data collected by scientific partners only, as we did not quantify leaf defences on leaves  
239 collected and sent by schoolchildren.

240 The tree-level response variables for each year and survey period ( $Y$ ) were either leaf damage (%  
241 of leaf area removed and mined by herbivores), the incidence of leaf-miners or gall-inducers

242 (proportions), mean bird attack rate (ratio of % attacked caterpillars on exposition period) or leaf  
243 chemistry (C:N ratio, N:P ratio, soluble sugar content [ $\text{g L}^{-1}$ ], cellulose content (g), concentrations  
244 of condensed or hydrolysable tannins, flavonoids or lignins [ $\text{mg g}^{-1}$  d.m.]). We scaled and centred  
245 every continuous predictor prior to modelling to facilitate comparisons of their effect sizes, and  
246 made sure that none of the explanatory variables were strongly correlated using the variance  
247 inflation factor (VIF) (all VIFs < 5). We used LMM with a Gaussian error distribution, with the  
248 exceptions of geographic, climatic and process-based models with the incidence of leaf-miners or  
249 gall-inducers as response variables. In these cases, we used Generalized LMM with a binomial  
250 error distribution and logit-link.

251 We analysed the data within the information theory framework (Burnham & Anderson, 2002).  
252 We first built a set of geographic and climatic models as well as nested models for each response  
253 variable separately. Biotic models (models including also leaf chemistry and bird attack rates as  
254 explanatory variables) were run on the subset of samples where all data were measured  
255 simultaneously. We then applied a procedure of model selection based on AIC corrected for small  
256 sample size (AICc). In the first step, we ranked the models according to the difference in AICc  
257 between a given model and the model with the lowest AICc ( $\Delta\text{AICc}$ ). Models within 2  $\Delta\text{AICc}$  units  
258 of the best model (*i.e.*, the model with the lowest AICc) are generally considered as equally likely.  
259 We also computed AIC weight ( $w_i$ ) that is the probability a given model to be the best model  
260 among the set of candidate models examined, as well as the relative variable importance (RVI) as  
261 the sum of  $w_i$  of every model including this variable. When several models competed with the  
262 best model (*i.e.*, when multiple models were such that their  $\Delta\text{AICc} < 2$ ), we applied a procedure  
263 of multimodel inference building a consensus model including the variables in the set of best  
264 models. We then averaged their effect size across all the models in the set of best models, using  
265 variable  $w_i$  as a weighting parameter (*i.e.*, model averaging). We considered that a given predictor  
266 had a statistically significant effect on the response variable when its confidence interval excluded  
267 zero.

268 In the results section, we report the degrees of freedom, log-likelihood, AICc,  $\Delta\text{AICc}$  (delta),  $w_i$   
269 and the variance explained by fixed and fixed plus random factors ( $R^2m$  and  $R^2c$ , respectively)  
270 (Nakagawa & Schielzeth, 2013) for every model, as well as averaged coefficient parameter  
271 estimates and RVI for all variables present in the set of competing best models. When  
272 appropriate, we plotted the relationship between raw data and explanatory variables together  
273 with the predictions of simplified models, holding undisplayed predictors constant. All analyses  
274 were run in the R language environment (Team, 2018) with packages *MuMIn* (Bartoń, 2018) and  
275 *lme4* (Bates *et al.*, 2018).

## 276 **Results**

### 277 *Latitudinal and climatic gradients in herbivory, leaf chemistry and bird attack rates*

278 Herbivores damaged on average ( $\pm$  se)  $8.7 \pm 0.4$  % of leaf area ( $n = 182$  trees, see **Table S1.1** for  
279 details). Model simplification identified the null model as the best model given the model set,  
280 indicating that none of the predictors had a consistent effect on leaf damage (**Figure 2j, k and l**,  
281 **Table S2.1**).

282 Insect galls were present on  $7.1 \pm 0.6$  % of the inspected leaves ( $n = 182$ , **Table S1.1**). In the set of  
283 best models (**Table S2.1; Figure S1.3**), the incidence of gall-inducers increased linearly with  
284 increasing spring temperature (**Figure 2e**) and peaked at intermediate levels of spring  
285 precipitation (**Figure 2f**). It was on average higher in 2018 than in 2019 (**Figure S1.2**). Other  
286 predictors had no significant effects on the incidence of gall-inducers (**Figure 2d**).

287 Leaf-miners were present on  $18.2 \pm 1.3$  % of the inspected leaves (**Table S1.1**). In the set of best  
288 models (**Table S2.1; Figure S1.3**), the incidence of leaf-miners peaked at intermediate mean  
289 spring temperatures (**Figure 2h**) and decreased linearly with increasing spring precipitation  
290 (**Figure 2i**). It was significantly higher in 2018 than in 2019 (**Figure S1.2**), and higher in leaves  
291 sampled by professional scientists than in those sampled by schoolchildren.

292 Some oak traits related to nutritional content, but not phenolic compounds, covaried with climate  
293 and latitude (**Table S1.1**). Specifically, leaf soluble sugar content ( $3.7 \pm 0.2$  g·L<sup>-1</sup>,  $n = 114$ , **Table**  
294 **S1.1**) decreased with increasing precipitation (**Figure 3a**). Leaf C:N ratio ( $18.6 \pm 0.2$ ,  $n = 114$ , **Table**  
295 **S1.1**) increased non-linearly with latitude (with concave up shape, **Figure 3b**) and was on average  
296 lower in leaves collected by professional scientists than those collected by schoolchildren. None  
297 of the predictors had a significant effect on N:P or cellulose content (**Table S1.1**).

298 From a total of 10,000 exposed dummy caterpillars, 2,390 had bird beak marks (*i.e.*, 23.9%).  
299 Model selection identified the null model as the best model, with no other competing model  
300 within two units of  $\Delta$ AICc of the best model.

### 301 *Mechanisms underlying latitudinal and climatic variation in herbivory*

302 Using a data subset limited to trees for which information on herbivory, leaf traits and bird  
303 predation rates was available, model selection identified the null model as the best model,  
304 indicating that none of the examined biotic and abiotic predictors had a significant effect on leaf  
305 damage (**Table S2.2**).

306 When leaf chemistry was included in the model, the incidence of gall-inducers increased with  
307 increasing soluble sugar concentration and decreased with increasing C:N ratio and lignin  
308 concentration (**Figure 4**), whereas the positive relationship between temperature and gall-  
309 inducers disappeared. When leaf traits were included in the models, the incidence of gall-inducers  
310 increased non-linearly with increasing latitude. The relative importance of leaf chemistry  
311 predictors (RVI = 0.65) was however higher than that of latitude (RVI = 0.05) or temperature (RVI  
312 = 0.30, **Figure S1.4**).

313 Leaf-miner incidence increased with increasing concentration of hydrolysable tannins. The  
314 relationship between temperature and leaf-miners remained significant, suggesting independent  
315 effects of leaf defences and temperature on leaf-miners. Other predictors had no significant  
316 effects on leaf-miners (**Figure 4; Table S2.2**).

## 317 Discussion

318 We found no evidence that either herbivory, oak chemical traits or bird attack rates varied with  
319 latitude linearly or non-linearly. Our work therefore supports the growing number of studies that  
320 have recently questioned the common view that biotic interactions are generally stronger at  
321 lower latitudes (Moles *et al.*, 2011; Mottl *et al.*, 2020). Several arguments may explain the absence  
322 of latitudinal patterns in herbivory or plant defences. On the one hand, Moles & Ollerton (2016)  
323 argued that the latitudinal herbivory-defence hypothesis should simply be dismissed because it is  
324 not convincingly supported despite decades of research. On the other hand, Anstett *et al.* (2016)  
325 called for a refinement of the concepts and methods in the field. We align with the latter  
326 perspective and henceforth discuss ecological sources of variation that may have obscured  
327 latitudinal patterns, and argue these could be real rather than methodological artefacts.

328 **Herbivory responded to climate rather than to latitude, but in a guild-specific manner—**  
329 Variation in the incidence of gall-inducers and leaf-miners, but not in leaf damage, was associated  
330 with variation in temperature and precipitation, rather than with latitude *per se* (Anstett *et al.*,  
331 2018; Moreira *et al.*, 2018; Loughnan & Williams, 2019). The absence of a climatic or geographic  
332 effect on leaf damage is in line with previous reviews and meta-analyses that have shown  
333 herbivory does not vary consistently along climatic or geographic gradients (Moles *et al.*, 2011;  
334 Moles & Ollerton, 2016). Our analysis suggests that different herbivore species or guilds may be  
335 differently affected by abiotic conditions, such that grouping different types of herbivores may  
336 prevent the detection of patterns for each herbivore type (Abdala-Roberts *et al.*, 2015; Moreira  
337 *et al.*, 2015; Anstett *et al.*, 2016).

338 In line with this explanation, we found that the incidence of gall-inducers and leaf-miners was  
339 associated with broad scale climatic conditions. Specifically, the incidence of both gall-inducers  
340 and leaf-miners increased with increasing temperature, but the shape of this relationship was  
341 accelerating for gall-inducers and decelerating (*i.e.*, convex) for leaf-miners (**Figure 5**). Although  
342 we did not identify species of leaf-miners, this result is in line with that of Kozlov *et al.* (2013) who  
343 found that in northern Europe, the diversity of leaf miners on birch trees increased linearly toward  
344 lower latitudes and was most likely associated with the direct impact of temperature, especially  
345 during cold years. We also found that the incidence of gall-inducers peaked at intermediate  
346 precipitation (Blanche & Ludwig, 2001; Leckey *et al.*, 2014) whereas leaf-miners decreased  
347 significantly with precipitation. It has been hypothesized that endophagous feeding modes such  
348 as galling and mining have evolved partly as adaptation to abiotic factors such as UV radiation

349 and desiccation (Fernandes & Price, 1992; Connor *et al.*, 1997; Danks, 2002). If so, gall-inducers  
350 and leaf-miners may be expected to be more common in the warmest and driest parts of the  
351 pedunculate oak range and at low latitudes where the light intensity is markedly higher  
352 (Fernandes & Price, 1992; Lara & Fernandes, 1996; Price *et al.*, 1998; Cuevas-Reyes *et al.*, 2004).  
353 However, even within the gall-inducer and leaf-miner groups, relationships to climate are highly  
354 variable among species and years (Blanche, 2000; Sinclair & Hughes, 2010; Kozlov *et al.*, 2013,  
355 2016), thus suggesting that other factors are also important in the incidence of gall-inducers and  
356 leaf-miner herbivores.

### 357 **Leaf chemical traits had an inconsistent response to latitude and climate—**

358 We did not find detectable latitudinal and climatic gradients in plant chemical defences. This  
359 contradicts the Latitudinal Herbivory Defence Hypothesis which predicts that plant species at  
360 lower latitudes experience higher mean rates of herbivory than their temperate counterparts  
361 (Coley & Barone, 1996; Schemske *et al.*, 2009; Lim *et al.*, 2015) and, for this reason, should have  
362 evolved higher levels of anti-herbivore defences (Rasmann & Agrawal, 2011; Pearse & Hipp,  
363 2012). However, the generality of this hypothesis is currently under debate (Moles & Ollerton,  
364 2016). Several studies found no evidence for a latitudinal gradient in herbivory and plant defences  
365 (Moles *et al.*, 2011) while others did (Salgado & Pennings, 2005; Woods *et al.*, 2012); there is also  
366 mixed evidence when comparing different herbivore species or plant defensive traits (Anstett *et al.*,  
367 2015; Moreira *et al.*, 2015, 2018). A plausible explanation for the lack of latitudinal gradients  
368 in oak defences may be that there is no latitudinal gradient in leaf damage, hence there is no  
369 reason for latitudinal gradient for defences to exist either. It could also be that we sampled leaves  
370 at the middle of the growing season rather than at the end, and we did not measure constitutive  
371 and induced defences separately. This is an insightful point because oak leaves may have  
372 differentially accumulated phenolics in response to herbivory (*i.e.*, induced defences) or have  
373 experienced marked differences in light intensity toward the end of the growing season  
374 (Karolewski *et al.*, 2013). Furthermore, despite attempts to synchronize phenology across sites,  
375 seasonal changes in oak chemical defences (Salminen & Karonen, 2011) might have masked  
376 latitudinal patterns in defences. Therefore, further studies should include measurements at  
377 multiple time points during the growing season and distinguish between different types of  
378 defences, including physical vs. chemical defences (Wang *et al.*, 2018) as well as constitutive vs.  
379 induced defences (Anstett *et al.*, 2018) in order to address latitudinal gradients in plant defence  
380 more comprehensively.

381 Some leaf traits related to leaf nutrient content were associated with latitude or climatic  
382 conditions, but their overall response was inconsistent. Leaf C:N ratio and sugar content varied  
383 along latitudinal and climatic gradients, respectively. The leaf C:N ratios were lowest at  
384 intermediate latitudes. This outcome may be due to temperature-related plant physiological  
385 stoichiometry and biogeographical gradients in soil substrate age (limitation of soil N at higher

386 latitudes) (Reich & Oleksyn, 2004). Leaf soluble sugar content decreased with increasing  
387 precipitation (Cao *et al.*, 2018). Soluble sugars, especially glucose and fructose, accumulate  
388 together with other osmolytes during drought (Nio *et al.*, 2011), resulting in high concentration  
389 in areas where precipitation is low.

#### 390 **Predation rate was not influenced by latitude or climatic conditions—**

391 We found no latitudinal or climatic gradients in bird attack rates on dummy caterpillars (**Figure**  
392 **5**). These results agree with the large-scale study performed by Roslin *et al.* (2017) who found an  
393 increase of the activity of predatory arthropods in several plant species toward the Equator, but  
394 no significant trend in avian predation. Several factors may explain the lack of response of avian  
395 predation to latitudinal or climatic gradients. First, some bird species are distributed through  
396 migration allowing them to breed at higher latitudes, resulting in a constant predation rate across  
397 climatic and geographical clines (Dufour *et al.*, 2020). In contrast, other predators with lower  
398 mobility such as arthropods (e.g. ants, ladybirds) are much more abundant at lower latitudes,  
399 resulting in a higher selection pressure toward the Equator (Roslin *et al.*, 2017). Second, bird  
400 communities are more influenced by forest habitat composition at lower latitudes, and more by  
401 food availability at higher latitudes (Charbonnier *et al.*, 2016) where the diet variability is lower  
402 (Barnagaud *et al.*, 2019), suggesting a stronger effect of local habitat features (e.g. resource  
403 availability and habitat suitability) than climatic gradients. Third, we cannot exclude that the lack  
404 of latitudinal trend in bird attack rates resulted from methodological limitations due to the fact  
405 that we only exposed green dummy caterpillars in low hanging branches. Birds depend more on  
406 food accessibility than abundance *per se*, so that the exact location of dummy caterpillars  
407 regarding factors such as edge, light contrast and shrubby understory may have modified the  
408 perception and the accessibility to the prey (Zvereva *et al.*, 2019).

#### 409 **Mechanisms underlying latitudinal and climatic variation in herbivory—**

410 We did not find any statistically significant relationship between leaf damage and leaf chemical  
411 traits. Although we cannot exclude that unmeasured traits (e.g. leaf toughness, inducible  
412 defences) may have correlated with herbivory, our results disqualify large-scale variation in  
413 bottom-up forces as important drivers of overall leaf damage at a continental scale. In contrast,  
414 the incidence of gall-inducers and leaf-miners was partially related to the variability in several leaf  
415 chemical traits (**Figure 5**). For instance, the incidence of gall-inducers increased with increasing  
416 leaf soluble sugars and N concentrations, which is consistent with gall-inducers being metabolic  
417 sinks (Huang *et al.*, 2014). However, the effect of precipitation on leaf-miners was likely indirectly  
418 mediated by climatic variation in defences, as such an effect became non-significant once  
419 hydrolysable tannins were included in the model. Similarly, the effects of temperature and  
420 precipitation on gall-inducers were indirectly mediated by climatic variation in defences, as such  
421 effects were also non-significant after soluble sugars, N concentrations and lignins were included



422 in the models. These results agree with previous studies reporting indirect effects (via leaf  
423 defences) of climate on herbivory (Anstett *et al.*, 2018; Moreira *et al.*, 2018). For instance, Anstett  
424 *et al.* (2018) found indirect effects of climate on herbivory in 80 species of evening primroses,  
425 which were mediated by leaf chemicals (total phenolics and oenothien A). However, these  
426 conclusions need to be considered with caution because the dataset used to test the effect of  
427 bottom-up and top-down forces on herbivory along large-scale latitudinal and climatic gradients  
428 of the biotic and abiotic models only consisted of a subset of the complete dataset used in the  
429 geographic and climatic models.

430 We found no evidence that bird attack rate drove large-scale variability in herbivory. This result  
431 is in line with a recent study by Zverev *et al.* (2020) who found that birds are unlikely to shape the  
432 spatial patterns of insect herbivory in an Arctic ecosystem. More generally, associations between  
433 bird insectivory and insect herbivores can be positive (Mäntylä *et al.*, 2014; Gunnarsson *et al.*,  
434 2018), negative (Maguire *et al.*, 2015; Kozlov *et al.*, 2017) or non-significant (Moreira *et al.*, 2019;  
435 Valdés-Correcher *et al.*, 2019), depending on the study and methods used. Arthropod predators  
436 (e.g. ants, ladybirds) play an important role in limiting herbivore populations and may respond to  
437 large-scale variation in climatic conditions at greater extent than vertebrate predators (Roslin *et al.*,  
438 2017; Zvereva *et al.*, 2019). For example, a meta-analysis conducted by Rodríguez-Castañeda  
439 (2013) found that ant predation on herbivores significantly increase at higher temperatures and  
440 precipitations, indicating that plants growing under warmer and wetter conditions exhibit lower  
441 levels of herbivory. Besides, birds are considered intraguild predators that not only eat insect  
442 herbivores but also arthropod predators (Gunnarsson, 2007) and intraguild predation may  
443 weaken herbivore suppression (Finke & Denno, 2005). Unfortunately, we were not able to  
444 quantify predation rates by such arthropods nor intraguild predation, which weakens our  
445 conclusions about the potential role of predators across climatic gradients.

446 **Conclusion** — By simultaneously investigating bottom-up and top-down forces driving herbivory  
447 along latitudinal and climatic clines in a widespread tree species in Europe, this study brings some  
448 new insights into the vivid debate about latitudinal variation in the direction and strength of biotic  
449 interactions (Schemske *et al.*, 2009; Moles *et al.*, 2013; Anstett *et al.*, 2016; Roslin *et al.*, 2017).  
450 We found no evidence that latitude or climate influenced insect herbivores feeding on oaks, but  
451 we found that climatic factors rather than latitude *per se* were the best predictors of the large-  
452 scale variation in the incidence of leaf-miner and gall-inducer herbivores as well as in variation in  
453 leaf nutritional content. In sharp contrast, we found no evidence that plant chemical defences  
454 and bird attack rates were influenced by latitude or climatic factors, which conflicts with the  
455 dominant view in ecology (Moles & Ollerton, 2016; Roslin *et al.*, 2017; Zvereva *et al.*, 2019).  
456 Because unravelling causes of latitudinal variation in the strength of biological interactions is one  
457 of the common approaches for the prediction of biotic interactions under global warming  
458 (Verheyen *et al.*, 2019), it is crucial that future studies simultaneously test for effects of latitude

459 *per se* and climate on herbivory by different feeding guilds (Kozlov *et al.*, 2017), as well as  
460 investigate the complexity of biotic interactions in which plant-herbivores interactions are  
461 embedded.

## 462 Data accessibility

463 The datasets supporting this article are available  
464 via an open-access repository (<https://doi.org/10.5061/dryad.18931zdw0>).

## 465 References

- 466 Abdala-Roberts, L., Mooney, K.A., Quijano-Medina, T., Campos-Navarrete, M.J., González-  
467 Moreno, A. & Parra-Tabla, V. (2015) Comparison of tree genotypic diversity and species  
468 diversity effects on different guilds of insect herbivores. *Oikos*, **124**, 1527–1535.
- 469 Abdala-Roberts, L., Moreira, X., Rasmann, S., Parra-Tabla, V. & Mooney, K.A. (2016) Test of  
470 biotic and abiotic correlates of latitudinal variation in defences in the perennial herb *Ruellia*  
471 *nudiflora*. *Journal of Ecology*, **104**, 580–590.
- 472 Adams, J.M., Rehill, B., Zhang, Y. & Gower, J. (2009) A test of the latitudinal defense hypothesis:  
473 Herbivory, tannins and total phenolics in four North American tree species. *Ecological*  
474 *Research*, **24**, 697–704.
- 475 Anstett, D.N., Ahern, J.R., Glinos, J., Nawar, N., Salminen, J.P. & Johnson, M.T.J. (2015) Can  
476 genetically based clines in plant defence explain greater herbivory at higher latitudes?  
477 *Ecology Letters*, **18**, 1376–1386.
- 478 Anstett, D.N., Ahern, J.R., Johnson, M.T.J. & Salminen, J.P. (2018) Testing for latitudinal  
479 gradients in defense at the macroevolutionary scale. *Evolution*, **72**, 2129–2143.
- 480 Anstett, D.N., Chen, W. & Johnson, M.T.J. (2016) Latitudinal gradients in induced and  
481 constitutive resistance against herbivores. *Journal of Chemical Ecology*, **42**, 772–781.
- 482 Barnagaud, J.Y., Mazet, N., Munoz, F., Grenié, M., Denelle, P., Sobral, M., Kissling, W.D.,  
483 Şekercioğlu, Ç.H. & Violle, C. (2019) Functional biogeography of dietary strategies in birds.  
484 *Global Ecology and Biogeography*, **28**, 1004–1017.
- 485 Bartoń, K. (2018) MuMIn: Multi-Model Inference. R package version 1.40. 4.
- 486 Bates, D., Maechler, M., Bolker, B. & Walker, S. (2018) lme4: Linear Mixed-Effects Models using  
487 “Eigen” and S4.
- 488 Blanche, K.R. (2000) Diversity of insect-induced galls along a temperature-rainfall gradient in the  
489 tropical savannah region of the Northern Territory, Australia. *Austral Ecology*, **25**, 311–318.
- 490 Blanche, K.R. & Ludwig, J.A. (2001) Species richness of gall-inducing insects and host plants

- 491 along an altitudinal gradient in Big Bend National Park, Texas. *The American Midland*  
492 *Naturalist*, **145**, 219–232.
- 493 Burnham, K.P. & Anderson, D.R. (2002) Model selection and multimodel interference, a practical  
494 information-Theoretic approach. *2nd ed. Springer, New York, 2.*
- 495 Cao, Y., Li, Y. & Chen, Y. (2018) Non-structural carbon, nitrogen, and phosphorus between black  
496 locust and chinese pine plantations along a precipitation gradient on the Loess Plateau,  
497 China. *Trees - Structure and Function*, **32**, 835–846.
- 498 Castagneyrol, B., Valdés-Correcher, E., Kaennel Dobbertin, M. & Gossner, M.M. (2019) Predation  
499 assessment on fake caterpillars and leaf sampling: Protocol for partner schools.  
500 *protocols.io.*
- 501 Charbonnier, Y.M., Barbaro, L., Barnagaud, J.Y., Ampoorter, E., Nezan, J., Verheyen, K. & Jactel,  
502 H. (2016) Bat and bird diversity along independent gradients of latitude and tree  
503 composition in European forests. *Oecologia*, **182**, 529–537.
- 504 Chen, Y., Han, W., Tang, L., Tang, Z. & Fang, J. (2013) Leaf nitrogen and phosphorus  
505 concentrations of woody plants differ in responses to climate, soil and plant growth form.  
506 *Ecography*, **36**, 178–184.
- 507 Coley, P. & Barone, J. (1996) Herbivory and plant defenses in tropical forests. *Annual Review of*  
508 *Ecology and Systematics*, **27**, 305–335.
- 509 Coley, P.D. & Aide, T.M. (1991) Comparison of herbivory and plant defences in temperate and  
510 tropical broad-leaved forests. *Plant–animal interactions: Evolutionary ecology in tropical*  
511 *and temperate regions*, 25–49.
- 512 Connor, E.F., Taverner, M.P., Oikos, S. & May, N. (1997) The evolution and adaptive significance  
513 of the leaf-mining habit. *Nordic Society Oikos*, **79**, 6–25.
- 514 Cuevas-Reyes, P., Quesada, M., Hanson, P., Dirzo, R. & Oyama, K. (2004) Diversity of gall-  
515 inducing insects in a Mexican tropical dry forest: The importance of plant species richness,  
516 life-forms, host plant age and plant density. *Journal of Ecology*, **92**, 707–716.
- 517 Danks, H. V (2002) Modification of adverse conditions by insects. *Oikos*, **99**, 10–24.
- 518 Del-Val, E. & Armesto, J.J. (2010) Seedling mortality and herbivory damage in subtropical and  
519 temperate populations: Testing the hypothesis of higher herbivore pressure toward the  
520 tropics. *Biotropica*, **42**, 174–179.
- 521 Dufour, P., Descamps, S., Chantepie, S., Renaud, J., Guéguen, M., Schiffers, K., Thuiller, W. &  
522 Lavergne, S. (2020) Reconstructing the geographic and climatic origins of long-distance bird  
523 migrations. *Journal of Biogeography*, **47**, 155–166.

- 524 Dyer, L.A. & Coley, P.D. (2002) *Tritrophic interactions in tropical versus temperate communities.*  
525 *Multitrophic Level Interactions*, pp. 67–88.
- 526 Eaton, E., Caudullo, G., Oliveira, S. & de Rigo, D. (2016) *Quercus robur* and *Quercus petraea*.  
527 *European Atlas of Forest Tree Species*, 160–163.
- 528 Fernandes, G.W. & Price, P.W. (1992) International association for ecology the adaptive  
529 significance of insect gall distribution: survivorship of species in xeric and mesic habitats.  
530 *Oecologia*, **90**, 14–20.
- 531 Finke, D.L. & Denno, R.F. (2005) Predator diversity and the functioning of ecosystems: The role  
532 of intraguild predation in dampening trophic cascades. *Ecology Letters*, **8**, 1299–1306.
- 533 De Frenne, P., Graae, B.J., Rodríguez-Sánchez, F., Kolb, A., Chabrierie, O., Decocq, G., De Kort, H.,  
534 De Schrijver, A., Diekmann, M., Eriksson, O., Gruwez, R., Hermy, M., Lenoir, J., Plue, J.,  
535 Coomes, D.A. & Verheyen, K. (2013) Latitudinal gradients as natural laboratories to infer  
536 species' responses to temperature. *Journal of Ecology*, **101**, 784–795.
- 537 Gaston, K.J., Genney, D.R., Thurlow, M. & Hartley, S.E. (2004) The geographical range structure  
538 of the holly leaf-miner. IV. Effects of variation in host-plant quality. *Journal of Animal*  
539 *Ecology*, **73**, 911–924.
- 540 Gely, C., Laurance, S.G.W. & Stork, N.E. (2019) How do herbivorous insects respond to drought  
541 stress in trees? *Biological Reviews*, **95**, 434–448.
- 542 Gunnarsson, B. (2007) Bird predation on spiders: Ecological mechanisms and evolutionary  
543 consequences. *Journal of Arachnology*, **35**, 509–529.
- 544 Gunnarsson, B., Wallin, J. & Klingberg, J. (2018) Predation by avian insectivores on caterpillars is  
545 linked to leaf damage on oak (*Quercus robur*). *Oecologia*, **188**, 733–741.
- 546 Hahn, P.G., Agrawal, A.A., Sussman, K.I. & Maron, J.L. (2019) Population variation,  
547 environmental gradients, and the evolutionary ecology of plant defense against herbivory.  
548 *American Naturalist*, **193**, 20–34.
- 549 Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G. & Jarvis, A. (2005) Very high resolution  
550 interpolated climate surfaces for global land areas. *International Journal of Climatology*, **25**,  
551 1965–1978.
- 552 Holopainen, J.K., Virjamo, V., Ghimire, R.P., Blande, J.D., Julkunen-Tiitto, R. & Kivimäenpää, M.  
553 (2018) Climate change effects on secondary compounds of forest trees in the Northern  
554 Hemisphere. *Frontiers in Plant Science*, **9**, 1–10.
- 555 Huang, M.Y., Huang, W.D., Chou, H.M., Lin, K.H., Chen, C.C., Chen, P.J., Chang, Y.T. & Yang, C.M.  
556 (2014) Leaf-derived cecidomyiid galls are sinks in *Machilus thunbergii* (Lauraceae) leaves.  
557 *Physiologia Plantarum*, **152**, 475–485.

- 558 Jamieson, M.A., Schwartzberg, E.G., Raffa, K.F., Reich, P.B. & Lindroth, R.L. (2015) Experimental  
559 climate warming alters aspen and birch phytochemistry and performance traits for an  
560 outbreak insect herbivore. *Global Change Biology*, **21**, 2698–2710.
- 561 Janzen, D.H. (1970) Herbivores and the number of tree species in Tropical forests. *American*  
562 *Naturalist*, **104**, 501–528.
- 563 Johnson, M.T.J. & Rasmann, S. (2011) The latitudinal herbivory-defence hypothesis takes a  
564 detour on the map. *New Phytologist*, **191**, 589–592.
- 565 Karolewski, P., Giertych, M.J., Zmuda, M., Jagodziński, A.M. & Oleksyn, J. (2013) Season and light  
566 affect constitutive defenses of understory shrub species against folivorous insects. *Acta*  
567 *Oecologica*, **53**, 19–32.
- 568 Kozlov, M. V., Lanta, V., Zverev, V., Rainio, K., Kunavin, M.A. & Zvereva, E.L. (2017) Decreased  
569 losses of woody plant foliage to insects in large urban areas are explained by bird  
570 predation. *Global Change Biology*, **23**, 4354–4364.
- 571 Kozlov, M. V., Lanta, V., Zverev, V. & Zvereva, E.L. (2015) Global patterns in background losses of  
572 woody plant foliage to insects. *Global Ecology and Biogeography*, **24**, 1126–1135.
- 573 Kozlov, M. V., van Nieuwerkerken, E.J., Zverev, V. & Zvereva, E.L. (2013) Abundance and diversity of  
574 birch-feeding leafminers along latitudinal gradients in northern Europe. *Ecography*, **36**,  
575 1138–1149.
- 576 Kozlov, M. V., Skoracka, A., Zverev, V., Lewandowski, M. & Zvereva, E.L. (2016) Two birch  
577 species demonstrate opposite latitudinal patterns in infestation by gall-making mites in  
578 northern Europe. *PLoS ONE*, **11**, 1–15.
- 579 Lara, C.F. & Fernandes, G. W. (1996) The highest diversity of galling insects: Serra do Cipo ,  
580 Brazil. *Biodiversity Letters*, **3**, 111–114.
- 581 Leckey, E.H., Smith, D.M., Nufio, C.R. & Fornash, K.F. (2014) Oak-insect herbivore interactions  
582 along a temperature and precipitation gradient. *Acta Oecologica*, **61**, 1–8.
- 583 Lim, J.Y., Fine, P.V.A. & Mittelbach, G.G. (2015) Assessing the latitudinal gradient in herbivory.  
584 *Global Ecology and Biogeography*, **24**, 1106–1112.
- 585 Loughnan, D. & Williams, J.L. (2019) Climate and leaf traits, not latitude, explain variation in  
586 plant–herbivore interactions across a species’ range. *Journal of Ecology*, **107**, 913–922.
- 587 Maguire, D.Y., Nicole, T., Buddle, C.M. & Bennett, E.M. (2015) Effect of fragmentation on  
588 predation pressure of insect herbivores in a north temperate deciduous forest ecosystem.  
589 *Ecological Entomology*, **40**, 182–186.
- 590 Mäntylä, E., Blande, J.D. & Klemola, T. (2014) Does application of methyl jasmonate to birch

- 591 mimic herbivory and attract insectivorous birds in nature? *Arthropod-Plant Interactions*, **8**,  
592 143–153.
- 593 Marković, C. & Stojanović, A. (2011) Phloemophagous and xylophagous insects, their  
594 parasitoids, predators and inquilines in the branches of the most important oak species in  
595 Serbia. *Biologia*, **66**, 509–517.
- 596 Martz, F., Peltola, Rainer, Fontanay, S., Duval, R.E., Riitta, J.T. & Stark, S. (2009) Effect of latitude  
597 and altitude on the terpenoid and soluble phenolic composition of juniper (*Juniperus*  
598 *communis*) needles and evaluation of their antibacterial activity in the boreal zone. *Journal*  
599 *of Agricultural and Food Chemistry*, **57**, 9575–9584.
- 600 Mattson, W.J. (1980) Herbivory in relation to plant nitrogen content. *Annual Review of Ecology*  
601 *and Systematics*, **11**, 119–161.
- 602 Moles, A.T., Bonser, S.P., Poore, A.G.B., Wallis, I.R. & Foley, W.J. (2011) Assessing the evidence  
603 for latitudinal gradients in plant defence and herbivory. *Functional Ecology*, **25**, 380–388.
- 604 Moles, A.T. & Ollerton, J. (2016) Is the notion that species interactions are stronger and more  
605 specialized in the tropics a zombie idea? *Biotropica*, **48**, 141–145.
- 606 Moles, A.T., Peco, B., Wallis, I.R., Foley, W.J., Poore, A.G.B., Seabloom, E.W., Vesk, P.A., Bisigato,  
607 A.J., Cella-Pizarro, L., Clark, C.J., Cohen, P.S., Cornwell, W.K., Edwards, W., Ejrnæs, R.,  
608 Gonzales-Ojeda, T., Graae, B.J., Hay, G., Lumbwe, F.C., Magaña-Rodríguez, B., Moore, B.D.,  
609 Peri, P.L., Poulsen, J.R., Stegen, J.C., Veldtman, R., von Zeipel, H., Andrew, N.R., Boulter,  
610 S.L., Borer, E.T., Cornelissen, J.H.C., Farji-Brener, A.G., Degabriel, J.L., Jurado, E., Kyhn, L.A.,  
611 Low, B., Mulder, C.P.H., Reardon-Smith, K., Rodríguez-Velázquez, J., De Fortier, A., Zheng,  
612 Z., Blendinger, P.G., Enquist, B.J., Facelli, J.M., Knight, T., Majer, J.D., Martínez-Ramos, M.,  
613 Mcquillan, P. & Hui, F.K.C. (2013) Correlations between physical and chemical defences in  
614 plants: Tradeoffs, syndromes, or just many different ways to skin a herbivorous cat? *New*  
615 *Phytologist*, **198**, 252–263.
- 616 Moles, A.T. & Westoby, M. (2003) Latitude , seed predation and seed mass. *Journal of*  
617 *Biogeography*, **30**, 105–128.
- 618 Moreira, X., Abdala-Roberts, L., Berny Mier y Teran, J.C., Covelo, F., de la Mata, R., Francisco, M.,  
619 Hardwick, B., Pires, R.M., Roslin, T., Schigel, D.S., ten Hoopen, J.P.J.G., Timmermans, B.G.H.,  
620 van Dijk, L.J.A., Castagnyrol, B. & Tack, A.J.M. (2018) Impacts of urbanization on insect  
621 herbivory and plant defences in oak trees. *Oikos*, **128**, 113–123.
- 622 Moreira, X., Abdala-Roberts, L., Henrik Bruun, H., Covelo, F., De Frenne, P., Galmán, A., Gaytán,  
623 Á., Jaatinen, R., ten Hoopen, J.P.J.G., Pulkkinen, P., Timmermans, B.G.H., Tack, A.J.M. &  
624 Castagnyrol, B. (2020) Latitudinal variation in seed predation correlates with latitudinal  
625 variation in seed defensive and nutritional traits in a widespread oak species. *Annals of*  
626 *Botany*, **126**, 881–890.

- 627 Moreira, X., Abdala-Roberts, L., Parra-Tabla, V. & Mooney, K.A. (2015) Latitudinal variation in  
628 herbivory: Influences of climatic drivers, herbivore identity and natural enemies. *Oikos*,  
629 **124**, 1444–1452.
- 630 Moreira, X., Castagneyrol, B., de la Mata, R., Fyllas, N.M., Galmán, A., García-Verdugo, C.,  
631 Larrinaga, A.R. & Abdala-Roberts, L. (2019) Effects of insularity on insect leaf herbivory and  
632 chemical defences in a Mediterranean oak species. *Journal of Biogeography*, **46**, 1226–  
633 1233.
- 634 Mottl, O., Fibich, P., Klimes, P., Volf, M., Tropek, R., Anderson-Teixeira, K., Auga, J., Blair, T.,  
635 Butterill, P., Carscallen, G., Gonzalez-Akre, E., Goodman, A., Kaman, O., Lamarre, G.P.A.,  
636 Libra, M., Losada, M.E., Manumbor, M., Miller, S.E., Molem, K., Nichols, G., Plowman, N.S.,  
637 Redmond, C., Seifert, C.L., Vrana, J., Weiblen, G.D. & Novotny, V. (2020) Spatial covariance  
638 of herbivorous and predatory guilds of forest canopy arthropods along a latitudinal  
639 gradient. *Ecology Letters*, **23**, 1499–1510.
- 640 Nakagawa, S. & Schielzeth, H. (2013) A general and simple method for obtaining R<sup>2</sup> from  
641 generalized linear mixed-effects models. *Methods in Ecology and Evolution*, **4**, 133–142.
- 642 Nio, S.A., Cawthray, G.R., Wade, L.J. & Colmer, T.D. (2011) Pattern of solutes accumulated  
643 during leaf osmotic adjustment as related to duration of water deficit for wheat at the  
644 reproductive stage. *Plant Physiology and Biochemistry*, **49**, 1126–1137.
- 645 Pearse, I.S. & Hipp, A.L. (2012) Global patterns of leaf defenses in oak species. *Evolution*, **66**,  
646 2272–2286.
- 647 Pennings, S.C., Ho, C., Salgado, C.S., Więski, K., Kunza, A.E., Wason, E.L., Ecology, S., Jan, N.,  
648 Pennings, S.C., Ho, C., Salgado, C.S., Wieski, K., Dave, N., Kunza, A.E. & Wason, E.L. (2009)  
649 Latitudinal variation in herbivore pressure in atlantic coast salt marshes. *Ecology*, **90**, 183–  
650 195.
- 651 Petit, R.J., Brewer, S., Bordács, S., Burg, K., Cheddadi, R., Coart, E., Cottrell, J., Csaikl, U.M., Van  
652 Dam, B., Deans, J.D., Espinel, S., Fineschi, S., Finkeldey, R., Glaz, I., Goicoechea, P.G.,  
653 Jensen, J.S., König, A.O., Lowe, A.J., Madsen, S.F., Mátyás, G., Munro, R.C., Popescu, F.,  
654 Slade, D., Tabbener, H., De Vries, S.G.M., Ziegenhagen, B., De Beaulieu, J.L. & Kremer, A.  
655 (2002) Identification of refugia and post-glacial colonisation routes of European white oaks  
656 based on chloroplast DNA and fossil pollen evidence. *Forest Ecology and Management*,  
657 **156**, 49–74.
- 658 Price, P.W., Wilson Fernandes, G., Lara, A.C.F., Brawn, J., Barrios, H., Wright, M.G., Ribeiro, S.P.  
659 & Rothcliff, N. (1998) Global patterns in local number of insect galling species. *Journal of*  
660 *Biogeography*, **25**, 581–591.
- 661 Rasmann, S. & Agrawal, A.A. (2011) Latitudinal patterns in plant defense: Evolution of  
662 cardenolides, their toxicity and induction following herbivory. *Ecology Letters*, **14**, 476–483.

- 663 Reich, P.B. & Oleksyn, J. (2004) Global patterns of plant leaf N and P in relation to temperature  
664 and latitude. *Proceedings of the National Academy of Sciences of the United States of*  
665 *America*, **101**, 11001–11006.
- 666 Rodríguez-Castañeda, G. (2013) The world and its shades of green: A meta-analysis on trophic  
667 cascades across temperature and precipitation gradients. *Global Ecology and*  
668 *Biogeography*, **22**, 118–130.
- 669 Rosenheim, J.A. (1998) Higher-order predators and the regulation of insect herbivore  
670 populations. *Annual Review of Entomology*, **43**, 421–447.
- 671 Roslin, T., Slade, A., Suchanková, A., Huang, S., Petry, W.K., Weissflog, A., Prokurat, A.,  
672 Laukkanen, L., Laird-Hopkins, B., Teder, T., Milne, S., Novotny, V., Dáttilo, W., Basset, Y.,  
673 Hopkins, T., Zhukovich, V., Koane, B., Hik, D.S., Donoso, D.A., Lewis, O.T., Mwesige, I.,  
674 Bonebrake, T.C., Nichols, E., Slade, E.M., Vandvik, V., Nell, C.S., Nakamura, A., Hill, S.J.,  
675 Hardwick, B., Gray, C.L., Drozd, P., Sam, K., Cameron, E.K., Schmidt, N.M., Slade, V., Asmus,  
676 A., Barrio, I.C., Andrew, N.R., van Nouhuys, S. & Boesing, A.L. (2017) Higher predation risk  
677 for insect prey at low latitudes and elevations. *Science*, **356**, 742–744.
- 678 Rößler, D.C., Pröhl, H. & Lötters, S. (2018) The future of clay model studies. *BMC Zoology*, **3**, 1–5.
- 679 Salgado, C.S. & Pennings, S.C. (2005) Latitudinal variation in palatability of salt-marsh plants:  
680 Are differences constitutive? *Ecology*, **86**, 1571–1579.
- 681 Salminen, J.P. & Karonen, M. (2011) Chemical ecology of tannins and other phenolics: We need  
682 a change in approach. *Functional Ecology*, **25**, 325–338.
- 683 Schemske, D.W., Mittelbach, G.G., Cornell, H. V., Sobel, J.M. & Roy, K. (2009) Is there a  
684 latitudinal gradient in the importance of biotic interactions? *Annual Review of Ecology,*  
685 *Evolution, and Systematics*, **40**, 245–269.
- 686 Schoonhoven, L.M., Van Loon, B., van Loon, J.J. & Dicke, M. (2005) *Insect-plant biology*. Oxford  
687 *University Press*, Oxford.
- 688 Sinclair, R.J. & Hughes, L. (2010) Leaf miners: The hidden herbivores. *Austral Ecology*, **35**, 300–  
689 313.
- 690 Smilanich, A.M., Fincher, R.M. & Dyer, L.A. (2016) Does plant apparency matter? Thirty years of  
691 data provide limited support but reveal clear patterns of the effects of plant chemistry on  
692 herbivores. *New Phytologist*, **210**, 1044–1057.
- 693 Southwood, T.R.E., Wint, W.G.R., Kennedy, C.E.J. & Greenwood, S.R. (2005) Composition of  
694 arthropod fauna in some species of *Quercus*. *European Journal of Entomology*, **102**, 65–72.
- 695 Stark, S., Julkunen-Tiitto, R., Holappa, E., Mikkola, K. & Nikula, A. (2008) Concentrations of foliar  
696 quercetin in natural populations of white birch (*Betula pubescens*) increase with latitude.



697 *Journal of Chemical Ecology*, **34**, 1382–1391.

698 Team, R.C. (2018) R: A Language and environment for statistical computing.

699 Valdés-Correcher, E., van Halder, I., Barbaro, L., Castagneyrol, B. & Hampe, A. (2019) Insect  
700 herbivory and avian insectivory in novel native oak forests: Divergent effects of stand size  
701 and connectivity. *Forest Ecology and Management*, **445**, 146–153.

702 Verheyen, J., Tüzün, N. & Stoks, R. (2019) Using natural laboratories to study evolution to global  
703 warming: contrasting altitudinal, latitudinal, and urbanization gradients. *Current Opinion in*  
704 *Insect Science*, **35**, 10–19.

705 Wang, J., Ding, J., Tan, B., Robinson, K., Michelson, I., Johansson, A., Nystedt, B., Scofield, D.,  
706 Nilsson, O., Jansson, S., Street, N. & Ingvarsson, P. (2018) A major locus controls local  
707 adaptation and adaptive life history variation in a perennial plant. *Genome Biology*, **19**, 1–  
708 17.

709 Woods, E.C., Hastings, A.P., Turley, N.E., Heard, S.B., Agrawal, A., Monographs, S.E., May, N.,  
710 Woods, E.C., Hastings, A.P., Turley, N.E., Heard, S.B. & Agrawal, A.A. (2012) Adaptive  
711 geographical clines in the growth and defense of a native plant. *Ecological Monographs*, **82**,  
712 149–168.

713 Zverev, V., Zvereva, E.L. & Kozlov, M. V. (2020) Bird predation does not explain spatial variation  
714 in insect herbivory in a forest-tundra ecotone. *Polar Biology*, **43**, 295–304.

715 Zvereva, E.L., Castagneyrol, B., Cornelissen, T., Forsman, A., Antonio, J., Klemola, H.T., Paolucci,  
716 L., Polo, V. & Salinas, N. (2019) Opposite latitudinal patterns for bird and arthropod  
717 predation revealed in experiments with differently colored artificial prey. *Ecology and*  
718 *Evolution*, **9**, 14273–14285.

719

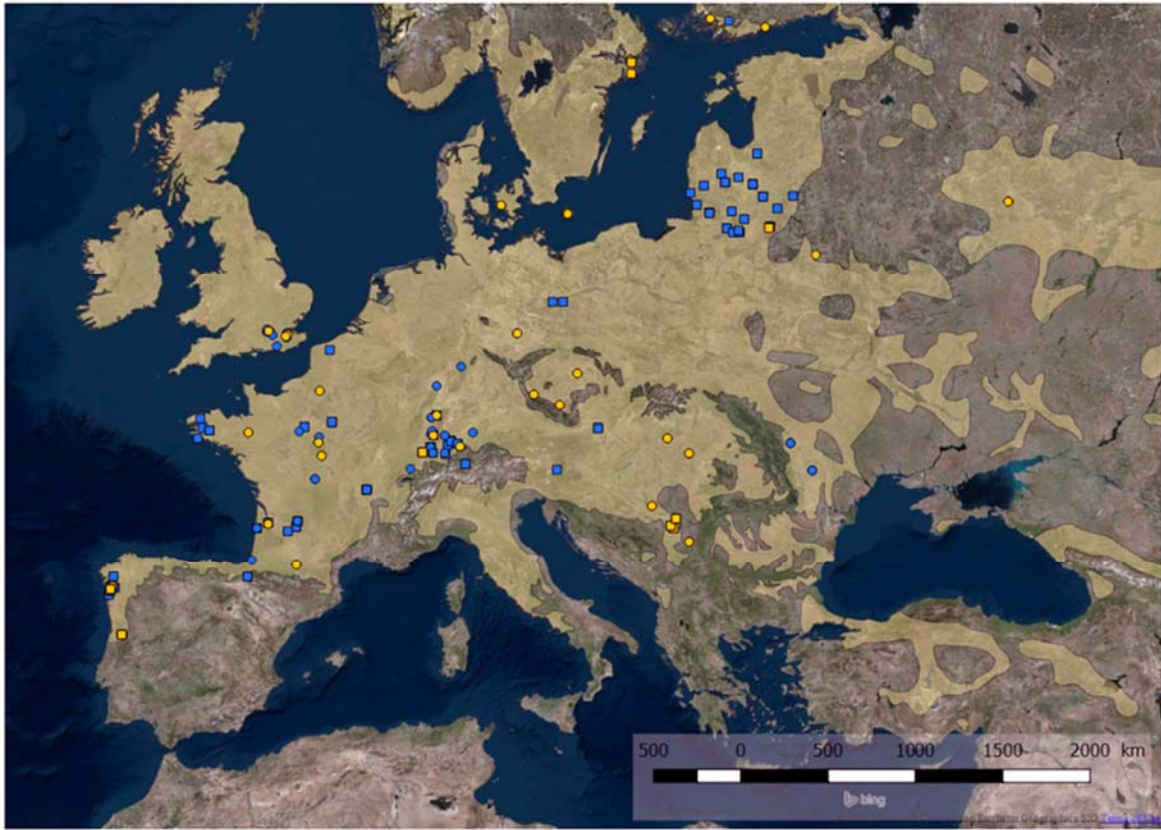
720

721

722

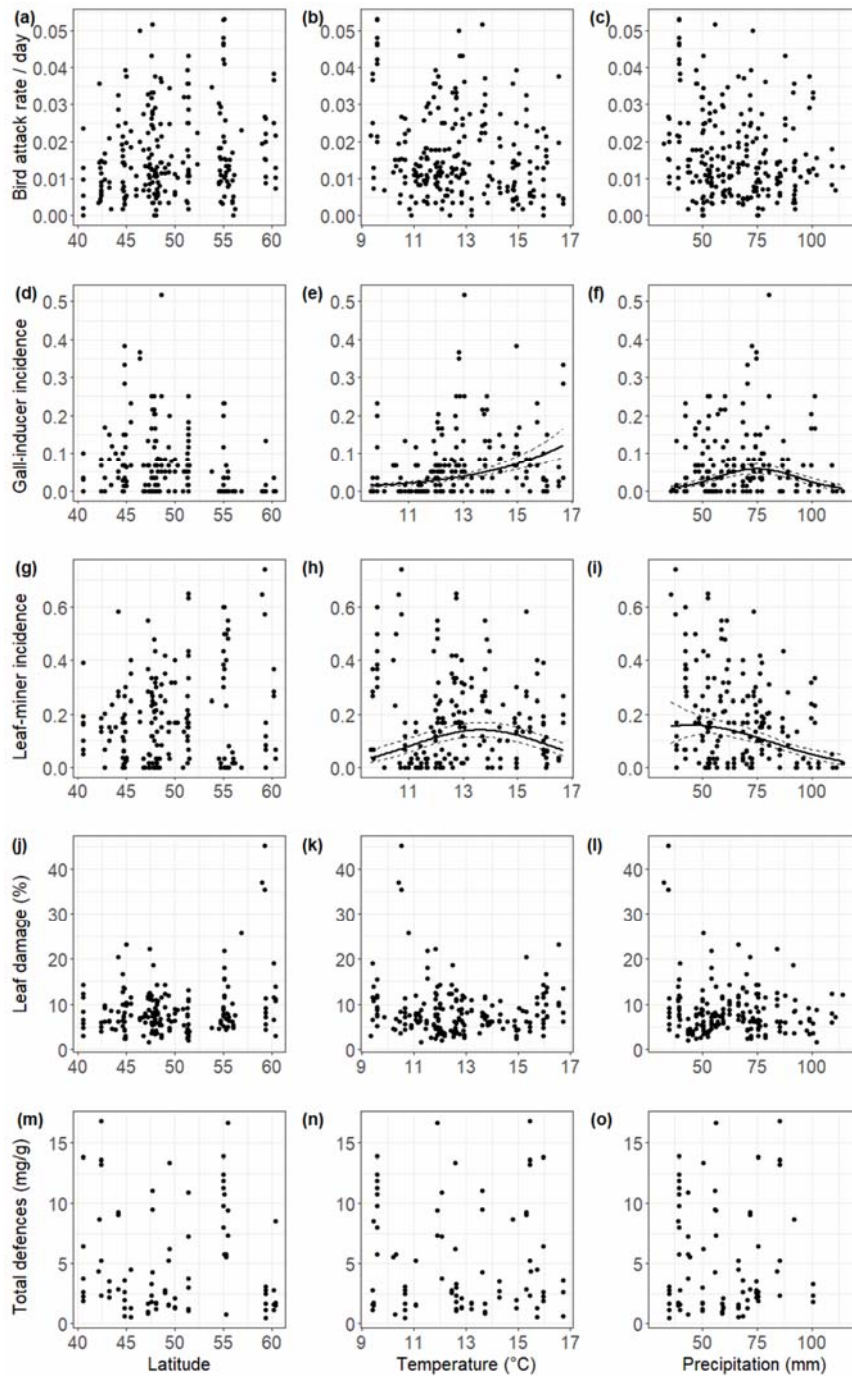
723

724



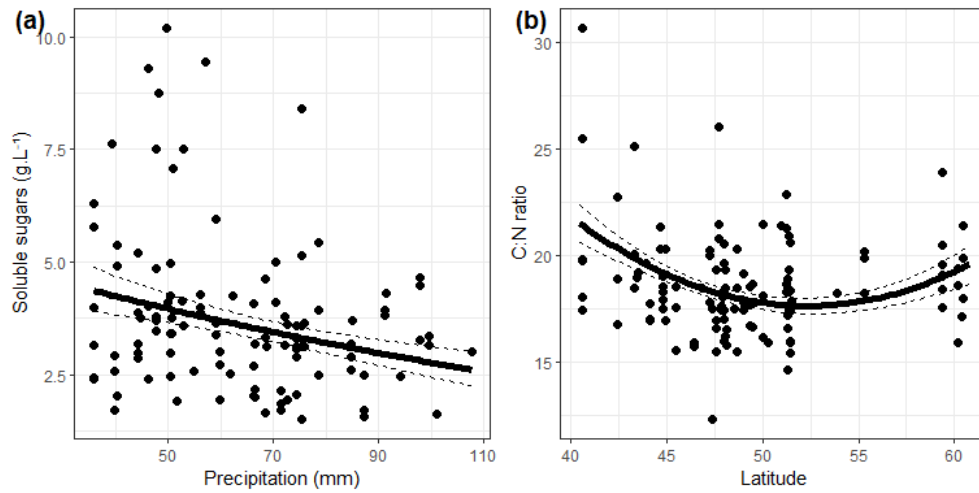
725

726 **Figure 1.** Distribution range of *Quercus robur* L. (shaded in yellow) and locations of trees sampled by professional  
727 scientists (orange symbols, 30 sites) and schoolchildren (blue symbols, 82 sites) in 2018 (circles, 57 sites) and 2019  
728 (squares, 55 sites). Additional maps showing oak trees used for estimating leaf herbivory, attack rates on dummy  
729 caterpillars and trait analyses are provided in supplementary material (Figure S1.1).



730

731 **Figure 2. Effects of Latitude, mean spring temperature and mean spring precipitation on bird attack rates (a, b, c),**  
 732 **gall-inducers incidence (d, e, f), leaf-miners incidence (g, h, i), leaf damage (j, k, l) and total phenolics (m, n, o). Dots**  
 733 **represent raw data averaged at the tree level. Solid and dashed lines represent model predictions (and corresponding**  
 734 **standard error) calculated after other significant variables (see Table S2.1) were set to their mean value. Only**  
 735 **statistically significant relationships are shown. Regression line equations are as follows: e,  $y = -3.32 + 0.44 \cdot x$ ; f,  $y =$**   
 736  **$-3.32 + 0.32 \cdot x - 0.51 \cdot x^2$ ; h,  $y = -1.98 + 0.37 \cdot x - 0.36 \cdot x^2$ ; i,  $y = -1.98 - 0.44 \cdot x$ .**



737

738 **Figure 3. Effect of mean spring precipitation and latitude on soluble sugar (a) and C:N ratio (b) on leaves,**  
 739 **respectively.** Dots represent raw data averaged at tree level. Solid and dashed lines represent model predictions  
 740 (and corresponding standard error) for temperature and latitude calculated after other significant variables (see  
 741 Table S2.2) were set to their mean value. Only significant relationships are shown. Regression line equations are as  
 742 follows: **a**,  $y = 1.53 - 0.10 \cdot x$ ; **b**,  $y = 17.9 - 0.86 \cdot x + 0.70 \cdot x^2$ .

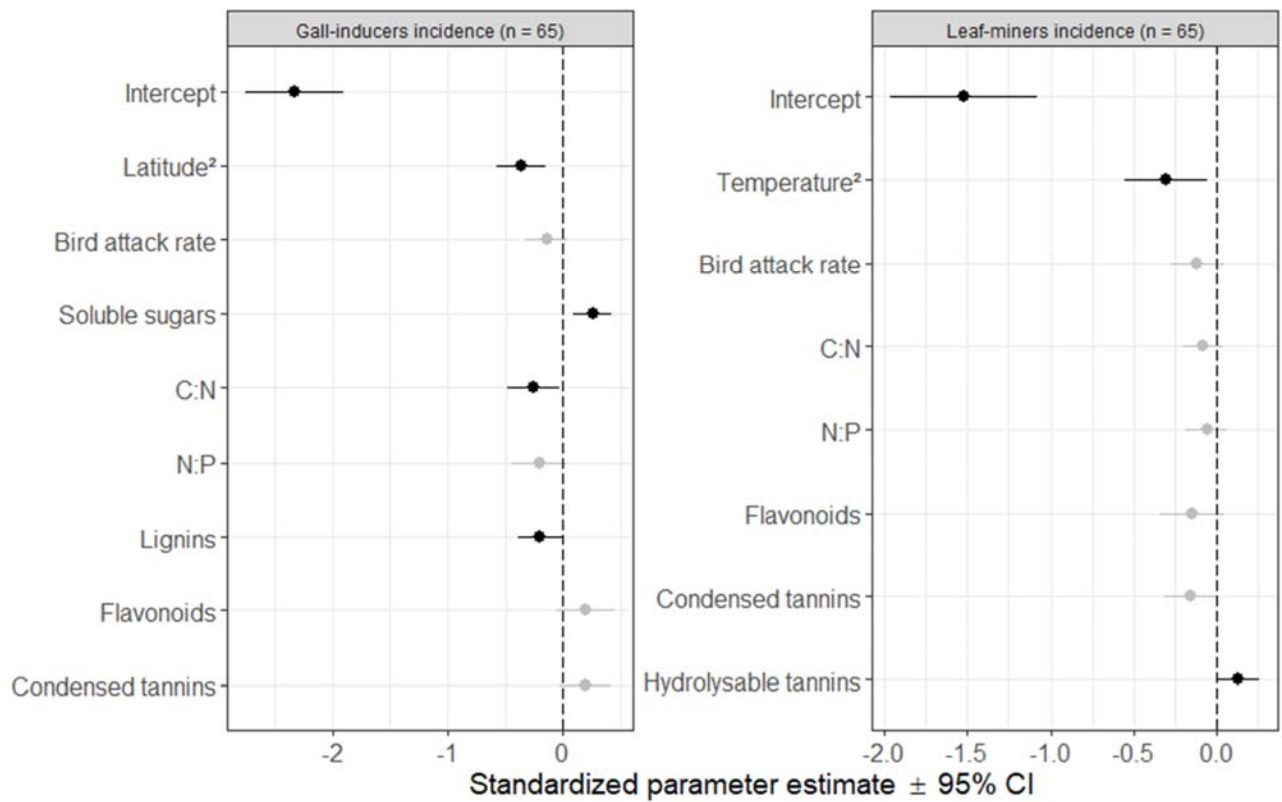
743

744

745

746

747



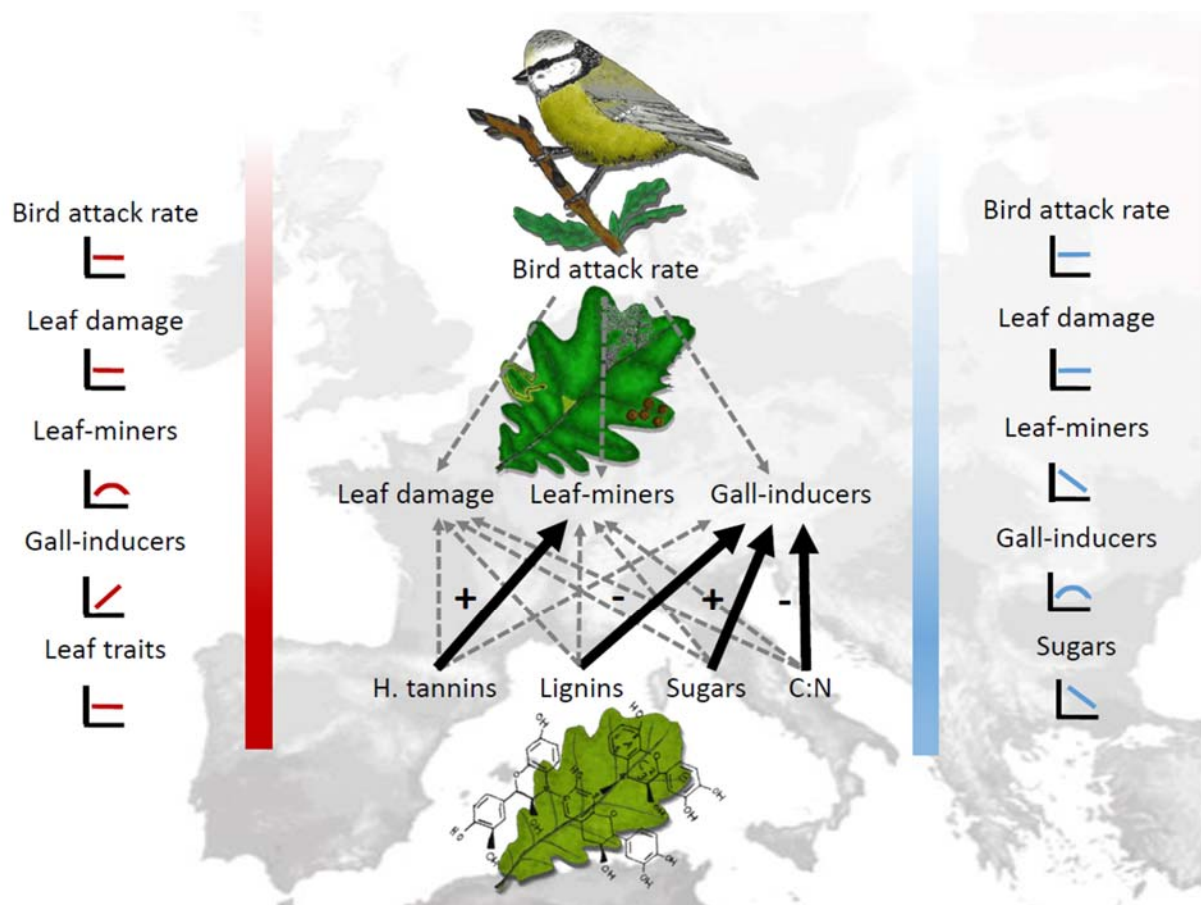
748

749

750 **Figure 4. Effects of latitude, mean spring temperature, mean spring precipitation and leaf chemistry on gall-inducer**  
 751 **(left) and leaf-miner (right) incidences.** Circles and error bars represent standardized parameter estimates and  
 752 corresponding 95% CI. The vertical dashed line centred on zero represents the null hypothesis. Black and grey circles  
 753 represent significant and non-significant effect sizes, respectively.

754

755



756

757

758 **Figure 5.** Summary illustrating plant-herbivore-predator relationships along a latitudinal gradient in Europe. The red  
 759 and blue bands denote the variation in mean spring temperature and precipitation, respectively. The five figures on  
 760 the left represent the correlation between the mean spring temperature and bird attack rate, leaf damage, the  
 761 incidence of gall-inducers and leaf-miners and leaf traits. The five figures on the right correspond with the correlation  
 762 between mean spring precipitation and bird attack rate, leaf damage, the incidence of leaf-miners and gall-inducers  
 763 and the concentration of soluble sugar in leaves. Solid black arrows represent significant positive (+) or negative (-)  
 764 relationships; dashed grey lines indicate non-significant relationships.

765

766 **Figure S1.1.** Location of the trees sampled for the assessment of herbivory (a), predation attack rate (b), leaf  
 767 nutritional content (c) and leaf defences (d). An interactive version of these maps are also included in the  
 768 supplementary material as Figures S1.1a, S1.1b, S1.1c and S1.1D.

769 Interactive version of the maps:

770 **Figure S1.1a.** Trees sampled for the assessment of herbivory.

771 **Figure S1.1b.** Trees sampled for the assessment of predation attack rate.

772 **Figure S1.1c.** Trees sampled for the assessment of leaf nutritional content.

773 **Figure S1.1d.** Trees sampled for the assessment of leaf defences.

774 **Figure S1.2.** Effects of partner type, year, mean spring temperature and mean spring precipitation on gall-inducers  
775 and leaf-miners incidences. Circles and error bars represent standardized parameter estimates and corresponding  
776 95% CI. The vertical dashed line centered on zero represents the null hypothesis. Black and grey circles indicate  
777 significant and non-significant effect sizes, respectively.

778 **Figure S1.3.** Importance of every variable (RVI) included in the geographic and climatic models that considered the  
779 effect of longitude, latitude, temperature and precipitation on herbivory (gall-inducers and leaf-miners incidences;  
780  $n = 182$ ) and on leaf chemistry (soluble sugar and C:N ratio;  $n = 114$ ).

781 **Figure S1.4.** Importance of every variable (RVI) included in the biotic model that considered the effect of leaf traits,  
782 bird attack rate, climatic variables on gall-inducers and leaf-miners incidence ( $n = 65$ ).

783 **Table S1.1.** Summary of the different variables measured.

784 **Table S2.1.** Summary of model coefficient parameter estimates (i.e., effect sizes), degrees of freedom (df), log-  
785 likelihood, AICc,  $\Delta$ AICc, AICc weight ( $w_i$ ) and the variance explained by fixed ( $R2m$ ) and fixed plus random factors  
786 ( $R2c$ ) of the different climatic models. The gradient of colours from red to green corresponds to the effect size,  
787 from large negative (red) to large positive (green) effect sizes. The set of models competing with the best model  
788 within 2 units of  $\Delta$ AICc is highlighted in bold font. *Partner type* is the effect of partner type (the estimate for  
789 schoolchildren being compared with the estimate for professional scientists that was included in the intercept) and  
790 *Year* is the effect of each year (2019 contrasted with 2018).

791 **Table S2.2.** Summary of model coefficient parameter estimates (i.e., effect sizes), degrees of freedom (df), log-  
792 likelihood, AICc,  $\Delta$ AICc, AICc weight ( $w_i$ ) and the variance explained by fixed ( $R2m$ ) and fixed plus random factors  
793 ( $R2c$ ) of the different climatic models. The gradient of colours from red to green corresponds to the effect size,  
794 from large negative (red) to large positive (green) effect sizes. The set of models competing with the best model  
795 within 2 units of  $\Delta$ AICc is highlighted in bold font.

796

797

798