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# 1 **Organic management and landscape heterogeneity sustain multiple bird**

## 2 **functions in European vineyards**

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23

24 **Running head** Multiple bird functions in vineyards

25

26 **Keywords** Bird communities – Bird cultural significance - Compositional heterogeneity –

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29

30 **Article impact statement:** In **≤140 characters** (including spaces and punctuation), provide a

31 statement that reveals the paper’s primary importance to conservation. See “Article Impact

32 Statement” below.

33 Wine-growing landscapes host a large diversity of bird functions enhanced by a combination of  
34 organic management and landscape heterogeneity

35 (140 characters)

36

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38

39

40

41

42 **Abstract**

43 Conserving functionally diverse bird communities in European farmland is becoming critical, with no  
44 exception for the regions of wine production. Management intensification combined with the loss of  
45 semi-natural habitats in wine-growing landscapes led to a long-term decline in birds of conservation  
46 concern, but also in once common insectivores and seed eaters. We investigated whether organic  
47 farming, inter-row grass management and landscape heterogeneity affected multiple bird functions  
48 in European vineyards. We analyzed taxonomic and functional diversity of 334 bird communities  
49 covering 12 vineyard regions of the three main wine-producing European countries (France, Italy and  
50 Spain). We found that organic management enhanced bird functional diversity but that its positive  
51 effect on bird functional groups depended on grass cover in vine inter-rows. For several bird  
52 functions, the positive effect of organic farming increased with landscape heterogeneity. Forest  
53 cover and landscape compositional heterogeneity increased both taxonomic and functional diversity  
54 of bird communities, especially functional insectivory. Landscape configurational heterogeneity also  
55 increased functional diversity and cultural significance, measured by song attractiveness of bird  
56 communities. However, mean bird specialization decreased with forest cover and configurational  
57 heterogeneity, meaning that open habitat specialists preferred open landscapes with large vineyard  
58 stands. Overall, both bird diversity and functions were enhanced by higher landscape heterogeneity  
59 and longer edges between vineyards and semi-natural habitats. Our study highlights the benefits of  
60 combining organic management and partial grass cover at the field level and maintenance of  
61 interfaces with semi-natural habitats at the landscape level to conserve multifunctional bird  
62 communities across European vineyards.

63 (247 words)

## 64 **Introduction**

65           Intensification of agricultural practices during the last decades has caused deep changes in  
66 bird communities worldwide, threatening endangered species as well as the provision of pest control  
67 or seed dispersal by major avian guilds (Bowler et al. 2019; Hendershot et al. 2020). The interplay  
68 between farming practices and landscape structure in farmland can mitigate, improve or dampen the  
69 synergies between multiple ecosystem functions and services provided by biodiversity (Martin et al.  
70 2019), including those provided by birds (Pejchar et al. 2018). Viticulture is of major economic  
71 importance in Europe and currently faces important environmental issues that lead wine growers to  
72 shift towards more environmentally friendly management, including agroecological practices such as  
73 organic management or use of permanent grass cover (Merot et al. 2019; Paiola et al. 2020). As  
74 vineyards are likely to expand more and more at the expense of semi-natural vegetation with climate  
75 warming (Hannah et al. 2013), it is critical to better understand the combined effects of field  
76 management and landscape structure on bird diversity across a large range of wine production areas.

77           In Europe, vineyards are managed with various intensification levels, but generally result in  
78 heterogeneous mosaics of semi-natural habitats interspersed with large areas of grape dedicated to  
79 wine production. However, European vineyards recently experienced a loss of landscape complexity  
80 under the combined effects of land use and climate changes, together with the use of agrochemicals  
81 and changes in soil management (Paiola et al. 2020). Landscape heterogeneity is a key driver of  
82 biodiversity dynamics in agricultural landscapes (Fahrig et al. 2011), and is of critical importance for  
83 the provision of multiple ecological functions and services, such as natural pest control delivered by  
84 multiple organisms (Winqvist et al. 2011). In vineyards, landscape diversity allows maintaining  
85 ecological functions provided by vertebrate insectivores such as bats and birds (Jedlicka et al. 2011;  
86 Assandri et al. 2016; Froidevaux et al. 2017; Rodriguez-San Pedro et al. 2019). Surprisingly, the  
87 conservation of vineyard bird communities has received little attention in Europe, mainly because  
88 vineyards are often considered as species-poor agroecosystems (Brambilla & Ronchi 2020). However,

89 vineyards have historically supported, and can still potentially host typical bird assemblages including  
90 threatened specialists such as lesser grey shrike *Lanius minor* or ortolan bunting *Emberiza hortulana*,  
91 now extirpated from most of these formerly used habitats (Isenmann & Debout 2000; Siervo &  
92 Arlettaz 2003; Brambilla et al. 2017a). Other species of conservation concern, such as great bustard  
93 *Otis tarda*, little bustard *Tetrax tetrax* or stone curlew *Burhinus oedicanus*, have also disappeared  
94 from most vineyard landscapes following management intensification and may not persist without  
95 sufficient scrub/grassland patches in the landscape (Pithon et al. 2016; Casas et al. 2020). More  
96 generally, the long-term decline of insectivorous birds in vineyards, as in other farmland, is directly  
97 related to the loss of semi-natural grasslands and crop intensification (Bowler et al. 2019, Hendershot  
98 et al. 2020). Birds provide multiple functions to agriculture and society, including biological control of  
99 insects and weeds, as well as cultural values (Cumming & Maciejewski 2017; Pejchar et al. 2018;  
100 Brambilla & Ronchi 2020). As a result, conserving functionally diverse bird assemblages in vineyards  
101 is becoming more and more critical, given the social and economic importance of this permanent  
102 crop production under Mediterranean-type climates (Muñoz-Sáez et al. 2020a; Paiola et al. 2020).

103         Such bird assemblages are highly dependent on multi-level habitat heterogeneity, both  
104 through sward management at the stand level creating various conditions of grass and bare ground  
105 cover (Duarte et al. 2014; Bosco et al. 2019) and habitat diversity at the landscape level (Assandri et  
106 al. 2016; Muñoz-Sáez et al. 2020b). At the stand level, organic farming or extensive grass cover  
107 management should benefit bird communities through higher availability of food resources (higher  
108 abundance and diversity of weeds and arthropods) or nesting sites (Fuller et al. 2005; Winqvist et al.  
109 2011). At the landscape scale, higher habitat heterogeneity should benefit bird communities through  
110 complementation or supplementation processes between key resources in the landscape (Dunning  
111 et al. 1992). The interplay between compositional and configurational landscape heterogeneity  
112 should modulate the effects of management practices at the stand level on bird communities, as  
113 found for other taxa or ecological functions (Tschardt et al. 2012; Martin et al. 2019). Following  
114 such hypothesis, the local effect of potentially beneficial management for birds, such as organic

115 farming or extensive grass cover management, should be maximal in more intensive landscapes  
116 compared to heterogeneous mosaics where bird communities are already diverse due to higher  
117 resource availability and habitat diversity (Tuck et al. 2014). The effects of organic farming for  
118 vineyard biodiversity is still seldom studied although organic vineyards are rapidly expanding in some  
119 countries such as France, Italy or Spain (Assandri et al. 2016; Muneret et al. 2019; Rollan et al. 2019).  
120 So far, most studies did not demonstrate a direct and consistent positive effect of organic farming on  
121 insectivorous birds or bats. Instead, several studies did report interacting effects of organic farming  
122 with grass cover, landscape composition and availability of arthropod prey (Froidevaux et al. 2017;  
123 Winter et al. 2018; Rodriguez-San Pedro et al. 2019). As studies on organic farming have shown a  
124 wide diversity of effects, we aimed here at exploring its interactions with other practices across a  
125 large range of wine-producing regions and landscapes, for a better understanding of the contribution  
126 of different management actions to vineyard bird diversity.

127         We hypothesized that the link between bird communities and vineyard habitats depends on  
128 the interaction between field-level management, including organic farming, and type, amount and  
129 spatial configuration of surrounding semi-natural habitats. We built a multi-regional dataset on  
130 vineyard bird communities from different wine-producing regions of southern Europe to test how  
131 the different components of bird communities display consistent and complementary responses to  
132 vineyard management and landscape heterogeneity across a wide biogeographic range. We  
133 computed a set of multiple community metrics to assess the effects of viticulture on birds as well as  
134 the effects of birds on viticulture, by integrating diversity metrics and functional measures, including  
135 trait diversity. We predicted that: (i) the effect of organic farming on bird communities would be  
136 stronger on bird functional than taxonomic diversity because of wider niche opportunities offered by  
137 organic practices, and would interact with both inter-row management and landscape heterogeneity;  
138 (ii) the effects of landscape composition and diversity would be more important than vineyard  
139 management for bird species diversity and the abundance of functional insectivores, by filtering the  
140 regional species pool able to use vineyards at the local scale; and (iii) the effects of landscape

141 configurational heterogeneity would be especially detectable on bird functional diversity and  
142 individual avian functions because it would primarily affect bird foraging opportunities by increasing  
143 complementation between vineyards and the wider landscape.

144

## 145 **Methods**

### 146 *Study areas*

147 We studied 12 wine-growing regions located in three countries of southern Europe that are  
148 the three main producers of wine worldwide: France (nine regions), Italy (two regions) and Spain  
149 (one region; Fig. 1). The proportion of organic vs conventional management and the extent of  
150 landscape complexity gradient covered varied among regions (Appendix S1). In each region, sites  
151 were selected along a landscape composition gradient based on the proportion of woodlands and  
152 semi-natural grasslands in the surrounding landscape. For each stand, we determined the type of  
153 management (organic or conventional) by local inquiries combined with information gathered by  
154 dedicated professional structures (e.g., DOQP bureau in Spain; see Puig-Montserrat et al. 2017). The  
155 proportion of organic stands was 31% in the overall dataset (N = 103 vs N = 231 for conventional  
156 stands; see Appendix S1). We also measured the proportion of grass cover in vine inter-rows, ranging  
157 between 100% (homogeneous grass cover within the entire stand), 50% (partial grass cover due to  
158 soil tillage in half of inter-rows) and 0% (bare ground over all the inter-rows). The mean grass cover  
159 over the 334 plots was 52% (see Appendix S1), and the distribution of continuous values of grass  
160 cover was well balanced between 0 and 100%.

161

### 162 *Bird sampling*

163 Bird communities were sampled in different years according to the study region, from 2010  
164 to 2018, but 85% (i.e., 285 among the 334 plots) were sampled either in 2013 or 2015, with no bias



165 towards organic or conventional vineyards in a particular year (Appendix S1). Birds were surveyed  
166 using point or transect counts by single trained observers per region. All birds heard and seen were  
167 recorded (except flyovers), within a distance of 50 to 100 m from the observer on each transect side,  
168 depending on the region (Appendix S1). Point counts were performed using a standard distance  
169 detection of 100 m. To account for differences in the area sampled between circular points and  
170 transect counts, we included the sampled area as a random model predictor. We assumed that  
171 variation in species detectability was limited among sampled vineyards due to the highly similar and  
172 homogeneous structure of vine rows. Bird counts were performed early in the morning (6.00 to  
173 10.00 am) only during days without heavy rain or wind. Bird counts were conducted twice, the first  
174 visit between mid-April (early-season breeders) and mid-May and the second visit between late May  
175 and mid-June (late-season breeders), except in Italy where a third visit was conducted between these  
176 two visits. For each species, the highest count among the two or three visits was further used as a  
177 standardized estimate of abundance (Appendix S3).

178

#### 179 *Bird functions and community metrics*

180 We computed nine community-level metrics, including taxonomic diversity, functional  
181 diversity and abundance-based avian functions. To account for both species abundance and richness  
182 of bird communities, we used the Shannon index of taxonomic diversity. To characterize individual  
183 avian functions within bird communities and analyse responses of bird functional composition  
184 beyond usual species diversity, we computed the cumulative abundance for several species groups  
185 that potentially benefit viticulture (pest control: functional insectivores FI; weed control: seed eaters  
186 SE) or may be considered as vine pests (grape eaters GE). We calculated an index of functional  
187 insectivory by cumulating the abundance of species sharing a similar combination of diet, foraging  
188 technique and habitat use (Barbaro et al. 2017). A bird species was considered a 'functional  
189 insectivore' in vineyards when at the same time: (i) it is insectivorous during the breeding period; (ii)

190 it predominantly forages by foliage gleaning or by hawking; and (iii) it uses vineyards as breeding  
191 and/or foraging habitats (N = 34 species). The abundance of seed eaters SE (N = 17 species) and  
192 grape eaters GE (N = 9 species) were calculated similarly (Appendix S3). Seed-eating birds were  
193 determined based on their diet preferences during the breeding season using authors' personal trait  
194 database (Jeliazkov et al. 2020), while potential grape consumers were established with literature  
195 inquiries and completed by expert knowledge based on personal field observations of the authors.

196 To go beyond individual avian functions, we computed three trait-based functional metrics  
197 expected to support the largest diversity of species functions, i.e. functional divergence FDiv,  
198 functional evenness FEve and functional entropy Rao's Q (Mouillot et al. 2013). We used a species-  
199 trait matrix of eight life-history traits, including six categorical traits (foraging method, adult diet,  
200 nesting site, migration strategy, mean laying date and mean home range size) and two continuous  
201 traits (clutch size and body mass; see Jeliazkov et al. 2020 and Appendix S3). Functional divergence  
202 (FDiv) measures trait abundance distribution within this volume and increases with extreme trait  
203 values, functional evenness (FEve) increases with the regularity of trait abundance distribution within  
204 the functional space, while Rao's Q measures functional entropy by characterizing species dispersion  
205 from the functional space centroid, i.e. indicates a community composed of species functionally  
206 different from the mean trait composition (Mouillot et al. 2013).

207 Following Blackburn et al. (2014) or Goodness et al. (2016), we considered mean bird song  
208 attractiveness to humans as the most effective proxy for bird cultural services, by calculating the  
209 number of individual species recording uploaded in XenoCanto.org online database, weighted by  
210 geographic range size (XCRw). Bird conservation concern was expressed by mean bird habitat  
211 specialization (Community Generalization Index, CGI), which can be considered as one among the  
212 main forms of ecological rarity (Godet et al. 2015; Sykes et al. 2020). The Community Generalization  
213 Index is the community-weighted mean value of all Species Generalization Indices (SGI) within a  
214 given community (Gaüzère et al. 2020). The Species Generalization Index (SGI) is a measure of

215 habitat niche width of a given species, or ecological rarity, and is computed as the coefficient of  
216 variation of the species density across 18 habitat classes at national scale and corresponds to the  
217 inverse value of the Species Specialization Index (Godet et al. 2015).

218

#### 219 *Computation of landscape variables*

220 Land cover maps were realized with ArcGIS 10.6 (ESRI, Redlands, CA, USA) for all regions  
221 using the following standard nomenclature: forests and hedgerows, grasslands, shrublands, crops,  
222 vineyards, orchards, roads, urban areas and bare ground, waters. For France, land cover maps were  
223 derived from two sources, the BD Topo version 2 of 2018 (Institut Géographique National IGN) and  
224 the Cesbio OSO2018 online database in vector format downloaded from [http://osr-cesbio.ups-](http://osr-cesbio.ups-tlse.fr/~oso/)  
225 [tlse.fr/~oso/](http://osr-cesbio.ups-tlse.fr/~oso/) (Inglada et al. 2018). For Italy and Spain, land cover maps were obtained from photo-  
226 interpretation of aerial photographs at 1:2000 combined with pre-established land cover maps using  
227 the same nomenclature as above; taken from [www.geoportale.regione.lombardia.it](http://www.geoportale.regione.lombardia.it) for Lombardy  
228 (Brambilla et al. 2017b); from Assandri et al. (2016) for Trentino; and from a DMAH land cover map  
229 of 2005 for Catalonia (Puig-Montserrat et al. 2017). All GIS maps obtained were visually corrected by  
230 one trained geomatician for each study area, especially when vineyards were partially overlapping  
231 with other land cover types such as wooded elements. For each land cover type, we computed the  
232 total and percent cover in circular 100m-radius buffers around all 334 stands and the Shannon  
233 diversity index of land cover types (hereafter referred to as 'landscape diversity') as a measure of  
234 landscape compositional heterogeneity. We further calculated the total length of interfaces between  
235 all vineyard patches and all types of semi-natural habitats (grasslands, shrublands and forests), as a  
236 measure of landscape configurational heterogeneity (Fahrig et al. 2011).

237

238

239 *Data analysis*

240 We performed a set of generalized linear mixed models of Gaussian, Poisson and quasi-  
241 Poisson families (LMMs and GLMMs) to test the relative effects of vineyard management (organic vs  
242 conventional), grass cover and landscape structure on nine bird community metrics and individual  
243 avian functions computed for 334 bird communities: species diversity SDiv, Community  
244 Generalization Index CGI, mean song attractiveness XCRw, functional insectivore abundance FI, seed  
245 eater abundance SE, grape eater abundance GE, functional divergence FDiv, functional evenness  
246 FEve and functional entropy RaoQ (Fig. 2). LMMs and GLMMs were built in R software v3.6.0 using  
247 the 'glmmTMB' package (Brooks et al. 2020). Three community metrics were count data (sum of  
248 species abundance for distinct foraging guilds), and therefore we modelled their responses using  
249 Poisson distribution (FI and GE) or quasi-Poisson (SE) distribution to handle over-dispersion. All other  
250 metrics (taxonomic and functional diversity, CGI and XCRw) were modelled using the Gaussian  
251 distribution after checking for normality and heteroscedasticity of residuals using Shapiro-Wilk tests.

252 We used the same full model structure for all response variables, including the following  
253 fixed effects: the interaction between vineyard management (organic vs conventional) and inter-row  
254 grass cover, the interactions between organic management and landscape compositional (Shannon  
255 diversity of land-use types) and configurational heterogeneity (length of vineyard – semi-natural  
256 habitat edges), and two variables of landscape composition (% of semi-natural open habitats and  
257 woodlands). The region of wine production (N = 12) was considered as a random effect to account  
258 for spatial gathering of sampled stands, biogeographical differences, and for the combination of year  
259 and observer effects. We also included the area sampled as a second, additive random effect to  
260 account for differences in sampling protocols among regions. Because we expected an effect of  
261 sampling protocol on abundance-based metrics (SDiv, FI, GE and SE), but not on integrative  
262 community indices (CGI, FDiv, FEve, Rao's Q and XCRw), we also tested the area sampled as an offset

263 in mixed models structure for bird guilds, and found no differences in predictor performance of  
264 model AICs (Brooks et al. 2020).

265 All continuous variables were standardized (i.e., rescaled to the same unit) to enable  
266 comparisons of effect magnitude. We evaluated multicollinearity among predictors with both the  
267 variance inflation factor (VIF) and the Spearman's correlation test; no strong correlation was found  
268 (VIF values < 3;  $|r| < 0.6$ ). Model validation was conducted using the 'DHARMA' package (Hartig 2020).  
269 We performed the Shapiro Wilk test on LMMs' residuals to ensure that normality assumption was  
270 met. Based on the full models, we generated a set of candidate models containing all possible  
271 variable combinations using 'MuMIn' package (Bartoń 2020). We applied an information theoretic  
272 approach to assess model parsimony and models were ranked based on their Akaike Information  
273 Criterion (AIC). To account for model selection uncertainties, we performed a model-averaged  
274 procedure of most parsimonious models (i.e. those with  $\Delta AIC < 2$ ), and further report the conditional  
275 model average estimates. We checked model residuals for spatial autocorrelation using bubble plots  
276 and variograms and drawn prediction biplots based on the best and most parsimonious models.

277

## 278 **Results**

279 Bird sampling of 334 vineyards across 12 wine-growing regions from three countries gave a  
280 total count of 11,472 individuals belonging to 131 species. Among the taxa of high conservation  
281 concern, we recorded the presence of *Tetrax tetrax*, *Burhinus oedicephalus*, *Galerida theklae*,  
282 *Calandrella brachydactyla*, *Oenanthe hispanica*, *Sylvia subalpina* and *Emberiza hortulana* (Appendix  
283 S3). The abundance of functional insectivores and the functional diversity of birds (both functional  
284 divergence and entropy) significantly increased with organic management (Fig. 2). For functional  
285 insectivores as well as grape eaters, the positive effect of organic management was contingent upon  
286 inter-row grass cover, with reverse patterns (Fig. 3). By contrast, the abundance of seed eaters  
287 decreased with grass cover (Fig. 2).

288 Forest cover significantly increased the abundance of functional insectivores, habitat  
289 generalists (i.e., it decreased mean bird specialization) and mean song attractiveness (Fig. 4). On the  
290 other hand, forest cover negatively affected the abundance of seed and grape eaters (Fig. 2 and 4).  
291 Landscape compositional heterogeneity (i.e. landscape diversity) had a positive effect on taxonomic  
292 diversity and the abundance of the three bird guilds (Fig. 2 and 3). Landscape configurational  
293 heterogeneity (i.e. edge length between vineyards and semi-natural habitats) significantly increased  
294 functional divergence and evenness, abundance of functional insectivores and habitat generalists,  
295 and mean song attractiveness (Fig. 5). Finally, we found significant interactions between organic  
296 management and landscape diversity for mean bird specialization, functional divergence and  
297 functional evenness, and between organic management and edge length for seed and grape eaters  
298 (Fig. 2).

299

## 300 **Discussion**

301 In the present work, we assessed simultaneously the effects of organic farming, grass cover  
302 management and landscape heterogeneity on the conservation of multiple bird functions in  
303 European vineyards. To account for multifaceted responses of bird communities to the vineyard  
304 agroecosystem, we assessed both taxonomic and functional diversity and a range of avian functions  
305 relevant to viticulture. We found that landscape composition and field-level management jointly  
306 contribute to shape vineyard bird communities. Landscape heterogeneity (both compositional and  
307 configurational) was also important for bird communities, and benefited most bird functional groups,  
308 taxonomic diversity and cultural significance. Organic management enhanced both functional  
309 diversity and the abundance of insectivorous birds without affecting species diversity *per se*.  
310 Moreover, the influence of organic farming interacted with inter-row grass cover for functional  
311 insectivores, potential grape consumers and seed eaters. Organic management also interacted with

312 landscape heterogeneity to increase bird functional diversity and the abundance of habitat  
313 generalists, and to decrease the abundance of seed and grape eaters.

314           It is now widely recognised that organic farming at the field level is not always sufficient to  
315 increase biodiversity in farmland, and that its effect depends on the taxa, the spatial scale, the crop  
316 type, and the landscape context (Fuller et al. 2005 ; Gabriel et al. 2010). The positive effect of organic  
317 management on species richness and abundance of most taxa is particularly noticeable in more  
318 homogeneous agricultural landscapes (Tuck et al. 2014). Birds are known to benefit from organic  
319 farming, both from sward management and release from pesticides that increases prey availability at  
320 both the field and landscape scale (Fuller et al. 2005; Rollan et al. 2019). However, several studies did  
321 not detect significant effects of organic vineyard management on birds or other insectivorous  
322 vertebrates such as bats (Assandri et al. 2016; Froidevaux et al. 2017; Puig-Montserrat et al. 2017). In  
323 vineyards as much as in other farmland landscapes, it is actually expected that organic management  
324 would interact with both fine and larger-scales habitat attributes (Gabriel et al. 2010), which may  
325 explain the variety of local responses observed in particular studies. Here, by gathering bird data  
326 obtained from a broad geographical scale in Europe, we show that the functional diversity of bird  
327 communities, and the abundance of target functional guilds, are enhanced in vineyard stands  
328 conducted under organic management. Moreover, the effect of organic viticulture was contingent  
329 upon both field-level management (i.e., grass cover in vine inter-rows) and the diversity and  
330 configuration of semi-natural cover in the landscape (Assandri et al. 2016; Rollan et al. 2019). Our  
331 results thus suggest that it is critical to maintain native vegetation within vineyard stands, as well as  
332 larger amounts of semi-natural cover – both open and wooded – within the landscape, to integrate  
333 production and conservation efforts in sustainable viticulture (Muñoz-Sáez et al. 2020b).

334           Management options mixing organic farming at the stand level and maintenance of semi-  
335 natural cover in the landscape is not only profitable to birds of conservation concern (Sierro &  
336 Arlettaz 2003; Brambilla et al. 2017a; Casas et al. 2020), but also to the diversity of bird functional

337 groups (Assandri et al. 2016; Barbaro et al. 2017). The effect of multi-level heterogeneity results into  
338 a complex interaction between field- and landscape-level management of vegetation cover likely to  
339 optimize ecosystem functions fulfilled by insectivorous vertebrates in vineyards (Froidevaux et al.  
340 2017). In particular, the interaction between organic management and grass cover showed that  
341 these two key viticultural practices are intrinsically related, and that applying a partial grass cover in  
342 vine ranks in organic stands will benefit endangered birds as well as functional diversity (Arlettaz et al  
343 2012; Rollan et al. 2019). However, we found an unexpected negative effect of grass cover on seed  
344 eaters, likely because many grass covers are often too intensively managed to have enough seeds  
345 available for specialists such as buntings or finches. Alternatively, uniform grass cover over the entire  
346 vineyards might decrease seed detectability for birds, which is likely to be higher in heterogeneous  
347 contexts, i.e. with patches of bare ground or low-density vegetation. Manipulating grass cover in vine  
348 rows also allows managing the abundance of potential grape-eating birds such as starlings or turdids,  
349 which, as ground probing foragers, are more favoured by a full than a partial grass cover. Like in  
350 other crop types, the conservation of biodiversity in vineyards has positive functional consequences  
351 for wine production by providing regulating services of natural pest control (Winqvist et al. 2011;  
352 Muneret et al. 2019). Such services might be considered as a biotic insurance against an expected  
353 increase in pest insect damage to vineyards with global change, through the diversity of bird  
354 functions and functional insectivory (Barbaro et al. 2017; Pejchar et al. 2018). How this effect  
355 cascades on other ecosystem services remains to be fully investigated, in particular those related to  
356 human well-being. Our study showed that this could be achieved by using new, exploratory indices of  
357 bird cultural significance to humans, including wine producers themselves, such as song or visual  
358 attractiveness (Blackburn et al. 2014; Goodness et al. 2016; Brambilla & Ronchi 2020).

359         The amount of semi-natural cover in the surrounding landscape is a key factor for bird  
360 communities in vineyards (Assandri et al. 2016; Pithon et al. 2016; Muñoz-Sáez et al. 2020a).  
361 Woodland cover increased the abundance of functional, often generalist insectivores with high song  
362 attractiveness to humans, but decreased the abundance of seed and grape eaters. In contrast,



363 species of conservation concern, such as woodlark *Lullula arborea* and ortolan bunting *Emberiza*  
364 *hortulana*, two among the most characteristic species of vineyard landscapes, strongly benefitted  
365 from a combination of stand and landscape-level heterogeneity (Arlettaz et al. 2012; Brambilla et al.  
366 2017a; Bosco et al. 2019). Interestingly, we found that both landscape configurational and  
367 compositional heterogeneity were important for the conservation of functionally diverse bird  
368 communities in vineyards, as predicted by ecological theory (Fahrig et al. 2011). Such positive  
369 responses to fragmentation are due to higher spatial complexity in mosaic landscapes, enhancing  
370 positive edge effects on insectivorous birds and their functional diversity, and allowing more  
371 complementation processes and spill-over movements between vineyards and adjacent semi-natural  
372 habitats (Barbaro et al. 2017; Muñoz-Sáez et al. 2020a). Overall, landscape heterogeneity had a  
373 positive effect on taxonomic diversity and allowed the coexistence of multiple avian functions in  
374 vineyard landscapes. Furthermore, it is likely that such heterogeneity would also benefit other  
375 functionally significant taxa, allowing pest regulation while contributing importantly to vineyard  
376 biodiversity (Caprio et al. 2015; Rodriguez-San Pedro et al. 2019). Managing the wider landscape  
377 matrix to conserve bird functions is therefore a valuable option for wine growers, as well as for  
378 conservationists and human societies inhabiting vineyard landscapes with high cultural significance  
379 (Assandri et al. 2018; Muñoz-Sáez et al. 2020b). Previous studies have highlighted that the local  
380 potential for biocontrol in vineyards is driven by the diversity of natural enemies and trait  
381 complementarities among predators of wine pests, and that this potential is narrowly linked to  
382 landscape complexity (Muneret et al. 2019; Rodriguez-San Pedro et al. 2019). Other services  
383 provided by biodiversity, such as pollination, are also favoured by the same type of landscape  
384 management (Kratschmer et al. 2019), although trade-offs may also occur between services  
385 (Brambilla et al. 2017b). Conserving a significant proportion of semi-natural cover in the landscape  
386 matrix is considered as necessary for biodiversity to provide these services, and to allow maintaining  
387 functional complementarity and redundancy across regions as a spatial insurance against global  
388 change (Tscharntke et al. 2012). In addition, there is a need for maintaining a diversity of wine-

389 growing techniques in the wider landscape to mitigate negative effects of climate and land use  
390 changes on vineyard biodiversity, together with the development of agroecological practices  
391 (Hannah et al. 2013; Assandri et al. 2018).

392 As a conclusion, our study advocates for encouraging a mixture of traditional and innovative  
393 practices in vineyard management, and to consider applying these options at multiple levels from the  
394 within-stand heterogeneity to the wider landscape. This offers promising strategies for wine growers  
395 to adapt their vineyards to global change, while avoiding further management intensification as  
396 observed in the last decades, that appeared to be detrimental to both biodiversity and wine  
397 production diversity (Merot et al. 2019; Morales-Castilla et al 2020). In other words, vineland, like  
398 other farmland, needs biodiversity to cope with global change and contribute to the critical aim of  
399 conserving vineyard-associated biodiversity in countries with Mediterranean-type climates (Muñoz-  
400 Sáez et al. 2020a; Paiola et al. 2020). A further step would be now to assess how the spatial  
401 expansion of agroecological farming practices in interaction with semi-natural habitats is affecting  
402 bundles of ecosystem functions and services in European vineyards, including cultural significance or  
403 aesthetic values.

404

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533 grant agreement No 311879. This is a contribution to the project LIFE+2009 BioDiVine.

534

535 **Supporting information**

536 **Appendix S1 to S5**

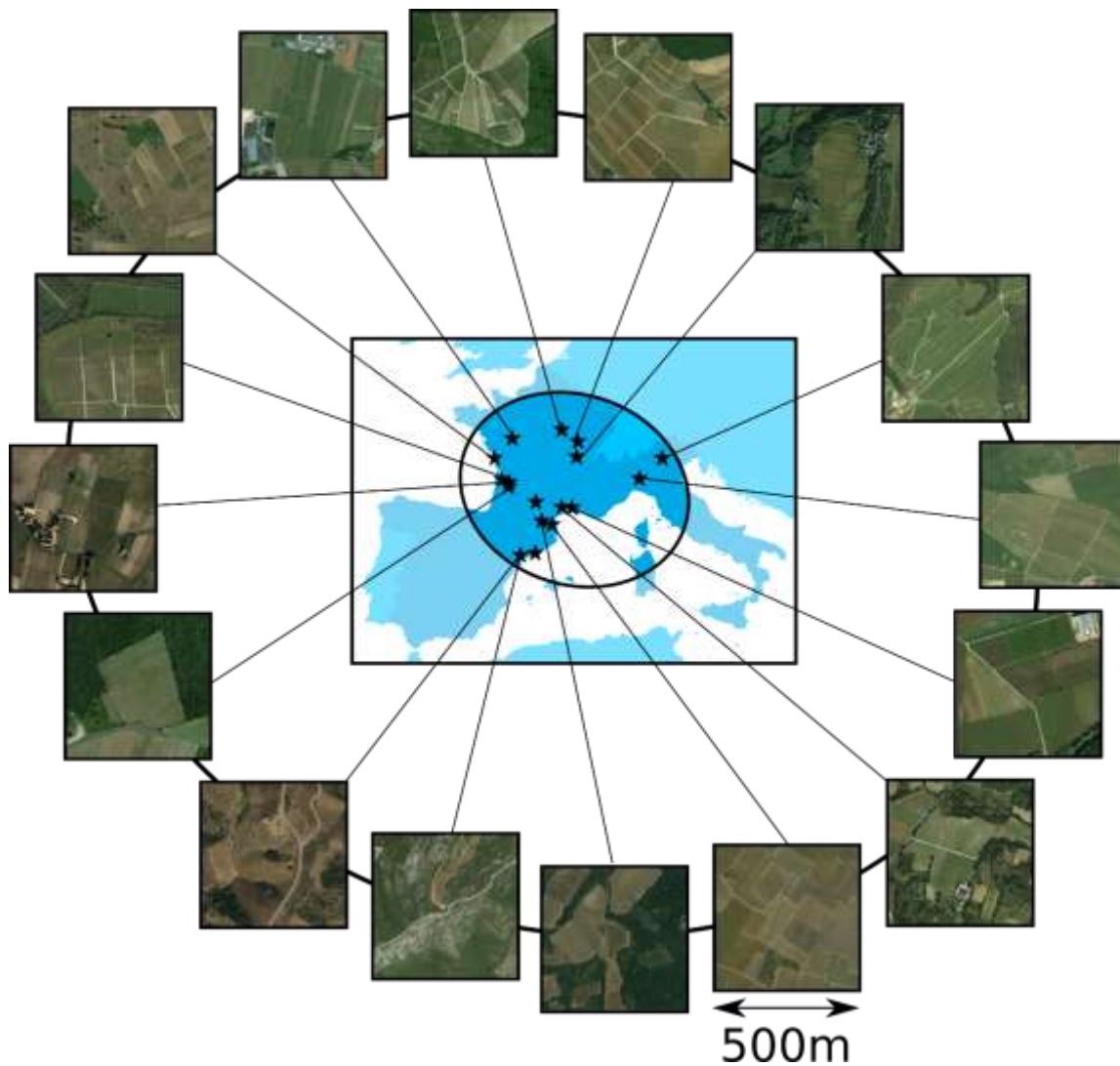
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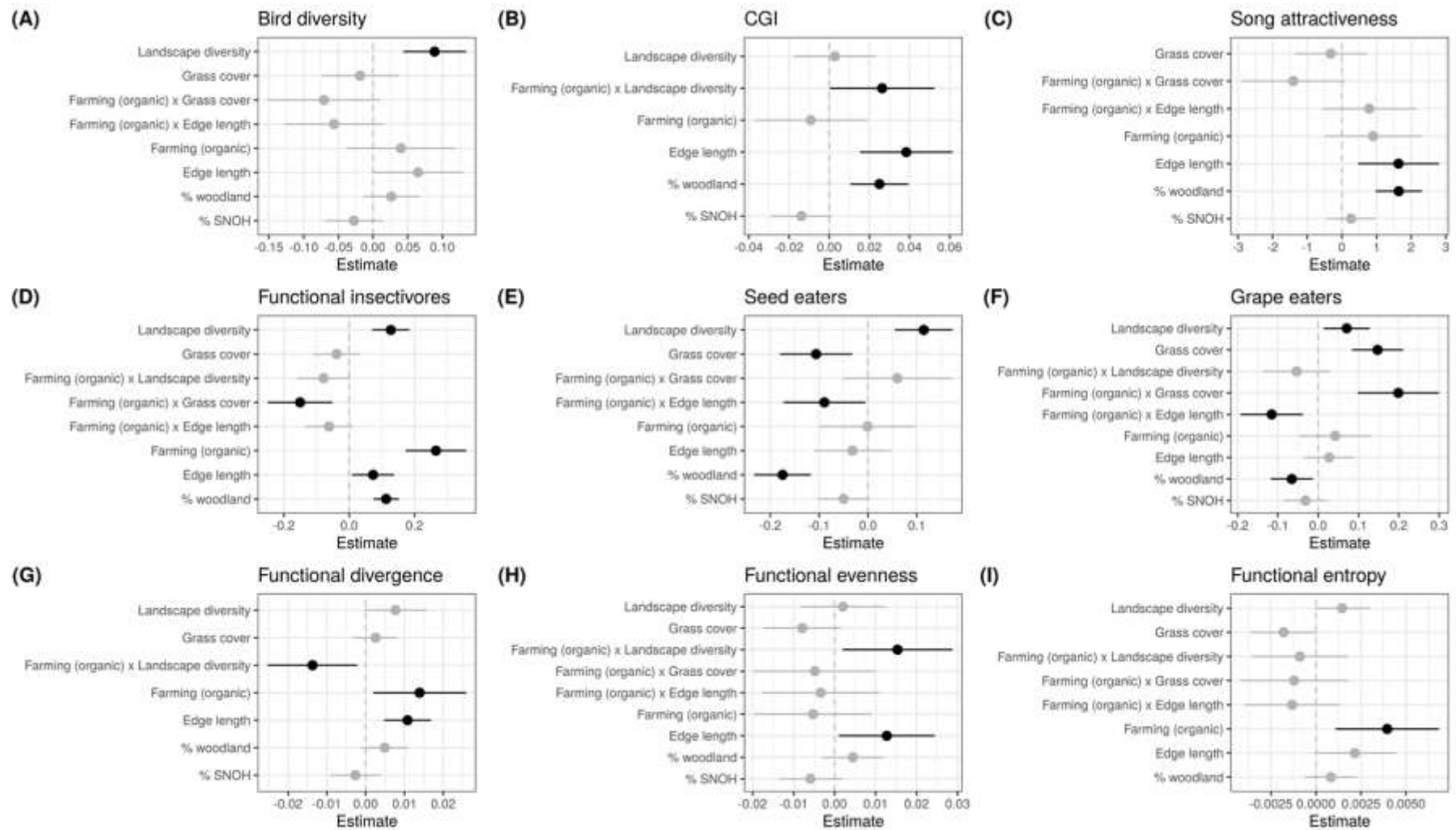


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545 Figure 1. Location map of sampled wine production regions. Clockwise from top: Bourgogne (3  
546 subregions), Trentino, Lombardy (Oltrepo), Costières de Nîmes, Terrasses du Larzac, Corbières,  
547 Limoux, Gaillac, Catalunya (2 subregions), Bordeaux (3 subregions: Entre-Deux-Mers, Saint Emilion,  
548 Medoc), Ile de Ré, Saumur.

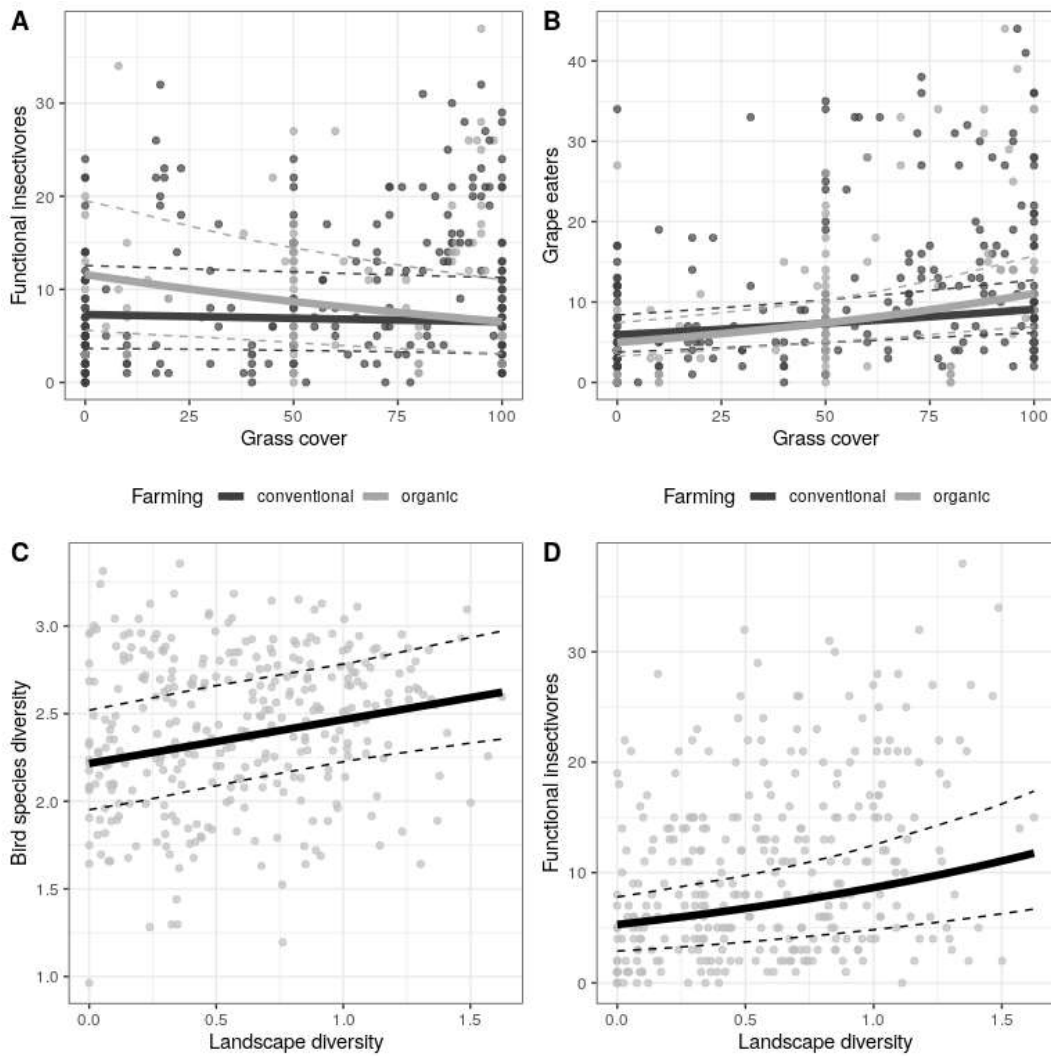


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551 Figure 2. Estimates of conditional averaged models for community metrics, avian functions and functional diversity.

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556 Figure 3. Effects of grass cover and landscape diversity on bird diversity and functions. See Fig. 2 for  
557 estimates and confidence intervals.

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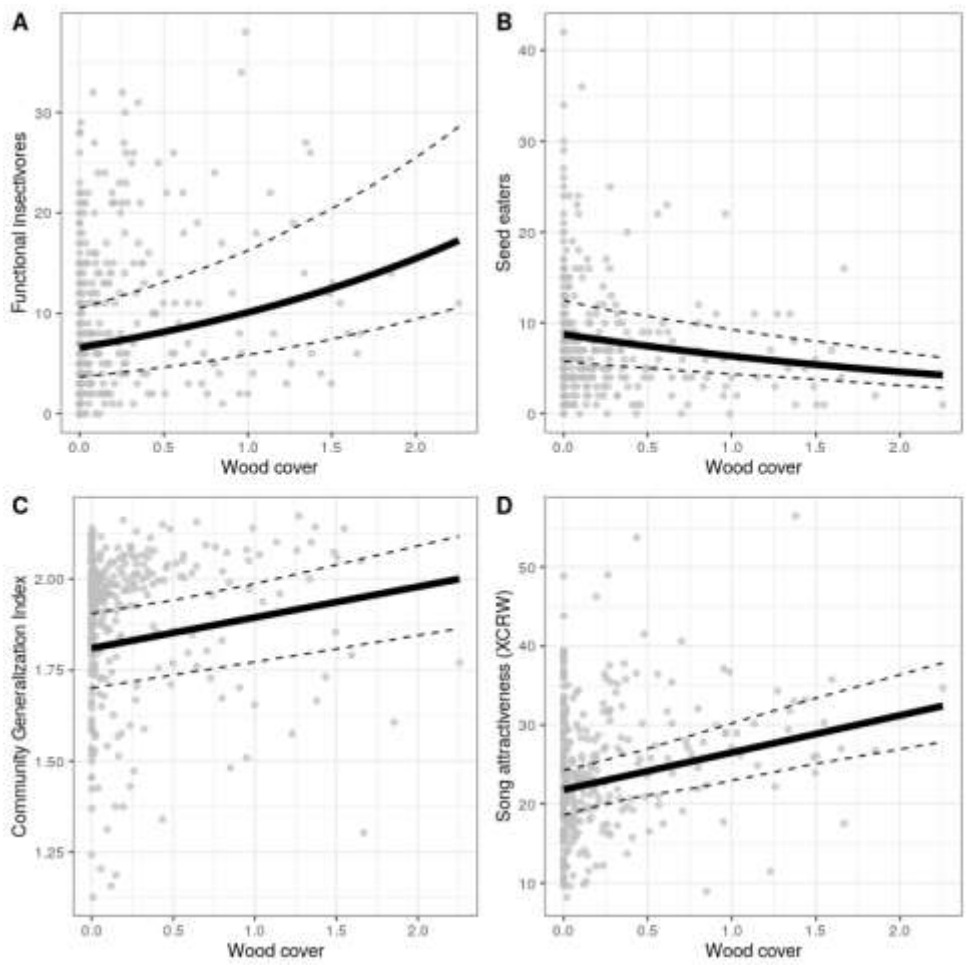
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565 Figure 4. Effects of woodland cover on bird community metrics. See Fig. 2 for estimates and  
 566 confidence intervals.

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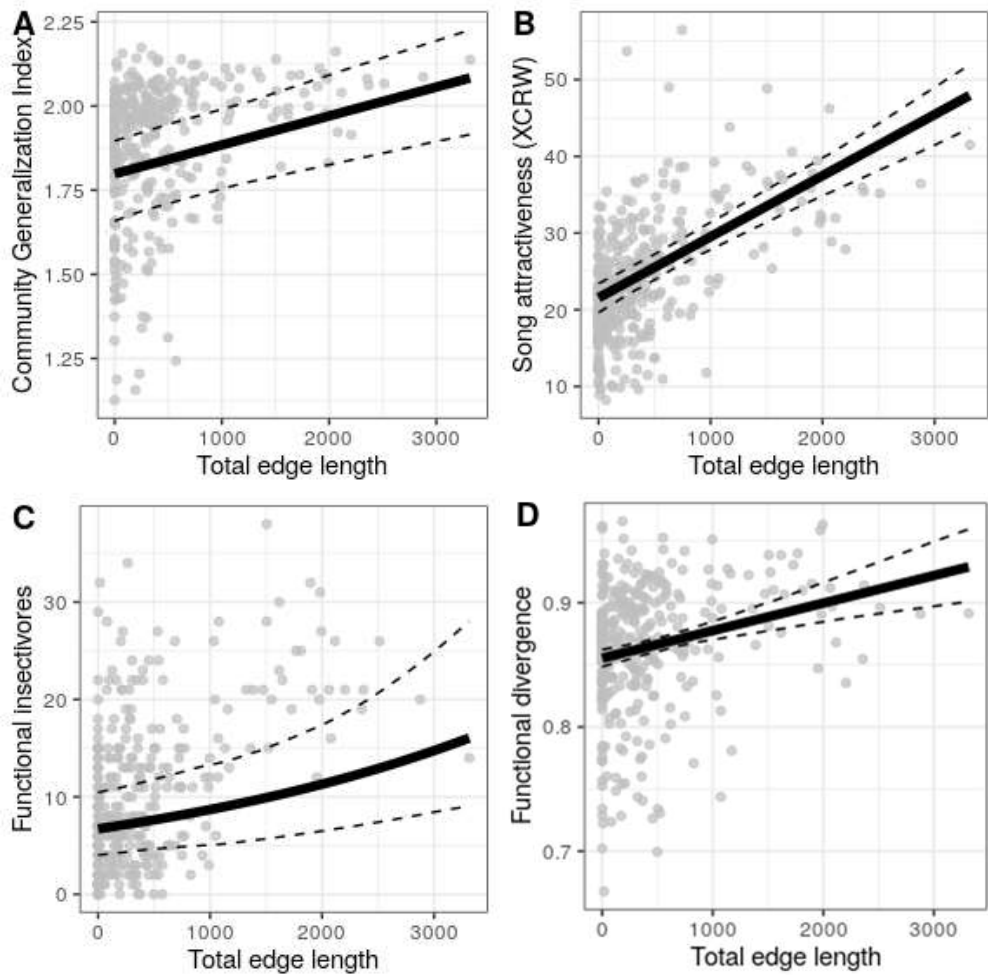


Figure 5. Effects of landscape configurational heterogeneity (edge length between vineyards and semi-natural habitats) on bird community metrics. . See Fig. 2 for estimates and confidence intervals.

592 **Appendix S1. Description of the 12 sampled regions of wine production and study areas**

593

594 Overall, the bioclimatic variation between wine production areas can be summarized as follows: (i)  
595 the northern part lies under more humid oceanic or semi-continental type climates and a mixture of  
596 limestones and alluvial soils (graves) where management has been traditionally very intensive and is  
597 now more and more organic with increasing use of full or partial inter-row grass cover (60% of vine  
598 inter-rows on average); (ii) the eastern part has a more continental climate and is managed  
599 intensively, although organic management is also rapidly developing, with increasing use of grass  
600 cover (75% on average); and (iii) the southern part has a large diversity of soils on limestones, schists  
601 or alluvial deposits under a dry and warm Mediterranean climate. The latter region is the less  
602 intensive where organic management is developing rapidly, while vine inter-rows are often kept as  
603 bare ground to avoid water competition with grapes (23% of grass cover on average).

604 The Bourgogne wine growing area extends to 190 000ha, including 25 000ha of AOC (Appellation  
605 d'Origine Contrôlée) vineyards, with a wide diversity of landscapes and climates. Vineyards are  
606 mainly concentrated in the plains and on the hillsides (12% of the total area) and interspersed with  
607 forests on the top of hills (50% of the total area) and cereal and other crops within the landscape  
608 mosaic. The climate is mainly oceanic to semi-continental, and the area is characterized by a low  
609 altitude variation (from 200 to 500m a.s.l) and limestone soils. In Saumur, part of the Loire Valley  
610 vineyards, the study area covered 5900 ha corresponding to the Saumur-Champigny wine production  
611 area (protected geographical indication, 'AOC'), of which 1600 ha were vineyards managed by  
612 around 120 viticulturists. The area is located on a low elevation limestone plateau submitted to an  
613 oceanic to semi-continental climate. Within this area, twelve 1 km square landscape units were  
614 selected to represent contexts varying in the proportion of vineyard cover, woodland, crops, built-up  
615 land and semi-natural areas (see Pithon et al. 2016 for more details).

616 The Ile de Ré wine production area covers 650 ha on the third main Atlantic coast island of western  
617 France. Ile de Ré vineyards are part of the Cognac production area and are established on sandy soils  
618 and maritime alluvions covering a limestone substrate, under an oceanic climate. Vineyards are  
619 interspersed with semi-natural grasslands, wetlands and pine forests, with a predominance of  
620 organic management. In Bordeaux, the study area is part of Nouvelle Aquitaine region, southwestern  
621 France, with a wine production currently covering 145 000 ha of vineyards. In Bordeaux, we sampled  
622 three subregions, namely Médoc (Margaux 'AOC' – 1500 ha), Saint Emilion (8000 ha) and Entre-Deux-  
623 Mers (1400 ha). The climate is oceanic but the soils are on quaternary alluvial deposits ('graves') in  
624 Medoc while they are located on silty and sandy limestones in Saint Emilion and Entre-Deux-Mers, at  
625 low elevations between 3 and 107 m a.s.l. In the three subregions, vineyards were selected along a  
626 landscape heterogeneity gradient based on the proportion of semi-natural habitats, including both  
627 woodlands and semi-natural grasslands.

628 The Gaillac area is located in southern France, in the Tarn district, where climate is between Oceanic  
629 and Mediterranean and the altitude is low, from 105 to 288 m a.s.l. The wine growing region covers  
630 8000 ha, with 2800 ha included in an AOC. Vineyards occur on sedimentary soils in the Tarn valley, in  
631 the East hillsides and the Cordes plateau. The study area included 17 vineyards along a landscape  
632 complexity gradient of SNH habitat cover, and about half of the stands are managed by organic  
633 farming. The Limoux area is located south of Carcassonne in Aude district, at the border between  
634 oceanic and Mediterranean climatic influences in southern France. The wine production area extends  
635 to 41150 ha from 150 m a.s.l to higher slopes of 750 m a.s.l. The selected area extends to 7800 ha,  
636 with two main land uses: forests (25%) and vineyards (26%).

637 Costières de Nîmes is an area situated in Gard district in southern France, close to the Mediterranean  
638 coast. The wine production area extends to 86291 ha at low elevations, from 80 to 100 m a.s.l. The  
639 sampled area include 4500 ha mainly composed of vineyards in steady slopes and in open areas,  
640 orchards and permanent crops. In the Terrasses du Larzac, the study area is located north-west of  
641 Montpellier not far from the Mediterranean coast and the mean altitude is low, lying between 57  
642 and 320 m a.s.l. Climate is Mediterranean and the sampled area extends to 459 ha and gather more  
643 than 80 cellars. For the study, 18 vineyards were selected among which 8 were organic. The  
644 Roussillon area covers a total area of 15600 ha located in the extreme south of France, under a warm  
645 and dry Mediterranean climate. Vineyards are established on various soils depending on topography:  
646 limestone slopes (southern Corbières), quaternary alluvial deposits ('graves' of Rivesaltes) and schists  
647 (Banyuls).

648 In Catalonia (Spain), there were two subregions sampled, the first one being in Penedès, in the  
649 Mediterranean coast of Spain, a large area of 15000 ha of vineyards (representing 80% of the  
650 cultivated area) and other crops such as cereals. It sits in a Tertiary sedimentary depression with a  
651 predominantly flat relieve and elevations below 250 m.a.s.l, and has seen in the last decades an  
652 intensification of the agricultural practices. The second one was in the Priorat Appellation of Origin, a  
653 wine-producing mountainous area covering ca 18000 ha, of which 1887 ha are covered by vineyards  
654 located at a mean elevation of  $472 \pm 250$  m.a.s.l. Both areas have a dry Mediterranean climate  
655 influenced by the proximity of the sea (Puig-Montserat et al. 2017).

656 In North Eastern Italy, in Trentino, the wine production area occurs between 65 m a.s.l and 750 m  
657 a.s.l. and covers about 10300 ha, concentrated in the valley bottoms and their hilly sides. Vineyard  
658 management is quite intensive and organic farming represented only less than 3% of the production  
659 area. In addition, there are two types of vineyard structure: *pergola* is the traditional and  
660 predominant one and consist in tall vines supported by a robust structure, while *spalliera* is the most  
661 widespread wire arrangement. The Oltrepò Pavese area is located in the southern extreme of  
662 Lombardy in the North of Italy, from the Po river (up to 50 m a.s.l) to the Apennines mountains (up to  
663 1724 m a.s.l). The study was conducted on the vineyard belt (from 70 to 500 m a.s.l), which covers  
664 about 15000 ha, mostly located on gently sloping hills. The area is dominated by intensively managed  
665 vineyards, broadleaved woodlands and heterogeneous farming systems.

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667

668 **Appendix S2. Bird sampling methods used in the 12 sampled wine-growing regions.**

669

Country	Region	N	Sampling year	Sampling method	Width or diameter (m)	Surface area sampled (ha)
France	Bordeaux	60	2013 and 2015	Transects and points	100	4
France	Languedoc	25	2015 and 2018	Transects	100	4
France	Gaillac	17	2017	Transects	100	4
France	Limoux	25	2013	Point counts	200	3
France	Costières	24	2013	Point counts	200	3
France	Bourgogne	24	2013	Point counts	200	3
France	Loire	22	2010	Transects	50	5
Spain	Catalunya	23	2013	Point counts	200	1
Italy	Trentino	46	2015	Transects	200	7
Italy	Lombardy	65	2015	Transects	200	7
France	Ile de Ré	3	2018	Transects	100	4

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673 **Appendix S3. Table of species abundance and traits**

674

675 The 10 most frequent species across all regions were in decreasing order *Turdus merula*, *Serinus*  
676 *serinus*, *Parus major*, *Corvus corone* (*corone/cornix*), *Sylvia atricapilla*, *Fringilla coelebs*, *Sturnus*  
677 *vulgaris*, *Carduelis carduelis*, *Lullula arborea* and *Columba palumbus* (see Appendix SX).

678

679 **TBC**

680

681 **Appendix S4. List of species used to calculate the abundance of functional insectivores, seed and**  
682 **grape eaters (see methods).**

683 **TBC**

684 **Functional insectivores (N = 34 species)**

685 AEGCAU  
686 CISJUN  
687 CLAGLA  
688 CUCCAN  
689 ERIRUB  
690 FICHYP  
691 FRICOE  
692 HIPICT  
693 HIPPOL  
694 LUSMEG  
695 MUSSTR  
696 PARCAE  
697 PARMAJ  
698 PHOOCH  
699 PHOPHO  
700 PHYBON  
701 PHYCOL  
702 PHYLUS  
703 PHYSIB  
704 PRUMOD  
705 REGIGN  
706 REGREG  
707 SAXRUB  
708 SAXTOR  
709 SYLATR  
710 SYLBOR  
711 SYLCAN  
712 SYLCOM  
713 SYLHOR  
714 SYLMEL  
715 SYLSUB  
716 SYLUND  
717 TROTRO  
718 UPUEPO  
719

720 **Seed eaters (N = 17 species)**

721 ALERUF  
722 CARCAN  
723 CARCAR  
724 CARCHL  
725 COCCOC  
726 COLOEN  
727 COLPAL  
728 COTCOT  
729 EMBCAL  
730 EMBCIA  
731 EMBCIR  
732 EMBCIT  
733 PASMOM  
734 PHACOL  
735 SERSER  
736 STRDEC  
737 STRTUR  
738

739 **Grape consumers (N = 9 species)**

740 CORCOR  
741 CORFRU  
742 CORMON  
743 GARGLA  
744 PASDOM  
745 STUVUL  
746 TURMER  
747 TURPHI  
748 TURPIL  
749

750 **Appendix S5. Multi-scale analysis of predicting performance for landscape metrics.**

751

752 Prior to inclusion of landscape predictors into final models, we used the French dataset (N = 200  
753 plots) to assess the relationship between each landscape variable calculated at different spatial  
754 scales (100, 500, 750, 1000m buffer around the sampling sites) and response variables to identify the  
755 scale best correlated with most bird community metrics. For the 200 French plots where land use data  
756 were fully homogeneous at larger distances around sampled vineyards (Inglada et al. 2018), we also  
757 computed the same landscape metrics, both compositional and configurational, for three larger  
758 buffer sizes of 500, 750 and 1000 m-radius, respectively. For this subsample of the data set, we ran  
759 all the models using the same full model structure as detailed in the Methods section for the full data  
760 set, with landscape predictors computed in increasing buffer scales of 100, 500, 750 and 1000 m to  
761 test their performances and consistencies across larger spatial scales. The best buffer scale selected  
762 was always 100 m for all metrics, with a strong consistency in predictor selection, except for species  
763 diversity, functional insectivores and functional entropy, for which the best scales were 500 m.  
764 However, the same predictors were selected at all scales, except open habitat cover and edge length  
765 that has increasing predicting performance with scale at the expense of landscape diversity for  
766 functional insectivores.

767

768