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Unveiling ecological assembly rules from trait distributions

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3 **1 Letters**

4
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2
3 26 **Author contributions**

4
5 27 NG, YLB-P. and PL developed the original idea. FM developed the model and the coding. NG
6
7 28 ran the simulations. All co-authors contributed to data interpretation. NG and YLB-P wrote the
8
9 29 first draft of the paper with major contribution from all co-authors.
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12
13 31 **Data accessibility statement:** This paper is based on simulations. The authors provide the R
14
15 32 code to generate the simulations with the submitted manuscript. The R code will be archived in
16
17 33 an appropriate public repository with the acceptance of the paper.
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3 43 **ABSTRACT**

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5 44 Deciphering the effect of neutral and deterministic processes on community assembly is critical
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7 45 to understand and predict diversity patterns. The information held in community trait
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9 46 distributions is commonly assumed as a signature of these processes, but empirical and
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11 47 modelling attempts have most often failed to untangle their confounding, sometimes opposing,
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13 48 impacts. Here, we simulated the assembly of trait distributions through stochastic (dispersal
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15 49 limitation) and/or deterministic scenarios (environmental filtering, niche differentiation). We
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17 50 characterized the shape of trait distributions through the skewness-kurtosis relationship. We
18
19 51 identified commonalities in the co-variation between the skewness and the kurtosis of trait
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21 52 distributions with a unique signature for each simulated assembly scenario. Our findings were
22
23 53 robust to variation in the composition of regional species pools, dispersal limitation, and
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25 54 environmental conditions. While ecological communities can exhibit a high degree of
26
27 55 idiosyncrasy, identification of commonalities across multiple communities can help to unveil
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29 56 ecological assembly rules in real-world ecosystems.
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34
35 58 **Key words:** dispersal limitation, community assembly rules, environmental filtering,
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37 59 functional diversity, niche differentiation, skewness-kurtosis relationship, stochasticity, trait
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39 60 distributions
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61 INTRODUCTION

62 A basic tenet in community ecology is that trait diversity can reveal the influence of
63 deterministic processes on species assemblages (e.g., competition and abiotic factors; McGill
64 *et al.* 2006; Weiher *et al.* 2011). Central to this research agenda is the hypothesis that high trait
65 diversity reflects niche differentiation processes, while low trait diversity reflects the effect of
66 environmental filtering selecting for species with similar trait values (Cornwell & Ackerly
67 2009; Maire *et al.* 2012; Keddy 1992; Grime 2006). Much research has been devoted to analyze
68 how patterns of trait diversity vary within and among communities (e.g., convergence or
69 divergence) with the aim to uncover general assembly rules that could apply across many
70 ecosystems (Diamond 1975; Weiher & Keddy 2001). However, hypothesizing a direct and
71 unequivocal linkage between patterns of trait diversity and assembly processes has proved too
72 simplistic and fed much debate (Weiher *et al.* 2011; Götzenberger *et al.* 2012; Münkemüller *et*
73 *al.* 2020).

74 Multiple assembly processes are likely to simultaneously influence local trait diversity,
75 sometimes in opposite directions (Mayfield & Levine 2010; Maire *et al.* 2012), making trait
76 diversity patterns often difficult to distinguish from randomness (Götzenberger *et al.* 2012;
77 Munoz & Huneman 2016). The signature of deterministic processes can be blurred by stochastic
78 processes such as demographic drift, or contingent variation in the regional species pool
79 composition (Hubbell 2001; Spasojevic *et al.* 2018). Furthermore, most trait-based approaches
80 have ignored the effect of species dispersal among communities (Spasojevic *et al.* 2014).
81 Dispersal limitation can strongly reduce species richness (Ricklefs 1987), increasing the imprint
82 of local demographic stochasticity (Leibold *et al.* 2004), and thereby yielding apparent
83 randomness in community-level patterns (Götzenberger *et al.* 2012; Munoz & Huneman 2016).
84 Deciphering how multiple stochastic and deterministic processes shape trait diversity is not
85 only crucial to expand our fundamental understanding of biodiversity patterns, but also to better
86 predict the response of communities and ecosystems to ongoing environmental changes.

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3 87 Trait diversity is usually assessed using synthetic indices derived from the distribution
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5 88 of trait values within communities (e.g., Mason *et al.* 2005; Laliberté & Legendre 2010; Enquist
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7 89 *et al.* 2015; Carmona *et al.* 2016). Most studies focus on the mean and variance/dispersion of
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9 90 trait distributions (Violle *et al.* 2012), or on related indices (e.g., FDis, Laliberté & Legendre
10
11 91 2010). Considering that mean and variance suffice for describing trait diversity relies on the
12
13 92 implicit assumption that trait values follow (approximately) a normal distribution. In essence,
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15 93 the normal distribution shows a bell shape, whose mean represents a local optimum that
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17 94 matches a given environment and the variance represents how trait values are constrained
18
19 95 around this mean (Enquist *et al.* 2015, Fig. 1a). However, trait distributions often deviate from
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21 96 a normal distribution in real-world communities, and can exhibit asymmetric, flat, peaky or
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23 97 multimodal shapes (Fig. 1b) (Enquist *et al.* 2017; Le Bagousse-Pinguet *et al.* 2017). In such
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25 98 cases, further information on the shape of trait distributions is required to characterize trait
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27 99 diversity.

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31 100 The shape of trait distributions can be quantified by the skewness and the kurtosis, which
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33 101 quantify the asymmetry and the evenness of the trait distributions, respectively. The skewness
34
35 102 and the kurtosis are increasingly used in ecological research because they can provide insights
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37 103 on how species assemble within communities, and how they respond to ongoing environmental
38
39 104 changes (Kraft *et al.* 2008; Enquist *et al.* 2015; Gross *et al.* 2017; Wiczyński *et al.* 2019). For
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41 105 instance, rapid environmental changes can simultaneously increase the skewness and kurtosis
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43 106 of trait distributions when a limited portion of the community with specific trait values benefit
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45 107 from the environmental change, and the recruitment of new species adapted to novel conditions
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47 108 is not immediate (see predictions in “Trait Driver Theory” in Enquist *et al.* 2015, 2017).
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49 109 Conversely, niche differentiation should decrease the kurtosis of trait distributions, yielding
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51 110 flatter or even bimodal distributions by promoting the coexistence of functionally contrasting
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53 111 species (Cornwell & Ackerly 2009; Maire *et al.* 2012).

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3 112 The skewness and the kurtosis are mathematically constrained through the Skewness-
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5 113 Kurtosis Relationship (SKR hereafter), which can be used to characterize a broad spectrum of
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7 114 distributions (Box 1) (Cullen & Frey 1999). Across a variety of dryland plant communities,
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9 115 Gross et al. (2017) reported that the distributions of plant height and specific leaf area followed
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11 116 non-random SKRs. These empirical SKRs suggested the existence of general assembly rules
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13 117 observable at the biome scale where traits organized according to specific families of trait
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15 118 distributions (see Box 1 and Fig. 1c;d for a definition of family of trait distributions). Yet, we
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17 do not know which ecological processes underpin such empirical assembly rules, how different
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19 120 deterministic and stochastic processes modulate SKR patterns, and whether the signature of
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21 121 these processes are distinguishable from one another. A theoretical evaluation of the SKR
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23 122 framework is therefore needed to assess the effects of multiple stochastic and deterministic
24
25 123 assembly processes on trait distributions, and more generally, to advance our ability to identify
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27 124 assembly rules from patterns of trait distributions.

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31 125 In a theoretical experiment, we simulated the effects of multiple stochastic and
32
33 126 deterministic processes on trait diversity patterns. Using the *ecolottery* package in R language
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35 127 (Munoz *et al.* 2018), we generated 800,000 communities spanning a broad spectrum of
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37 128 ecological processes and of resulting trait distributions. Specifically, we considered four
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39 129 community assembly scenarios: (i) a purely neutral scenario including stochastic processes
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41 130 only; and (ii) three different “trait-based” scenarios, each combining stochastic processes with
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43 131 a distinct outcome of deterministic processes entailing either convergence or divergence in trait
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45 132 distributions (Loranger *et al.* 2018). Our goals were:

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48 133 (1) to assess the extent to which the SKR and standard metrics of trait diversity capture
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50 134 distinguishable signatures of assembly processes;

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53 135 (2) to evaluate how the importance of stochastic processes (through the manipulation of
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55 136 the regional species pool richness, and dispersal limitation among communities) and

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3 137 environmental variations across communities affect our ability to identify unambiguous
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5 138 signatures of the four community assembly scenarios.

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8 9 140 **METHODS**

10 11 141 **Community assembly scenarios**

12
13 142 *ecolottery* is a modeling platform simulating the assembly of ecological communities (Munoz
14
15 143 *et al.* 2018). It relies on a spatially implicit framework in which communities are assembled
16
17 144 from an external pool of potential immigrants (e.g., a regional species pool) through dispersal.
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19 145 Stochastic and deterministic processes can be parameterized in *ecolottery*. They both influence
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21 146 the establishment success of immigrants and survival of their descendants in the model
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23 147 (Loranger *et al.* 2018; Munoz *et al.* 2018). Stochastic processes influence local diversity
24
25 148 through demographic drift, dispersal limitation, and species richness in the regional pool
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27 149 (Etienne & Alonso 2005). Deterministic processes modulate the success of immigrants based
28
29 150 on how their trait values allow establishment and persistence in local environments.
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33 151 We simulated three different outcomes of deterministic processes in *ecolottery*
34
35 152 henceforth “trait-based filtering”. The first trait-based filtering operates around a single optimal
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37 153 value defined by its matching with the local environmental conditions (“stabilizing filtering”
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39 154 hereafter), consistently with the classical view of the environmental filtering (Keddy 1992;
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41 155 Kraft *et al.* 2015). Stabilizing filtering typically generates normal or more leptokurtic
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43 156 distributions (e.g. hyperbolic distributions, Fig. 1 a,b), depending on the strength of the filtering
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45 157 (Enquist *et al.* 2015; Le Bagousse-Pinguet *et al.* 2017). The second trait-based filtering
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47 158 generates assemblages where alternative optimal trait values can confer greater species
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49 159 performance (“disruptive filtering”; Rolhauser & Pucheta 2017; Loranger *et al.* 2018). This
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51 160 trait-based filtering produces patterns that reflect either niche differentiation among
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53 161 functionally contrasting species (Cornwell & Ackerly 2009; Maire *et al.* 2012; Rolhauser &
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3 162 Pucheta 2017). Disruptive filtering typically produces uniform or bimodal distributions (Fig.
4
5 163 1b). Finally, the third trait-based filtering generates assemblages in which species performance
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7 164 varies monotonically with trait values (“Directional filtering”). This produces asymmetric trait
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9 165 distributions (Loranger *et al.* 2018) (e.g. exponential distribution in Fig. 1b), e.g. in the case of
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11 166 asymmetric light competition (Schamp *et al.* 2007) or directional environmental changes
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13 167 (Enquist *et al.* 2015).
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169 **In silico assembly experiment**

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20 170 We considered four theoretical scenarios to assemble communities: (i) a purely neutral scenario
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22 171 that only considers the effect of stochastic processes on assembly without influence of trait
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24 172 differences on species performance and on trait distribution in assemblages (Hubbell 2001);
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26 173 and (ii) three contrasting scenarios in which deterministic processes (aforementioned trait-
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28 174 based filtering) are combined with stochastic processes. We modeled the four scenarios using
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30 175 different values of the parameter “*filt*” in *ecolottery* to represent the “neutral”, “stabilizing”,
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32 176 “disruptive” and “directional” scenarios (see R code in Appendix S1 in Supporting
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34 177 Information). In addition, we draw species trait values i between -3 to +3.
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37 178 In a first simulation experiment, we simulated all communities with the same
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39 179 environmental conditions (“fixed environment”, hereafter), under low or high dispersal
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41 180 limitation. Dispersal limitation was manipulated using a migration parameter “ m ” that was set
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43 181 to either low ($m = 0.95$) or high dispersal limitation ($m = 0.05$). In a second simulation
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45 182 experiment, we simulated communities along an environmental gradient. The environmental
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47 183 gradient was simulated by adjusting the parameter “*Env*” from -2 to + 2, corresponding to the
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49 184 two extremes of a gradient (e.g., from cold to hot environments), and where 0 represented the
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51 185 mild environment (see Table S1 for model parameters). We chose the range of environmental
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53 186 variation “*Env*” (ranging from -2 to + 2) to be lower than the range of possible trait values
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3 187 (ranging from -3 to +3) to avoid edge effects (Denelle *et al.* 2019). We set “*m*” to either low or
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5 188 high dispersal limitation as explained in the first simulation experiment.

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7 189 The total number of individuals in the external species pool was set to 25,000. This pool
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9 190 delineates the ‘regional species pool’ in our simulations. To evaluate how the richness of the
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11 191 species pool influenced community assembly, the pool of immigrants varied from 10 to 500
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13 192 species. Richness at the regional level was fixed for each simulation run. The total number of
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15 193 individuals per community was set to 250 (Table S1), and we simulated 100 communities in
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17 194 each run. We chose these parameter values to approximate “realistic” vegetation sampling in
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19 195 the context of dryland (Gross *et al.* 2017, Table S1). Within communities, the number of
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21 196 individuals per species could vary according to the simulated assembly processes. The trait
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23 197 values of the regional species pool followed a uniform distribution, such that all trait values had
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25 198 equal probability of being selected. We randomly assigned trait values to each species. During
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27 199 this procedure, the model allowed different species to exhibit similar trait values, therefore
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29 200 allowing functional redundancy to occur among species (Munoz *et al.* 2018). For simplicity,
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31 201 we did not consider intraspecific trait variability in our simulations. Conspecific individuals
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33 202 thus displayed the same trait value.

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38 39 204 **Analysis of simulated trait distributions**

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41 205 We generated 800,000 communities: 500 runs where species richness varied from 10 to 500
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43 206 species in the species pool \times 100 communities per run \times 4 scenarios \times 2 levels of dispersal
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45 207 limitation (low vs. high) \times 2 environmental contexts (a fixed environment vs. an environmental
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47 208 gradient). For each run, we recorded the total number of species in a given community (species
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49 209 richness). We quantified species relative abundance as the relative frequency of each species in
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51 210 each community (i.e., the number of individual of a species divided by the total number of
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53 211 individual in the community).

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3 212 We calculated the four moments of the trait distribution associated to each simulated
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5 213 community: the mean, variance, skewness and kurtosis. Then, we calculated the parameters of
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7 214 the SKRs across the 100 simulated distributions obtained from each run (R^2 , slope, Y-intercept
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9 215 and the distance to the lower boundary; see Box 1). For each community, we also computed
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11 216 classical indices of trait diversity using the dbFD function in the “FD” package in R [functional
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13 217 dispersion (FDis), the Rao index, functional evenness (FEve); Laliberté and Legendre 2010].
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15 218 The four distribution moments and **all trait diversity metrics were abundance-weighted**. While
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17 219 the SKR parameters were estimated across the 100 communities generated by each simulation
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19 220 run, the four distribution moments, and taxonomic and trait diversity indices were calculated at
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21 221 the community level, and thus generated 100 values for each run.

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24 222 We evaluated whether the parameters of the SKRs, the four distribution moments and
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26 223 all diversity metrics can be used to discriminate assembly processes from trait distributions. For
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28 224 instance, lower variance, FDis and Rao values compared to the neutral scenario are expected
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30 225 for the stabilizing and directional filtering, i.e., trait convergence. Conversely, higher variance,
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32 226 FDis and Rao values compared to the neutral scenario are expected for the disruptive filtering,
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34 227 i.e., trait divergence. To test for significant differences among scenarios, we used the “overlap”
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36 228 package in R (Ridout & Linkie 2009). For each diversity metric, we calculated the overlap
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38 229 between each pair of scenarios and the mean overlap considering all scenarios together. Overlap
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40 230 values ranged from 0 to 1. An overlap < 0.05 indicated that the metric under consideration
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42 231 significantly discriminate assembly scenarios.

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47 233 **RESULTS**

48 234 **Discriminating scenarios using functional diversity metrics**

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51 235 The neutral scenario generated a wide range of trait diversity patterns, which challenges the
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53 236 inference of ecological processes from trait distributions (Figs. 2, 3). Significant differences
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3 237 among scenarios were observed under low dispersal limitation. For instance, we observed
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5 238 significant differences between the stabilizing and the neutral scenarios for FDis and for Rao
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7 239 under low dispersal limitation (Figs. 2, 3). However, under high dispersal limitation, all trait
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9 240 diversity indices - with the exception of the variance - simulated under the neutral scenario
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11 241 systematically overlapped with those of the deterministic scenarios, meaning a lack of
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13 242 differences among them.

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15 243 The variance of trait distributions best discriminated assembly scenarios. We observed
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17 244 a low mean overlap (M_o) when dispersal limitation was low and under fixed environment (M_o
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19 245 = 0.01, Fig. 2a, left panel). The stabilizing and directional scenarios exhibited significantly
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21 246 lower variances than those of the neutral scenario. However, the variance could not discriminate
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23 247 the stabilizing from the disruptive and directional scenarios, and the neutral from the disruptive
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25 248 scenario along an environmental gradient combined with high dispersal limitation ($M_o = 0.16$,
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27 249 Fig. 2b). Overall, our results showed limited ability of these metrics of trait diversity to identify
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29 250 unambiguous signatures of the four community assembly scenarios.

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34 35 252 **SKR parameters simulated in a fixed environment**

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37 253 Each studied scenario generated contrasting SKR patterns under fixed environmental
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39 254 conditions and low dispersal limitation (Fig. 4, see also SKR plots in Fig. S1). The neutral and
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41 255 stabilizing scenarios both generated weaker SKRs than the disruptive and directional scenarios
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43 256 (significantly lower R^2 , Fig. 4b). These first two scenarios produced distributions that clustered
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45 257 in the skewness-kurtosis space, and tended to overall converge toward a single skewness-
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47 258 kurtosis coordinate as species richness in the regional pool increased (from dark to bright blue
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49 259 dots, Fig. 4a). In contrast, both the directional and the disruptive scenarios yielded strong SKRs,
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51 260 i.e. characterized by significantly higher R^2 than neutral and stabilizing scenarios (Fig. 4b).
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53 261 Specifically, the disruptive scenario generated SKRs with the highest R^2 , ranging from 0.6 to
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3 262 0.99 (Fig. 4b), indicating that all communities assembled under this scenario belong to a unique
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5 263 family of trait distributions (skew-bimodal distribution, Fig. 1c, d).
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7 264 The SKRs generated from the four scenarios had also clear and significantly distinct Y-
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9 265 intercepts (Fig. 4b). Therefore, the SKR not only discriminated the three scenarios containing
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11 266 deterministic processes from a purely neutral scenario, but also clearly differentiated these three
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13 267 scenarios from one another. The neutral scenario had a Y-intercept = 1.86 (Fig. 4b) suggesting
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15 268 that it converged toward uniform distribution (Fig. 1b). This pattern of trait diversity observed
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17 269 at the Y-intercept of the SKR thus mirrored the trait distribution of the regional pool. The
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19 270 stabilizing scenario produced communities that converged toward a normal distribution (Y-
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21 271 intercept = 2.82 Fig. 4b). This scenario generated a peakier and therefore less even trait
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23 272 distribution than the neutral one. The directional scenario generated distributions that were
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25 273 mostly asymmetric, and that converged toward a skewness value ~ 1 as species richness
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27 274 increased (Fig. 4a) and a Y-intercept = 2.1 (Fig. 4b). The disruptive scenario had the lowest Y-
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29 275 intercept = 1.33 (Fig. 4b). This scenario generated a family of trait distributions approaching
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31 276 the lower boundary of the skewness-kurtosis space, corresponding to highly platykurtic
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33 277 bimodal distributions (see the skew-bimodal distribution in Fig. 1c). Therefore, the disruptive
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35 278 scenario was the one with the highest trait diversity in our simulations.
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39 279 Increasing the influence of stochasticity, through high dispersal limitation, reduced the
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41 280 species richness at the community level (Fig. S2). It consistently increased the asymmetry of
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43 281 trait distributions for all scenarios, thereby increasing the scatter of distributions in the
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45 282 skewness-kurtosis space (Fig. 4a). Higher scattering of skewness and kurtosis values produced
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47 283 SKRs with higher R^2 for all scenarios (Fig. 4b). For instance, the R^2 of the neutral scenario
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49 284 ranged from 0.25 to 0.75 (Fig. 4b), indicating that significant SKRs can be observed and
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51 285 produced by a high stochasticity even under a strictly neutral scenario. Yet, irrespective of the
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53 286 degree of stochasticity, the disruptive filtering always produced the strongest SKRs (Fig. 4b).
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3 287 Furthermore, increasing dispersal limitation marginally influenced the Y-intercept of the SKRs.
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5 288 The mean overlap among scenarios remained extremely low under high dispersal limitation (Y-
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7 289 intercept: $M_o = 0.03$, Fig. 4b). We observed significant differences between the stabilizing,
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9 290 neutral and disruptive scenarios, although the directional scenario overlapped with the neutral
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11 291 one (overlap = 0.18). These results suggested that the SKR approach is therefore able to
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13 292 differentiate a neutral assembly scenario from the effect of environmental filtering (stabilizing
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15 293 scenario) and niche differentiation (disruptive scenario) even under a high stochasticity.
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20 295 **SKR parameters simulated along an environmental gradient**

22 296 Changing environmental conditions across communities affected local mean trait value for all
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24 297 the studied scenarios except the neutral one (Fig. S3). Furthermore, the variance was constant
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26 298 and low for the stabilizing scenario while all moments varied under the disruptive scenario (Fig.
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28 299 S3). For the disruptive scenario, increasing the environmental constraint (by changing the *Env*
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31 300 parameter from 0 to -2 or from 0 to +2) reduced the variance and increased the skewness and
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33 301 the kurtosis (Fig. S3a). Nevertheless, the SKR parameters calculated across communities along
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35 302 the gradient, and particularly the Y-intercept, consistently produced similar patterns than those
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37 303 observed under a fixed environment with significant differences among scenarios (Y-intercept
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39 304 $M_o = 0.005$ with low dispersal limitation, Y-intercept $M_o = 0.04$ with high dispersal limitation
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41 305 Fig. 5b). The R^2 of the SKRs also differed among scenarios with the stabilizing scenario
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43 306 showing the lowest range of R^2 and the disruptive scenario showing the highest R^2 (Fig. 5b).
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48 308 **DISCUSSION**

50 309 **Commonalities in trait distributions across multiple communities unveil assembly rules**

52 310 Our reference neutral scenario generated a wide range of trait diversity patterns, following the
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54 311 view that “*simple stochastic models can reproduce natural diversity patterns*” (Hubbell 2005).
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3 312 Likewise, deterministic assembly scenarios generated a wide range of trait distributions
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5 313 (Enquist *et al.* 2015; Le Bagousse-Pinguet *et al.* 2017). The ambiguity among scenarios further
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7 314 increased in our simulations as stochasticity increased or when the environmental conditions
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9 315 changed (e.g., sampling performed along environmental gradients). Therefore, our study
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11 316 therefore demonstrates that distinguishing deterministic from purely neutral patterns based on
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13 317 standard trait-based metrics, or by the four moments of trait distributions taken separately, is
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15 318 particularly challenging.

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18 319 The SKR framework provided far less ambiguous insights into underlying assembly
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20 320 processes, as contrasting assembly scenarios exhibited distinct SKR patterns. The SKR
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22 321 approach did not only well discriminate the neutral from trait-based filtering scenarios, but also
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24 322 clearly differentiated each of the filtering scenarios from one another, irrespective of the degree
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26 323 of dispersal limitation, environmental condition or the size of the species pool. The SKR
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28 324 approach appears therefore as a promising tool to unveil assembly processes from trait
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30 325 distributions, even under high stochasticity, a yet important limitation of existing approaches
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32 326 aiming at interpreting trait diversity patterns of real-world communities (Weiher *et al.* 2011;
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34 327 Münkemüller *et al.* 2020).

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36
37 328 The SKR framework shifts our perspective to focus on commonalities in trait
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39 329 distributions - rather than differences – to unveil assembly rules. For instance, the disruptive
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41 330 filtering generated a set of symmetric and asymmetric distributions. Under this scenario,
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43 331 increasing environmental constraints simultaneously reduced the variance and increased both
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45 332 the skewness and the kurtosis of trait distributions (Enquist *et al.* 2015; Le Bagousse-Pinguet
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47 333 *et al.* 2017), consistently with the environmental filtering hypothesis (Keddy 1992). However,
48
49 334 the trait distributions all aligned along a specific SKR. The observed SKR pattern thus described
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51 335 a family of trait distributions, in which each single distribution represents an instance of a more
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53 336 general trait distribution operating at larger spatial scale (e.g., at the biome scale in Gross *et al.*

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3 337 2017). All distributions within this family shared a common property: trait distributions are
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5 338 bimodal (Fig. S5), thus reflecting the effect of niche differentiation processes promoting the
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7 339 coexistence of functionally contrasting species (Maire *et al.* 2012). In other words, even if
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9 340 different environmental conditions along a gradient (and stochastic processes) can affect the
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11 341 shape parameters of distributions by increasing the skewness and the kurtosis separately, a
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13 342 bimodality persists as the signature of niche differentiation. **By characterizing commonalities**
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15 343 **in trait distributions, the SKR represents a critical step toward the identification of general rules**
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17 344 **governing the functional organization of species assemblages.**

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19
20 345 By identifying families of trait distributions, the SKR framework scales up trait
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22 346 distributions from the community to broader geographical scales (e.g. landscape / regional /
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24 347 continental / biome scales). The SKR framework differs from other trait-based approaches
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26 348 aiming at addressing biogeographical questions by aggregating trait patterns across spatial
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28 349 scales (Violle *et al.* 2014; Carmona *et al.* 2016). While the SKR framework scales up trait
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30 350 distributions from local to wider geographical scales, it also keeps the information on the shape
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32 351 of trait distributions observed at the community scale. In other words, the SKR framework
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34 352 increases the spatial extent by quantifying trait distributions at large spatial scale but does not
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36 353 reduce the grain size. Therefore, the SKR framework represents an opportunity to merge
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38 354 community ecology and biogeography and to “*achieve a deeper understanding of biodiversity*
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40 355 *and its change across communities*” (Chase *et al.* 2018).

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44 356 The skewness and the kurtosis of trait distributions are increasingly used to detect the
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46 357 short term effects of global change on communities, and may be used as a signal of
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48 358 compositional changes (Enquist *et al.* 2017; Griffin-Nolan *et al.* 2019; Wieczynski *et al.* 2019;
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50 359 Zhang *et al.* 2019; Aguirre-Gutiérrez *et al.* 2020). **Yet, variations in skewness and kurtosis are**
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52 360 **often considered independently, i.e. without accounting for their mathematical dependency**
53
54 361 **(Box 1). Observing a decrease in trait evenness (higher kurtosis) may entirely rely on changes**

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3 362 in the degree of skewness. Accounting for the SKR when assessing the impact of global change
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5 363 on ecological communities should provide additional insights on how the ecological reassembly
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7 364 under environmental change, and may provide a more robust validation of theoretical
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9 365 predictions (e.g. the “Trait Driver Theory”: Enquist *et al.* 2015, 2017). Focusing on the distance
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11 366 to the lower boundary of the SKR (see Box 1) (Liu *et al.* 2020) or the deviation of individual
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13 367 communities from empirical to random SKR (Gross *et al.* 2017) – rather than raw variations in
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15 368 kurtosis - may help diagnosing shifts in community evenness and community disassembly
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18 369 under ongoing environmental change.
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21 22 371 **Linking theoretical predictions to empirical patterns of trait diversity**

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24 372 Our simulations suggest that niche differentiation and dispersal limitation jointly shape the trait
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26 373 diversity in real-world communities. Combining the disruptive scenario with high dispersal
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28 374 limitation was the only case that generated a strong family of trait distributions (i.e., a SKR
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30 375 with a $R^2 > 0.90$). This family of trait distributions exhibited lower kurtosis than under the
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32 376 neutral scenario, a pattern similar to the empirical SKRs observed in drylands worldwide (Gross
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34 377 *et al.* 2017). When simulating along an environmental gradient, the disruptive scenario also
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36 378 reproduced the impact of environmental filtering by reducing variance and increasing the
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38 379 kurtosis within communities (Fig. S3), and the shift in dominance of contrasting functional
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40 380 groups (e.g. the shrub-to-grass shift in dominance commonly observed in drylands; Fig. S4,
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42 381 Bestelmeyer *et al.* 2018). Altogether, our results corroborate the view that dryland communities
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44 382 are shaped by a combination of these multiple drivers.
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48 383 The modelling platform *ecolottery* provides a simple and flexible environment to
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50 384 simulate the effect of multiple stochastic and deterministic processes on trait distributions
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52 385 (Munoz *et al.* 2018). Further implementations are needed to provide reliable and quantitative
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3 386 predictions when comparing theoretical predictions with observed trait distributions in real-
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5 387 world communities:

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7 388 First, different assembly processes have been shown to act differently on traits
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9 389 describing contrasted axes of functional specialization (e.g. size vs. resource use related-traits,
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11 390 Kraft *et al.* 2008; Cornwell & Ackerly 2009; Maire *et al.* 2012). On the one hand, the disruptive
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13 391 scenario may well apply to resource-use traits as contrasting strategies for resources acquisition
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15 392 and utilization may favor species coexistence (Maire *et al.* 2012). On the other hand, the
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17 393 directional scenario could constitute a valuable hypothesis for size-related traits in light limited
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19 394 environments where asymmetric competition predominates (Schamp *et al.* 2007; Gross *et al.*
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21 395 2009). Considering the effect of multiple assembly processes acting on independent functional
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23 396 dimensions would allow a more realistic representation of community assembly and species
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25 397 coexistence (Maire *et al.* 2012).

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28 398 Second, the distributions of trait values at large spatial scale (e.g. at the scale of the
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30 399 regional species pool) are important drivers determining local assembly processes and trait
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32 400 diversity (Carmona *et al.* 2016; Le Provost *et al.* 2017; Spasojevic *et al.* 2018). To evaluate
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34 401 how regional trait pools influence SKR patterns, we conducted a sensitivity analysis using
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36 402 different trait distributions at the regional scale (see Fig. S5 for details). Our results show that
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38 403 the regional trait pool influenced the SKR parameters for each scenario. It supports the need to
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40 404 account for the regional trait pool to provide reliable predictions on trait diversity within and
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42 405 across communities (Carmona *et al.* 2016; Spasojevic *et al.* 2018). Nevertheless, we also
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44 406 observed that the relative differences among contrasting assembly scenarios remained
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46 407 consistent despite variation in the composition of the regional trait pool, highlighting the
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48 408 robustness of the SKR approach.

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51 409 Third, downscaling from a global / regional trait pool to the local community scale may
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53 410 require the development of a spatially-explicit framework, as contrasting assembly processes
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3 411 may act simultaneously but at different spatial scale (Keddy 1992; Spasojevic *et al.* 2014; Le
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5 412 Provost *et al.* 2017). Such framework would allow evaluating how environmental filtering
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7 413 acting at broad spatial scale (Kraft *et al.* 2015; Le Bagousse-Pinguet *et al.* 2017) interacts with
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9 414 local biotic processes (Maire *et al.* 2012), e.g. by simulating the stabilizing and disruptive
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11 415 scenarios within a single framework but acting at contrasted spatial scale. In this context, further
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13 416 implementations of the model would allow for a more explicit representation of biotic
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15 417 interactions. For instance, the disruptive scenario considers a selection of two optima due to
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17 418 alternative suitable strategies (e.g., coexisting guilds, Cornwell & Ackerly 2009; Maire *et al.*
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19 419 2012). However, the disruptive scenario does not account for interactions among individual
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21 420 species through density-dependent mechanisms (MacArthur & Levins 1967) as well as for other
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23 421 types of biotic interactions (e.g., facilitation and indirect interactions) that have been shown to
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25 422 impact trait diversity in plant communities (Gross *et al.* 2009; Saiz *et al.* 2019).
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31 424 **Conclusions**

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33 425 Here we show that the skewness-kurtosis relationship (SKR) offers a powerful mean to evaluate
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35 426 the effect of multiple stochastic and deterministic processes on the assembly of ecological
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37 427 communities. By focusing on the co-variation between the skewness and the kurtosis across
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39 428 multiple communities, the SKR framework identifies commonalities in the shape of the trait
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41 429 distributions that can serve as a basis to infer assembly processes. The SKR framework is robust
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43 430 to stochastic processes such as variation in regional species and trait pools, dispersal limitation.
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45 431 Finally, our simulations suggest that the joint effect of local-scale processes such as niche
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47 432 differentiation and regional-scale dispersal limitation can have key implications for shaping
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49 433 biodiversity. Our study offers promising avenue for identifying ecological assembly rules in
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51 434 real-world ecosystems.
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18 590 **SUPPORTING INFORMATION**

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BOX 1: The Skewness-Kurtosis Relationship (SKR)

We examined whether the relationship between the skewness and the kurtosis of trait distributions (Fig. 1a) can help deciphering the signatures of contrasting assembly processes. The approach is inspired by optimization procedures increasingly used in physics, climatology and economy (e.g. Cristelli *et al.* 2012). We apply the approach to diagnose assembly rules from trait distributions (the Skewness-Kurtosis Relationship [SKR] approach). While the mean and variance reflect the location and the scale of a distribution (the latter being the dispersion of trait values within a community), the skewness and kurtosis inform on its shape. The degree of skewness quantifies the asymmetry of a given distribution. For instance, a skew distribution indicates the dominance of extreme trait values (see exponential distribution in Fig. 1b), which can typically arise from asymmetric competition for light (Schamp *et al.* 2007). The kurtosis quantifies the relative peakiness of a trait distribution and the relative density of its tails. Low kurtosis values reflect an even distribution of trait values within a given community, a definition of a high trait diversity (Gross *et al.* 2017). Low Kurtosis may reflect the coexistence of functionally contrasting species (see uniform and bimodal distributions in Fig. 1b) (Enquist *et al.* 2017; Gross *et al.* 2017). In contrast, peaked distributions characterized by high kurtosis value reflect a low trait diversity, and may typically occur under strong environmental filtering (*sensu* Keddy 1992) selecting for a limited range of trait value (see hyperbolic distribution in Fig. 1b).

Skewness (S) and kurtosis (K) are related through the following inequality (Fig. 1b):

$$K \geq \beta S^2 + \alpha \quad (1)$$

This inequality generates a mathematically constrained triangle in which all possible trait distributions can be represented and characterized (Gross *et al.* 2017), i.e. the skewness-kurtosis space. For instance, the normal distribution is defined by a unique combination of skewness and kurtosis values of 0 and 3 respectively. It can therefore be represented as a single coordinate

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3 623 in the skewness-kurtosis space (red dot, in Fig. 1d). Families of trait distributions can also be
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5 624 represented by a skewness-kurtosis relationship (SKR, hereafter) with a slope β and a Y-
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7 625 intercept α . A SKR implies that when trait distributions become more skewed, they also become
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9 626 more peaked, resulting in a decrease in evenness. The slope β of the SKR measures the strength
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11 627 of the relationship, i.e., the extent to which evenness decreases as trait distributions become
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13 628 more skewed. The Y-intercept α indicates the lowest kurtosis value at skewness = 0, and
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15 629 corresponds to the highest trait diversity predicted by a given SKR. Distributions belonging to
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17 630 a family of distributions share common properties. For instance, in the case of a skew-bimodal
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19 631 distribution (Fig. 1c, green dashed line in Fig. 1d), all distributions are bimodal although their
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21 632 degree of skewness and kurtosis can vary across communities. This would be the case when
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23 633 two distinct functional groups coexist within communities (e.g. grass and shrub species) but
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25 634 their relative abundance can vary across communities.

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29 635 The inequality (1) has a lower boundary that sets a limit to the minimal kurtosis value
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31 636 predicted for any degree of skewness, i.e. the potential maximum trait diversity for a given
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33 637 skewness (black dash line, $K = S^2 + 1$, Fig. 1b) (see Gross et al. 2017 for a mathematical
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35 638 demonstration). The distance to the lower boundary for a given distribution - exemplified with
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37 639 the black arrow in the case of the exponential distribution (Fig. 1b) - thus quantifies the extent
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39 640 to which trait diversity departs from the potential maximum trait diversity independently from
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41 641 the degree of skewness. Although skewness and kurtosis individually provide valuable
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43 642 information on community trait distributions, the SKR approach helps to diagnose complex
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45 643 trait distributions (Cullen & Frey 1999) and to reveal the extent to which trait diversity is
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47 644 maximized within communities. Applying the SKR framework to ecological communities may
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49 645 allow identifying assembly rules through the identification of commonalities in the shape of the
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51 646 trait distributions observed across multiple communities.

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3 648 **Figures Legend:**

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5 649 **Figure 1** Characterizing complex trait distributions using the Skewness-Kurtosis Relationship
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7 650 (SKR). **(a)** A trait distribution is a density function representing the relative frequency or
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9 651 abundance of trait values within a community. Examples of trait distributions include normal
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11 652 (panel a), uniform, bimodal, hyperbolic or exponential distributions (panel b). (c) Example of
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13 653 a family of distributions: a skew-bimodal distribution (see results, and Box 1). (d) The
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15 654 Skewness-Kurtosis space. Trait distributions are characterized by distinct skewness (S) -
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17 655 kurtosis (K) coordinates. Families of trait distributions can be characterized by a specific
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19 656 Skewness-Kurtosis Relationship (SKR) such as $K = \beta S^2 + \alpha$. The distance to the lower boundary
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21 657 (i.e. potential minimum kurtosis value) of a given distribution (exemplified with the black arrow
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23 658 in the case of the exponential distribution) quantifies the extent to which trait evenness is
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25 659 maximized.
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31 661 **Figure 2** The four moments of the trait distributions simulated under the four theoretical
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33 662 scenarios (Neutral [Neu], Stabilizing [Sta], Disruptive [Dis], Directional [Dir]) under **(a)** fixed
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35 663 environment and varying species pools, **(b)** changing environment and fixed species pool, and
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37 664 under low / high dispersal limitation. We represent the mean, variance, skewness and kurtosis
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39 665 of the trait distributions simulated under each scenario using violin plots. For each panel, we
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41 666 provide the mean overlap (Mo) among the four moments of the trait distributions. Different
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43 667 letters indicate significant differences between scenarios (overlap < 0.05; NS for Not-
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45 668 Significant).
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51 670 **Figure 3** Species richness (Sp. Rich.) and commonly-used trait diversity indices (FDis, Rao,
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53 671 FEve) simulated under the four theoretical scenarios (Neutral [Neu], Stabilizing [Sta],
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55 672 Disruptive [Dis], Directional [Dir]) in **(a)** Fixed environment varying species pools, **(b)**
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3 673 changing environment fixed species pool, and under low / high dispersal limitation. We
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5 674 represent each predicted parameter using violin plots. For each panel, we provide the mean
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7 675 overlap (Mo) between the four-parameter distributions. Different letters indicate significant
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9 676 differences between scenarios (overlap < 0.05; but NS for Not-Significant).

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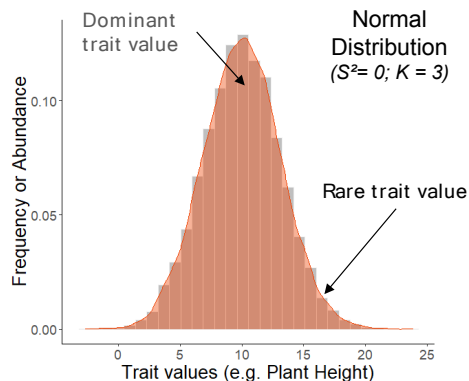
12
13 678 **Figure 4** Effect of the different assembly scenarios on the Skewness-Kurtosis Relationship
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15 679 (SKR) in a fixed environment and varying regional species pool. **(a)** Coordinates of the
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17 680 simulated communities in the skewness-kurtosis space under the different scenarios. We
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19 681 simulate 100 communities (number of run = 500) assembled in a constant environment under
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21 682 low / high dispersal limitation. Dark and light blue dots represent poor and rich regional pools,
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23 683 respectively. **(b)** We represent the parameters of the SKRs (R^2 , Y-intercept (alpha), slope of the
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25 684 SKRs (beta), distance to the lower boundary) for each scenario using violin plots. We provide
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27 685 the mean overlap (Mo) among the four scenarios. Different letters indicate significant
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29 686 differences between scenarios (overlap < 0.05; but NS for Not-Significant).

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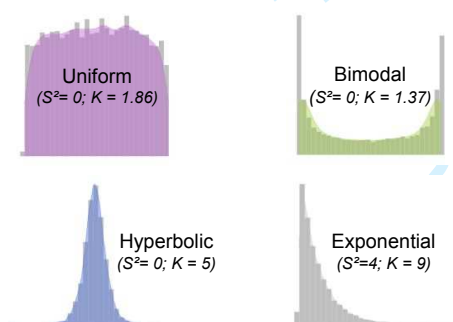
32
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34
35 688 **Figure 5** Effect of the different assembly scenarios on the Skewness-Kurtosis Relationship
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37 689 (SKR) along an environmental gradient. **(a)** Coordinates of the simulated communities in the
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39 690 skewness-kurtosis space under the different scenarios. We simulated 100 communities (number
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41 691 of run = 500) assembled along an environmental gradient (from Env = -2 to Env = +2) and a
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43 692 fixed regional species pool (n = 150 species) under low / high dispersal limitation. Dark and
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45 693 light blue dots represent the environmental gradient ranging from -2 to +2. **(b)** We represent
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47 694 the parameters of the SKRs (R^2 , Y-intercept (alpha), slope of the SKRs (beta), distance to the
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49 695 lower boundary) for each scenario using violin plots. We provide the mean overlap (Mo) among
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51 696 the four scenarios. Different letters indicate significant differences between scenarios (overlap
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54 697 < 0.05; but NS for Not-Significant).

698 **FIGURES**

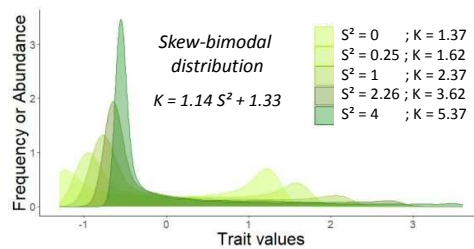
(a) Trait distribution



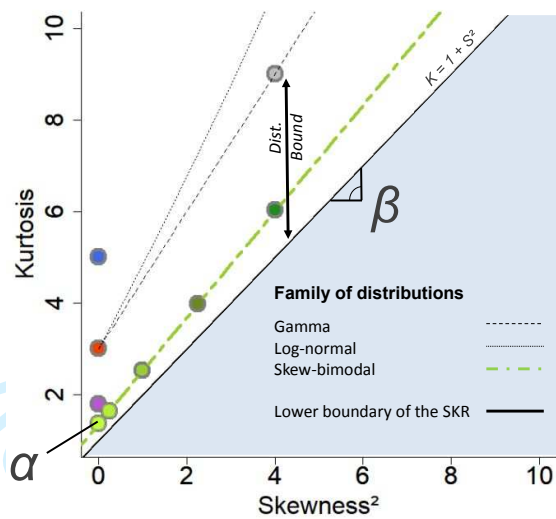
(b) Example of trait distributions



(c) Example of a family of trait distributions



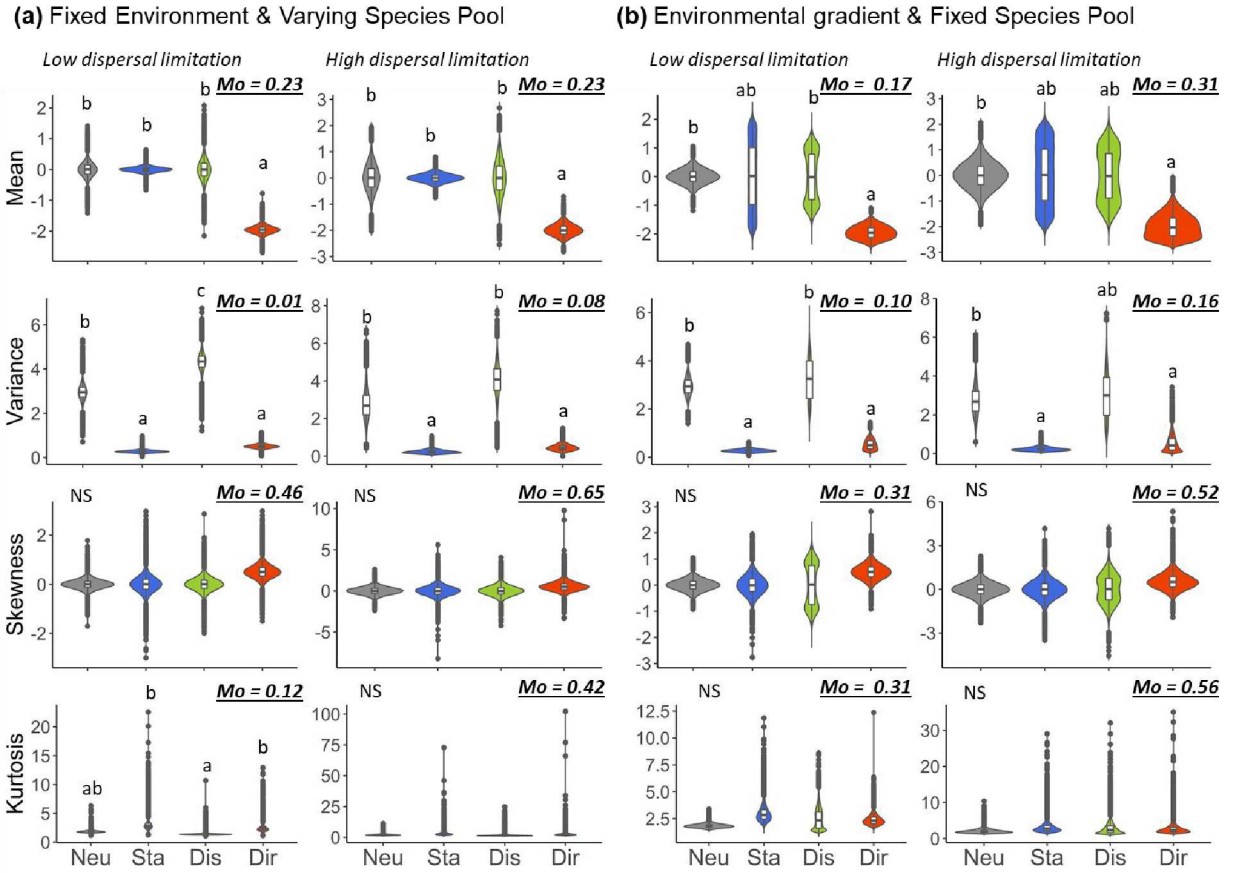
(d) The Skewness-Kurtosis Relationship ($K = \alpha + \beta S^2$)



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700 **Figure 1**

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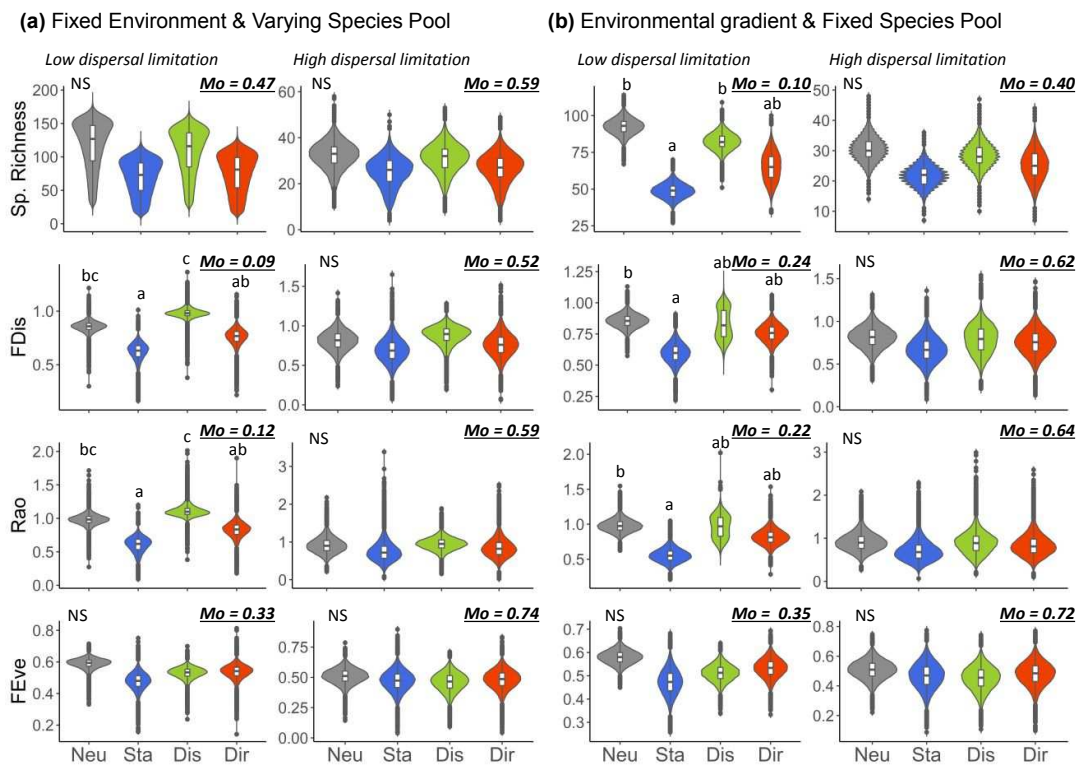


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703 **Figure 2**

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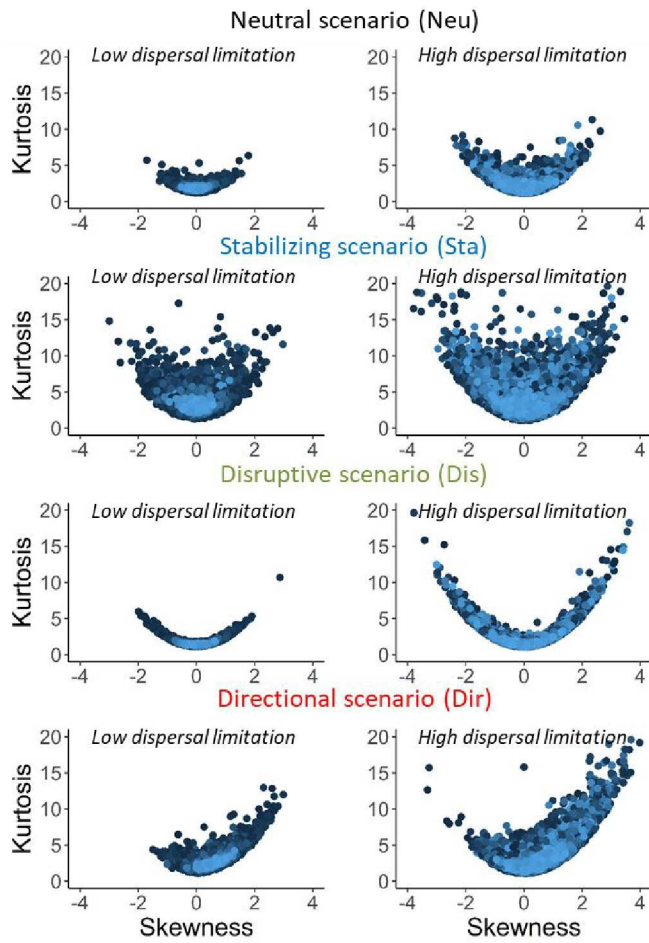
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707 **Figure 3**

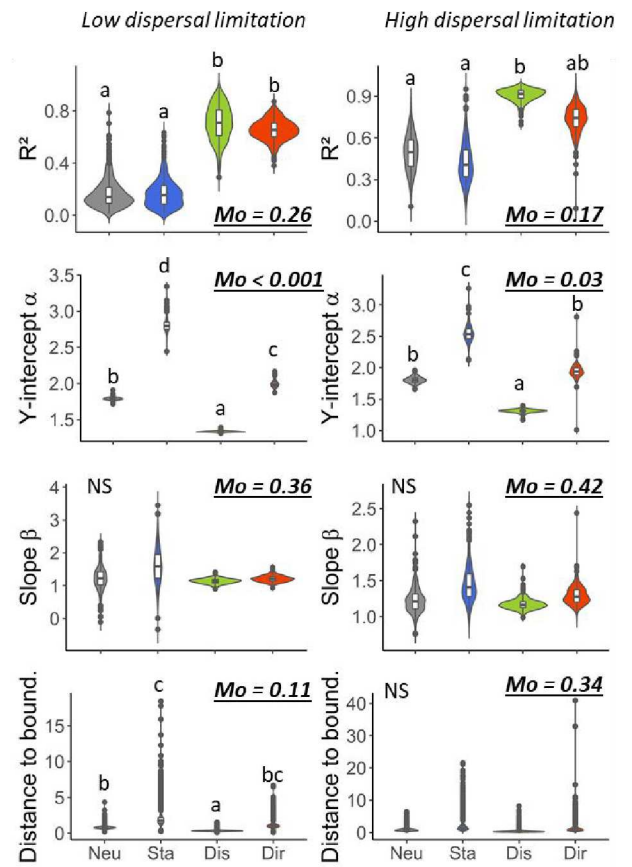
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(a) Skewness-Kurtosis space



(b) SKR parameters

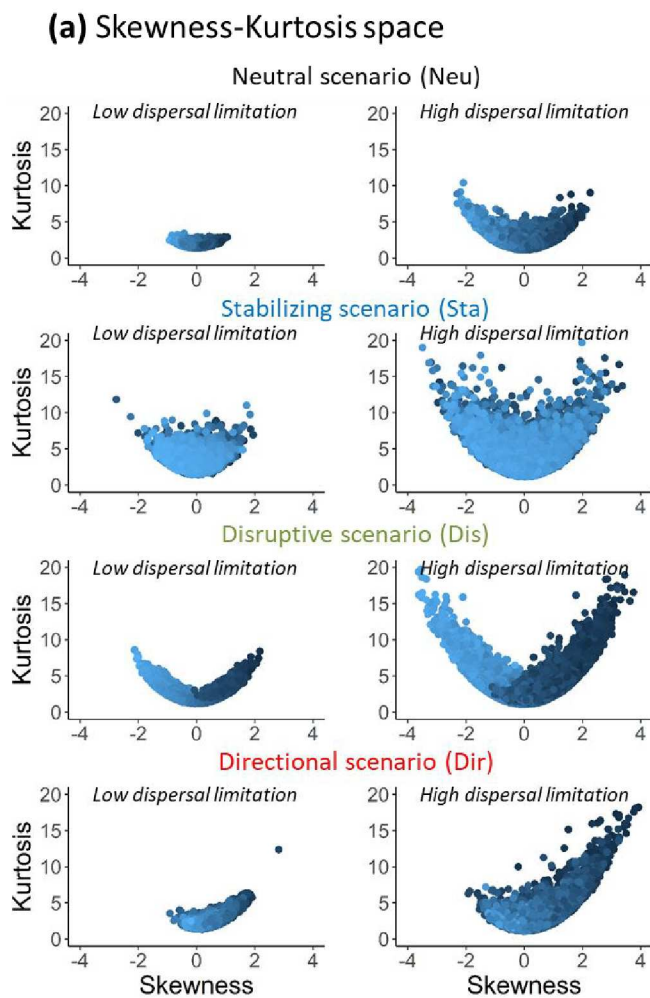


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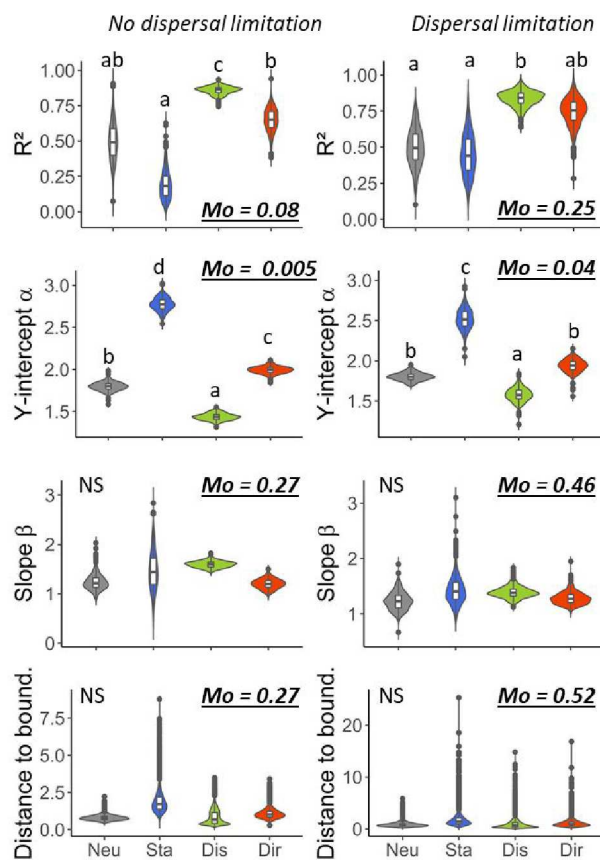
710 **Figure 4**

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(b) SKR parameters



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714 **Figure 5**

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