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1 Normalization in LCA: How to ensure consistency?

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9 1 Introduction

10 1.1 Overview

11 In this discussion paper, we focus on a topic that seems probably not sufficiently addressed in Life
12 Cycle Assessment (LCA): the normalization (Andreas et al. 2020). More precisely, we look again at the need for
13 the normalization step in LCA and inconsistencies caused by this step. We highlight the importance of having
14 the same data source for the system under study and the normalization references. While this is possible in a
15 conventional normalization way for Economic Input-Output LCA (EIO-LCA), this remains a challenge for
16 process-based LCA. We show how to overcome this limitation and discuss how a normalization by the
17 geometric mean of the inventory database (Hélias et al. 2020) ensures consistency.

18 1.2 Context

19 In recent decades, the interconnections between people and nature have been increasingly highlighted.
20 Many concepts and approaches have been developed to address these relationships, as the "One Health" concept
21 to bring together three of them: humans, animals and their environments (Gibbs 2014), the seventeen Sustainable
22 Development Goals for a single planet (UNEP 2020), the global biodiversity affected by five drivers
23 (Millennium Ecosystem Assessment 2005), the product environmental footprint to display sixteen impacts
24 (Fazio et al. 2018) or the three areas of protection (Verones et al. 2017). Preserving the environment as a whole
25 does not elude the recurring problem of most of the decision-making processes: the multicriteria.

26 Arguing the relevance of impacts on a case-by-case basis, contextualizing and putting them into
27 perspective with regard to the options available for the system under study, are all attempts to avoid a multi-
28 attribute decision-making (MADM). Nevertheless, a generic, structured, repeatable, automatable, and objective

1 decision-making process is an expectation. In the field of LCA, this justifies the normalization and weighting
2 steps. The weight represents the importance of the criterion and its design is a real research topic in itself, but we
3 are interested here in normalization.

4 2 **Method**

5 2.1 **Meaning of the normalization**

6 The normalization in the LCA framework is the division of the impact computed for the system under
7 study, by the impact of a reference value named the normalization reference. LCA normalization has been
8 identified as a leading driver in the aggregation process, with high consequences on the results (Myllyviita et al.
9 2014; Prado et al. 2019; Muhl et al. 2021).

10 Normalization has three main purposes (Pizzol et al. 2017): (1) to compare the results in order to check
11 plausibility, (2) to facilitate communication, and (3) to be free of unit constraints for weighting, with the impacts
12 expressed on a common scale. This last point focuses our attention here.

13 In LCA, normalization is also twofold (Laurent and Hauschild 2015): (1) internal if the normalization
14 reference is defined on the basis of case studies, as the result of a system “A” expressed in proportion of the
15 result of a system “B”, and (2) external if an independent reference is used, as the results of systems “A” and
16 “B” expressed in proportion to the average citizen year impact.

17 An internal normalization ensures, by definition, consistency between the system under study and the
18 reference. However, it remains context-dependent, which is contrary to the objectives of LCA, and cannot be
19 used with generic weighting (Norris 2001). An external normalization makes possible an aggregation process
20 that is reproducible and independent of the study. However, two modelling processes are used, one for the
21 system and the other for the reference. This raises questions about the representativeness and the consistency
22 between both representations (Heijungs et al. 2007).

23 The normalization task (sometimes named standardization in some scientific communities) is not
24 unique to LCA and is encountered by anyone who deals in data processing. For decision, not all, but most
25 MADM models require a normalization step. The choice of the normalization method is often part of the
26 MADM process and this affects the result: A different method of normalization will lead to a different result.
27 There is no consensus on this step in the MADM community and Jahan and Edwards (2015) reviewed 31
28 methods (sum, norm, max-min, Z-transformation...). All of them are based on the systems under study and from
29 LCA viewpoint, they are internal normalizations. LCA was initially defined outside the framework of the
30 MADM (Pizzol et al. 2017) and the choice was made to mainly use external normalization: As a result, the

1 aggregation process is not determined by the specifics of the study, contrary to most MADM methods. This is
2 undoubtedly an advantage for the genericity, repeatability, and independence of the process.

3 Global external normalizations are now recommended in the guidelines (Verones et al. 2017). This is
4 the most relevant system of reference: In our globalized economy, any functional unit involves processes across
5 all continents and the problem has to be taken as a whole, using a normalization reference encompassing the
6 entire world.

7 2.2 Consistency issue

8 LCA is a modelling work: For a functional unit, we draw up the associated impacts. This process is
9 built with rules, and one of the most important is additivity: the impact of two functional units is the sum of the
10 impacts of each. In this sense, external normalization means dividing the impact of the product or service
11 (corresponding to the functional unit of the study) by the sum of all the products and services on a global scale
12 (the reference). To ensure consistency in the modelling work, the way in which the products and services are
13 represented should be the same. As an illustration, Figure 1 offers a simplistic schematic view, with the system
14 under study being an apple and the reference for normalization being the world. From real systems, the
15 modelling processes allow having representations, here blurred or partial. For the normalization, having a
16 blur/blur or partial/partial ratio is preferable to a blur/partial or partial/blur one, whatever the soundness of the
17 modelling.

18 *Figure 1 about here*

19 The raw data and their processing must be identical both for the system under study and the
20 normalization references (unless it can be proved that there is no bias or that the bias is identical). In other
21 words, we can say that an LCA is a model of a part of the world and the reference is a model of the world.
22 Models and choices must be consistent.

23 2.2.1 Economic Input-Output LCA

24 An EIO-LCA uses aggregated sectoral data and their interdependencies. Fractions of the impacts
25 associated to the sectors of the economy are then attributed to the system under study with the Leontief's well-
26 known equation. With this top-down approach, we have the same model between the world (the totality of all
27 economic sectors) and the system under study (parts of economic sectors). We can ensure consistency in the
28 normalization process (subject to the calculating of this value from the EIO database). Unfortunately, the
29 process-based LCA differs.

2.2.2 Process-based LCA

A process-based LCA is more like a Lego® construction. The LCA practitioners take “bricks” from a Life Cycle Inventory (LCI) database and assemble them to build the system under study. Sometimes, they create a “brick”, but this is far from being their primary activity. For each type of “brick” (a component or a process, called a dataset by LCI database providers) that is required, the practitioners take a certain number of them (i.e. the final demand). With this bottom-up approach, we do not have the information to shape the world as a whole. All the datasets into the LCI database describe what constitutes the world, but the quantity of each is missing. As an example, we can consider that the dataset "market for apple | apple (GLO)" from ecoinvent (Wernet et al. 2016) allows us to have the average impact of an apple at the world level, but it does not allow us to know how many apples are produced in all the orchards of the world.

With only a process-based LCA database, we cannot calculate the LCA of human activity at the global level. The consistency of the normalization cannot be argued.

3 Results

3.1 Normal versus log-normal laws

In LCA, we normalize by the whole impact (or by a proportion of it when reduced to a citizen, but the reasoning remains the same). This is undeniably a relevant solution, the result being a fraction (studied system divided by the whole) that is easy to interpret and communicate. Divided by the sum of elements is one of the normalization identified by Jahan and Edwards (2015) but there are many others.

In data processing, looking at the law of data distribution to select the descriptors is always relevant. The best known is the normal distribution law, described by the arithmetic mean and the standard deviation. When in LCA we divide by the totality of the impact as normalization, up to a factor (the number of elements that make up the whole), we divide by the arithmetic mean. Implicitly, this says that we consider this arithmetic mean is a good descriptor of impacts, that we are in a "normal" world.

But the LCA practitioner is not working in a "normal" world. Uncertainty calculations are generally done with log-normal distribution laws. In an LCI database, the distribution of substance emissions commonly follows a log-normal law (Qin and Suh 2017). Some LCA results are interpreted in orders of magnitude (such as toxicities (Frischknecht and Jolliet 2019)) and therefore on a logarithmic scale. An LCA is the combination of elements following log-normal distributions. If no formal solution has been found for a sum of log-normal variables, there is a general agreement this is well approximated by a log-normal distribution (Beaulieu et al. 1995). Although the log-normal law is less intuitive than the normal law, this distribution frequently occurs in

1 nature (abundance of species, concentrations of minerals in the earth's crust, concentrations of pollutants in the
 2 atmosphere, dose-response curves, etc.) (Limpert et al. 2001). We can reasonably assume the impacts of the set
 3 of products and services follow log-normal distributions.

4 3.2 Geometric mean as the normalization reference

5 This leads to thinking about normalization in a different way. The log-normal distribution is described
 6 by the geometric mean, not by the arithmetic one. The geometric mean is more robust to extreme values, but
 7 more importantly, it results from a multiplication, while the arithmetic mean is constructed from a sum. It
 8 therefore becomes possible to factorize and thus simplify the result.

9 As an illustration, let a reference system resulting from n processes ($i \in \{1, \dots, n\}$), with m impacts ($j \in$
 10 $\{1, \dots, m\}$), let f_i the final demand of the process i and $h_{i,j}$ the impact j of the process i ($h_{i,j}$ results from the
 11 multiplication of all elementary flows of the process by the associated characterisation factors.). The arithmetic
 12 mean of each impact category for the reference system is

$$13 M_j^A = \frac{1}{n} \sum_{i=1}^n f_i h_{i,j}$$

14 and the geometric mean

$$15 M_j^G = \left(\prod_{i=1}^n f_i h_{i,j} \right)^{\frac{1}{n}} = \left(\prod_{i=1}^n f_i \right)^{\frac{1}{n}} \times \left(\prod_{i=1}^n h_{i,j} \right)^{\frac{1}{n}}$$

16 The term $\left(\prod_{i=1}^n f_i \right)^{\frac{1}{n}}$ is constant whatever the impact, therefore it can be removed without changing the
 17 proportions between the normalized impacts.

18 With $\left(\prod_{i=1}^n h_{i,j} \right)^{\frac{1}{n}}$, the geometric mean of $h_{i,j}$ instead of $f_i h_{i,j}$, we have a normalization reference
 19 value that we only calculate with the LCI database, without having to deal with the final demand for the
 20 reference. This approach ensures consistency between the modelling of the system under study and the
 21 normalization references. This approach is especially relevant for process-based LCA, but can also be used with
 22 EIO-LCA.

23 4 Discussion

24 Normalizing by the geometric mean of the inventory database raises several considerations that are
 25 worthy of discussion. We present below elements of responses to several remarks that can be made.

4.1 About the geometric mean

First remark : Even if there are no data bias, normalizing with the geometric mean gives a different result than normalizing with the overall impact. Indeed, normalizing by geometric mean will not lead to the same conclusions as the "conventional" normalization, each normalization choice plays a key role in the aggregation/decision process as shown in MADM. However, it is worth noting that the results are identical with the assumption of a constant $\frac{M_j^1}{M_j^0}$ ratio regardless of j (Hélias et al. 2020). They prove the results are highly correlated in some cases, despite data biases that can be explained. More precisely, assuming lognormal distribution, $\frac{M_j^1}{M_j^0} = \sqrt{1 + cv_j}$ (see equation (10-11) in Limbrunner et al. (2000)) and the hypothesis of a constant ratio becomes a constant coefficient of variation $cv_j, \forall j$ (or at least small changes in the coefficient of variations). Note in addition that the variability of the ratio $\frac{M_j^1}{M_j^0}$ is reduced by the square root with respect to cv_j . It would be interesting to investigate further the values of these ratios with the EIO-LCA databases where M_j^1 and M_j^0 are then computable from the same data for all impact categories.

Second remark: The geometric mean does not make the quantities commensurable. Indeed, for the purposes of normalization, the objective of being able to compare and communicate the results is not achieved by this normalization. This approach is interesting from the perspective of aggregating impacts. With some practice, reasoning with the geometric mean would be possible for an LCA practitioner but it would obviously be uncomfortable to communicate, the citizen-equivalent remaining the best solution for this purpose.

4.2 About the inventory database as data for normalization

Third remark: As the LCI database is used, this is not a true external normalization. In this approach, there is a link between the system studied and the references, but the building of both from the same source does not make it an internal normalization. The normalization here is not based on the case under study, but on the used database. It therefore goes beyond the internal normalization and can therefore be considered as an external normalization.

Fourth remark: No LCI database is perfect, some elements are missing and none represents the world as a whole. Over time, LCI databases are improved, but they are still perfectible. Using them to represent the world implies that the world is imperfectly modelled. But it is as imperfectly modelled as the system under study. Consider one modelled with a database where infrastructures are not represented, as the agrifootprint database (e.g. see Corrado et al. (2018)). Using a reference including infrastructure is not really relevant: the

1 biases are potentially different according to the impacts (Heijungs et al. 2007), which will play a role in
2 aggregation and may lead to biased conclusions. The best solution would obviously be a model of the system
3 under study with infrastructures, but if missing they have to be missing in the reference as well. Consistency of
4 the modelling is more important than the completeness of the reference for normalization purpose.

5 *Fifth remark: The LCA practitioner uses several databases / the LCA practitioner modifies the*
6 *database.* It is always risky to make an LCA with heterogeneous databases that do not necessarily have the same
7 perimeters. But in this case, it is enough to take as normalization value the geometrical mean of the union of the
8 databases. If LCA practitioners substantially modify their databases, it is indeed necessary to recalculate the
9 reference values and they will therefore have to do so (practitioners could also make their data public, these data
10 would then be directly integrated into the inventory databases and the reference values would be calculated for
11 them).

12 *Sixth remark: The results will change if another LCI database is used.* This is a necessity to ensure
13 consistency between the system under study and the normalization reference. In this case, there is no longer a
14 single set of normalization values linked to a life cycle impact assessment (LCIA) method, but rather a set for
15 each database-LCIA method pair. An alternative would be to define a single reference resulting from the union
16 of all the inventory databases, but this deserves further consideration.

17 4.3 About planetary boundaries as normalization

18 *Seventh remark: With a normalization by planetary boundaries, the problem of consistency between*
19 *inventory and reference seems solved.* We have a real enthusiasm for the use of planetary boundaries in LCA
20 (see Bjørn et al. (2020) for a review), Bjørn and Hauschild (2015) being probably the most emblematic
21 corresponding work. The principle is to use the carrying capacity as a normalization value instead of the global
22 impact. Planetary boundaries are biophysical characteristics, not determined by human activities as it is the case
23 for the system under study. Consequently, one might consider that it is not necessary to ascertain the consistency
24 of the modelling. It is then possible to argue that it is not the planetary limit that is the value of interest, but
25 whether or not it is exceeded, or close to it. In Bjørn and Hauschild (2015), this is addressed considering the
26 normalization by planetary boundaries as a distance-to-target: Let G_j , the global impact j , C_j the corresponding
27 carrying capacity according to the planetary boundaries and S_j the impact of the system under study. The
28 normalization by the planetary boundary $\frac{S_j}{C_j}$ corresponds to the simplification of $\frac{S_j}{G_j} \times \frac{G_j}{C_j}$, with $\frac{S_j}{G_j}$ the result of the
29 conventional normalization and $\frac{G_j}{C_j}$ the distance-to-target, named the reduction factor by Vargas-Gonzalez et al.

1 (2019). This result makes it possible to be free of the conventional normalization reference (G_j) and at first
2 glance, S_j and C_j do not therefore need to be modelled in the same way. However, two limitations can be raised.
3 The first one is a distance-to-target expressed in another way, with non-linearity (as squared with $\left(\frac{G_j}{C_j}\right)^2$ as it is
4 done in the ecoscarcity method (Frischknecht and Büsler Knöpfel 2013)), which makes this simplification
5 unfeasible. The second is that we only represent these elements by means of models. We try to determine S_j but,
6 with our data, we only obtain an approximation, we will name \hat{S}_j . The same happens with C_j , our knowledge of
7 the involved mechanisms and our hypotheses only allow us to obtain a representation \check{C}_j . In $\frac{S_j}{G_j}$ the modelling of
8 the numerator must be consistent with the modelling of the denominator (i.e. $\frac{\hat{S}_j}{\hat{G}_j}$) and the same is true for $\frac{G_j}{C_j}$ with
9 $\frac{\check{C}_j}{\check{G}_j}$. As the system under study and the carrying capacity do not have the same processes of modelling, the
10 simplification (\hat{G}_j versus \check{G}_j) cannot be achieved. Unless to prove the similarities of the modelling, the problem
11 of consistency between the studied system and normalization remains with planetary boundaries. The
12 geometrical mean solves this problem.

13 5 Conclusion

14 Multi-criteria decision-making is a complex field of research and there is no universal method that can
15 be applied in all cases. The LCA presents an interesting particularity, with normalization we generally use
16 external information to express the attributes (i.e. impacts) into a common scale. This makes the multi-criteria
17 decision process more generic, independent of the case study. The other side of the coin is the introduction of an
18 additional bias via different modelling between the system under study and the reference, especially for process-
19 based LCA. Normalization by the geometric mean of the inventory database avoids this bias. That is argued
20 considering that log-normal laws govern the phenomena and processes under investigation. In addition to this
21 normalization issue, describing the database by its geometric means and comparing them to the cumulative
22 impact values is a way to see the exhaustiveness of the database (subject to the exhaustiveness of the
23 normalization value, see Hélias et al. (2020)). This can be an additional step towards more robust and relevant
24 LCAs, that best represent environmental impacts.

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References

- 1
2 Andreas R, Serenella S, Jungbluth N (2020) Normalization and weighting: the open challenge in LCA.
3 Int J Life Cycle Assess. <https://doi.org/10.1007/s11367-020-01790-0>
- 4 Beaulieu NC, Abu-Dayya AA, McLane PJ (1995) Estimating the distribution of a sum of independent
5 lognormal random variables. IEEE Trans Commun 43:2869. <https://doi.org/10.1109/26.477480>
- 6 Bjørn A, Chandrakumar C, Boulay A-M, et al (2020) Review of life-cycle based methods for absolute
7 environmental sustainability assessment and their applications. Environ Res Lett 15:083001.
8 <https://doi.org/10.1088/1748-9326/ab89d7>
- 9 Bjørn A, Hauschild MZ (2015) Introducing carrying capacity-based normalisation in LCA: framework
10 and development of references at midpoint level. Int J Life Cycle Assess 20:1005–1018.
11 <https://doi.org/10.1007/s11367-015-0899-2>
- 12 Corrado S, Castellani V, Zampori L, Sala S (2018) Systematic analysis of secondary life cycle
13 inventories when modelling agricultural production: A case study for arable crops. J Clean Prod 172:3990–4000.
14 <https://doi.org/10.1016/j.jclepro.2017.03.179>
- 15 Fazio S, Castellani V, Sala S, et al (2018) Supporting information to the characterisation factors of
16 recommended EF Life Cycle Impact Assessment method. EUR 28888 EN, European Commission, JRC109369,
17 Ispra
- 18 Frischknecht R, Büsler Knöpfel S (2013) Swiss Eco-Factors 2013 according to the Ecological Scarcity
19 Method. Methodological fundamentals and their application in Switzerland. Environmental studies no. 1330.
20 Bern
- 21 Frischknecht R, Jolliet O (eds) (2019) Global Guidance for Life Cycle Impact Assessment Indicators:
22 Volume 2. UNEP/SETAC Life Cycle Initiative
- 23 Gibbs EPJ (2014) The evolution of one health: A decade of progress and challenges for the future. Vet
24 Rec 174:85–91. <https://doi.org/10.1136/vr.g143>
- 25 Heijungs R, Guinée J, Kleijn R, Rovers V (2007) Bias in normalization: Causes, consequences,
26 detection and remedies. Int J Life Cycle Assess 12:211–216. <https://doi.org/10.1065/lca2006.07.260>
- 27 Hélias A, Esnouf A, Finkbeiner M (2020) Consistent normalization approach for Life Cycle
28 Assessment based on inventory databases. Sci Total Environ 703:134583.
29 <https://doi.org/10.1016/j.scitotenv.2019.134583>

1 Jahan A, Edwards KL (2015) A state-of-the-art survey on the influence of normalization techniques in
2 ranking: Improving the materials selection process in engineering design. *Mater Des* 65:335–342.
3 <https://doi.org/10.1016/j.matdes.2014.09.022>

4 Laurent A, Hauschild MZ (2015) Normalisation. In: Hauschild MZ, Huijbregts MAJ (eds) *Life cycle*
5 *impact assessment*. Springer Netherlands, Dordrecht, pp 271–300

6 Limbrunner JF, Vogel RM, Brown LC (2000) Estimation of Harmonic Mean of a Lognormal Variable.
7 *J Hydrol Eng* 5:59–66. [https://doi.org/10.1061/\(ASCE\)1084-0699\(2000\)5:1\(59\)](https://doi.org/10.1061/(ASCE)1084-0699(2000)5:1(59))

8 Limpert E, Stahel WA, Abbt M (2001) Log-normal Distributions across the Sciences: Keys and Clues.
9 *Bioscience* 51:341–352. [https://doi.org/10.1641/0006-3568\(2001\)051\[0341:LNDATS\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2001)051[0341:LNDATS]2.0.CO;2)

10 Millennium Ecosystem Assessment (2005) *Ecosystems and human well-being: Biodiversity synthesis*.
11 World Resources Institute, Washington D.C.

12 Muhl M, Berger M, Finkbeiner M (2021) Distance-to-target weighting in LCA—A matter of
13 perspective. *Int J Life Cycle Assess* 26:114–126. <https://doi.org/10.1007/s11367-020-01837-2>

14 Myllyviita T, Leskinen P, Seppälä J (2014) Impact of normalisation, elicitation technique and
15 background information on panel weighting results in life cycle assessment. *Int J Life Cycle Assess* 19:377–386.
16 <https://doi.org/10.1007/s11367-013-0645-6>

17 Norris GA (2001) The requirement for congruence in normalization. *Int J Life Cycle Assess* 6:85.
18 <https://doi.org/10.1007/BF02977843>

19 Pizzol M, Laurent A, Sala S, et al (2017) Normalisation and weighting in life cycle assessment: quo
20 vadis? *Int J Life Cycle Assess* 22:853–866. <https://doi.org/10.1007/s11367-016-1199-1>

21 Prado V, Cinelli M, Ter Haar SF, et al (2019) Sensitivity to weighting in life cycle impact assessment
22 (LCIA). *Int J Life Cycle Assess*. <https://doi.org/10.1007/s11367-019-01718-3>

23 Qin Y, Suh S (2017) What distribution function do life cycle inventories follow? *Int J Life Cycle*
24 *Assess* 22:1138–1145. <https://doi.org/10.1007/s11367-016-1224-4>

25 UNEP (2020) United Nation, Department of Economic and Social Affairs, The 17 Goals.
26 <https://sdgs.un.org/goals>. Accessed 29 Jul 2020

27 Vargas-Gonzalez M, Witte F, Martz P, et al (2019) Operational Life Cycle Impact Assessment
28 weighting factors based on Planetary Boundaries: Applied to cosmetic products. *Ecol Indic* 107:105498.
29 <https://doi.org/10.1016/j.ecolind.2019.105498>

1 Verones F, Bare J, Bulle C, et al (2017) LCIA framework and cross-cutting issues guidance within the
2 UNEP-SETAC Life Cycle Initiative. *J Clean Prod* 161:957–967. <https://doi.org/10.1016/j.jclepro.2017.05.206>
3 Wernet G, Bauer C, Steubing B, et al (2016) The ecoinvent database version 3 (part I): overview and
4 methodology. *Int J Life Cycle Assess* 21:1218–1230. <https://doi.org/10.1007/s11367-016-1087-8>

5

6 **Figure Caption**

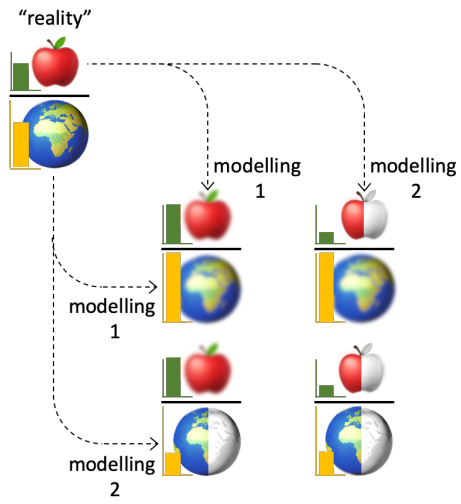
7

8 **Figure 1. Schematic representation of the consistency issue in the normalization step**

9

1

Figure



2

3

Figure 1