

Decreased burrowing activity of endogeic earthworms and effects on water infiltration in response to an increase in soil bulk density

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1	Decreased burrowing activity of endogeic earthworms and effects on water infil-
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24 Abstract

25 Endogeic earthworms live and burrow in the soil to find their food. They burrow by pushing the soil 26 aside or ingesting it and are thus sensitive to soil compaction. However there is a scarcity of data 27 regarding the effects of soil bulk density on the burrowing behavior and activity of endogeic earth-28 worms. We carried out laboratory experiments using repacked soil cores with various levels of bulk 29 density (from 1.18 to 1.38 g cm⁻³) in which individuals of Aporrectodea caliginosa or Allolobophora 30 chlorotica were incubated for six weeks. The burrow systems inside the soil cores and the compac-31 tion around the burrows were then analyzed using X-ray tomography. Soil water infiltration meas-32 urements were also carried out. The increase in bulk density had a negative impact on all burrow 33 system characteristics (length, volume, diameter, continuity, number of burrows). When bulk densi-34 ty increased from 1.18 to 1.38 g cm⁻³, volume, diameter, continuity and the number of burrows de-35 creased on average by 77%, 21%, 81% and 58%, respectively. The increase in density due to com-36 paction around the burrows was similar whatever the species and the bulk density. Increasing soil 37 bulk density from 1.18 to 1.38 g cm⁻³ also greatly decreased water infiltration (-89% for both species) 38 and increased breakthrough time (10 and 25-fold for A. chlorotica and A. caliginosa respectively). 39 However, compared to a control without earthworms, water infiltration in cores incubated with en-40 dogeic species was only increased significantly at 1.18 and 1.23 g cm⁻³. This illustrates that burrows 41 made by endogeic earthworms moderately increase water infiltration and only when the soil bulk 42 density is low. Data provided in this study could be used to refine simulation models of earthworm 43 burrowing behavior where burrowing is assumed to be mainly governed by soil water content, tem-44 perature and soil bulk density.

45

Keywords: compaction; burrow system; *Aporrectodea caliginosa; Allolobophora chlorotica*; behaviour

48 Abbreviations

49 BD: soil bulk density

50

51 Introduction

52 Compaction, partly due to an increase of engine loads, is a major threat to agricultural and forest 53 soils and their resilience (Hansen 1996). Earthworms are particularly sensitive to soil compaction as 54 they burrow their way in the soil either by pushing the soil aside or ingesting it (Beylich et al., 2010; 55 Keller et al., 2017). Indeed, ingestion is the only possible way of moving in highly compacted soils 56 (Dexter, 1978) and often results in higher quantities of excreted surface casts (Joschko et al., 1989). 57 A high BD also means that more energy is spent burrowing (Ruiz et al., 2010). Although earth-58 worms generally avoid highly compacted zones (Stovold et al., 1994), epi-anecic such as Lumbricus 59 *terrestris*, earthworms have been observed to burrow in them under semi-field conditions in the 60 presence of a plough pan (Capowiez et al., 2009). Epi-anecics and other species have also been 61 found to burrow in compacted zones below wheels tracks under field conditions (Capowiez et al., 62 2012).

63 Thus bulk density (BD) is one important factor indirectly governing earthworm abundance in 64 croplands (Chan and Barchia, 2007; Ouellet et al., 2008; Capowiez et al., 2009), forests (Jordan et 65 al., 1999; Bottinelli et al., 2014), tallgrass prairies (Althoff et al., 2009) and urban soils (Smetak et 66 al., 2007). Among earthworms, anecic and epi-anecic earthworms (sensu Bouché, 1972) are thought 67 to be more powerful axially (Keudel and Schrader, 1999) and thus less sensitive to compaction 68 (Beylich et al., 2010). Paradoxically, much more information has been reported on the effects of soil 69 compaction on anecic or epi-anecic earthworm behavior than on other ecological categories. Rush-70 ton (1986) and Joschko et al. (1989) demonstrated that burrow length was decreased when L. 71 terrestris was challenged with increased BD. Stovold et al. (1994) made a similar observation for 72 Aporrectodea nocturna in a choice experiment. Kretzschmar (1991) estimated that Aporrectodea

73 *longa* burrowing activity, as assessed by surface cast production, decreased for higher BD. This bias 74 towards the anecic or epi-anecic species is explained by the fact that these species have a greater 75 influence on water infiltration since their burrows are more vertical, more continuous (because less 76 refilled) and have surface openings (Capowiez et al., 2015). In fact, compaction impacts not only 77 burrow numbers but also burrow continuity (Schäffer et al., 2007; Capowiez et al., 2012) and this in 78 turn can drastically reduce water infiltration (Kim et al., 2010). In comparison, information regar-79 ding the effects of compaction on endogeic earthworms is scarce. Langmaack et al (1999) studied A. 80 caliginosa burrowing behavior in compacted vs non-compacted soils and concluded that in response 81 to compaction this species increased burrow continuity but no clear trend was observed for burrow 82 length. Dexter (1978) also found that high BD did not have a negative impact on A. caliginosa bur-83 rowing. In contrast, Söchtig and Larink (1992) determined that A. caliginosa burrow length de-84 creased linearly when pore volume decreased (and thus BD increased).

85 In most of these studies, the effects of high BD on earthworms were measured indirectly, i.e. either 86 by assessing surface cast production (assuming most of the casts were produced at the soil surface) 87 or the burrows on the outer borders of soil cores (assuming these burrows were representative of 88 burrows in the cores). This latter case is due to the absence of non-invasive techniques for estimat-89 ing burrows in soil cores in past decades. Nevertheless the first application of X-ray tomography to 90 analyze earthworm systems was in the early 90s (Joschko et al., 1991) and medical scanners are 91 now readily available to solve this problem. Microtomography with an higher resolution can now 92 also be used (Balseiro-Romano et al., 2020). Another point that sometimes limits the relevance of 93 past studies is the fact that they were observed using choice experiments: earthworms were placed 94 in mesocosms containing soils with different BD and their burrowing activity was estimated in the 95 different situations (Stovold et al., 1994; Söchtig and Larink, 1992). This kind of study provides in-96 formation on earthworm preferences but not really on their ability to burrow and live in compacted 97 soils. Furthermore, in most studies only two to three BD were tested (Joschko et al., 1989; Joschko

et al., 1993; Langmaack et al., 1999; Stovold et al., 1994; Jégou et al., 2002; Capowiez et al., 2012)
whereas recent studies suggested that depending on the soil texture, earthworm responses in terms
of burrowing are not always monotonic (see Figure 5 in Pöhlitz et al., 2020).

101 Overall, there is an increasing demand for accurate estimates of the effects of BD on earthworm 102 burrowing (Roeben et al., 2020) since this kind of information is required, for example, to simulate 103 earthworm behavior and better assess their real exposure to pesticides. However, there is a scarcity 104 of useable data in the literature, especially for endogeic species.

105 Thus, in the present study, we carried out a lab experiment with soil cores compacted at five differ-

106 ent BD currently observed in orchard soils close to Avignon, France. We then studied the resulting

107 effects on the burrow system characteristics of two very common endogeic species (Aporrectodea

108 *caliginosa* and *Allolobophora chlorotica*). We also determined how BD and earthworm species

109 modified water infiltration. Furthermore, it is known that earthworms can compact the soil when

110 they burrow (Rogasik et al., 2014; Capowiez et al., 2011; Milleret et al., 2009) and thus we assessed

111 whether the lateral compaction around the burrows was influenced by soil BD for these species.

112

113 Material and Methods

114 Soil and earthworms

115 Soil (0-20 cm depth) and earthworms were sampled in an orchard abandoned since at least ten years

116 in Avignon, in the south east of France. The soil has the following characteristics: pH = 8.2, clay =

117 19.7%, silt = 50.8%, sand = 29.5%, OM = 3.4%, CEC = 8.48 cmol C kg⁻¹. The soil was sieved at 3

118 mm and kept at constant humidity (20% on a mass basis, i.e. about 80% of its water holding capaci-

119 ty). In this orchard, we found high densities of earthworms (> 300 individuals m^{-2}) and the most

120 dominant species were two endogeic species: Allolobophora chlorotica (leucotypic form according

121 to Bouché (1972) but often called pink morph) and Aporrectodea caliginosa. However, these two

122 species presumably do not occupy exactly the same niche as these species were indeed described as

123 an 'intermediate' and 'endogeic' species by Bouché (1972), respectively. Recently, Bottinelli et al.

124 (2020) demonstrated that each lumbricid earthworm species can be categorized by three percentages

125 into the three main ecological categories defined by Bouché (1977). According to that study, A.

126 chlorotica (leucotypic form) is 97% endogeic and 3% anecic, whereas A. caliginosa is 80%

127 endogeic, 16% epigeic and 4% anecic.

128

129 Mesocosms and experimental design

130 Mesocosms were PVC cores (height: 30 cm, diameter:16 cm) filled with sieved soil. For each soil 131 layer (600 g of soil at 80% WHC), the soil was pressed to the desired bulk density using a hydraulic 132 press with increasing pressures. For this experiment, 50 cores were prepared with ten cores for each 133 soil bulk density (BD): four with A. caliginosa, four with A. chlorotica and two control cores with-134 out earthworms. The five levels of BD were 1.18, 1.23, 1.29, 1.34 and 1.38 g cm⁻³. This range was chosen to encompass the BD observed in orchards, from tilled zones under the row to compacted 135 zones below wheel tracks (personal observation). In each core, we introduced three individuals of 136 either A. caliginosa or A. chlorotica (thus a density of 150 individuals m⁻²) or no earthworms. These 137 endogeic densities can be found in most agricultural soils (excepted vineyards). Only adults were 138 139 used in the experiment and the mean (+/-SD) biomass for each species was 0.57 (+/-0.04) and 0.44 140 (+/- 0.04) g for A. caliginosa and A. chlorotica, respectively. At the end of the experiment (5 weeks), 141 5 ml of chloroform was poured into each core to kill earthworms and prevent further burrowing. In 142 each core, surface casts were sampled, dried for 48 h at 110°C and weighed.

143

144 X-ray computed tomography and burrow system characterization

145 Macroporosity within the soil cores was analyzed by X-ray computed tomography using a medical

146 scanner (BrightSpeed Exel 4, General Electric) in the INRA Nancy research centre. The scanner

147 settings were 130 kV and 50 mA. DICOM images were transformed into 8-bit images using the fol-

148 lowing parameters WL = -1000 and WW = 2000 HU. The greylevel histograms of the images 149 showed two well separated peaks, one for the soil matrix and one for the porosity. We decided to 150 choose a binary threshold value set at 1/3 of the distance between the two peaks. Since the peak 151 value is dependent on the density of the soil matrix, this ensures that differences in density between 152 BD treatments did not influence the total macroporosity volume. Images were filtered using a mean 153 filter of radius 2 to remove some noise. Burrow system 3D reconstructions and skeletonizations 154 were obtained using Avizo software (ThermoScientific). Other computations were carried out using 155 ImageJ (Schindelin et al., 2012) and the 'SoilJ' and 'BoneJ' plugins. We thus estimated burrow sys-156 tem volume. Mean burrow diameter was estimated by selecting the most circular pores ('circulari-157 ty' > 0.8 in ImageJ) on 2D images, assessing their area and computing an equivalent diameter. To 158 assess burrow system continuity, as no standardized measure exists, we compute the number of bur-159 rows whose vertical extension was larger than 30% of the core length (i.e. 9 cm) was counted. We 160 also computed the degree of anisotropy of the burrow system using BoneJ; this algorithm assesses whether objects are oriented in the same direction (degree of anisotropy close to 0) or randomly in 161 162 3D (degree of anisotropy close to 1). To assess whether the endogeic species compacted the soil 163 around their respective burrows, we selected the 8 mm spatial zones around the burrows (as as-164 sessed visually on the images) and simply computed the mean greylevel in these zones. For cores 165 without earthworms, we randomly selected zones of similar size in the images to compute the mean 166 greylevel value.

167

168 Water infiltration

To study the effects of the presence of an earthworm burrow system and its characteristics on water infiltration, we used the same protocols in Capowiez et al. (2015). Soil cores were placed on a funnel to be able to observe water drainage below the cores. In brief, 150 ml of water was poured onto a soil core, the time was recorded until the water was completely absorbed and then a new volume

173 was added. We also noted the time at which water began to drain ('breakthrough time'). This proce-174 dure was repeated until a linear trend was observed (i.e. same time intervals between applied water 175 volumes), which was typically after around eight volumes. This single-ring protocol allowed the 176 steady-state infiltration rate to be estimated by linear regression for each core. The first data points 177 of the relationship between time and the number of water volumes poured were removed.

178

179 Statistical analysis

180 Our experimental design was balanced for comparing the responses of the two species (if we ex-

181 cluded cores without earthworms). For count data (burrow continuity), we used a GLM with a qua-

182 si-poisson family. For other burrow system characteristics and cast production, we used ANOVA

183 since no large deviations from normality (Shapiro-Wilk test: all p-values > 0.061) or

homoscedasticity (Levene test: all p-values > 0.047) were detected. In both cases, we first ran the

185 analyses with two factors (BD and earthworm species) and then if the interaction was not signifi-

186 cant (the most common case), we ran two tests (GLM or ANOVA) for each species separately. For

187 infiltration data (infiltration rate and breakthrough time), we applied the same analysis but we also

188 computed the infiltration gain, i.e. the difference between infiltration of each species compared to

189 the cores at the same BD but without earthworms. These gains were also analyzed using a two-way

190 ANOVA and then two one-way ANOVA. Linear regressions were carried out for burrow volume

191 with the 'lm' function. All analyses were made in R (R core team, 2019).

192

193 **Results**

194 Effects of bulk density on the earthworm burrow system

195 Examples of the 3D burrow systems are shown in Figure 1. The two-way ANOVA indicated that

196 BD had a highly significant effect on burrow volume whereas species and the interaction between

197 BD and species were not significant (Table S1). One-way ANOVA for each species separately

198 showed that burrow volumes were significantly larger at 1.18 g cm⁻³than at other BD for both spe-

199 cies (Fig. 2). At 1.38 g cm⁻³, the burrow volume was significantly lower than at 1.18 and 1.23 g cm⁻³

200 for both species. The decrease in burrow volume between the minimal (1.18 g cm⁻³) and maximal

201 (1.38 g cm⁻³) tested BD was high with -75% and -80% for A. caliginosa and A. chlorotica, respec-

202 tively. The linear regressions between BD and burrow volume were significant (p = 0.027 and

203 0.0071) and the R² values were 0.79 and 0.91 and for A. caliginosa and A. chlorotica respectively,

even if the linear trend was more obvious for the latter species. We observed very similar decreasesand significant differences in burrow lengths (Table 1).

BD and species (p < 0.001) also had a significant influence on burrow diameter, but no interaction

207 was observed between these factors (p = 0.21). For. A. caliginosa, burrow diameter was only sig-

nificantly different at the lowest BD compared to the highest BD. For *A. chlorotica* diameter values
at the two lowest BD were significantly different from those at the other BD (Table 1).

210 BD also has a significant effect on burrow continuity, assessed as the number of burrows with a ver-

211 tical elongation larger than 9 cm (Table S1). At the lowest BD, a mean of 6 and 7 such burrows

212 were observed for *A. caliginosa* and *A. chlorotica* respectively. At the highest BD, only one such

213 burrow was observed on average for both species (Fig. 3). Thus, BD had a highly significant effect

on burrow continuity (p < 0.001), whereas species and interaction did not (p = 0.17 and 0.16, re-

215 spectively). For A. caliginosa, only the continuity for the lowest and the highest BD were different,

whereas for *A. chlorotica*, continuity at the three highest BD was significantly different from that atthe lowest BD.

The degree of anisotropy of the burrow systems increased with increasing BD whereas species and the interaction were not significant (Table S1 and Fig. 4). For both species, the degree of anisotropy for the two lowest BD was significantly lower than those observed at the highest BD which sug-

221 gests that increased BD leads to less randomly orientated burrow systems. Finally BD or species

(results not shown) did not affect burrow system verticality (mean deviation from the vertical of thesegments of the skeleton).

Surface cast production was low (less than 10 g of dry soil) and not influenced by BD for *A. chlo- rotica* (Table 1). In contrast, for *A. caliginosa*, cast production was higher (15-32 g of dry soil; Table
1) and was the highest at a BD of 1.34 g cm⁻³ and significantly different from the amount produced
at the two lowest tested BD.

228

229 Effects on water infiltration

BD (p < 0.001) had a significant impact on infiltration rates and breakthrough times, but the earth-

231 worm factor (with three modalities: A. caliginosa, A. chlorotica and no earthworm) and the interac-

- tion were not significant (p = 0.36 and 0.55 respectively for infiltration rate ; p = 0.30 and 0.21 res-
- 233 pectively for breakthrough time). When BD increased, infiltration rates decreased and breakthrough
- times increased (Table 2).

235 When infiltration was expressed relative to the control without earthworms, BD has a significant

236 impact (p < 0.001) but species (with two modalities: A. caliginosa and A. chlorotica) and the inter-

action were not significant (p = 0.26 and 0.17 respectively). Significant differences due to earth-

worm's presence were only observed at the lowest BD for *A. caliginosa* and at the two lowest BD

239 for A. chlorotica (Fig. 5). No significant effect was observed for breakthrough times expressed rela-

- tive to the control (data not shown).
- 241

242 Effects on soil bulk density around the burrows

The mean soil BD around the burrows increased when BD of the soil core increased (Fig. S2). Soil BD had a highly significant effect (p < 0.001) as did the presence of earthworms (p < 0.001), but there was no significant interaction between these two factors (p = 0.83). For each BD, the differ-

ence in BD around the burrows appeared to be similar (around five greylevels corresponding to 59

HU) in the presence or absence of earthworms.

248

249 **Discussion**

250 Effects of bulk density on earthworm burrowing and casting behavior

251 BD strongly and negatively influenced all the geometrical characteristics of the burrow systems of 252 both endogeic species. First of all in terms of magnitude, the volumes (and lengths) of the burrow 253 systems decreased almost linearly with increasing BD. Thus with our soil which lies between loam 254 and silt loam soil, we did not observe a non-monotonic relationship between BD and soil 255 macroporosity as reported by Pöhlitz et al. (2020) using the epi-anecic earthworm L. terrestris. This 256 earthworm species has a mono-typical burrowing behavior with the creation of an often single ver-257 tical burrow, eventually with a few lateral branches close to the surface resulting in a Y-shaped burrow system (Shipitalo and Butt, 1999). These features make this species less prone to drastically 258 259 decreasing burrow length since their burrow is also a vital shelter.

260 Burrow diameter gradually decreased when BD increased for both species, but this change was 261 more marked for A. chlorotica for which we already observed a significant difference between 1.18 262 and 1.29 g cm⁻³. As the earthworms were of similar weight at the beginning of the experiment, this 263 suggests that they may have created thinner burrows in denser soil, probably to save energy. Other burrow system characteristics were also modified. The most drastic change was burrow sys-264 265 tem continuity. Its decrease with increasing BD was almost linear for A. caliginosa, but for A. chlo-266 rotica the most important changes occurred between 1.18 and 1.29 g cm⁻³ with a sharp decrease of -267 60%. This suggests that the burrows of this species were less vertically continuous as soon as the 268 BD increases. Moreover, the degree of anisotropy of the burrow systems increased significantly 269 with BD for both species. This showed that the burrows had no preferential orientation at low BD 270 but then a trend for a non-preferential orientation appeared for higher BD. Taken together, these ob-

- 271 servations further suggests that A. chlorotica burrows less vertically in soil with intermediate to
- 272 higher BD values, even if we did not observe significant differences for verticality.

Overall, we did not observe large differences between the responses of the two species to the increase in BD. However, for most of the burrow system characteristics (except the degree of anisotropy) the decreases observed at 1.34 compared to 1.18 g cm⁻³ were less marked for *A. caliginosa* than for *A. chlorotica*, even if the values observed for a BD of 1.38 g cm⁻³ were similar for both species. Thus although a BD of 1.38 g cm⁻³ was a strongly limiting factor for both species, *A.caliginosa* appeared to have a better resistance to lower but still high BD (1.34 g cm⁻³).

280

281 Another striking difference between the two species was the production of casts at the surface with 282 A. chlorotica producing very few casts regardless of the BD. In contrast, A. caliginosa produced more casts and showed a tendency to cast more at the soil surface when BD increased (except for 283 the highest BD but in that case the burrow volume was very limited). This corroborates previous 284 285 findings (Rushton, 1986; Kretzschmar, 1991) showing that when the soil is compacted the earth-286 worms tend to produce more surface casts in order to increase the air-filled porosity and aeration 287 within the soil. If we assume that the casts had the same BD than the soil from which they were 288 created, we could compute the ratio of the volume of macroporosity (burrows) that was deposited at 289 the soil surface as casts. In this case, the proportion of burrow volume increased with SBD for both 290 species (Fig. S3); this increase was however higher for A. caliginosa with casts representing 100% 291 of the burrow volume at the highest BD tested. The difference between the two species may partly 292 be explained by the fact that A. caliginosa is epigeic at 16% (A. chlorotica at 0%) and thus is more 293 prone to defecate at the soil surface.

294

295 Consequences for water infiltration

296 Breakthrough time characterizes the transient regime of infiltration while the infiltration rate is a 297 measure of the stationary flow regime. In control samples without earthworms, infiltration rates de-298 creased by a factor of 4.4 and and breakthrough time increased by a factor of 7.6 when BD in-299 creases from 1.18 to 1.38 g cm⁻³. The occurence of macropores (earthworm burrows) in the homo-300 geneous porous sample tended to mitigate the effect but only for the lowest BD. Infiltration rates 301 and breakthrough times increased when BD increased. This finding was expected since an increased 302 BD means lower macroporosity and generally a decreased continuity of such pores as illustrated in 303 the present study with the decreased burrow continuity.

304 Endogeic burrows are often discontinuous, since they can be refilled by casts, and often less con-305 nected to the surface (Capowiez et al., 2014). Thus their involvement in increased water transfer 306 efficiency is still open for debate. In the present study, we demonstrated that the effect of the activity of endogeic species on infiltration was only significant for low BD (1.18 g cm⁻³ for A. caliginosa 307 and 1.18 and 1.23 g cm⁻³ for A. chlorotica). This may appear to be counterintuitive: when the soil is 308 compacted the burrows could have been preferential pathways for air and water transfer 309 310 (Mossadeghi-Björklund et al., 2016). However these pathways are less vertical and only partial 311 since the burrow continuity rapidly decreased when BD increased. In addition, the burrow systems 312 were less extensive at moderate and high BD and thus their functional role was limited. 313 We observed a small but significant increase in the soil matrix greylevels surrounding earthworm 314 burrows (for both species). Thus, regardless of the initial BD, earthworm burrowing appears to 315 cause a constant increase in the BD around the burrows. This also limit the roles of earthworm bur-316 rows in water transfer (Pelisek et al., 2018) since the two species produced discontinuous burrows 317 and thus water would need to enter and go out of several burrows to reach the bottom of the cores. 318 Rogasik et al. (2014) previously showed that L. terrestris burrowing caused a lateral compaction of the soil around the burrows. We further demonstrate that this is true for the two studied endogeic 319

320 species. However, we were unable to detect an influence of BD on the intensity of this lateral com-

321 paction even though the burrow diameter significantly decreased at the highest tested BD.

322

323 Conclusions

324 We demonstrated that increasing BD impaired the burrowing behavior of two common endogeic 325 species, and that total burrow volume and length decreased linearly with increasing BD. Although 326 the earthworms promoted the soil infiltration capacity compared to control samples without earth-327 worms, their effect was limited and only significant at low BD. These results were obtained using a 328 single soil and only two species. The main findings could be compared with previous results ob-329 tained in other soils by computing an effective BD (Beylich et al., 2010) or by using a reference BD 330 (Keller and Hakansson, 2010). Our results could be used in simulation models of earthworm behav-331 iour, such as the model recently presented by Roeben et al. (2020). In this model, an energy budget is applied for each earthworm and our data could help parameterize the energy costs of burrowing 332 333 as a function of bulk density.

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339 References

- Althoff, P.S., Todd, P.C., Thien, S.J., Callaham, M.A., 2009. Response of soil microbial and invertebrate
 communities to tracked vehicle disturbance in tallgrass prairie. Appl. Soil Ecol. 43, 122-130.
- Beylich, A., Oberholzer, H.R., Schrader, S., Hoper, H., Wilke, B.M., 2010. Evaluation of soil compaction
 effects on soil biota and soil biological processes in soils. Soil Till. Res. 109, 133-143.
- Bottinelli, N., Capowiez, Y., Ranger, J., 2014. Slow recovery of earthworm populations after heavy traffic in
 two forest soils in northern France. Appl. Soil Ecol. 73, 130-133.
- Bottinelli, N., Hedde, M., Jouquet, P., Capowiez, Y., 2020. An explicit definition of earthworm ecological
 categories Marcel Bouché's triangle revisited. Geoderma 372, 114361.
- 348 Bouché, M.B., 1972. Lombriciens de France. Ecologie et systématique. INRA, Paris.
- Bouché, M.B., 1977. Stratégies lombriciennes. In: Lohm, U., Persson, T., (Eds.) Soil organisms as components of ecosystems. Stockholm, Ecol. Bull. 25, 122–132.
- Capowiez, Y., Cadoux, S., Bouchand, P., Ruy, S., Roger-Estrade, J., Richard, G., Boizard, H., 2009. The in fluence of tillage type and compaction on earthworm communities and the consequences for

353 macroporosity and water infiltration in crop fields. Soil Till. Res. 105, 209-216.

- Capowiez, Y., Sammartino, S., Michel, E., 2011. Using X-ray tomography to quantify earthworm bioturbation non-destructively in repacked soil cores. Geoderma 162, 124–131.
- Capowiez, Y., Sammartino, S., Cadoux, S., Bouchant, P., Richard, G., Boizard, H., 2012. Role of earthworm
 in regenerating soil structure after compaction in reduced tillage systems. Soil Biol. Biochem. 55, 93103.
- Capowiez, Y., Bottinelli, N., Jouquet, P., 2014. Quantitative estimates of burrow construction and destruction
 by anecic and endogeic earthworms in repacked soil cores. Appl. Soil Ecol. 74, 46-50.
- 361 Capowiez, Y., Bottinelli, N., Sammartino, S., Michel, E., Jouquet, P., 2015. Morphological and functional
- 362 characterisation of the burrow systems of six earthworm species (Lumbricidae). Biol. Fertil. Soils 51,
 363 869-877.
- 364 Chan, K.Y., Barchia, I., 2007. Soil compaction controls the abundance, biomass and distribution of earth-
- 365 worms in a single dairy farm in south-eastern Australia. Soil Till. Res. 94, 75-82.

- 366 Dexter, A.R., 1978. Tunnelling in soil by earthworms. Soil Biol. Biochem. 10, 447-449.
- Hansen, S., 1996. Effects of manure treatment and soil compaction on plant production of a dairy farm system converting to organic farming practice. Agr. Ecosyst. Environ. 56, 173-186.
- 369 Jégou, D., Brunotte, J., Rogasik, H., Capowiez, Y., Diestel, H., Schrader, S., Cluzeau, D., 2002. Impact of
- 370 soil compaction on earthworm burrow systems using X-ray computed tomography. Preliminary study.
- 371 Eur. J. Soil Biol. 38, 329-336.
- Jordan, D., Hubbard, V.C., Ponder, F., Berry, E.C., 1999. Effect of soil compaction and organic matter re-
- moval on two earthworm populations and some soil properties in a hardwood forest. Pedobiologia 43,
 802-807.
- Joschko, M., Diestel, H., Larink, O., 1989. Assessment of earthworm burrowing efficiency in compacted soil
 with a combination of morphological and soil physical measurements. Biol. Fertil. Soils 8, 191-196.
- 377 Joschko, M., Graff, O., Muller, P.C., Kotzke, K., Lindner, P., Pretschner, D.P., Larink, O., 1991. A nonde-
- 378 structive method for the morphological assessment of earthworm burrow systems in 3 dimensions by
 379 X-ray computed-tomography. Biol. Fertil. Soils 11, 88-92.
- Joschko, M., Müller, P.C., Kotzke, K., Döhring, W., Larink, O., 1993. Earthworm burrow system development assessed by means of X-ray computed tomography. Geoderma 56, 201–209.
- 382 Keller, T., Colombi, T., Ruiz, S., Manalili, M.P., Rek, J., Stadelman, V., Wunderli, H., Breitenstein, D., Reiser,
- 383 R., Oberholzer, H., Schymanski, S., Romero-Ritz, A., Linde, N., Weisskopf, P., Walter, A., Or, D.,
- 2017. Long-term soil structure observatory for monitoring post-compaction evolution of soil structure.
 Vadose Zone J. 16, vzj2016.11.0118.
- Keller, T., Hakasson, I., 2010. Estimation of reference bulk density from soil particle size distribution and
 soil organic matter content. Geoderma 154, 398-406.
- Kim, H., Anderson, S.H., Motavalli, P.P., Gantzer, C.J., 2010. Compaction effects on soil macropore geometry and related parameters for an arable field. Geoderma 160,244-251.
- 390 Keudel, M., Schrader, S., 1999. Axial and radial pressure exerted by earthworms of different ecological
- 391 groups. Biol. Fertil. Soils 29, 262-269.

- Kretzschmar, A., 1991. Burrowing ability of the earthworm *Aporrectodea longa* limited by soil compaction
 and water potential. Biol. Fertil. Soils 11, 48-51.
- Langmaack, M., Schrader, S., Rapp-Bernhardt, U., Kotzke, K., 1999. Quantitative analysis of earthworm
 burrow systems with respect to biological soil-structure regeneration after soil compaction. Biol. Fertil.
 Soils 28:219-222.
- Milleret, R., Le Bayon, C., Lamy, F., Gobat, J.M., Boivin, P., 2009. Impact of roots, mycorrhizas and earthworms on soil physical properties as assessed by shrinkage analysis. J. Hydrol. 373, 499-507.
- Mossadeghi-Björklund, M., Arvidsson, J., Keller, T., Koestel, J., Lamandé, M., Larsbo, M., Jarvis N., 2016.
 Effects of subsoil compaction on hydraulic properties and preferential flow in a Swedish clay soil. Soil
 Till. Res. 156, 91-98.
- 402 Ouellet, G., Lapen, D.R., Topp, E., Sawada, M., Edwards, M., 2008. A heuristic model to predict earthworm
- 403 biomass in agroecosystems based on selected management and soil properties. Appl. Soil Ecol. 39, 35404 45.
- Pelisek, I., 2018. Investigation of soil water infiltration at a scale of individual earthworm channels. Soil Water Res. 13, 1-10.
- 407 Pöhlitz, J., Rucknagel, J., Schlütter, S., Vogel, H.J., Christen, O., 2020. Computed tomography as an exten408 sion of classical methods in the analysis of soil compaction, exemplified on samples from two tillage
 409 treatments and at two moisture tensions. Geoderma 346, 52-62.
- R Core Team, 2019. R: A language and environment for statistical computing. R Foundation for Statistical
 Computing, Vienna, Austria. URL https://www.R-project.org/
- 412 Roeben, V., Oberdoerster, S., Capowiez, Y., Ernst, G., Preuss, T.G., Gergs, A., Oberdoerster, C., 2020. To413 wards a spatiotemporally explicit toxicokinetic-toxicodynamic model for earthworm toxicity. Sci. To414 to 15 diagram (202) 127(72)
- 414 tal Environ. 722, 137673.
- 415 Rogasik, H., Schrader, S., Onasch, I., Kiesel, J., Gerke, H.H., 2014. Micro-scale dry bulk density variation
 416 around earthworm (*Lumbricus terrestris* L.) burrows based on X-ray computed tomography. Ge-
- 417 oderma 213, 471-477.

- 418 Ruiz, S., Straub, I., Schymanski, S.J., Or, D., 2010. Experimental evaluation of earthworm and plant root soil
- 419 penetration-cavity expansion models using cone penetrometer analogs. Vadose Zone J. 15,
 420 vzj2015.09.0126.
- Rushton, S.P., 1986. The effects of soil compaction on *Lumbricus terrestris* and its possible implications for
 populations on land reclaimed from open-cast coal mining. Pedobiologia 29, 85-90.
- Schäffer, B., Stauber, M., Müller, R., Schulin, R., 2007. Changes in the macro-pore structure of restored soil
 caused by compaction beneath heavy agricultural machinery: a morphometric study. Eur. J. Soil Sci.
 58, 1062-1073.
- 426 Schindelin, J., Arganda-Carreras, I., Frise, E., Kaynig, V., Longair, M., Pietzsch, T., Preibisch, S., Rueden, C.,
- 427 Saalfeld, S., Schmid, B., Tinevez, J.-Y., White, D.J., Hartenstein, V., Eliceiri, K., Tomancak, P.,
- 428 Cardona, A., 2012. Fiji: an open-source platform for biological-image analysis. Nat. Methods 9, 676-
- 429 682.
- Shipitalo, M.J., Butt, K.R., 1999. Occupancy and geometrical properties of *Lumbricus terrestris* L. burrows
 affecting infiltration. Pedobiologia 43,782-794.
- Smetak, K.M., Johnson-Maynard, J.L., Lloyd, J.E., 2007. Earthworm population density and diversity in different-aged urban systems. Appl. Soil Ecol. 37, 161-168.
- 434 Söchtig, W., Larink, O., 1992. Effect of soil compaction on activity and biomass of endogeic lumbricids in
 435 arable soil. Soil Biol. Biochem. 24, 1595-1599.
- 436 Stovold, R.J., Whalley, W.R., Harris, P.J., White, R.P., 2004. Spatial variation in soil compaction, and the
- 437 burrowing activity of the earthworm *Aporrectodea caliginosa*. Biol. Fertil. Soils 39, 360-365.

- 439 **Figure legends:**
- 440 **Fig. 1** Examples of 3D burrow systems made by three individuals of either *A. chlorotica* or *A.*
- 441 *caliginosa* for 6 weeks in soil cores with increasing bulk densities.
- 442 **Fig. 2** Volume of burrows (means + SE) created by three individuals of either *A. chlorotica* or *A.*
- 443 caliginosa for 6 weeks in soil cores with increasing bulk densities. Bars bearing different letters are
- 444 significantly different (each species was tested separately).
- 445 Fig. 3 Continuity of the burrow systems (means + SE) made by three individuals of either A. chlo-
- 446 rotica or A. caliginosa for 6 weeks in soil cores with increasing bulk densities. Bars bearing differ-
- 447 ent letters are significantly different (each species was tested separately).
- 448 Fig. 4 Degree of anisotropy of the burrow systems (means + SE) made by three individuals of either
- 449 A. chlorotica or A. caliginosa for 6 weeks in soil cores with increasing bulk densities. Bars bearing
- 450 different letters are significantly different (each species was tested separately).
- 451 Fig. 5 Gain in water infiltration (means + SE) compared to control without earthworms, when three
- 452 individuals of either A. chlorotica or A. caliginosa were incubated for 6 weeks in soil cores with
- 453 increasing bulk densities. Stars indicate values significantly different from zero (each species was
- 454 tested separately).





Bulk density (g cm⁻³)







Table 1. Effects of an increase of soil bulk density on mean (SE) burrow diameter, total length andsurface cast production in soil cores incubated with three individuals of either *Aporrectodea*caliginosa or Allolobophora chlorotica. ANOVA were carried out for each species separately.Values bearing different letters are different at the 5% significance threshold.

	Soil bulk density (g cm ⁻³)					
	1.18	1.23	1.29	1.34	1.38	
A. caliginosa						
Burrow diameter (mm)	2.97 ^a (0.06)	2.96 ^{ab} (0.07)	2.87 ^{ab} (0.07)	2.82 ^{ab} (0.06)	2.65 ^b (0.14)	
Burrow length (cm)	58.82 ^a (5.66)	38.52 ^{ab} (3.52)	34.74 ^b (6.73)	35.02 ^b (5.93)	15.73 ^c (3.95)	
Surface cast production (g)	17.17 ^b (3.65)	15.17 ^b (1.84)	24.36 ^{ab} (4.97)	32.11 ^a (5.43)	24.85 ^{ab} (6.33)	
A. chlorotica						
Burrow diameter (mm)	2.71 ^a (0.09)	2.62 ^a (0.09)	2.36 ^b (0.06)	2.30 ^b (0.06)	2.59 ^b (0.14)	
Burrow length (cm)	73.89 ^a (9.85)	50.00 ^{ab} (4.48)	30.05 ^b (4.08)	21.50 ^b (4.52)	15.02 ^c (6.02)	
Surface cast production (g)	11.12 (2.13)	10.42 (0.78)	8.49 (0.25)	11.59 (2.44)	6.03 (1.88)	

Table 2. Mean (SE) infiltration rates and breakthrough times in the soil cores incubated with three individuals of either *A. caliginosa* or *A. chlorotica* or no earthworm (control) for six weeks. Values bearing different letters are statistically different (each species tested separately).

	Soil bulk density (g cm ⁻³)							
	1.18	1.23	1.29	1.34	1.38			
Infiltration rate $(L h^{-1})$								
A. caliginosa	9.95 ^a (1.58)	3.70 ^b (0.96)	$4.00^{b}(1.00)$	1.66 ^b (0.64)	1.12 ^b (0.52)			
A. chlorotica	8.72 ^a (1.48)	4.91 ^{ab} (1.11)	2.48 ^b (0.79)	1.15 ^b (0.54)	0.84 ^b (0.46)			
Control	4.61 (1.07)	2.50 (0.79)	2.45 (0.78)	1.71 (0.65)	1.04 (0.51)			
Breakthrough times (mn)								
A. caliginosa	0.55 ^b (0.27)	6.89 ^a (1.31)	6.73 ^a (1.29)	9.08 ^a (1.51)	14.28 ^a (1.89)			
A. chlorotica	1.93 ^b (0.69)	2.58 ^b (0.80)	9.60 ^{ab} (1.55)	18.80 ^a (2.17)	12.63 ^a (1.78)			
Control	2.23 (0.75)	6.93 (1.32)	11.35 (1.68)	15.72 (1.98)	16.96 (2.06)			