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## Enantiomeric fraction of hexabromocyclododecanes in foodstuff from the Belgian market

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27 **Abstract**

28 Diet is considered a major route of human exposure to hexabromocyclododecane, a chiral  
29 environmental contaminant. A previous study reported on the occurrence of  
30 hexabromocyclododecane diastereoisomers in food items of animal origin collected in Belgium. The  
31 present study reports further results on corresponding enantiomeric fractions of the same samples.  
32 None of the samples could be considered as racemic for the  $\alpha$ -isomer suggesting that foodstuff  
33 contamination occurred prior to death of the corresponding producing animal and was not the result  
34 of the food item being in contact with technical HBCDD. Non-racemic chiral signatures were also  
35 observed for  $\beta$ - and  $\gamma$ -isomers. We conclude that, depending on their dietary habits, different  
36 individuals might be overall exposed to non-racemic profiles. Considering that toxicological effects are  
37 enantiomer-dependent, this could modulate potential adverse effects.

38

39 **Keywords**

40 Environmental contaminant; HBCDD; Chiral compound; Enantiomer; Food.

41

## 42 1. Introduction

43 1,2,5,6,9,10-Hexabromocyclododecane (HBCDD) is a brominated flame retardant exhibiting  
44 developmental neurotoxicity (Maurice et al., 2015) listed in the Annex A of the Stockholm Convention  
45 (UNEP) since 2014. Although 16 stereoisomers are possible, technical HBCDD is a racemic mixture of  
46 mainly 3 enantiomer pairs:  $\gamma$ -HBCDD (~80%),  $\beta$ -HBCDD (~10%) and  $\alpha$ -HBCDD (~10%) (Heeb et al., 2005).  
47 While HBCDD contaminating surface soils close to suspected emission sources (e-waste recycling  
48 areas) remains racemic (Gao et al., 2011; Zhu et al., 2017),  $\alpha$ -HBCDD dominates in biota and most  
49 processed food due to differential biodegradation, bioisomerisation and biomagnification rates (Law  
50 et al., 2014; Barghi et al., 2016).

51 Further, due to stereospecific biological processes depending on species, enantiomeric  
52 enrichments have been reported in biota, including plant (Zhu et al., 2016), earthworm (Li et al., 2016),  
53 marine mollusc, crustacean and fish (Janák et al., 2005; Köppen et al., 2010; Ortiz et al., 2011; Du et  
54 al., 2012; Zhang et al., 2014; Zhu et al., 2017; Ruan et al., 2018a), bird (Janák et al., 2008; Esslinger et  
55 al., 2011; Sun et al., 2012; Jondreville et al., 2017; Omer et al., 2017), marine mammal (Vorkamp et al.,  
56 2012; Ruan et al., 2018b) and human (Roosens et al., 2009; Abdallah et al., 2011).

57 Diet is considered a major route of human exposure to HBCDD through foodstuffs of animal origin  
58 (EFSA, 2011). In addition to the diastereoisomeric profile, chiral signature appears as a valuable  
59 information to determine whether such food was contaminated ante- or post-mortem of the  
60 producing animal, thereby contributing to elucidation of the contamination source (Omer et al. 2017).

61 A previous study reported on the occurrence of HBCDD diastereoisomers among other halogenated  
62 flame retardants in food items available on the Belgian market (Poma et al., 2018). Edible parts of 1289  
63 individual food samples collected in 2015–2016 were pooled, ground and homogenized to create 183  
64 composite food samples (further referred to as “samples”) representative of 15 relevant sub-  
65 categories. Results revealed the presence of HBCDDs above limits of quantification (LOQs, 5–250 pg g<sup>-1</sup>  
66 wet weight – ww) in 22 samples of food from animal origin belonging to 7 sub-categories, and up to

67 5.5 ng g<sup>-1</sup> ww in eel. Those specific results are gathered in **Table S1**. Authors provided complete results  
68 and discussed diastereomeric patterns thoroughly.

69 The present study aims at complementing the study of [Poma et al. \(2018\)](#) with information on  
70 enantiomeric fractions of HBCDD isomers, to discuss chiral signatures.

71

## 72 **2. Material and methods**

### 73 *2.1. Sample preparation and diastereoisomeric analysis*

74 Since the original extracts from [Poma et al. \(2018\)](#) (assay A) were no longer available, twenty-one  
75 out of the 22 samples from this previous study exhibiting quantified levels of at least one HBCDD  
76 diastereoisomer (**Table S1**) were extracted again from stored (-20 °C) lyophilised matter and analysed  
77 according to the same procedure and using the chemicals (assay B) ([Malysheva et al., 2018](#)). Briefly,  
78 2.5 g lyophilised sample (or 1 g oil) fortified with <sup>13</sup>C-labelled internal standards for quantification  
79 according to the isotopic dilution principle were extracted with a mixture of hexane/dichloromethane  
80 1:1 (v/v), purified with acidic silica and reconstituted in 100 µL acetonitrile prior to analysis. Reversed  
81 phase liquid chromatography coupled to tandem mass spectrometry (Xevo TQ-S, Waters, Milford, MA,  
82 USA) fitted with an electrospray ionisation source operating in the negative mode was used for  
83 analysis. Only diastereoisomer occurrences previously quantified by assay A were considered for  
84 assay B.

85

### 86 *2.2. Enantiomeric fraction*

87 Extracts were further reconstituted in a mixture of acetonitrile/water 4:1 (v/v, 50 µL) and HBCDD  
88 enantiomers analysed according to [Omer et al. \(2017\)](#) with minor modifications (assay C). Briefly,  
89 enantiomers were separated on a cellulose tris-(3,5-dimethylphenylcarbamate) chiral column  
90 (ACQUITY UPC<sup>2</sup> Trefoil CEL1, 2.1 mm × 150 mm, 2.5 µm of granulometry, Waters) and analysed using  
91 an Orbitrap Q-Exactive mass spectrometer fitted with an electrospray ionisation source (Thermo  
92 Scientific, San Jose, CA, USA). Data were acquired in negative mode over the *m/z* range [620–680] at a

93 resolving power of about 78,000 full width at half-maximum for the [M – H]<sup>-</sup> monitored ions. Peak  
94 areas from extracted ion chromatograms (± 5 ppm) were used to calculate enantiomer concentrations.  
95 **Figure S1** illustrates the typically obtained chromatographic separation for a selected sample (fresh  
96 oysters MC-01). Further analytical and QA/QC details are available in the Supplementary Data.  
97 Enantiomeric fraction (EF) of each considered enantiomeric pair was determined according to the  
98 following equation:

$$99 \quad EF = \frac{\frac{(+A)}{(+A)_{IS}}}{\frac{(+A)}{(+A)_{IS}} + \frac{(-A)}{(-A)_{IS}}}$$

100 where (+)A and (-)A represent the peak areas for the native pair, (+)A<sub>IS</sub> and (-)A<sub>IS</sub> the peak areas for  
101 the corresponding internal standards, assuming that internal standards are racemic mixtures and that  
102 relative response factors are equal within each pair of enantiomers, as described in the literature  
103 (Marvin et al., 2007; Gao et al., 2011). Four standard calibration curves (25 injections) including 2  
104 independent preparations (assays B and C, different manufacturers and different operators in distinct  
105 laboratories) were used to determine racemic ranges as mean ± 3 standard deviations (Table S2).

106

### 107 **3. Results and discussion**

#### 108 *3.1. Consistency between assays for diastereoisomers*

109 Assay B confirmatory results of diastereoisomer concentrations were generally consistent with  
110 assay A original results. Indeed, mean absolute deviation was 21% and the linear regression curve slope  
111 was 0.91 (R<sup>2</sup> = 0.997), excluding four cases. The first case was an increase from 3.8 to 10.5 ng g<sup>-1</sup> ww  
112 of α-HBCDD in organic chicken eggs (EGC-02), making it the most contaminated sample, instead of the  
113 fresh eel sample as initially reported. This divergence could originate from heterogeneity of lyophilised  
114 sample or varying amounts of extracted fat which was not checked, according to the procedure.  
115 However, no influence on the EF was expected. The three other cases showed results below LOQs  
116 (β-HBCDD in fresh low fat milk LC-07 and chorizo MEC-03, and α-HBCDD in duck rillettes MEC-12  
117 samples), likely due to slightly higher LOQs. Detailed results are available in **Table S3** and **Figure S2**.

118 Results for assays B and C, obtained from the same sample extracts but with different instrumental  
119 set-ups, were generally consistent. Indeed, mean absolute deviation was 14% and the linear regression  
120 curve slope was 0.99 ( $R^2 = 0.999$ ). Moreover, while confirmatory assay B failed to provide values above  
121 LOQs for 3 diastereoisomer values quantified with original assay A, assay C provided 2 consistent  
122 values above LOQs. Thus, assays B and C did not confirm the presence of  $\beta$ -HBCDD in fresh low fat milk  
123 which was the only sample with  $\beta$ -HBCDD as only isomer. Detailed results are presented in **Table S3**  
124 and **Figure S3**.

125

### 126 3.2. Enantiomeric fractions

127 Enantiomeric fractions observed are provided in **Table S4** and displayed in Figure 1.

128  **$\alpha$ -HBCDD.** Diastereoisomer profiles of the 19 considered samples were all dominated by  $\alpha$ -isomer  
129 (70–100%), suggesting a preferential accumulation through the food chain rather than migration from  
130 food contact material. It appears that none of the samples could be considered as racemic for the  $\alpha$ -  
131 isomer (**Figure 1a**), suggesting that foodstuff contamination occurred prior to death of the  
132 corresponding producing animal and was not the result of the food item being in contact with technical  
133 HBCDD. No trend was observed between  $EF_\alpha$  and  $\alpha$ -HBCDD concentration, suggesting that enrichment  
134 processes do not depend on the contamination levels. Interestingly, the samples from terrestrial bird  
135 or mammal ( $n = 6$ ) were enriched in  $(-)\alpha$ -isomer while most samples from aquatic mollusc, crustacean  
136 or fish ( $n = 11$ ) were enriched in  $(+)\alpha$ -isomer; sole and anchovies were enriched in  $(-)\alpha$ -isomer.

137 Grey fresh shrimp exhibited the highest  $EF_\alpha$  value (0.606), which is consistent with enrichment in  
138  $(+)\alpha$ -HBCDD reported by [Zhu et al. \(2017\)](#) and [Zhang et al. \(2013\)](#) in mantis shrimp ( $EF_\alpha = 0.545$  and  
139 0.588, respectively).  $EF_\alpha$  values of the three herring-based foodstuffs (matjes, herring in "roll mops"  
140 process and prepared herring) were consistently within the 0.543-0.562 range. Although less  
141 pronounced, enrichment in  $(+)\alpha$ -HBCDD observed in eel (0.518) was also consistent with  $EF_\alpha$  reported  
142 by [Janák et al. \(2005\)](#) (0.54), [Bester and Vorkamp \(2013\)](#) in sand eel oil (0.554) and [Zhang et al. \(2013\)](#)  
143 in ricefield eel (0.516).

144 The general trend of most samples from aquatic mollusc, crustacean and fish toward relative  
145 enrichment in (+) $\alpha$ -HBCDD was consistent with some previous studies on distinct species: bib and  
146 whiting liver (Janák et al., 2005), zebrafish (Du et al., 2012), mirror carp (Zhang et al., 2014) and bartial  
147 flathead (Zhu et al., 2017). Opposite  $EF_{\alpha}$  tendencies of sole was also in line with Janák et al. (2005)  
148 findings for the same species and Köppen et al. (2010) results in mackerel, codfish, thorny skate,  
149 pollack and flounder. Actually, Ruan et al. (2018a) reported a small but significant decrease in the  $EF_{\alpha}$   
150 with the trophic level in a Hong Kong waters (5 mollusc species, 6 crustacean species, and 19 fish  
151 species) but the species-specific enrichments remained diverse. Intrinsic (species) and extrinsic (diet)  
152 factor contributions to the observed enrichments remain undetermined.

153 The lowest  $EF_{\alpha}$  value (0.404) was observed in organic chicken eggs, this pronounced enrichment in  
154 (+) $\alpha$ -enantiomer being consistent with findings of Omer et al. (2017), Jondreville et al. (2017) and  
155 Zheng et al. (2017) on chicken egg and various tissues. In the literature, reported  $EF_{\alpha}$  in eggs of other  
156 bird species were dominated by either the (-) $\alpha$ -enantiomer in peregrine falcon, common tern and  
157 herring gull or the (+) $\alpha$ -enantiomer in white-tailed sea eagle and guillemot (Janák et al., 2008; Esslinger  
158 et al., 2011).

159  **$\beta$ -HBCDD.** Among the 6 considered samples of animal source, 4 were of aquatic origin and 2 of  
160 terrestrial origin (**Figure 1b**). Only chicken eggs and eel seemed to be enriched in (+) or (-) $\beta$ -  
161 enantiomer, respectively, likely due to a higher uncertainty as regards the racemic range but also due  
162 to concentration levels relatively close to LOQs. Jondreville et al. (2017) previously reported  
163 enrichment in (+) $\beta$ -enantiomer in eggs of hens after ingestion of HBCDD-containing extruded  
164 polystyrene.

165 Interestingly, enrichment was more pronounced for (-) $\beta$ -isomer than (+) $\alpha$ -enantiomer in eel. At the  
166 opposite, enrichment was more pronounced for (-) $\alpha$ - than for (+) $\beta$ -enantiomer in eggs. For these two  
167 relatively highly contaminated samples,  $\beta$ -isomer contributed to less than 1% of the diastereoisomer  
168 contamination profile. Pronouncement of enantiomer enrichment also appeared to vary from one  
169 isomer to another.

170  $\gamma$ -HBCDD. The 4 considered samples were from animals of aquatic origin (**Figure 1c**). For oysters,  
171 trout and eel, enrichments were significant but also appeared to be possibly opposite depending on  
172 the animal species. Also, enriched enantiomer was independent from other diastereoisomers  
173 behaviours.

174 The most pronounced enrichment was observed for (-) $\gamma$ -enantiomer in fresh oysters ( $EF_{\gamma} = 0.293$ ).  
175 In eel, the slight enrichment observed for (+) $\gamma$ -enantiomer was consistent with findings of [Bester and](#)  
176 [Vorkamp \(2013\)](#). In scampi, the only sample with  $\gamma$ -HBCDD as only isomer according to original assay A,  
177 the diastereoisomer profile suggests food contamination by direct contact; but the significant  
178 enrichment in (-) $\gamma$ -isomer suggests the opposite. However, the  $EF_{\gamma}$  value was very close to the decision  
179 limit. Eventually, assays revealed the presence of  $\alpha$ - and  $\beta$ -isomers in a way that the diastereoisomeric  
180 profile was 42-4-54% for  $\alpha$ - $\beta$ - $\gamma$ -isomers, suggesting that, in that case, contamination arose from  
181 cumulated sources.

182

#### 183 **4. Concluding remarks**

184 The present study corroborates preferential accumulations of selected HBCDD enantiomers in food  
185 from animal origin intended for human consumption. Enantiomer enrichments suggest contamination  
186 of the human food chain through biological processes related to feedstock of animal origin rather than  
187 food contact contamination. Observed extent and opposite enrichments (depending on species) might  
188 lead to average dietary exposure to racemic mixtures, as observed by [Roosens et al. \(2009\)](#) in most  
189 duplicate diet samples from their sampling plan (except those containing fish, meat or cheese).  
190 However, in this particular study, most samples were dominated by  $\gamma$ -HBCDD, which was not  
191 consistent with [Poma et al.](#) findings. Thus, depending on their dietary habits, probabilistic total diet  
192 study considering consumption habits could reveal that Belgian sub-populations might be overall  
193 exposed to non-racemic mixtures. This could exacerbate (or mitigate) enantiomer enrichments  
194 occurring in the human body by enantioselective metabolism after consumption. Considering that  
195 enantiomeric pairs mostly display different toxicological activities ([Müller and Kohler, 2004](#); [Zhang et](#)

196 al., 2008), risk assessors should consider chiral profiles of HBCD diastereomers. Consequently, it is  
197 advised, on the one hand, to also include EF values in reporting the HBCDD occurrence levels, and on  
198 the other hand, to put focus of toxicological studies on investigating enantiomer-specific adverse  
199 effects.

200

#### 201 **Supporting Information**

202 The following file is available free of charge (PDF file): overview of the analysed samples, including  
203 quantification and EF results; extracted ion chromatogram illustrating achieved chromatographic  
204 separation; Comparisons between assay results.

205

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209

#### 210 **Conflict of interest**

211 The authors declare no financial competing interest.

212

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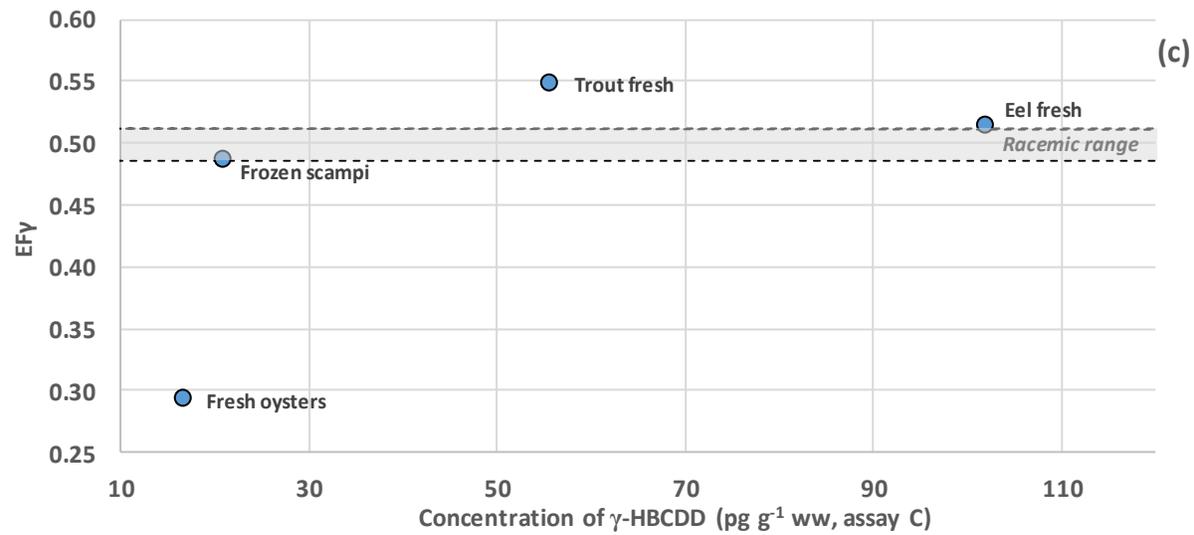
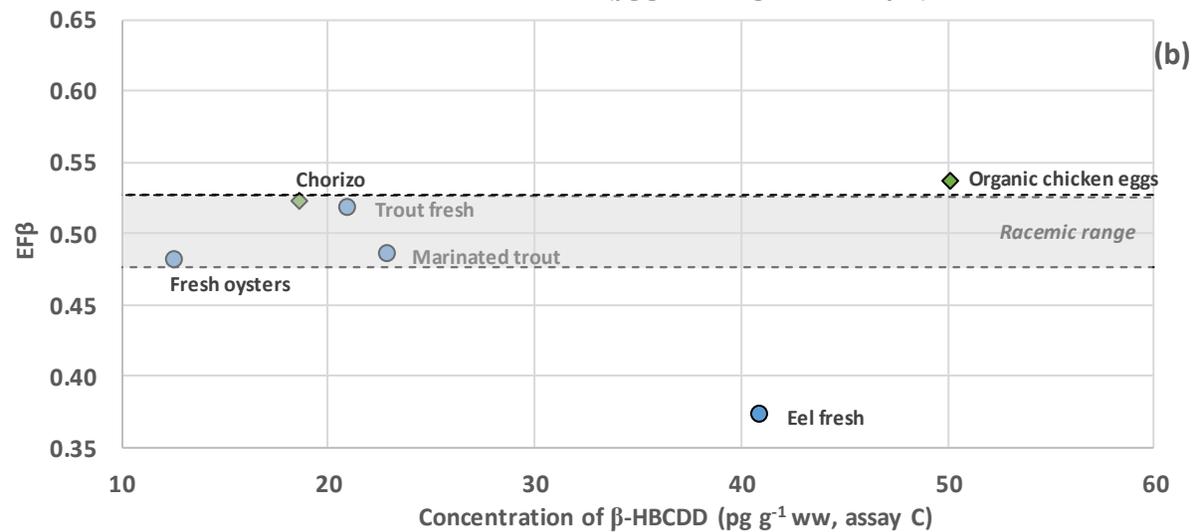
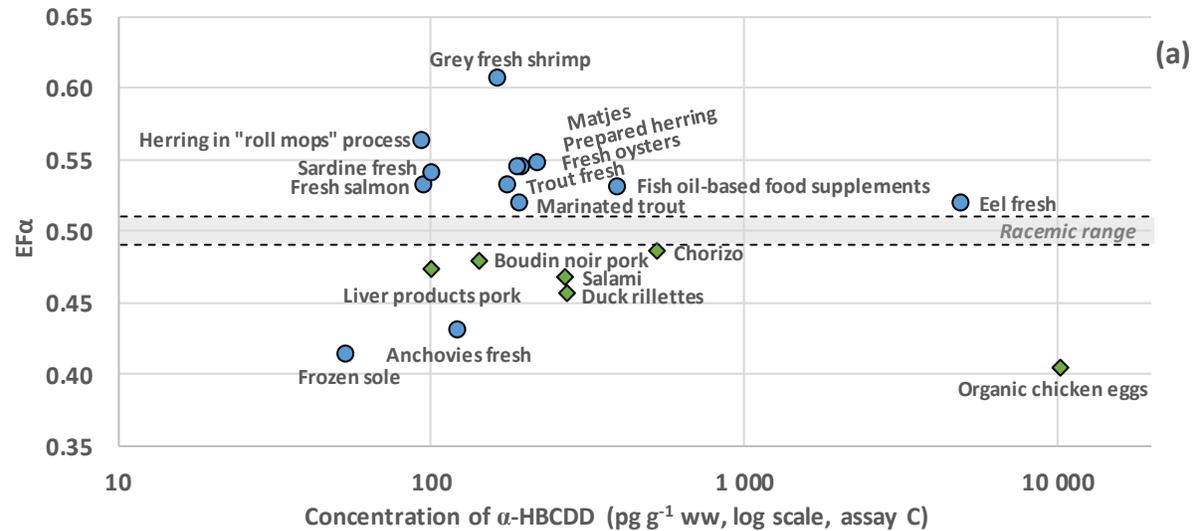
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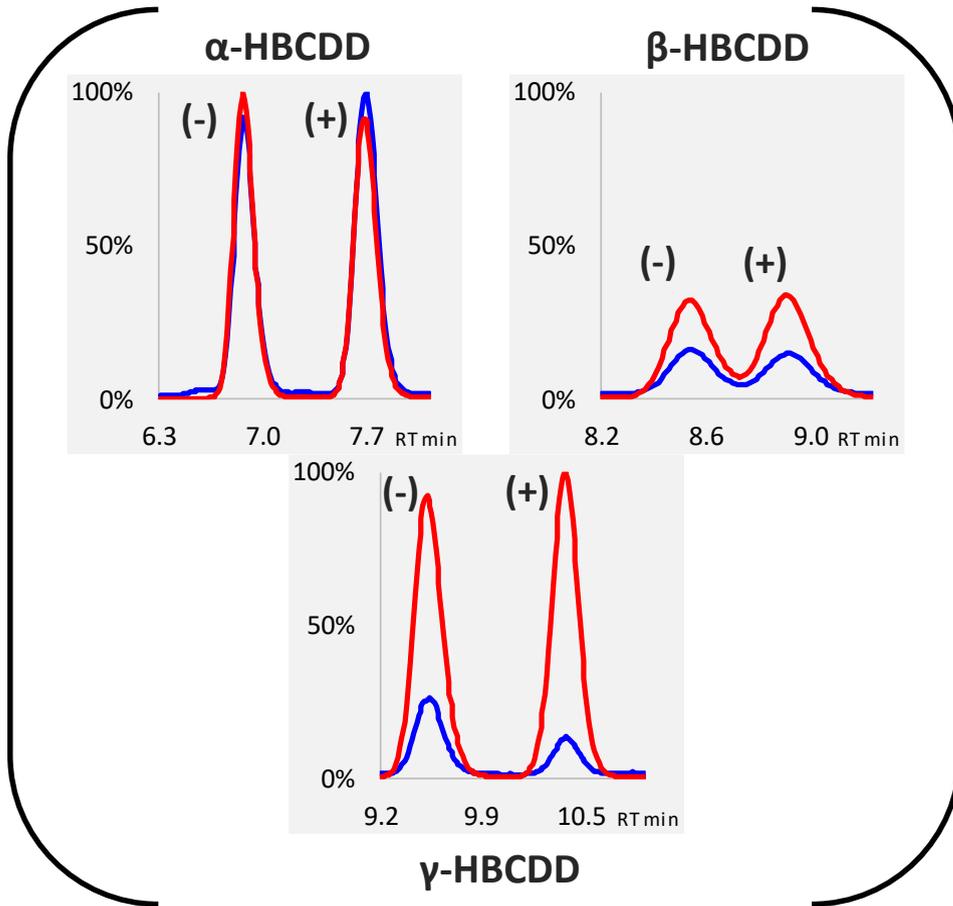
323 **Figure captions**

324 **Figure 1.** Enantiomeric fractions of  $\alpha$ - (a),  $\beta$ - (b) and  $\gamma$ - (c) HBCDDs according to diastereoisomer  
325 concentration (log scale for  $\alpha$ -) resulting from assay C. Sample from **aquatic** mollusc, crustacean or  
326 fish (●) or **terrestrial** bird or mammal (◆) food producing animal; racemic ranges delimited by  
327 dashed line.



# ENANTIOMERIC FRACTION

**FOOD**  
of animal origin



ante-mortem

**CONTAMINATION SOURCE**

post-mortem