



**HAL**  
open science

# Nutrient deficiency enhances the rate of short-term belowground transfer of nitrogen from *Acacia mangium* to *Eucalyptus* trees in mixed-species plantations

I.R. Oliveira, B. Bordron, J.P. Laclau, R.R. Paula, A.V. Ferraz, J.L.M. Gonçalves, G. Le Maire, J.P. Bouillet

► **To cite this version:**

I.R. Oliveira, B. Bordron, J.P. Laclau, R.R. Paula, A.V. Ferraz, et al.. Nutrient deficiency enhances the rate of short-term belowground transfer of nitrogen from *Acacia mangium* to *Eucalyptus* trees in mixed-species plantations. *Forest Ecology and Management*, 2021, 491, pp.119192. 10.1016/j.foreco.2021.119192 . hal-03264026

**HAL Id: hal-03264026**

**<https://hal.inrae.fr/hal-03264026v1>**

Submitted on 24 Apr 2023

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial 4.0 International License

1 **Nutrient deficiency enhances the rate of short-term belowground transfer of nitrogen from**  
2 ***Acacia mangium* to *Eucalyptus* trees in mixed-species plantations**

3

4 I.R. Oliveira<sup>a</sup>, B. Bordron<sup>b,c</sup>, J.P Laclau<sup>b,c,d</sup>, R.R Paula<sup>e</sup>, A.V. Ferraz<sup>f</sup>, J.L.M Gonçalves<sup>a</sup>, G. le Maire<sup>b,c</sup>,  
5 J.P. Bouillet<sup>a,b,c\*</sup>

6

7 <sup>a</sup> USP, ESALQ, Forest Science Department, 13418-900, Piracicaba, Brazil

8 <sup>b</sup> Eco&Sols, INRA, CIRAD, IRD, Montpellier SupAgro, University of Montpellier, Montpellier,  
9 France

10 <sup>c</sup> CIRAD, UMR Eco&Sols, F-34398, Montpellier, France

11 <sup>d</sup> UNESP, Departamento de Solos e Recursos Ambientais, Universidade Estadual Paulista ‘Julio de  
12 Mesquita Filho’, Botucatu, 18610-300, Brazil

13 <sup>e</sup>Department of Forest Science and Wood, UFES, 29550-000, Jeronimo Monteiro, Brazil

14 <sup>f</sup> IPEF, Instituto de Pesquisas e Estudos Florestais, 13415-000, Piracicaba, Brazil

15

16 (\*) Corresponding author: [jpbouillet@cirad.fr](mailto:jpbouillet@cirad.fr)

17

18

19

20

21

22

23

24

25

26

27

28

29

30 **Abstract**

31 While a recent study showed that significant amounts of the nitrogen (N) requirements of young  
32 *Eucalyptus* trees can be provided by nitrogen-fixing trees (NFTs) in mixed-species plantations through  
33 short-term belowground N transfer, the consequences of soil fertility on this facilitation process  
34 remain unknown. We assessed the effect of fertilization on the percentage of N derived from transfer  
35 (%NDFT) from *Acacia mangium* trees to *Eucalyptus* trees in mixed-species plantations. A complete  
36 randomized block design with two treatments (fertilized vs unfertilized) and three blocks was set up in  
37 mixed-species plantations of *A. mangium* and *Eucalyptus* in Brazil, with 50% of each species at 2.5 m  
38 x 2.5 m spacing. Collection of litterfall and forest floor made it possible to estimate the annual N  
39 release from forest floor decomposition between 46 and 58 months after planting, close to harvest age.  
40  $^{15}\text{N-NO}_3^-$  was injected into the stem of one dominant *Acacia* tree in each plot, 58 months after  
41 planting. The  $x(^{15}\text{N})$  values of *Acacia* and *Eucalyptus* fine roots sampled within 1.8 m of the labelled  
42 *A. mangium* tree were determined at 7, 14, 30 and 60 days after labelling. The  $x(^{15}\text{N})$  values in wood,  
43 bark, branch and leaf samples were also determined for the 6 labelled *Acacia* trees and their two  
44 closest *Eucalyptus* neighbours, just before and 60 days after labelling. The amount of N released from  
45 forest floor decomposition was 31% higher in fertilized (F+) than in unfertilized (F-) plots. Sixty days  
46 after labelling, the aboveground compartments of *Eucalyptus* trees were significantly  $^{15}\text{N}$  enriched in  
47 both treatments. The  $x(^{15}\text{N})$  values of *Acacia* fine roots were higher than background values from 7  
48 days after labelling onwards in F+ and 30 days after labelling in F-. The  $x(^{15}\text{N})$  values of *Eucalyptus*  
49 fine roots were higher than background values in both treatments, from 30 days after labelling  
50 onwards. Mean %NDFT values were 18.0% in F+ and 33.9% in F- over the first 60 days after  
51 labelling, and 22.8% in F+ and 67.7% in F- from 30 to 60 days after labelling. Fertilization decreased  
52 short-term transfer belowground of N from *Acacia* trees to *Eucalyptus* trees. Our study suggests that  
53 belowground facilitation processes providing N from NFTs to *Eucalyptus* trees in mixed-species  
54 plantations are more pronounced in low-fertility soils than in nutrient-supplied stands.

55 **Keywords:**  $^{15}\text{N}$ , fertilization, facilitation, mature trees, nitrogen-fixing trees, Brazil

56

57

## 58 **1. Introduction**

59 Most tropical forest plantations are established in nutrient-poor soils (Stape et al., 2010;  
60 Mareschal et al., 2011; Keenan et al., 2015). Large amounts of biomass harvested every 6-7 years in  
61 commercial *Eucalyptus* plantations can lead to unbalanced input-output nutrient budgets (Laclau et al.,  
62 2010; Voigtlaender et al., 2019). Nitrogen (N) fertilizers are commonly applied in commercial  
63 *Eucalyptus* plantations to enhance tree early growth and to balance the N budget in the soil for  
64 sustainable plantation management (Gonçalves et al., 2013; Koutika et al., 2014). However, the use of  
65 N fertilizers may be limited in the future because of their rising cost (Brunelle et al., 2015) and their  
66 environmental impact through the use of fossil energy for their production (Elser, 2011), nitrate  
67 leaching, NH<sub>3</sub> volatilization or nitrogen oxide emissions (Binkley and Fisher, 2019).

68 *Acacia mangium* (Willd.) is a fast-growing tree species largely planted in South-East Asia for  
69 the pulp industry that stands out among the nitrogen-fixing tree (NFT) species of high silvicultural  
70 interest (Yamashita et al., 2008). The association of *A. mangium* with *Eucalyptus* can be an alternative  
71 to the use of N fertilizers. *A. mangium* can fix large amounts of atmospheric N<sub>2</sub>, as shown on 4 sites in  
72 Brazil with about 250 kg N ha<sup>-1</sup> fixed in mixed-species plantations of *A. mangium* and *Eucalyptus*  
73 *grandis* (Hill ex Maiden) over a 6-year rotation (Voigtlaender et al., 2019). N<sub>2</sub> fixation by NFTs can  
74 improve the nitrogen status of companion species in mixed plantations, through the decomposition of  
75 N-rich above-ground litter (Munroe and Isaac, 2014; Santos et al., 2017; Tchichelle et al., 2017a) and  
76 fine roots (Bachega et al., 2016). Moreover, non-NFTs may benefit from short-term belowground  
77 transfer of N from NFTs, as observed in pot experiments (He et al., 2004, 2005; Yao et al., 2019), and  
78 in the field between *A. mangium* and *Eucalyptus* trees within a radius of 6.2 m around *Acacias* trees  
79 (Paula et al., 2015). However, as far as we are aware, the effects of soil fertility on this facilitation  
80 process have never been investigated.

81 Facilitation occurs when at least one species benefits from another. The balance between  
82 facilitation and competition between plant species depends on resource availability. According to the  
83 stress-gradient hypothesis, competitive interactions decrease and facilitation increases under stressful  
84 environmental conditions (Callaway and Walker, 1997; Maestre et al., 2009; Holmgren and Scheffer,

85 2010; Kikvidze et al., 2011). Consistently, some studies in forest ecosystems have shown that  
86 competition between species is lower in nutrient-poor soils than in nutrient-rich soils (Baribault and  
87 Kobe, 2011; Coates et al., 2013). However, the opposite results were also found with high interspecific  
88 competition in low-fertility sites (Newmann, 1973; Trinder et al., 2012). In addition, asymmetric  
89 competition between tree species can be highly dependent on soil fertility. In a mixed-species  
90 plantation of *Falcataria mollucana* and *Eucalyptus saligna* in Hawaii, *Falcataria* tree growth was  
91 reduced by *Eucalyptus* neighbours on phosphorus-rich soils and facilitated on phosphorus-poor soils  
92 (Boyden et al., 2005). Conversely, *Eucalyptus* tree growth was reduced by *Falcataria* neighbours on  
93 low-phosphorus soils, but increased on high-phosphorus soils.

94         The N<sub>2</sub> fixation rate of *Acacia mangium* trees decreased when soil fertility increased in *Acacia*  
95 monocultures (Galiana et al., 2002), and was higher in association with *Eucalyptus* than in *Acacia*  
96 monocultures, likely due to *Eucalyptus* competition for soil N (Paula et al., 2018). In a pot experiment,  
97 N addition reduced the N<sub>2</sub> fixation rate of *Dalbergia odorifera* and led to lower N transfer from  
98 *Dalbergia* to *Eucalyptus* seedlings (Yao et al., 2019). Comparing fertilized and non-fertilized mixed-  
99 species plantations, fine root monitoring in young mixed-species plantations suggested that soil  
100 nutrient deficiency could promote belowground facilitation in mixed-species plantations of *Eucalyptus*  
101 and *A. mangium*, through an increase in the density of *Eucalyptus* fine roots close to *Acacia* trees  
102 where *Eucalyptus* trees could take advantage of high soil N availability (Bordron et al., 2021). Short-  
103 term belowground transfer of N from *Acacia* trees to *Eucalyptus* trees can be another facilitation  
104 process, as shown 2 years after planting in a nearby experiment (Paula et al., 2015). However, no  
105 experimental evidence supported the hypothesis of higher belowground transfer of N in non-fertilized  
106 than in fertilized plots.

107         Our study, conducted in the same trial as Bordron et al. (2021), set out to gain insights into the  
108 effect of soil fertility on the short-term belowground N transfer from NFTs to non-NFTs in mixed-  
109 species plantations. We estimated the short-term belowground transfer of N from *Acacia mangium* to  
110 *Eucalyptus grandis* x *E. urophylla* S.T Blake neighbours under two contrasting levels of NPK  
111 fertilization. We hypothesized that: (1) N contents in litterfall, N stocks in the forest floor and the rates  
112 of N release during forest floor decomposition are higher in fertilized than in non-fertilized plots, and

113 (2) the rate of short-term belowground N transfer between *Acacia* and *Eucalyptus* is higher in non-  
114 fertilized plots than in fertilized plots.

115

## 116 2. Material and methods

### 117 2.1 Study site

118 The study was carried out at the Itatinga experimental station of São Paulo University, Brazil  
119 (23°02'S, 48°38'W), at 860 m above mean sea level. The total rainfall over the study period from April  
120 2017 to May 2018 was 2300 mm and the mean temperature was 19.6 °C, with an average of 13.6 °C  
121 for the coldest month (July 2017) and 22.9 °C for the hottest month (March 2018). The soils were deep  
122 Ferralsols (FAO classification), acidic and of low fertility. In the 0-1.0 m layer, soil pH<sub>H2O</sub> was about  
123 4.0, clay content ranged from 18 to 24% and CEC from 3.2 to 7.6 cmol<sub>c</sub> kg<sup>-1</sup> (KCl extraction), the sum  
124 of base cations was around 0.4 cmol<sub>c</sub> kg<sup>-1</sup>, organic matter ranged from 6 to 16 g kg<sup>-1</sup> and total N from  
125 0.6 to 0.9 g kg<sup>-1</sup> (Bordron et al., 2021).

126

### 127 2.2 Experimental layout

128 In May 2013, a complete randomized block design was set up with two treatments  
129 (fertilization vs non-fertilization) and three blocks in mixed-species plantations of *Eucalyptus* hybrid  
130 (*E. grandis* x *E. urophylla*) and *A. mangium*. Each plot consisted of 10 x 10 plants at a spacing of 2.5  
131 m x 2.5 m with two buffer rows. Mixed-species stands were established in a proportion of 1:1 between  
132 *Eucalyptus* and *A. mangium*, with the two species planted alternately in the row, and between adjacent  
133 rows. *A. mangium* seeds originated from Papua New Guinea, and were inoculated with *Rhizobium*  
134 strains (BR 3609T and BR6009 provided by EMBRAPA Agrobiologia, Seropédica - Rio de Janeiro  
135 state) selected for their high levels of nodulation in nursery and high N<sub>2</sub> fixation efficiency. *Eucalyptus*  
136 cuttings (H13 clone) were provided by the Instituto de Pesquisa e Estudos Florestais (IPEF - São Paulo  
137 state).

138 In fertilized plots (F+), 2000 kg ha<sup>-1</sup> of dolomite limestone was applied at planting, as well as  
139 32.0 kg ha<sup>-1</sup> of FTE-BR (Fritted Trace Element, micronutrients and Borogran), 150 kg ha<sup>-1</sup> of K, 35 kg  
140 ha<sup>-1</sup> of P, and 24 kg ha<sup>-1</sup> of N only close to the *Eucalyptus* plants, to not reduce N<sub>2</sub> fixation by *Acacia*

141 trees (Paula et al., 2018). One year after planting, 166 kg ha<sup>-1</sup> of K and 37 kg ha<sup>-1</sup> of P were broadcast  
142 at the soil surface. In non-fertilized plots (F-), no mineral fertilizer was applied.

143

### 144 2.3 *Litterfall and forest floor*

145 Litterfall was collected every month from April 2017 (47 months after planting) to March  
146 2018 (58 months after planting). Leaf litterfall was collected in 12 traps (50 cm x 50 cm) per plot  
147 installed at different distances from the trees (Fig. S1). Bark and dead branches were collected in an  
148 area of 6.25 m<sup>2</sup> delimited between four trees in each plot in the 3 blocks. All compartments were  
149 separated by species and dried at 65°C to constant weight. For a given species, the components were  
150 gathered for each season (autumn from April to June 2017, winter from July to September 2017,  
151 spring from October to December 2017, summer from January to March 2018) and ground for N  
152 analysis.

153 The forest floor was sampled in all plots at the beginning (April 2017) and at the end of the  
154 study period (March 2018). Forest floor material was collected in twelve quadrats (50 cm x 50 cm)  
155 representing the spatial variability within the plot (Fig. S1). For each position, the forest floor was  
156 divided into two components: Lf (intact material or coarse fragments) and Hf (highly fragmented  
157 material). For each component and each plot, the samples were manually homogenized and dried at  
158 65°C to constant weight. A composite sample was then ground for N analysis. For both litter floor and  
159 forest floor, the ash content was determined by heating sub-samples at a 500 °C in an oven for 4 h and  
160 the ash content was used as a correction to determine the ash-free dry mass and N concentration.

161

### 162 2.4 *Rate of N release from forest floor decomposition*

163 The amount of N released from the decomposition of the forest floor was estimated in each  
164 plot as:

$$165 N_{\text{release}} = N_{\text{Forest-floor}_{2017}} + N_{\text{Litterfall}_{2017-2018}} - N_{\text{Forest-floor}_{2018}} \quad (1)$$

166 Where:

167 -  $N_{\text{Forest-floor}_{2017}}$  is the amount of N in the forest floor in April 2017 (in kg N ha<sup>-1</sup>).

168 -  $N_{\text{Forest-floor}_{2018}}$  is the amount of N in the forest floor in March 2018 (in kg N ha<sup>-1</sup>).

169 -  $N_{\text{Litterfall}_{2017-2018}}$  is the total amount of N in the litterfall between April 2017 and March 2018 (in kg  
170 N ha<sup>-1</sup>).

171

## 172 2.5 <sup>15</sup>N labelling of *Acacia* trees

173 In March 2018 (58 months after planting), a <sup>15</sup>N-labelled solution was injected into the stem of  
174 one *Acacia* tree in each plot, following the methodology described in Paula et al. (2015). The labelled  
175 *Acacia* trees were dominant trees, and did not show any disease or damage. The mean height of  
176 *Acacia* trees across the 3 blocks was 16.0 m in F+ and 15.6 m in F-. No tree mortality was observed  
177 within a radius of 2.5 m around the labelled *Acacia* trees. A hole was drilled into the stem (6 mm in  
178 diameter and 20 mm in depth). The trees with a single stem were drilled at 1.30 m in height and the  
179 trees with two stems were drilled below the fork. The drill was lubricated using distilled water to  
180 prevent damage to the xylem vessels. After removing the drill, a polyethylene tube (6 mm in diameter)  
181 attached to a bottle containing 500 mL of distilled water was pushed 20 mm into the drilled hole. As a  
182 preliminary experiment had shown possible reflux of the injected solution, 3 mL of acetic acid was  
183 added to the distilled water to prevent vascular clogging (Johansen, 1940). This bottle was then  
184 connected to a second one, containing 2.0 g of N (98 atom% <sup>15</sup>N-NO<sup>3</sup>) as potassium nitrate, dissolved  
185 in 500 mL of distilled water. <sup>15</sup>N contamination was avoided by packing the bark around the tube with  
186 non-toxic mineral putty (Terostat®) and placing a plastic bag around the trees before labelling. The  
187 solution was absorbed by the stem between 28 and 60 days after labelling depending on the treatments  
188 and trees. Over the 60 days, the few *Acacia* leaves that fell were removed within 24 h after the fall.

189

## 190 2.6 Sampling

191 In each plot, samples of leaves, living branches, stem wood, stem bark and fine roots of one  
192 acacia and one eucalypt tree were collected before labelling to measure the corresponding <sup>15</sup>N  
193 background values. Seven, 14, 30 and 60 days after the first day of <sup>15</sup>N labelling, four soil samples  
194 were collected from each plot using a PVC tube (5 cm in diameter and 10 cm in length), at four  
195 positions randomly located within a radius of 1.8 m around the <sup>15</sup>N-labelled *Acacia* trees (Fig. 1). The  
196 mean height of the four *Eucalyptus* trees neighbouring the labelled *Acacia* tree across the 3 blocks was



20.6 m in F+ and 18.7 m in F-. For each plot, the soil samples were then bulked and rapidly brought to the laboratory. Living fine roots (diameter < 2 cm) of *Acacia* and *Eucalyptus* were then carefully separated. The colour, thickness and branching patterns were good indications of the species to which they belonged, with *Acacia* roots that are brighter, rougher, thicker and less branched than *Eucalyptus* roots (Germon et al., 2018). At 60 days after labelling, the labelled *Acacia* tree and two *Eucalyptus* trees neighbouring the labelled *Acacia* tree, in the row and in the adjacent row (Fig. 1), were destructively sampled in each plot (6 *Acacia* trees and 12 *Eucalyptus* trees in total). The N concentration and  $x(^{15}\text{N})$  values of leaves, living branches, stem wood, stem bark and fine roots were determined for all the sampled trees (see below, equation 2).

206

### 207 2.7 Isotopic analyses

The fine roots were gently washed in tap water. All samples were dried at 65°C to constant weight. The bark, wood, leaves and living branches were ground in Retsch mill Zm 200 (120 micra) and the fine roots were ground in a porcelain mortar. The  $x(^{15}\text{N})$  value and N concentration of the samples were determined using a Hydra 20-20 mass spectrometer coupled to an automatic N analyzer (ANCA-GSL, SERCON Co., Crewe, UK), using 10 mg of dry mass of dry plant material, with a precision of 0.0001  $^{15}\text{N}$  atom%. The  $x(^{15}\text{N})$  value of a given sample was expressed as:

$$214 \quad x(^{15}\text{N})_{\text{sample}} (\%) = \left[ \left( \frac{^{15}\text{N}}{^{14}\text{N}} \right)_{\text{sample}} - \left( \frac{^{15}\text{N}}{^{14}\text{N}} \right)_{\text{air}} \right] / \left( \frac{^{15}\text{N}}{^{14}\text{N}} \right)_{\text{air}} * 1000 \quad (2)$$

215

### 216 2.8 N derived from transfer

The proportion of *Eucalyptus* N derived from transfer from *Acacia* was estimated at the collection date from fine root  $x(^{15}\text{N})$  values of the two species using the equation (Jalonen et al., 2009; Isaac et al., 2012; Paula et al., 2015):

$$220 \quad \% \text{NDFT} = \left( x(^{15}\text{N})_{\text{Euca}}(0) - x(^{15}\text{N})_{\text{Euca}}(t) \right) / \left( x(^{15}\text{N})_{\text{Euca}}(0) - x(^{15}\text{N})_{\text{Acacia}}(t) \right) * 100 \quad (3)$$

221 Where:

222 -  $x(^{15}\text{N})_{\text{Euca}}$  is the  $x(^{15}\text{N})$  value of fine *Eucalyptus* roots collected at 0 - 1.8 m from the  $^{15}\text{N}$  labelled  
223 *Acacia*.

224 -  $x(^{15}\text{N})_{\text{Acacia}}$  is the  $x(^{15}\text{N})$  value of fine *Acacia* roots collected at 0 - 1.8 m from the  $^{15}\text{N}$  labelled *Acacia*.  
225 -  $x(^{15}\text{N})_{\text{Euca}}(0)$  and  $x(^{15}\text{N})_{\text{Euca}}(t)$  are the  $x(^{15}\text{N})$  values of fine *Eucalyptus* roots before *Acacia* labelling  
226 and at the end of each collection date.  
227 -  $x(^{15}\text{N})_{\text{Acacia}}(t)$  is the  $x(^{15}\text{N})$  of fine *Acacia* roots at the end of each collection date.

228

## 229 2.9 Statistical analyses

230 For each collection date and treatment, the  $^{15}\text{N}$  enrichment of *Acacia* and *Eucalyptus* material  
231 was tested against  $^{15}\text{N}$  background values using a one-tailed paired t test. Differences between  
232 treatments and blocks in dry matter, N concentration and N content in litterfall and forest floor,  $x(^{15}\text{N})$   
233 of *Acacia* and *Eucalyptus* materials and %NDFt were tested using two-way ANOVA. The  
234 homogeneity of variances was tested using Levene's test. When the variances were unequal, the values  
235 were log-transformed. When ANOVA indicated significant effects, the means were compared with  
236 Bonferroni's multiple range test. Statistical analyses were carried out using R 3.5.2 (R Core Team  
237 2018). The significance level was 0.05.

238

## 239 3 Results

### 240 3.1 Litterfall and forest floor

241 Fertilization greatly influenced litterfall in our mixed-species stands (Fig. 2). Litterfall dry  
242 matter was 38% higher in F+ than in F- with values of 7.8 and 5.6 Mg ha<sup>-1</sup> yr<sup>-1</sup>, respectively. The  
243 proportion of *Eucalyptus* dry matter in the total litterfall was 81.9% on average in F+ and 70.3% in F-.  
244 Leaves accounted for 92.4% of the dry matter of *Acacia* litterfall and 45.3% of that of *Eucalyptus*  
245 litterfall. Mean N concentrations in litterfall were 17.3 g kg<sup>-1</sup> in F+ and 14.8 g kg<sup>-1</sup> in F- for *Acacia*  
246 leaves, and 6.5 g kg<sup>-1</sup> in F+ and 6.4 g kg<sup>-1</sup> in F- for *Eucalyptus* leaves (data not shown). On average,  
247 branches accounted for 7.6% of the dry matter of *Acacia* litterfall and 50.0% of *Eucalyptus* litterfall.  
248 Mean N concentrations in branch litterfall were 6.6 g kg<sup>-1</sup> in F+ and 8.1 g kg<sup>-1</sup> in F- for *Acacia*, and 2.0  
249 g kg<sup>-1</sup> in F+ and 1.9 g kg<sup>-1</sup> in F- for *Eucalyptus* (data not shown). The total amount of N in litterfall  
250 was 26% higher in F+ than in F- with values of 49 and 39 kg N ha<sup>-1</sup> yr<sup>-1</sup>, respectively (Fig. 2). *Acacia*  
251 material accounted for 46.3% of the amount of N in litterfall in F+ and 62.4% in F-. On average,

252 leaves accounted for 72% of the total N content in *Eucalyptus* litterfall and 97% of the total N content  
253 in *Acacia* litterfall.

254 The dry matter of the forest floor amounted to 6.7 Mg ha<sup>-1</sup> in F+ and 6.0 Mg ha<sup>-1</sup> in F- on  
255 average for the two sampling dates (Table 1). For a given layer, N contents in the forest floor were not  
256 significantly different between treatments, except for the Hf layer in 2017 with significantly higher N  
257 content in F- than in F+. The amount of N in the forest floor was not significantly influenced by the  
258 treatments, with average values of 65 kg N ha<sup>-1</sup> in F+ and 62 kg N ha<sup>-1</sup> in F-. Between 46 and 58  
259 months after planting, the amount of N released from forest floor decomposition was 31% higher in  
260 F+ than in F-, with values of 43 and 33 kg N ha<sup>-1</sup> yr<sup>-1</sup>, respectively (Table 1).

261

### 262 3.2 $\delta^{15}\text{N}$ in labelled *Acacia* trees

263  $\delta^{15}\text{N}$  values sharply increased in *Acacia* fine roots over the first 60 days after labelling in  
264 both fertilization treatments. In F+, mean  $\delta^{15}\text{N}$  values of *Acacia* fine roots were multiplied by 3  
265 relative to the mean background value of 2.73‰ from 7 days after labelling onwards, with average  
266 values of 10.78‰ over the 4 sampling dates after labelling (Fig. 3a). In F-, *Acacia* fine roots had  
267 higher values of  $\delta^{15}\text{N}$  than the mean background value of 1.91‰ from 30 days after labelling  
268 onwards, with average values of 4.18‰ over the study period of 60 days after labelling. *Acacia* fine  
269 roots were significantly <sup>15</sup>N-enriched at 30 days in F-, and at 60 days in F+ and in F- (Fig. 3a).

270 Sixty days after labelling, the aboveground compartments of the labelled *Acacia* trees  
271 exhibited higher  $\delta^{15}\text{N}$  values than background values, with significant differences for all  
272 compartments in both treatments (Table 2).

273

### 274 3.3 $\delta^{15}\text{N}$ in *Eucalyptus* trees

275 Labelling *Acacia* trees increased the  $\delta^{15}\text{N}$  values of the fine roots of neighbouring *Eucalyptus*  
276 trees. From 30 days after labelling onwards, mean  $\delta^{15}\text{N}$  values of *Eucalyptus* fine roots were slightly  
277 higher than the mean background values of 1.27‰ in F+ and 0.36‰ in F-, with a significant  
278 enrichment in <sup>15</sup>N at 60 days after labelling in F+ and in F-. Over the 4 sampling dates after labelling,  
279 the average  $\delta^{15}\text{N}$  values of *Eucalyptus* fine roots were 1.73‰ in F+ and 1.41‰ in F- (Fig. 3b).  $\delta^{15}\text{N}$

280 values of *Eucalyptus* fine roots were highest at 60 days after labelling with values of 2.82‰ in F+ and  
281 4.28‰ in F-. Fertilization did not change  $x(^{15}\text{N})$  values of *Eucalyptus* fine roots, except at 7 days after  
282 labelling, with significantly higher  $x(^{15}\text{N})$  values in F+ than in F-.

283 Sixty days after labelling,  $x(^{15}\text{N})$  values in the aboveground compartments of *Eucalyptus*  
284 neighbouring trees were significantly higher than background values for all compartments, in both  
285 treatments (Table 2).

286

### 287 3.4 Rates of N transfer from *Acacia* to *Eucalyptus* trees

288 The percentage of N of *Eucalyptus* trees derived from transfer (%NDFT) from *Acacia* trees  
289 estimated using Eq. (3) ranged from 9.5% to 29.5% in F+ and from 0 to 68.7% in F-, depending on the  
290 sampling dates (Fig. 4). Over the study period, the average values of %NDFT were 18.0% in F+ and  
291 33.9% in F-.

292

## 293 4. Discussion

### 294 4.1 Fertilization increases N availability for *Eucalyptus* trees

295 In agreement with our first hypothesis, the amount of N released from forest floor  
296 decomposition was slightly higher in fertilized than in non-fertilized plots. The first 3 years after  
297 planting in the same experiment, fertilization increased the *Eucalyptus* fine root foraging strategy for  
298 nutrients in the topsoil (Bordron et al., 2021). The competition with *Eucalyptus* fine roots led to a  
299 partial exclusion of *Acacia* fine roots from the topsoil more marked in F+ than in F-, and a higher  
300 proportion of *Eucalyptus* fine roots at the vicinity of *Acacia* trees in F- (Bordron et al., 2021).

301 N concentrations were lower in *Eucalyptus* litterfall than in *Acacia* litterfall. Therefore, a  
302 higher percentage of *Eucalyptus* than *Acacia* material in litterfall in F+ than in F- led to only a 26%  
303 higher N content in F+, while the amount of dry matter was 38% higher in fertilized plots. Nitrogen  
304 contents in litterfall of 39 kg N ha<sup>-1</sup> yr<sup>-1</sup> in F- and 49 kg N ha<sup>-1</sup> yr<sup>-1</sup> in F+ were lower than in other  
305 mixed-species plantations of *Eucalyptus* and *A. mangium* in Brazil (from 55 to 85 kg N ha<sup>-1</sup> yr<sup>-1</sup>)  
306 (Voigtlaender et al., 2012, 2019; Santos et al., 2017), which could be consistent with the particularly  
307 low nutrient availability in the soil of this experiment, especially in non-fertilized stands.

308           The higher proportion of N-poor *Eucalyptus* litterfall in F+ than in F- did not prevent the  
309 decomposition of the forest floor. Nutrient concentrations, stoichiometry and C quality are major  
310 drivers of litter decomposition and nutrient release in forest ecosystems (Hobbie, 2000;  
311 Hättenschwiler et al., 2011). The decomposition rates are commonly positively correlated with the N  
312 concentrations in litter fractions (Cornwell et al., 2008). However, some negative correlations are also  
313 reported (Berg, 2000). High lignin contents can lower the decomposition of forest residues (Freschet et  
314 al., 2012). However, high N concentrations can delay the processes of degrading lignin and lower the  
315 rates of litter decomposition (Santos et al., 2017). In a nearby experiment, decomposition rates in  
316 litterbags were faster for *Eucalyptus* leaves than *Acacia* leaves, with initial higher water-soluble  
317 carbon and lower lignin concentrations in *Eucalyptus* leaves (Bachega et al., 2016). Fertilization can  
318 increase litter decomposition rates and nutrient availability in forest ecosystems (Aslam et al., 2015;  
319 Keuskamp et al., 2015). Litter decomposition can be stimulated at low levels of N addition (Zhang et  
320 al., 2018). The small amounts of N applied at planting might therefore have contributed to an increase  
321 in the rates of forest floor decomposition and N release in F+ relative to F-.

322

#### 323 4.2 Belowground transfer of N from *Acacia* to *Eucalyptus* trees

324           The <sup>15</sup>N enrichment of *Acacia* and *Eucalyptus* fine roots in our study was lower than in 26-  
325 month-old trees in a nearby experiment (Paula et al., 2015). The totality of <sup>15</sup>N solution was absorbed  
326 in the *Acacia* stem between 28 and 60 days after the start of labelling in our study, instead of 12-36  
327 hours in Paula et al. (2015). This difference could be partly explained by a double volume of solution  
328 to be absorbed in the present study, as well as the difference in tree age between the two experiments.  
329 In the ecological conditions of São Paulo state, *Acacia* trees are much more suppressed by *Eucalyptus*  
330 trees at the end of stand rotation than in young stands (Bouillet et al., 2013; le Maire et al., 2013),  
331 which led to lower transpiration rates of mature *Acacia* trees than young *Acacia* trees in a nearby  
332 experiment (unpublished data). Other factors might have delayed the absorption of the <sup>15</sup>N solution by  
333 *Acacia* stems, in particular a partial clogging of xylem vessels after stem drilling, or narrow  
334 conductive sapwood width in mature trees (Pallardy, 2008; Debell and Lachenbruch, 2009) that would  
335 have needed holes drilled less than 20 mm into the stem. However, this slow <sup>15</sup>N absorption reduced

336 the risk of artificial flush of  $^{15}\text{N}$  in N root exudates after labelling, which is likely to overestimate the  
337 short-term belowground transfer (Paula et al., 2015). The limited number of replicates may have  
338 accounted for the high  $x(^{15}\text{N})$  variability in *Acacia* and *Eucalyptus* fine roots between blocks and the  
339 little number of significant differences with background values over the study period.

340  $^{15}\text{N}$  labelling of *Acacia* was heterogeneous between tree compartments and individuals (Table 2,  
341 Figure 3). Such heterogeneity was also found during the first weeks / months after  $^{15}\text{N}$  stem injection for  
342 other tree species (Horwath et al. 1992; Swanston and Myrold 1998; Augusto et al. 2011). However,  
343 despite marked variability in the  $x(^{15}\text{N})$  of fine roots between labelled acacias, the use of Equation 3  
344 based on the relative differences in  $x(^{15}\text{N})$  between fine roots of a given *Acacia* tree and *Eucalyptus*  
345 neighbours made it possible to reliably estimate the effect of nutrient deficiency on the belowground  
346 transfer of N.

347 Short-term belowground N transfer may occur directly via common mycorrhizal networks  
348 (CMNs) (Simard and Durall, 2004; Selosse et al., 2006; He et al., 2019) or root exudates of N  
349 compounds (Marschner and Dell, 1994; Fustec et al., 2010), or indirectly through rapid decomposition  
350 of very fine roots and microbial tissues (May and Attiwill, 2003; Staddon et al., 2003). Montesinos-  
351 Navarro et al. (2016) showed that short-term belowground N transfer between adult plants can be  
352 more effective by CMNs than via root exudates. CMNs can be formed by arbuscular mycorrhizal  
353 fungi (AMF) (Montesinos-Navarro et al., 2012) or ectomycorrhizal fungi (He et al., 2005).  
354 Mycorrhizal roots of both *Eucalyptus* and *A. mangium* were observed in our experiment (Bordron et  
355 al., 2021) and both species may potentially form CMNs. AMF are closely associated with *A. mangium*  
356 (Tawaraya et al., 2003; Dhar and Mridha, 2012) as well as with *Eucalyptus* species (Adjoud-Sadadou  
357 and Halli-Hargas, 2000). Intercropping *A. mangium* and *Eucalyptus* enhanced AMF colonization of  
358 *Eucalyptus* roots in the 0-10 cm layer (Bini et al., 2018) in a nearby stand, where 16 AMF species  
359 (across 6 AMF genera) were observed down to a depth of 8 m (Pereira et al., 2018). Ectomycorrhizas  
360 are associated with *Eucalyptus* (Horton et al., 2017; Robin et al., 2019) and *A. mangium* (Founoune et  
361 al., 2002; Diagne et al., 2013). *Pisolithus*, *Scleroderma*, *Thelephora* and *Boletellus* genera were found  
362 under *A. mangium* and *E. urophylla* trees (Aggangan et al., 2015), and *E. grandis* and *A. mangium*

363 roots can both be colonized by *Pisolithus* sp. and *Scleroderma* sp. (Founoune et al., 2002; Ducouso et  
364 al., 2012).

365

### 366 4.3 Nutrient deficiency increases the rate of belowground transfer of N from *Acacia* to *Eucalyptus*

367 In agreement with our second hypothesis, the values of N transfer rates from *Acacia* to  
368 *Eucalyptus* trees were higher in non-fertilized than in fertilized mixed-species stands, with an average  
369 %NDFT of 33.9% in F- and 18.0% in F+-. The difference in %NDFT between the two treatments  
370 could be higher as a result of a faster enrichment in <sup>15</sup>N of *Acacia* roots in F+ than in F-. *Acacia* fine  
371 roots had higher  $x(^{15}\text{N})$  values than  $x(^{15}\text{N})$  background values from 7 days after labelling onwards in  
372 F+, and 30 days after labelling in F-. Eq. (3) was therefore applicable at 7 and 14 days after labelling  
373 in F+, but not in F-. Considering the last 30 days of the experiment, when the  $x(^{15}\text{N})$  values of the fine  
374 roots of both species were higher than the  $x(^{15}\text{N})$  background values in the two treatments, the average  
375 %NDFT would be 67.7% in F- and 22.8% in F+. %NDFTs were probably overestimated using Eq. 3  
376 due to <sup>15</sup>N discrimination during this process (Paula et al., 2015). However, coherent results at  
377 successive sampling dates using the same equation for both treatments are consistent with higher  
378 belowground transfer of N in non-fertilized stands than in fertilized stands.

379 The average %NDFT was 43% in a nearby 26-month-old fertilized plantation of *Eucalyptus*  
380 and *A. mangium* (Paula et al., 2015). The higher rate of belowground transfer of N in Paula et al.  
381 (2015) relative to our study could be explained by the high N demand of young *Eucalyptus* trees to  
382 build the crown (Laclau et al., 2010). At the end of stand rotation, *Eucalyptus* trees are less dependent  
383 on soil N availability since a large share of the N requirements is provided by internal retranslocation  
384 (Laclau et al., 2010). The lower N release from forest floor decomposition in non-fertilized plots than  
385 in fertilized plots could be compensated for *Eucalyptus* trees through higher N belowground transfer  
386 rates from *Acacia* trees. In F-, *Eucalyptus* trees could proportionally benefit from higher belowground  
387 transfer of N from *Acacia* trees than in F+, through a higher exploration of *Eucalyptus* fine roots in the  
388 vicinity of *Acacia* trees. Bordron et al. (2021) showed at 34 months after planting in the same  
389 experiment that *Eucalyptus* fine root mass density in the topsoil (0-0.15 m) was higher in F- than in F+  
390 near *Acacia* trees. The specific root length (ratio between length and dry mass of fine roots) of

391 *Eucalyptus* fine roots was also higher in F- than in F+ close to *Acacia* trees, which might enhance the  
392 belowground transfer of N between the two species by increasing the length of fine roots in this area  
393 of nitrogen-enriched soil.

394

#### 395 4.4 Perspectives

396 Higher rates of belowground transfer of N from *A. mangium* to *Eucalyptus* trees in non-  
397 fertilized than in fertilized mixed-species stands suggest that this facilitation process increases under  
398 harsh conditions. This finding reinforces the interest of associating *Acacia* with *Eucalyptus* in forest  
399 plantations to limit N deficiencies of *Eucalyptus* trees when plantations are established in low-fertility  
400 soils and/or when fertilizers are not applied, as commonly observed in smallholder plantations in  
401 tropical regions (Verhaegen et al., 2014; Nambiar, 2015). The evidence of significant short-term  
402 belowground transfer of N between NFT and companion trees could also promote association in  
403 temperate regions of NFT to non-NFT in short-rotation forests (Georgiadis et al., 2017) or agroforestry  
404 systems (López-Díaz et al., 2017). Higher values of belowground N transfer from NFT to companion  
405 trees in harsh conditions than in fertilized stands suggest that mixed-species plantations including  
406 NFTs could be particularly interesting to restore degraded soils (Du et al., 2019; Jourgholami et al.,  
407 2019).

408 Mixed-species plantations of *Eucalyptus* and *Acacia* can outperform *Eucalyptus* monocultures,  
409 as observed in Brazil (Santos et al., 2016), Congo (Bouillet et al., 2013; Tchichelle et al., 2017b) and  
410 Australia (Forrester et al., 2006). However, this pattern is not general, depending on the balance  
411 between competitive and facilitative processes (Forrester et al., 2006; Bouillet et al., 2013). Further  
412 studies dealing with belowground N transfer from NFTs to non-NFTs would be of interest for other  
413 tree species, silvicultural practices, soil types and ecological conditions. Such insights would be  
414 worthwhile for the management of both rural and commercial plantations, in particular in marginal  
415 zones where the extension of fast-growing forest species will mainly occur in the future (Booth, 2013).

416

#### 417 Acknowledgements

418 We are grateful for the support of SOERE F-ORE-T granted by Ecofor, Allenvi and the



419 French National Research Infrastructure ANAEEF (<http://www.anaee-france.fr/fr/>), the IN-SYLVA  
420 French network, FAPESP Thematic Project (2010/16623-9) and Intens&fix Project (ANR-2010-  
421 STRA-004-03). We gratefully acknowledge the staff of IPEF, particularly Edison Luis da Fonseca and  
422 José Cardoso de Araújo, the staff of the Itatinga Experimental Station, particularly Rildo Moreira e  
423 Moreira and Eder Araujo da Silva (<http://www.floragroapoio.com.br>), and the staff of the USP-Cena  
424 and USP-Esalq laboratories for their technical support.

425

## 426 **References**

- 427 Adjoud-Sadadou, D., Halli-Hargas, R., 2000. Occurrence of arbuscular mycorrhiza on aged  
428 Eucalyptus. *Mycorrhiza* 9, 287–290. doi:10.1007/PL00009993
- 429 Aggangan, N.S., Pampolina, N.M., Cadiz, N.M., Raymundo, A.K., 2015. Assessment of plant  
430 diversity and associated mycorrhizal fungi in the mined-out sites of atlas mines in Toledo City,  
431 Cebu for bioremediation. *J. Environ. Sci. Manag.* 18, 71–86. ISSN 0119-1144
- 432 Aslam, T.J., Benton, T.G., Nielsen, U.N., Johnson, S.N., 2015. Impacts of eucalypt plantation  
433 management on soil faunal communities and nutrient bioavailability: trading function for  
434 dependence? *Biol. Fertil. Soils* 51, 637–644. doi:10.1007/s00374-015-1003-6
- 435 Augusto, L., Zeller, B., Midwood, A.J., Swanston, C., Dambrine, E., Scheinder, A., Bosc, A., 2011.  
436 Two-year dynamics of foliage labelling in 8-year-old *Pinus pinaster* trees with <sup>15</sup>N, <sup>26</sup>Mg and  
437 <sup>42</sup>Ca - simulation of Ca transport in xylem using an upscaling approach. *Ann. For. Sci.* 68, 169-  
438 178. doi: 10.1007/s13595-011-0018-x
- 439 Bacheга, L.R., Bouillet, J.-P., Piccolo, M.C., Saint-André, L., Bouvet, J.-M., Nouvellon, Y.,  
440 Gonçalves, J.L.M, Robin, A., Laclau, J.-P., 2016. Decomposition of *Eucalyptus grandis* and  
441 *Acacia mangium* leaves and fine roots in tropical conditions did not meet the Home Field  
442 Advantage hypothesis. *For. Ecol. Manage.* 359, 33–43. doi: 10.1016/j.foreco.2015.09.026
- 443 Baribault, T.W., Kobe, R.K., 2011. Neighbour interactions strengthen with increased soil resources in  
444 a northern hardwood forest. *J. Ecol.* 99, 1358–1372. doi:10.1111/j.1365-2745.2011.01862.x
- 445 Berg, B., 2000. Litter decomposition and organic matter turnover in northern forest soils. *For. Ecol.*  
446 *Manage.* 133, 13-22.

447 Bini, D., Santos, C.A. dos, Silva, M.C.P. da, Bonfim, J.A., Cardoso, E.J.B.N., Andreote, F.D., 2018.  
448 Intercropping *Acacia mangium* stimulates AMF colonization and soil. *Sci. Agric.* 75, 102–110.  
449 doi:10.1590/1678-992X-2016-0337

450 Binkley, D., Fisher, R.F., 2019. *Ecology and Management of Forest Soils*. 5<sup>th</sup> Edition, Wiley-  
451 Blackwell, 456 p.

452 Booth, T.H., 2013. Eucalypt plantations and climate change. *For. Ecol. Manage.* 301, 28–34.  
453 doi:10.1016/j.foreco.2012.04.004

454 Bordron, B., Germon, A., Laclau, J.-P., Oliveira, I.R., Robin, A., Jourdan, C., Paula, R.R., Pinheiro,  
455 R.C., Guillemot, J., Gonçalves, J.L.M., Bouillet, J.-P., 2021. Nutrient supply modulates species  
456 interactions belowground: dynamics and traits of fine roots in mixed plantations of *Eucalyptus*  
457 and *Acacia mangium*. *Plant Soil*, in press. doi: 10.1007/s11104-020-04755-2

458 Bouillet, J.-P., Laclau, J.-P., Gonçalves, J.L.M., Voigtlaender, M., Gava, J.L., Leite, F.P., Hakamada,  
459 R., Mareschal, L., Mabilia, A., Tardy, F., Levillain, J., Deleporte, P., Epron, D., Nouvellon, Y.,  
460 2013. *Eucalyptus* and *Acacia* tree growth over entire rotation in single- and mixed-species  
461 plantations across five sites in Brazil and Congo. *For. Ecol. Manage.* 301, 89–101.  
462 doi:10.1016/j.foreco.2012.09.019

463 Boyden, S., Binkley, D., Senock, R., 2005. Competition and facilitation between *Eucalyptus* and  
464 nitrogen-fixing *Falcataria* in relation to soil fertility. *Ecology* 86, 992–1001. doi:10.1890/04-  
465 0430

466 Brunelle, T., Dumas, P., Souty, F., Dorin, B., Nadaud, F., 2015. Evaluating the impact of rising  
467 fertilizer prices on crop yields. *Agric. Econ. (United Kingdom)* 46, 653–666.  
468 doi:10.1111/agec.12161

469 Callaway, J.C., Walker, L.R., 1997. Competition and Facilitation. *Ecology* 78, 1958–1965.

470 Coates, D.K., Lilles, E.B., Astrup, R., 2013. Competitive interactions across a soil fertility gradient in  
471 a multispecies forest. *J. Ecol.* 101, 806–818. doi:10.1111/1365-2745.12072

472 Cornwell, W.K., Cornelissen, J.H.C., Amatangelo, K., Dorrepaal, E., Eviner, V.T., Godoy, O.,  
473 Hobbie, S.E., Hoorens, B., Kurokawa, H., Pérez-Harguindeguy, N., Quested, H.M., Santiago,  
474 L.S., Wardle, D.A., Wright, I.J., Aerts, R., Allison, S.D., Van Bodegom, P., Brovkin, V.,

475 Chatain, A., Callaghan, T. V., Díaz, S., Garnier, E., Gurvich, D.E., Kazakou, E., Klein, J.A.,  
476 Read, J., Reich, P.B., Soudzilovskaia, N.A., Vaieretti, M.V., Westoby, M., 2008. Plant species  
477 traits are the predominant control on litter decomposition rates within biomes worldwide. *Ecol.*  
478 *Lett.* 11, 1065–1071. doi:10.1111/j.1461-0248.2008.01219.x

479 Debell, J.D., Lachenbruch, B., 2009. Heartwood / sapwood variation of western red cedar as  
480 influenced by cultural treatments and position in tree. *For. Ecol. Manage.* 258, 2026–2032.  
481 doi:10.1016/j.foreco.2009.07.054

482 Dhar, P.P., Mridha, M.A.U., 2012. Arbuscular mycorrhizal associations in different forest tree species  
483 of Hazarikhil forest of Chittagong, Bangladesh. *J. For. Res.* 23, 115–122. doi:10.1007/s11676-  
484 012-0241-9

485 Diagne, N., Thioulouse, J., Sanguin, H., Prin, Y., Krasova-Wade, T., Sylla, S., Galiana, A., Baudoin,  
486 E., Neyra, M., Svistoonoff, S., Lebrun, M., Duponnois, R., 2013. Ectomycorrhizal diversity  
487 enhances growth and nitrogen fixation of *Acacia mangium* seedlings. *Soil Biol. Biochem.* 57,  
488 468–476. doi:10.1016/j.soilbio.2012.08.030

489 Du, B., Pang, J., Hu, B., Allen, D.E., Bell, T.L., Pfautsch, S., Netzer, F., Dannenmann, M., Zhang, S.,  
490 Rennenberg, H., 2019. N<sub>2</sub>-fixing black locust intercropping improves ecosystem nutrition at the  
491 vulnerable semi-arid Loess Plateau region, China. *Science of the Total Environment* 688, 333-  
492 345. doi: 10.1016/j.scitotenv.2019.06.245

493 Ducousso, M., Duponnois, R., Thoen, D., Prin, Y., 2012. Diversity of Ectomycorrhizal Fungi  
494 Associated with *Eucalyptus* in Africa and Madagascar. *Int. J. For. Res.* 2012, 1–10.  
495 doi:10.1155/2012/450715

496 Elser, J. J., 2011. A world awash with nitrogen. *Science* 334 (6062), 1504-1505. doi:  
497 10.1126/science.1215567

498 Forrester, D.I., Bauhus, J., Cowie, A.L., Vanclay, J.K., 2006. Mixed-species plantations of *Eucalyptus*  
499 with nitrogen-fixing trees: A review. *For. Ecol. Manage.* 233, 211–230.  
500 doi:10.1016/j.foreco.2006.05.012

501 Founoune, H., Duponnois, R., Bâ, A.M., 2002. Ectomycorrhization of *Acacia mangium*, Willd. and  
502 *Acacia holosericea*, A. Cunn. ex G. Don in Senegal. Impact on plant growth, populations of

503 indigenous symbiotic microorganisms and plant parasitic nematodes. *J. Arid Environ.* 50, 325–  
504 332. doi:10.1006/jare.2001.0800

505 Freschet, G.T., Aerts, R., Cornelissen, J.H.C., 2012. Multiple mechanisms for trait effects on litter  
506 decomposition: Moving beyond home-field advantage with a new hypothesis. *J. Ecol.* 100, 619–  
507 630. doi:10.1111/j.1365-2745.2011.01943.x

508 Fustec, J., Lesuffleur, F., Mahieu, S., Cliquet, J.B., 2010. Nitrogen rhizodeposition of legumes. A  
509 review. *Agron. Sustain. Dev.* 30, 57–66. doi: 10.1051/agro/2009003

510 Galiana, A., Balle, P., N'Guessan Kanga, A., Domenach, A.M., 2002. Nitrogen fixation estimated by  
511 the <sup>15</sup>N natural abundance method in *Acacia mangium* Willd. inoculated with *Bradyrhizobium*  
512 sp. and grown in silvicultural conditions. *Soil Biol. Biochem.* 34, 251–262. doi:10.1016/S0038-  
513 0717(01)00179-1

514 Georgiadis, P., Taeroe, A., Stupak, I., Kepfer-Rojas, S., Zhang, W., Pinheiro Bastos, R. Raulund-  
515 Rasmussen, K., 2017. Fertilization effects on biomass production, nutrient leaching and budgets  
516 in four stand development stages of short rotation forest poplar. *For. Ecol. Manage.* 397, 18-26.  
517 doi:10.1016/j.foreco.2017.04.020

518 Germon, A., Guerrini, I.A., Bordron, B., Bouillet, J.-P., Nouvellon, Y., Gonçalves, J.L.M., Jourdan,  
519 C., Paula, R.R., Laclau, J.-P., 2018. Consequences of mixing *Acacia mangium* and *Eucalyptus*  
520 *grandis* trees on soil exploration by fine-roots down to a depth of 17 m. *Plant Soil* 424, 203–220.  
521 doi:10.1007/s11104-017-3428-1

522 Gonçalves, J.L.M., Alvares, C.A., Higa, A.R., Silva, L.D., Alfenas, A.C., Stahl, J., Ferraz, S.F. de B.,  
523 Lima, W. de P., Brancalion, P.H.S., Hubner, A., Bouillet, J.-P., Laclau, J.-P., Nouvellon, Y.,  
524 Epron, D., 2013. Integrating genetic and silvicultural strategies to minimize abiotic and biotic  
525 constraints in Brazilian eucalypt plantations. *For. Ecol. Manage.* 301, 6–27.  
526 doi:10.1016/j.foreco.2012.12.030

527 Hättenschwiler, S., Coq, S., Barantal, S., Handa, I.T., 2011. Leaf traits and decomposition in tropical  
528 rainforests: Revisiting some commonly held views and towards a new hypothesis. *New Phytol.*  
529 189, 950–965. doi:10.1111/j.1469-8137.2010.03483.x

530 He, X., Critchley, C., Ng, H., Bledsoe, C., 2004. Reciprocal N (<sup>15</sup>NH<sub>4</sub><sup>+</sup> or <sup>15</sup>NO<sub>3</sub><sup>-</sup>) transfer between non

531 N<sub>2</sub>-fixing *Eucalyptus maculata* and N<sub>2</sub>-fixing *Casuarina cunninghamiana* linked by the  
532 ectomycorrhizal fungus *Pisolithus* sp. *New Phytol.* 163, 629–640. doi:10.1111/j.1469-  
533 8137.2004.01137.x

534 He, X., Critchley, C., Ng, H., Bledsoe, C., 2005. Nodulated N<sub>2</sub>-fixing *Casuarina cunninghamiana* is  
535 the sink for net N transfer from non-N<sub>2</sub>-fixing *Eucalyptus maculata* via an ectomycorrhizal  
536 fungus *Pisolithus* sp. using <sup>15</sup>NH<sub>4</sub><sup>+</sup> or <sup>15</sup>NO<sub>3</sub><sup>-</sup> supplied as ammonium nitrate. *New Phytol.* 167,  
537 897–912. doi:10.1111/j.1469-8137.2005.01437.x

538 He, Y., Cornelissen, J.H.C., Wang, P., Dong, M., Ou, J., 2019. Nitrogen transfer from one plant to  
539 another depends on plant biomass production between conspecific and heterotrophic species via  
540 a common arbuscular mycorrhizal network. *Environ. Sci Pollut. Res.* 26, 8828–8837  
541 doi:10.1007/s11356-019-04385-x

542 Hobbie, S.E., 2000. Interactions between litter lignin and soil nitrogen availability during leaf litter  
543 decomposition in a Hawaiian montane forest. *Ecosystems* 3, 484–494.  
544 doi:10.1007/s100210000042

545 Holmgren, M., Scheffer, M., 2010. Strong facilitation in mild environments: The stress gradient  
546 hypothesis revisited. *J. Ecol.* 98, 1269–1275. doi:10.1111/j.1365-2745.2010.01709.x

547 Horton, B.M., Glen, M., Davidson, N.J., Ratkowsky, D.A., Close, D.C., Wardlaw, T.J., Mohammed,  
548 C., 2017. An assessment of ectomycorrhizal fungal communities in Tasmanian temperate high-  
549 altitude *Eucalyptus delegatensis* forest reveals a dominance of the Cortinariaceae. *Mycorrhiza*  
550 27, 67-74. doi: 10.1007/s00572-016-0725-0

551 Horwath, W. R., Paul, E.A., Pregitzer, K.S., 1992. Injection of Nitrogen-15 into Trees to Study  
552 Nitrogen Cycling in Soil. *Soil Sci. Soc. Am. J.* 56, 316-319.

553 Isaac, M.E., Hinsinger, P., Harmand, J.-M., 2012. Nitrogen and phosphorus economy of a legume tree-  
554 cereal intercropping system under controlled conditions. *Sci. Total Environ.* 434, 71–78.  
555 doi:10.1016/j.scitotenv.2011.12.071

556 Jalonen, R., Nygren, P., Sierra, J., 2009. Transfer of nitrogen from a tropical legume tree to an  
557 associated fodder grass via root exudation and common mycelial networks. *Plant Cell Environ.*  
558 32, 1366–1376. doi:10.1111/j.1365-3040.2009.02004.x

559 Jourgholami, M., Ghassemi, T., Labelle, E.R., 2019. Soil physio-chemical and biological indicators to  
560 evaluate the restoration of compacted soil following reforestation. *Ecological Indicators* 101,  
561 102-110. doi:10.1016/j.ecolind.2019.01.009

562 Keenan, R.J., Reams, G.A., Achard, F., Freitas, J.V. de, Grainger, A., Lindquist, E., 2015. Dynamics  
563 of global forest area: Results from the FAO Global Forest Resources Assessment 2015. *For.*  
564 *Ecol. Manage.* 352, 9–20. doi:10.1016/j.foreco.2015.06.014

565 Keuskamp, J.A., Hefting, M.M., Dingemans, B.J.J., Verhoeven, J.T.A., Feller, I.C., 2015. Effects of  
566 nutrient enrichment on mangrove leaf litter decomposition. *Sci. Total Environ.* 508, 402–410.  
567 doi:10.1016/j.scitotenv.2014.11.092

568 Kikvidze, Z., Suzuki, M., Brooker, R., 2011. Importance versus intensity of ecological effects: Why  
569 context matters. *Trends Ecol. Evol.* 26, 383–388. doi:10.1016/j.tree.2011.04.003

570 Koutika, L.S., Epron, D., Bouillet, J.-P., Mareschal, L., 2014. Changes in N and C concentrations, soil  
571 acidity and P availability in tropical mixed acacia and eucalypt plantations on a nutrient-poor  
572 sandy soil. *Plant Soil* 379, 205–216. doi:10.1007/s11104-014-2047-3

573 Laclau, J.-P., Ranger, J., Gonçalves, J.L.M, Maquère, V., Krusche, A. V., M'Bou, A.T., Nouvellon,  
574 Y., Saint-André, L., Bouillet, J.-P., Piccolo, M.C., Deleporte, P., 2010. Biogeochemical cycles of  
575 nutrients in tropical Eucalyptus plantations. Main features shown by intensive monitoring in  
576 Congo and Brazil. *For. Ecol. Manage.* 259, 1771–1785. doi:10.1016/j.foreco.2009.06.010

577 le Maire, G., Nouvellon, Y., Christina, M., Ponzoni, F.J., Gonçalves, J.L.M., Bouillet, J.-P., Laclau, J.-  
578 P., 2013. Tree and stand light use efficiencies over a full rotation of single- and mixed-species  
579 Eucalyptus grandis and Acacia mangium plantations. *For. Ecol. Manage.* 288, 31–42.  
580 doi:10.1016/j.foreco.2012.03.005

581 López-Díaz, M.L., Benítez, R., Moreno, G., 2017. How do management techniques affect carbon  
582 stock in intensive hardwood plantations? *For. Ecol. Manage.* 389, 228-239. doi:  
583 10.1016/j.foreco.2016.11.048

584 Maestre, F.T., Callaway, R.M., Valladares, F., Lortie, C.J., 2009. Refining the stress-gradient  
585 hypothesis for competition and facilitation in plant communities. *J. Ecol.* 97, 199–205.  
586 doi:10.1111/j.1365-2745.2008.01476.x

587 Mareschal, L., Nzila, J.D.D., Turpault, M.P., Thongo M'Bou, A., Mazoumbou, J.C., Bouillet, J.-P.,  
588 Ranger, J., Laclau, J.-P., 2011. Mineralogical and physico-chemical properties of Ferralic  
589 Arenosols derived from unconsolidated Plio-Pleistocenic deposits in the coastal plains of Congo.  
590 *Geoderma* 162, 159–170. doi:10.1016/j.geoderma.2011.01.017

591 Marschner, H., Dell, B., 1994. Nutrient uptake in mycorrhizal symbiosis. *Plant Soil* 159, 89–102.  
592 doi:10.1007/BF00000098

593 May, B.M., Attiwill, P.M., 2003. Nitrogen-fixation by *Acacia dealbata* and changes in soil properties 5  
594 years after mechanical disturbance or slash-burning following timber harvest. *For. Ecol. Manage.*  
595 181, 339–355. doi:10.1016/S0378-1127(03)00006-9

596 Montesinos-Navarro, A., Segarra-Moragues, J. G., Valiente-Banuet, A., Verdú, M., 2012. Plant  
597 facilitation occurs between species differing in their associated arbuscular mycorrhizal fungi.  
598 *New Phytol.* 196, 835–844. doi:10.1111/j.1469-8137.2012.04290.x

599 Montesinos-Navarro, A., Verdú, M., Querejeta, J.I., Sortibrán, L., Valiente-Banuet, A., 2016. Soil  
600 fungi promote nitrogen transfer among plants involved in long-lasting facilitative interactions.  
601 *Perspect. Plant Ecol. Evol. Syst.* 18, 45–51. doi:10.1016/j.ppees.2016.01.004

602 Munroe, J.W., Isaac, M.E., 2014. N<sub>2</sub>-fixing trees and the transfer of fixed-N for sustainable  
603 agroforestry: A review. *Agron. Sustain. Dev.* 34, 417–427. doi:10.1007/s13593-013-0190-5

604 Nambiar, E.K.S., 2015. Forestry for rural development, poverty reduction and climate change  
605 mitigation: we can help more with wood. *Aust. For.* 78, 55-64.  
606 doi:10.1080/00049158.2015.1050776

607 Newman, E., 1973. Competition and Diversity in Herbaceous Vegetation. *Nature* 244, 310.  
608 doi:10.1038/244310a0

609 Pallardy G.S., 2008. *Physiology of woody plants*. 3rd Edition. 464 p. Academic Press. ISBN 978-0-  
610 12-088765-1

611 Paula, R.R., Bouillet, J.-P., Trivelin, P.C.O., Zeller, B., Gonçalves, J.L.M, Nouvellon, Y., Bouvet, J.-  
612 M., Plassard, C., Laclau, J.-P., 2015. Evidence of short-term belowground transfer of nitrogen  
613 from *Acacia mangium* to *Eucalyptus grandis* trees in a tropical planted forest. *Soil Biol.*  
614 *Biochem.* 91, 99–108. doi:10.1016/j.soilbio.2015.08.017

615 Paula, R.R., Bouillet, J.-P., José, J.L., Trivelin, P.C.O., de C. Balieiro, F., Nouvellon, Y., de C.  
616 Oliveira, J., de Deus Júnior, J.C., Bordron, B., Laclau, J.-P., 2018. Nitrogen fixation rate of  
617 *Acacia mangium* Wild at mid rotation in Brazil is higher in mixed plantations with *Eucalyptus*  
618 *grandis* Hill ex Maiden than in monocultures. *Ann. For. Sci.* 75. doi:10.1007/s13595-018-0695-9

619 Pereira, A.P.A, Zagatto, M.R.G., Brandani, C.B., Mescolotti, D.L., Cotta, S.R., Gonçalves, J.L.M.,  
620 Cardoso, E.J.B.N., 2018. *Acacia* changes microbial indicators and increases C and N in soil  
621 organic fractions in intercropped *Eucalyptus* plantations. *Frontiers in Microbiology*, 9, art. 655.  
622 doi:10.3389/fmicb.2018.00655

623 Robin, A., Pradier, C., Sanguin, H., Mahé, F., Lambais, G.R., de Araujo Pereira, A.P., Germon, A.,  
624 Santana, M.C., Tisseyre, P., Pablo, A.L., Heuillard, P., Sauvadet, M., Bouillet, J.-P., Andreote,  
625 F.D., Plassard, C., Gonçalves, J.L.M, Cardoso, E.J.B.N., Laclau, J.-P., Hinsinger, P., Jourdan, C.,  
626 2019. How deep can ectomycorrhizas go? A case study on *Pisolithus* down to 4 meters in a  
627 Brazilian eucalypt plantation. *Mycorrhiza* 29, 637–648. doi:10.1007/s00572-019-00917-y

628 Santos, F.M., Balieiro, F. de C., Ataíde, D.H. dos S., Diniz, A.R., Chaer, G.M., 2016. Dynamics of  
629 aboveground biomass accumulation in monospecific and mixed-species plantations of  
630 *Eucalyptus* and *Acacia* on a Brazilian sandy soil. *For. Ecol. Manage.* 363, 86–97.  
631 doi:10.1016/j.foreco.2015.12.028

632 Santos, F.M., Chaer, G.M., Diniz, A.R., Balieiro, F. de C., 2017. Nutrient cycling over five years of  
633 mixed-species plantations of *Eucalyptus* and *Acacia* on a sandy tropical soil. *For. Ecol. Manage.*  
634 384, 110–121. doi:10.1016/j.foreco.2016.10.041

635 Selosse, M.A., Richard, F., He, X., Simard, S.W., 2006. Mycorrhizal networks: des liaisons  
636 dangereuses? *Trends Ecol. Evol.* 21, 621–628. doi:10.1016/j.tree.2006.07.003

637 Simard, S.W., Durall, D.M., 2004. Mycorrhizal networks: A review of their extent, function, and  
638 importance. *Can. J. Bot.* 82, 1140–1165. doi: 10.1139/B04-116

639 Staddon, P.L., Ramsey, C.B., Ostle, N., Ineson, P., Fitter, A.H., 2003. Rapid Turnover of Hyphae of  
640 Mycorrhizal Fungi Determined by AMS Microanalysis of <sup>14</sup>C. *Science*. 300, 1138–1141.

641 Stape, J.L., Binkley, D., Ryan, M.G., Fonseca, S., Loos, R.A., Takahashi, E.N., Silva, C.R., Silva,  
642 S.R., Hakamada, R.E., Ferreira, J.M. de A., Lima, A.M.N., Gava, J.L., Leite, F.P., Andrade,



643 H.B., Alves, J.M., Silva, G.G.C., Azevedo, M.R., 2010. The Brazil Eucalyptus Potential  
644 Productivity Project: Influence of water, nutrients and stand uniformity on wood production. For.  
645 Ecol. Manage. 259, 1684–1694. doi:10.1016/j.foreco.2010.01.012

646 Swanston, C.W., Myrold, D.D., 1998. Evaluation of the stem injection technique and subsequent 15 N  
647 partitioning in red alder crowns. Plant Soil, 198, 63-69.

648 Tawarayama, K., Takaya, Y., Turjaman, M., Tuah, S.J., Limin, S.H., Tamai, Y., Cha, J.Y., Wagatsuma,  
649 T., Osaki, M., 2003. Arbuscular mycorrhizal colonization of tree species grown in peat swamp  
650 forests of Central Kalimantan, Indonesia. For. Ecol. Manage. 182, 381–386. doi:10.1016/S0378-  
651 1127(03)00086-0

652 Tchichelle, S.V., Epron, D., Mialoundama, F., Koutika, L.S., Harmand, J.M., Bouillet, J.-P.,  
653 Mareschal, L., 2017a. Differences in nitrogen cycling and soil mineralisation between a eucalypt  
654 plantation and a mixed eucalypt and *Acacia mangium* plantation on a sandy tropical soil. South.  
655 For. 79, 1–8. doi:10.2989/20702620.2016.1221702

656 Tchichelle, S.V., Mareschal, L., Koutika, L.S., Epron, D., 2017b. Biomass production, nitrogen  
657 accumulation and symbiotic nitrogen fixation in a mixed-species plantation of eucalypt and  
658 acacia on a nutrient-poor tropical soil. For. Ecol. Manage. 403, 103-111. doi:  
659 10.1016/j.foreco.2017.07.041

660 Trinder, C.J., Brooker, R.W., Davidson, H., Robinson, D., 2012. A new hammer to crack an old nut:  
661 Interspecific competitive resource capture by plants is regulated by nutrient supply, not climate.  
662 PLoS One 7. doi:10.1371/journal.pone.0029413

663 Verhaegen, D., Randrianjafy, H., Andriatsitohaina, H.R., Rakotonirina, T.M.C., Andriamampianina,  
664 N., Montagne, P., Rasamindisa, A., Chaix, G., Bouillet, J.-P., Bouvet, J.-M., 2014. Eucalyptus  
665 robusta for sustainable fuelwood production in Madagascar: Review of knowledge and future  
666 prospects. Bois et Forêts des Trop. 320, 15-30.

667 Voigtlaender, M., Laclau, J.-P., de Gonçalves, J.L.M., de Piccolo, M.C., Moreira, M.Z., Nouvellon,  
668 Y., Ranger, J., Bouillet, J.-P., 2012. Introducing *Acacia mangium* trees in *Eucalyptus grandis*  
669 plantations: Consequences for soil organic matter stocks and nitrogen mineralization. Plant Soil  
670 352, 99–111. doi:10.1007/s11104-011-0982-9

671 Voigtlaender, M., Brandani, C.B., Caldeira, D.R.M., Tardy, F., Bouillet, J.-P., Gonçalves, J.L.M.,  
672 Moreira, M.Z., Leite, F.P., Brunet, D., Paula, R.R., Laclau, J.-P., 2019. Nitrogen cycling in  
673 monospecific and mixed-species plantations of *Acacia mangium* and *Eucalyptus* at 4 sites in  
674 Brazil. *For. Ecol. Manage.* 436, 56–67. doi:10.1016/j.foreco.2018.12

675 Yamashita, N., Ohta, S., Hardjono, A., 2008. Soil changes induced by *Acacia mangium* plantation  
676 establishment : Comparison with secondary forest and *Imperata cylindrica* grassland soils in  
677 South Sumatra , Indonesia. *For. Ecol. Manage.* 254, 362–370. doi:10.1016/j.foreco.2007.08.012

678 Yao, X., Li, Y., Liao, L., Sun, G., Wang, H., Ye, S., 2019. Enhancement of nutrient absorption and  
679 interspecific nitrogen transfer in a *Eucalyptus urophylla* × *Eucalyptus grandis* and *Dalbergia*  
680 *odorifera* mixed plantation. *For. Ecol. Manage.* 449, 117465. doi:10.1016/j.foreco.2019.117465

681 Zhang, T., Luo, Y., Chen, H.Y.H., Ruan, H., 2018. Responses of litter decomposition and nutrient  
682 release to N addition: A meta-analysis of terrestrial ecosystems. *Appl. Soil Ecol.* 128, 35–42.  
683 doi:10.1016/j.apsoil.2018.04.004

684

685

686 **Figure captions**

687

688 **Fig. 1.** Sampling scheme. Fine roots of *A. mangium* and *Eucalyptus* trees were sampled at 0, 7, 14, 30  
689 and 60 days after <sup>15</sup>N-labelling of *Acacia* trees. At 60 days after labelling, the labelled *Acacia* tree and  
690 two neighbouring *Eucalyptus* trees (in the row and in the adjacent row) were harvested in each plot (6  
691 *Acacia* trees and 12 *Eucalyptus* trees in total).

692

693 **Fig. 2.** Litterfall dry matter collected from April 2017 to March 2018 in mixed-species stands  
694 (50A:50E) with fertilization (F+) (a), and without fertilization (F-) (b). Corresponding N contents in  
695 litterfall in F+ (c) and in F- (d). For each treatment, total dry matter and total N content in litterfall, of  
696 bark, leaves and branches of *Acacia* trees (*Acacia*) and of *Eucalyptus* trees (*Eucalyptus*) are shown.  
697 Standard errors between blocks are indicated (n=3).

698

699 **Fig. 3.** Mean  $x(^{15}\text{N})$  values of *A. mangium* fine roots (a) and *Eucalyptus* fine roots (b) collected within  
700 1.8 m of the <sup>15</sup>N-labelled *A. mangium*, at 0, 7, 14, 30 and 60 days after labelling in mixed-species  
701 stands with (F+) or without fertilization (F-). Vertical bars indicate standard errors between blocks  
702 (n=3). At a given collection date, \* indicates  $x(^{15}\text{N})$  values significantly higher than background values  
703 ( $P < 0.05$ ). Different letters indicate significant differences between treatments ( $P < 0.05$ ).

704

705 **Fig. 4.** Estimates of the percentage of *Eucalyptus* nitrogen derived from *A. mangium* (%NDFT) at 7,  
706 14, 30 and 60 days after *Acacia* labelling in mixed-species stands with (F+) or without fertilization (F-  
707 ). Vertical bars indicate standard errors between blocks (n=3). The average %NDFT value over the  
708 study period of 60 days was 17.96% in F+ and 33.85% in F-.

709

710

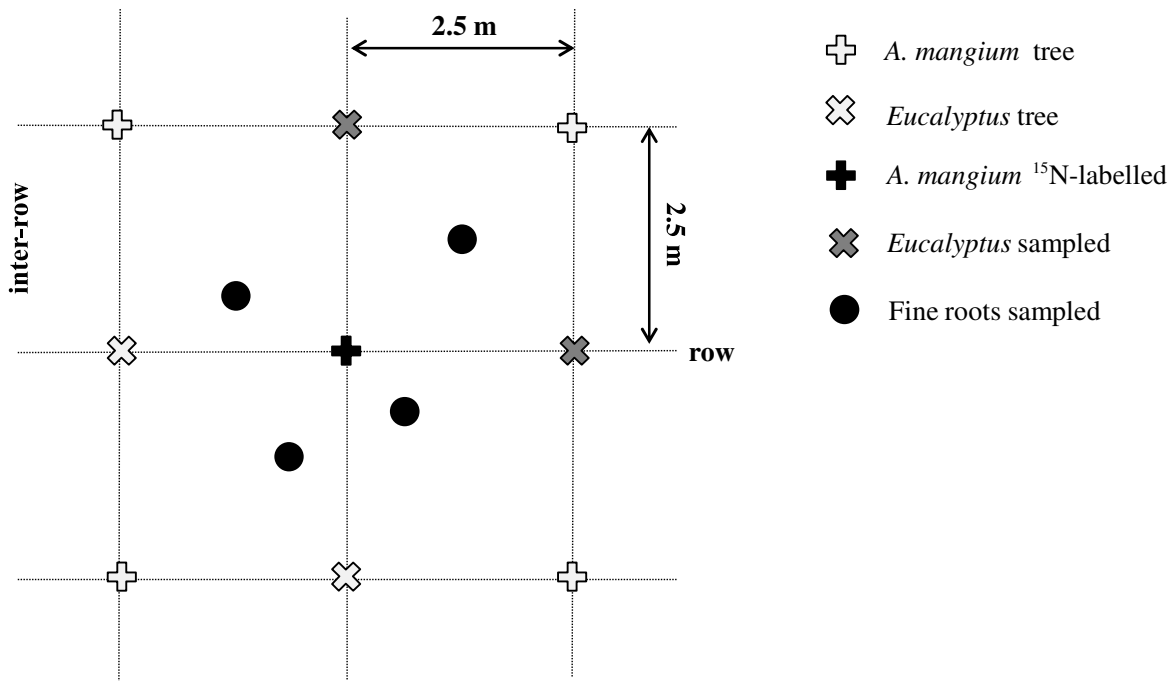
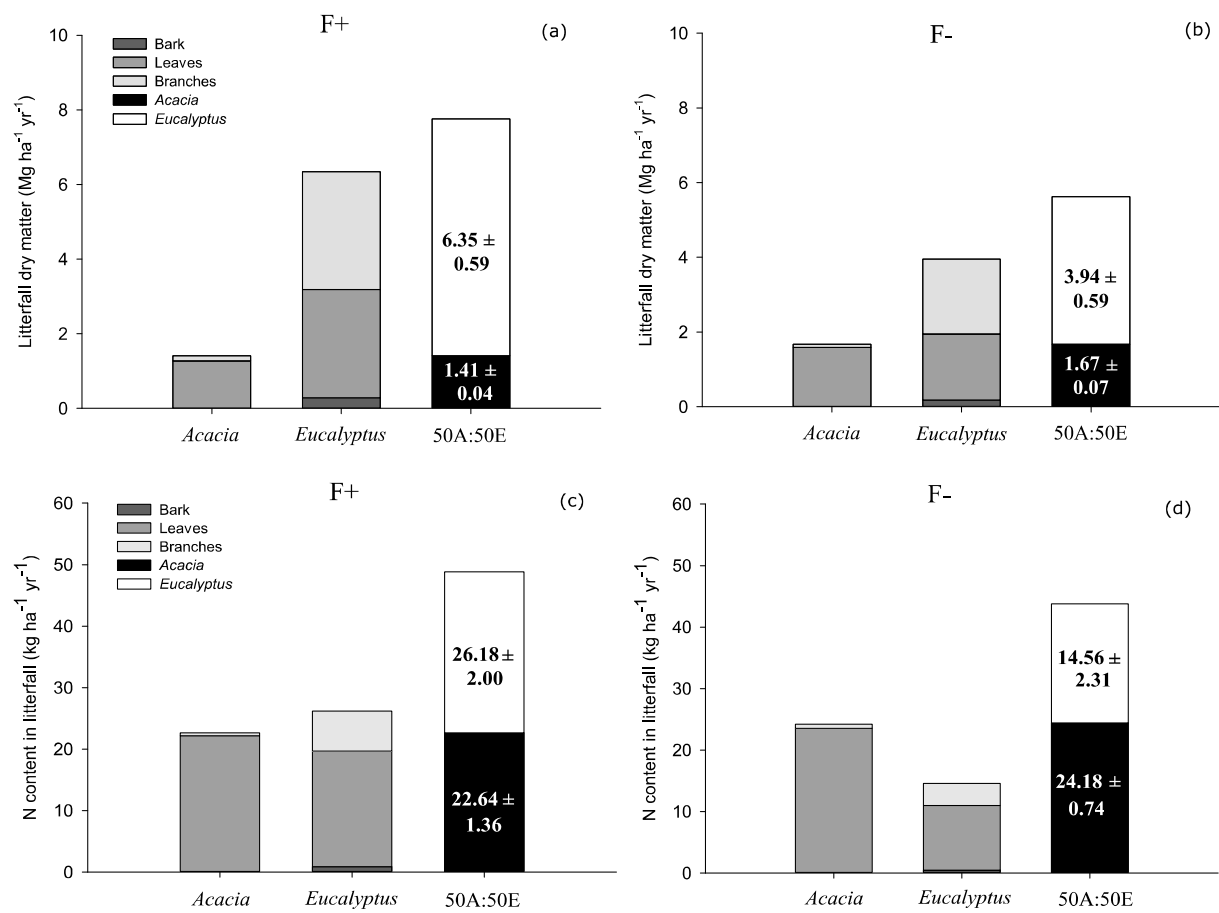
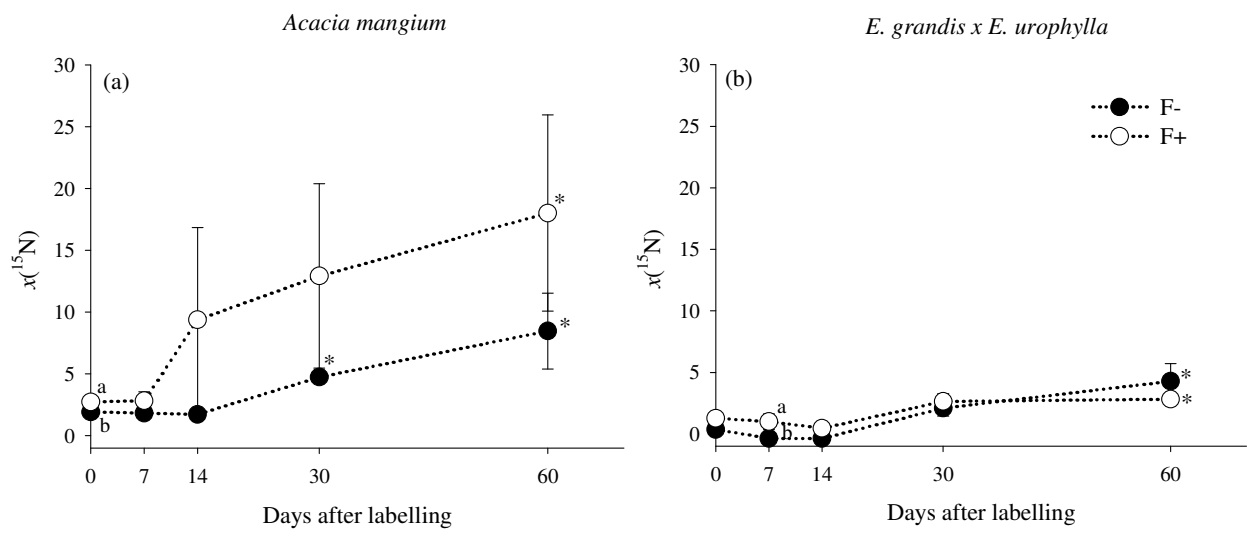


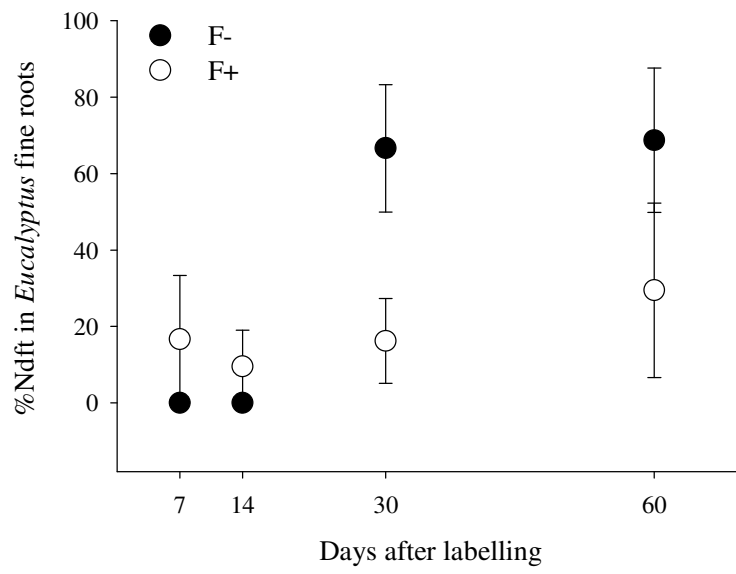
Fig. 1.



**Fig. 2.**



**Fig. 3.**



**Fig. 4.**

**Table 1**

Dry matter and N content in the Lf and HF components of the forest floor collected in April 2017 and March 2018 in mixed-species stands (50A:50E) with (F+) or without (F-) fertilization. For a given plot, the values of dry matter and N content near *Acacia* and near *Eucalyptus* correspond to the mean of the 6 positions sampled close to 6 different *A. mangium* trees and 6 positions sampled close to 6 different *Eucalyptus* trees, respectively. Dry matter and N content in the litterfall collected between April 2017 and March 2018 in the same treatments are indicated, as well as the corresponding amounts of N released during forest floor decomposition during the same period. Standard errors between blocks are indicated (n=3). Different Latin uppercase letters indicate significant differences ( $P < 0.05$ ) between treatments (F+ vs F-) for each species, and different Latin lowercase letters indicate significant differences between *Acacia* and *Eucalyptus* in each treatment. Different Greek letters indicate significant differences between treatments for the whole stand.

	Near <i>Acacia</i>	F+ Near <i>Eucalyptus</i>	Whole stand	Near <i>Acacia</i>	F- Near <i>Eucalyptus</i>	Whole stand
Forest floor 2017						
Dry matter (Mg ha <sup>-1</sup> )						
Lf	4.44±0.34 Aa	4.08±0.54 Aa	4.26±0.35 α	3.46±0.32 Aa	3.61±0.19 Aa	3.53±0.25 α
Hf	1.75±0.05 Ba	1.80±0.09 Aa	1.77±0.05 □	2.37±0.21 Aa	1.94±0.16 Aa	2.15±0.03 α
Total	6.19±0.39 Aa	5.89±0.46 Aa	6.04±0.34 α	5.83±0.34 Aa	5.55±0.28 Aa	5.69±0.31 α
N content (kg N ha <sup>-1</sup> )						
Lf	46.81±1.08 Aa	37.20±6.79 Aa	42.01±3.80 α	36.52±1.45 Ba	30.43±5.23 Aa	33.47±3.14 α
Hf	19.93±0.27 Ba	20.02±3.00 Aa	19.97±1.41 □	27.26±1.17 Aa	24.35±2.76 Aa	25.81±1.56 α
Total	66.74±1.35 Aa	57.22±3.81 Aa	61.98±2.46 α	63.78±2.54 Aa	54.78±2.73 Aa	59.28±2.43 α
Forest floor 2018						
Dry matter (Mg ha <sup>-1</sup> )						
Lf	4.38±0.39 Aa	5.21±0.29 Aa	4.80±0.24 α	4.25±0.57 Aa	3.97±0.19 Ba	4.11±0.34 α
Hf	2.54±0.12 Aa	2.66±0.27 Aa	2.60±0.19 α	2.08±0.27 Aa	2.33±0.25 Aa	2.20±0.26 α
Total	6.92±0.36 Aa	7.88±0.56 Aa	7.40±0.37 α	6.32±0.37 Aa	6.30±0.29 Aa	6.31±0.33 α
N content (kg N ha <sup>-1</sup> )						
Lf	48.90±4.58 Aa	36.89±2.59 Aa	42.89±3.19 α	44.22±8.17 Aa	35.33±0.99 Aa	39.78±4.18 α
Hf	26.42±1.94	23.54±1.64	24.98±1.76	22.94±2.60	27.90±1.05	25.42±1.82



Total	Aa 75.32±3.39 Aa	Aa 60.42±3.71 Ab	α 67.87±3.12 α	Aa 67.16±7.11 Aa	Aa 63.23±0.36 Aa	α 65.20±3.38 α
Litterfall						
Dry matter (Mg ha <sup>-1</sup> yr <sup>-1</sup> )			7.76±0.62 α			5.62±0.65 α
N content (kg N ha <sup>-1</sup> yr <sup>-1</sup> )			48.81±2.96 α			38.73±3.00 α
N release (kg N ha <sup>-1</sup> yr <sup>-1</sup> )			42.92±7.34 α			32.82±4.18 α

**Table 2**

Mean  $x(^{15}\text{N})$  values in aboveground compartments of *A. mangium* and *Eucalyptus* neighbours before (background values) and 60 days after *Acacia*  $^{15}\text{N}$ -labelling. Standard errors between blocks are indicated (n=3). For a given tree compartment and a given treatment (F+ and F-), \* indicates significantly higher  $x(^{15}\text{N})$  values than background values ( $P < 0.05$ ).

Compartments	<i>A. mangium</i>				<i>Eucalyptus</i>			
	Background		60 days		Background		60 days	
	F+	F-	F+	F-	F+	F-	F+	F-
Bark	1.91±0.16	1.09±0.55	274.91*±85.65	219.77*±58.71	-0.73±0.40	-1.27±0.09	2.40*±0.79	2.28*±0.66
Wood	2.73±0.16	2.00±0.33	287.20*±37.01	347.07*±197.41	-0.27±0.42	-0.36±0.24	4.31*±0.78	2.87*±0.57
Branches	1.82±0.64	0.91±0.64	156.52*±29.07	168.90*±32.93	-1.37±0.42	-1.46±0.18	1.53*±0.41	0.77*±0.09
Leaves	3.00±0.55	2.09±0.74	165.89*±18.84	109.75*±38.46	-0.09±0.46	-0.18±0.36	0.96*±0.49	0.96*±0.08