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1 **How much inundation occurs in the Amazon River basin?**

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40 **Abstract**

41 The Amazon River basin harbors some of the world's largest wetland complexes, which are of major
42 importance for biodiversity, the water cycle and climate, and human activities. Accurate estimates of
43 inundation extent and its variations across spatial and temporal scales are therefore fundamental to
44 understand and manage the basin's resources. More than fifty inundation estimates have been generated for
45 this region, yet major differences exist among the datasets, and a comprehensive assessment of them is
46 lacking. Here we present an intercomparison of 29 inundation datasets for the Amazon basin, based on
47 remote sensing only, hydrological modeling, or multi-source datasets, with 18 covering the lowland
48 Amazon basin (elevation < 500 m, which includes most Amazon wetlands), and 11 covering individual
49 wetland complexes (subregional datasets). Spatial resolutions range from 12.5 m to 25 km, and temporal
50 resolution from static to monthly, spanning up to a few decades. Overall, 31% of the lowland basin is

51 estimated as subject to inundation by at least one dataset. The long-term maximum inundated area across
52 the lowland basin is estimated at $599,700 \pm 81,800$ km² if considering the three higher quality SAR-based
53 datasets, and $490,300 \pm 204,800$ km² if considering all 18 datasets. However, even the highest resolution
54 SAR-based dataset underestimates the maximum values for individual wetland complexes, suggesting a
55 basin-scale underestimation of ~10%. The minimum inundation extent shows greater disagreements among
56 datasets than the maximum extent: $139,300 \pm 127,800$ km² for SAR-based ones and $112,392 \pm 79,300$ km²
57 for all datasets. Discrepancies arise from differences among sensors, time periods, dates of acquisition,
58 spatial resolution, and data processing algorithms. The median total area subject to inundation in medium
59 to large river floodplains (drainage area > 1,000 km²) is 323,700 km². The highest spatial agreement is
60 observed for floodplains dominated by open water such as along the lower Amazon River, whereas
61 intermediate agreement is found along major vegetated floodplains fringing larger rivers (e.g., Amazon
62 mainstem floodplain). Especially large disagreements exist among estimates for interfluvial wetlands
63 (Llanos de Moxos, Pacaya-Samiria, Negro, Roraima), where inundation tends to be shallower and more
64 variable in time. Our data intercomparison helps identify the current major knowledge gaps regarding
65 inundation mapping in the Amazon and their implications for multiple applications. In the context of
66 forthcoming hydrology-oriented satellite missions, we make recommendations for future developments of
67 inundation estimates in the Amazon and present a WebGIS application ([https://amazon-](https://amazon-inundation.herokuapp.com/)
68 [inundation.herokuapp.com/](https://amazon-inundation.herokuapp.com/)) we developed to provide user-friendly visualization and data acquisition of
69 current Amazon inundation datasets.

70 **Key words:** flooding, surface water, floodplains, interfluvial wetlands

71

72 **1. Introduction**

73 Aquatic ecosystems cover extensive areas of the Amazon basin, and are associated with temporally
74 and spatially dynamic habitats such as floodable forests, savannas, grasslands, large and small
75 rivers, and lakes (Hess et al., 2015; Junk et al., 2011; Melack and Coe, 2021; Reis et al., 2019a).
76 These systems, hereafter called wetlands, support plants and animals that are adapted to the flood
77 pulse (Junk et al., 1989), play key roles in regional and global biogeochemical cycles, especially
78 the carbon cycle (Richey et al 1990; Dunne et al., 1998; Abril et al., 2014; Melack et al., 2004;
79 Pangala et al., 2017; Martínez-Espinosa et al., 2020), and regulate the riverine transport of
80 dissolved and particulate material, including sediment and organic matter (Armijos et al., 2020;
81 Fassoni-Andrade and Paiva, 2019; Melack and Forsberg, 2001; Ward et al., 2017). Additionally,
82 human settlements along Amazon wetlands (Blatrix et al., 2018; Denevan, 1996) benefit from
83 ecosystem services, including food provision from native plants and animals as well as crop and
84 livestock production (Coomes et al., 2016; Jardim et al., 2020).

85 Many of the wetlands of the Amazon basin are considered floodplain because they are subject to
86 seasonal or periodic inundation by river overflow (i.e., the flood pulse; Junk et al., 1989). The
87 region also hosts large interfluvial wetlands, which unlike fringing floodplains along large rivers,
88 are flooded mainly by local rainfall and runoff and characterized by shallow water (Belger et al.,
89 2011; Bourrel et al., 2009; Junk et al., 2011). Water sources, inundation patterns, and
90 geomorphology interact to determine the structure and function of these biodiverse ecosystems
91 (Junk et al., 2011; Latrubesse, 2012; Park and Latrubesse, 2017).

92 The extent of inundated land (also called flooded land or surface water extent), and its temporal
93 variation, are core variables to understand wetland processes and are of interest for multiple
94 scientific disciplines, including ecology (Silva et al., 2013; Hawes et al., 2012; Luize et al. 2015),
95 land-atmosphere interactions (Prigent et al., 2011; Santos et al., 2019; Taylor et al., 2018), carbon

96 cycling and greenhouse gas emissions (Guilhen et al., 2020; Melack et al., 2004; Richey et al.,
97 2002), and natural hazard management (Restrepo et al., 2020; Trigg et al., 2016). The Amazon
98 basin has been a focus for remote sensing developments and applications in hydrology (Fassoni-
99 Andrade et al., 2021), especially for inundation estimation, given the basin’s large scale and global
100 environmental relevance, relatively pristine landscape, and technical challenges posed by
101 persistent cloud cover (Asner, 2001) and dense vegetation. This resulted in the development of
102 more than 50 inundation maps and datasets for this region in recent decades. Tables 1 (datasets
103 used in this study) and S1 (datasets not used due to redundancy or unavailability) summarize most
104 of the datasets developed for mapping inundation in the Amazon basin.

105 Digital wetland maps were first produced for the Amazon basin by Matthews and Fung (1987)
106 from aeronautical charts. Optical remote sensing systems in the visible or thermal spectral range,
107 such as Landsat, are of limited value for most Amazon wetlands, since inundation under persistent
108 cloud cover and dense vegetation canopies can be difficult to detect. Because of this, microwave
109 systems have been employed. Large-scale inundation mapping was pioneered in the region through
110 analysis of Scanning Multi-channel Microwave Radiometer (SMMR) and Special Sensor
111 Microwave/Imager (SSM/I) passive microwave observations, which provided all-weather
112 capability and sensitivity to inundation even in the presence of partial vegetative cover (Hamilton
113 et al., 2002; Prigent et al., 2001; Sippel et al., 1998). Meanwhile, research demonstrated the all-
114 weather capability and superior spatial resolution of synthetic aperture radar (SAR) systems. L-
115 band SAR that can penetrate forest canopies and reveal underlying water through the “double
116 bounce” effect was shown to be promising for mapping inundation in the Amazon (Hess et al.,
117 2003). More specifically, the high-resolution, dual-season classification of the Japanese Earth
118 Resources Satellite-1 (JERS-1) L-band SAR data for the entire lowland Amazon basin by Hess et

119 al. (2015), validated with airborne videography images, has been used as a benchmark for the
120 inundation extent of Amazon wetlands. Since these initial studies, and with the availability of other
121 imagery (e.g., Advanced Land Observing Satellite (ALOS) 1 and 2 missions), the remote sensing
122 community seeking to map and characterize inundation employed various combinations of active
123 and passive microwave data to benefit from the higher spatial resolution of the former and the
124 higher temporal resolution of the latter (Aires et al., 2013; Jensen and McDonald, 2019; Papa et
125 al., 2010; Parrens et al., 2019, 2017; Prigent et al., 2007, 2020; Schroeder et al., 2015).

126 Besides the basin-scale mappings (which, in our context, refer to both basin-scale datasets and
127 those that cover only the lowland areas below 500 m.a.s.l. elevation) of annual maximum and
128 minimum inundation (Chapman et al., 2015; Hess et al., 2015; Rosenqvist et al., 2020), dynamic
129 datasets with high spatial and temporal resolution are mainly based on satellite passive microwave
130 observations of coarse spatial resolution (Global Inundation Extent Multi-Satellite – GIEMS),
131 Surface Water Microwave Product Series (SWAMPS), Surface Water Fraction (SWAF), Wetland
132 Area and Dynamics for Methane Modeling (WAD2M) datasets; see Table 1), which can be
133 downscaled using ancillary data (Aires et al., 2017, 2013; Parrens et al., 2019). Basin-scale,
134 dynamic inundation estimates based on the ALOS satellite are limited given its low temporal
135 resolution (repeat cycle of 46 days). Thus, some studies have analyzed time series of ALOS-
136 Phased Array L-band Synthetic Aperture Radar (PALSAR) (Arnesen et al., 2013; Ferreira-Ferreira
137 et al., 2015) and ALOS-2 PALSAR-2 backscatter retrievals (Jensen et al., 2018) for subsets of
138 Amazon wetlands. However, with a few exceptions using subregional datasets (Arnesen et al.,
139 2013; Ferreira-Ferreira et al., 2015; Hess et al., 2003; Jensen et al., 2018; Resende et al., 2019), in
140 situ validation of the basin-scale estimates has seldom been performed, given the remoteness of

141 much of the Amazon basin and the often dense forest cover, which hampers airborne monitoring
142 of below-canopy inundation.

143 Complementary to the remotely sensed datasets, process-based hydrological models estimating
144 variables such as river discharge and flood extent have been developed and assessed from basin to
145 local scales in the major rivers of the basin (Beighley et al., 2009; Coe et al., 2008; Getirana et al.,
146 2017, 2012; Hoch et al., 2017; Luo et al., 2017; Miguez-Macho and Fan, 2012; Paiva et al., 2013;
147 Yamazaki et al., 2011), thanks to the advent of new computational and modeling capabilities.
148 Local-scale hydraulic models with coarse (Trigg et al., 2009; Wilson et al., 2007; Fleischmann et
149 al., 2020) and detailed input data (Ji et al., 2019; Pinel et al., 2019; Rudorff et al., 2014; Fassoni-
150 Andrade, 2020) have further developed model capabilities for mapping inundation dynamics,
151 especially for the floodplains fringing the Amazon mainstem. These models complement satellite-
152 based flood mapping due to their higher temporal and spatial resolution, and capability to estimate
153 long-term time series, for both past and future (e.g., due to climate change) scenarios. The
154 understanding of their uncertainties can lead to optimal data fusion with satellite-based estimates,
155 such as considering multiple constraints within the water cycle representation (Pellet et al., 2021).

156 Among these numerous inundation datasets for the Amazon basin (Tables 1 and S1), divergences
157 can be substantial due to the differences in sensor systems, timing, and data processing algorithms
158 (Aires et al., 2018; Fleischmann et al., 2020; Parrens et al., 2019; Pham-Duc et al., 2017;
159 Rosenqvist et al., 2020), and a comprehensive assessment of inundation estimates for the Amazon
160 is lacking. The need to compare different hydrological datasets for the Amazon has been recently
161 highlighted in the context of river discharge (Towner et al., 2019), precipitation (Wongchuig et al.,
162 2017; Zubieta et al., 2019) and evapotranspiration (Paca et al., 2019; Wu et al., 2020). Meanwhile,
163 rapid environmental changes in the basin underscore the urgency for a better understanding of

164 Amazon water resources (Fassoni-Andrade et al., 2021), for which management and planning can
165 be hindered by the discrepancies among datasets. These questions regarding current data
166 limitations in the largest basin in the world are also timely in anticipation of forthcoming
167 hydrological satellite missions such as Surface Water and Ocean Topography (SWOT) and NASA-
168 ISRO SAR (NISAR).

169 To better understand and quantify the state of understanding of inundation patterns in the Amazon
170 wetlands, we address the following questions: 1) How much Amazon land area is subject to
171 seasonal or permanent flooding, and how accurate are the estimates? 2) Which areas are in
172 particular disagreement and thus deserve further attention? 3) How do basin-scale estimates with
173 coarser resolution and less calibrated classification methods differ from those for individual
174 wetland complexes, with independent validation? 4) How do the various inundation estimation
175 approaches (optical imagery, SAR, passive microwave, hydrologic models) differ in terms of
176 inundation mapping and for different wetland types (e.g., floodplains and interfluvial areas)? In
177 order to answer these questions, we gathered 29 inundation datasets for the Amazon basin,
178 spanning a wide range of spatial (12.5 m to 25 km) and temporal (static, dual-season, monthly,
179 daily) resolutions, and coverages from the whole basin to individual wetland complexes (Table 1),
180 into a framework that provides a comprehensive assessment of current knowledge of Amazon
181 inundation.

182

183 Table 1. List of 29 studies that mapped inundation over areas ranging from the entire Amazon basin to individual
184 wetland complexes. These data sources were selected based on data availability and relevance for this intercomparison.
185 In the case of hydrological models, time resolutions are the values assessed or provided by the models, which can be
186 provided at finer time resolution if necessary, since many of them compute flood maps at daily or sub-daily time steps

187 and report time-integrated results. The column “Data type” refers to: OS: optical sensor; SAR: synthetic aperture
 188 radar; HM: hydrological model; HR: multiple datasets at high resolution; CR: multiple datasets at coarse resolution.
 189 The column “Type of inundation estimated” has three classes: “All”, meaning both open water and vegetated wetlands,
 190 “Open water”, and “Wetland only (no open water)”.

Data type	Dataset name and main mission/ model associated (if applicable)	Spatial resolution	Temporal resolution	Time period	Region	Type of inundation estimated	Reference
CR	GIEMS-2	25 km	Monthly	1992-2015	Basin	All	Prigent et al., 2020
CR	SWAMPS	25 km	Monthly	1992-2020	Basin	All	Jensen and McDonald, 2019
CR	WAD2M	25 km	Monthly	2000-2018	Basin	Wetland only (no open water)	Zhang et al., 2020
HR	GIEMS-D3	90 m	Monthly	1993-2007	Basin	All	Aires et al., 2017
HR	CIFOR	232 m	Static (max inundation)	1950-2000	Basin	All	Gumbrecht et al., 2017
HR	ESA-CCI	300 m	Annual	1992-2015	Basin	All	Bontemps et al., 2013
HR	GIEMS-D15	500 m	Monthly climatology	1993-2004	Basin	All	Fluet-Chouinard et al., 2015
HR	GLWD	1 km	Static	1992-2004	Basin	All	Lehner and Döll, 2004

HR	SWAF-HR / SMOS mission	1 km	Weekly to monthly	2010-2020	Basin	All	Parrens et al., 2019
HM	THMB model	5-min	Monthly	1961-2010	Basin	All	Coe et al., 2008
HM	CaMa-Flood model	500 m	Monthly	1980-2014	Basin	All	Yamazaki et al., 2011
HM	MGB model	500 m	Monthly	1980-2015	Basin	All	Siqueira et al., 2018
HM	Bonnet model	180 m	Monthly	2006-2019	Janauacá	All	Bonnet et al., 2017
HM	TELEMAC-2D model	30 m	Monthly	2006-2015	Janauacá	All	Pinel et al., 2019
HM	LISFLOOD-FP model	90 m	Monthly	1994-2015	Curuai	All	Rudorff et al., 2014
OS	G3WBM / Landsat mission	30 m	Static (open water areas)	1990-2010	Basin	Open water	Yamazaki et al., 2015
OS	GLAD / Landsat mission	30 m	Annual and monthly climatology	1999-2018	Basin	Open water	Pickens et al., 2020
OS	GSWO / Landsat mission	30 m	Monthly (cloud cover may occur)	1984-2019	Basin	Open water	Pekel et al., 2016
OS	Ovando / MODIS mission	500 m	8 days	2001-2014	Llanos de Moxos	Open water	Ovando et al., 2016
OS	Park / MODIS mission	230 m	Monthly climatology	2000-2015	Amazon River down-	Open water	Park and Latrubesse, 2019

					stream of Manaus		
SAR	Hess / JERS-1 mission	90 m	Max. and min. annual inundation (dual season)	1995- 1996	Basin (lowlands)	All	Hess et al., 2003, 2015
SAR	Chapman / ALOS- PALSAR mission	90 m	Monthly	2006- 2011	Basin	All	Chapman et al., 2015
SAR	Rosenqvist / ALOS-2 PALSAR- 2	50 m	Max. and min. annual inundation (dual season)	2014- 2017	Basin	All	Rosenqvist et al., 2020
SAR	Jensen / ALOS-2 PALSAR-2 mission	50 m	Irregular (26 images)	2014- 2018	Pacaya- Samiria	All	Jensen et al., 2018
SAR	Arnesen / ALOS- PALSAR mission	90 m	Irregular (12 images)	2006- 2010	Curuai	All	Arnesen et al., 2013
SAR	Ferreira-Ferreira / ALOS-PALSAR mission	12.5 m	Flood frequency only	2007- 2010	Mamirauá	All	Ferreira-Ferreira et al., 2015
SAR	Ovando-2 / ALOS- PALSAR mission	100 m	Irregular (6 images)	2006- 2010	Llanos de Moxos	All	Ovando et al., 2016
SAR	Pinel-2 / ALOS- PALSAR mission	30 m	Irregular (16 images)	2007- 2011	Janauacá	All	Pinel et al., 2019
SAR	Resende / ALOS- PALSAR mission	25 m	Static (max inundation)	2006- 2011	Uatumã	All	Resende et al., 2019

191

192

193 **2. Methodology**

194 **2.1 Study area**

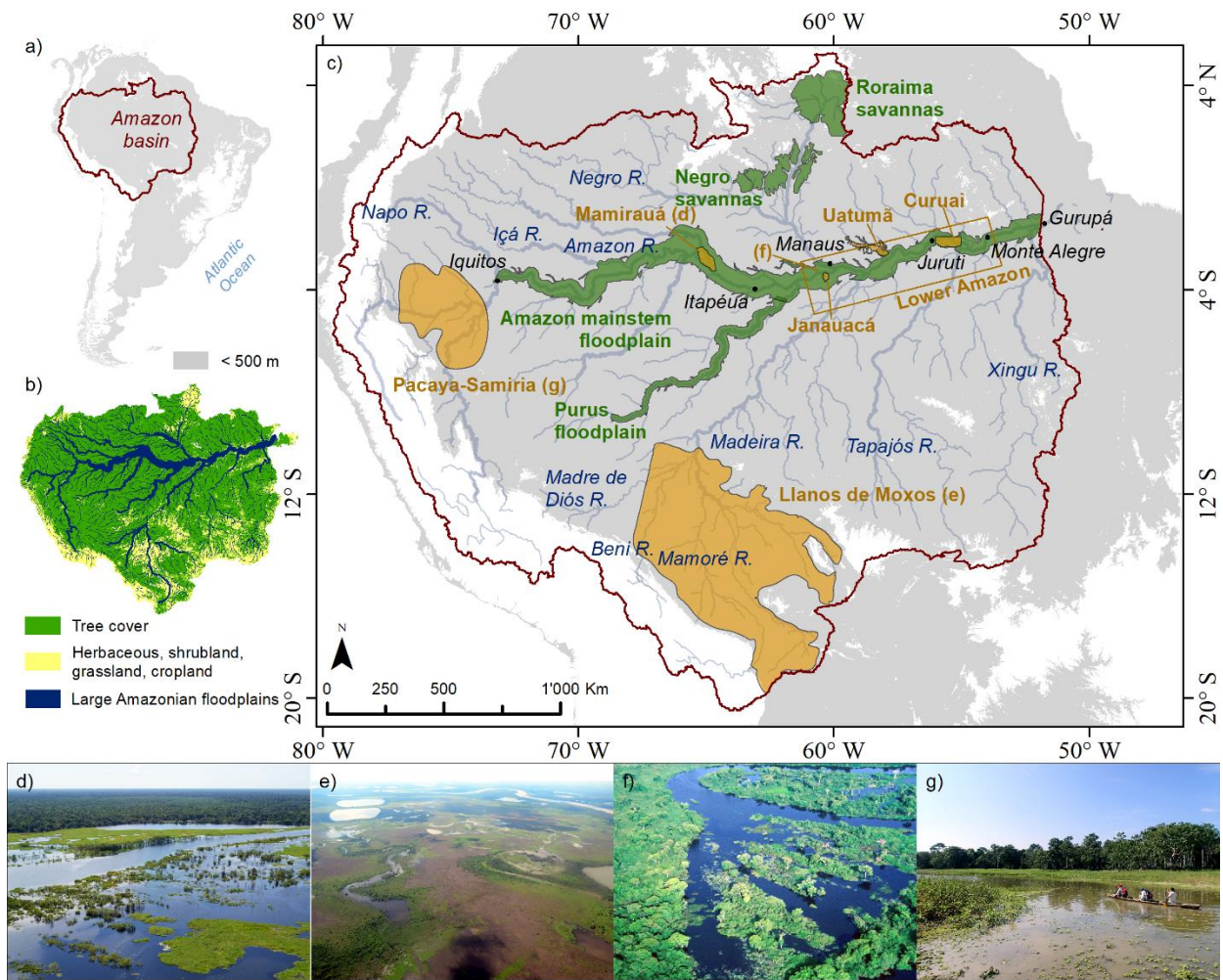
195 The Amazon basin spans around 6 million km² in nine South American countries (Figure 1), with
196 high annual rainfall ($\sim 2,200$ mm year⁻¹), and the Amazon River discharge makes a major
197 contribution to global freshwater and sediment exports to the ocean (Fassoni-Andrade et al., 2021).
198 We delineated the catchment area upstream from Gurupá city, within the tidal river ~ 390 km from
199 the ocean; hence not including the Tocantins-Araguaia basin and parts of the Amazon estuary and
200 Marajó Island. We selected the 5.11×10^6 km² of Amazon lowlands defined as areas lower than
201 500 m elevation based on the Shuttle Radar Topography Mission Digital Elevation Model (SRTM
202 DEM) for the area of dataset comparisons in our study. This decision is consistent with several
203 studies limited to lowlands because of the limitations of certain methods in estimating flooding in
204 mountainous terrain (Hess et al., 2015).

205 In addition to basin-scale datasets, estimates of inundated areas for 11 individual wetland
206 complexes (also referred to as “subregional”) in the Amazon basin were analyzed, including seven
207 areas for which more detailed estimates were available. This was performed to understand how the
208 basin-scale datasets may vary in accuracy across different wetland types (Figure 1): Curuai
209 floodplain lake (Arnesen et al., 2013; Rudorff et al., 2014), Janauacá floodplain lake (Bonnet et
210 al., 2017; Pinel et al., 2019), Uatumã river floodplain (Resende et al., 2019), Mamirauá Reserve
211 (Ferreira-Ferreira et al., 2015), Pacaya-Samiria wetlands (Jensen et al., 2018), Llanos de Moxos

212 wetlands (Ovando et al., 2016), lower Amazon floodplain (Park and Latrubesse, 2019), Amazon
213 mainstem floodplain (from Iquitos to Gurupá), Purus floodplain, Roraima savannas, and Negro
214 savannas. A brief summary of these wetlands is provided in supplementary Table S2, and their
215 main features are summarized in the following. Curuai is representative of the shallow lakes in the
216 lower Amazon floodplain. It is separated from the river by narrow levees (Rudorff et al., 2014)
217 and has a high suspended sediment concentration. Janauacá is typical of the middle Amazon River
218 floodplain, and is composed of a ria lake (i.e., a blocked valley lake with relatively sediment-free
219 waters; Latrubesse (2012)) and “várzea” environments (white-water floodplains) in its northern
220 part (Pinel et al., 2019). Uatumã River is an Amazon tributary with black-water floodplain
221 (“igapó”), and includes the Balbina hydroelectric reservoir, operating since 1987, which affects
222 the river’s hydrological regime (Schöngart et al., 2021). The Uatumã floodplain reach assessed
223 here is the 300-km reach between Balbina dam and the confluence with the Amazon River. The
224 Mamirauá Sustainable Development Reserve is located in the confluence between Solimões and
225 Japurá rivers, and is characterized by a mosaic of “chavascal”, herbaceous, and low and high
226 várzea vegetation (Ferreira-Ferreira et al., 2015). The Purus River is a major tributary, and its
227 floodplain was chosen because of its large floodplain to river width ratio. Pacaya-Samiria wetlands
228 are composed of flooded forests, palm swamps and peatlands in the upper Solimões River (Draper
229 et al., 2014; Lähteenoja et al., 2012). The Llanos de Moxos floodable savannas occupy the
230 interfluvial areas between the Beni, Mamoré and Madre de Dios rivers in the upper Madeira basin
231 (Hamilton et al., 2004). The Negro savannas, locally known as “campina wetlands” and
232 “campinarana wetlands”, depending on the vegetation density, are thought to have formed from
233 regional neotectonic depressions and were called the “Septentrional Pantanal” given their large
234 area (Rossetti et al., 2017a, 2017b; Santos et al., 1993). The Roraima floodable savannas extend

235 from Roraima State in Brazil to the Rupununi savannas in Guyana, and comprise mainly smaller
 236 river floodplains interspersed with poorly drained interfluvial savannas subject to flooding by local
 237 rainfall (Hamilton et al., 2002); here we only considered the Roraima wetlands in the upper Branco
 238 River basin, which is within the Amazon basin.

239



240

241 Figure 1. The Amazon basin and its major wetland systems: (a) Amazon basin delineation (red lines) over the countries
 242 of South America (black lines). (b) Land cover based on a 2010 map from the European Space Agency Climate Change
 243 Initiative (ESA-CCI) (Bontemps et al., 2013), showing the distribution of forest and savanna across the basin, as well
 244 as large floodplains (see methodology section 2.3). (c) Basin distribution of major wetland systems showing locations

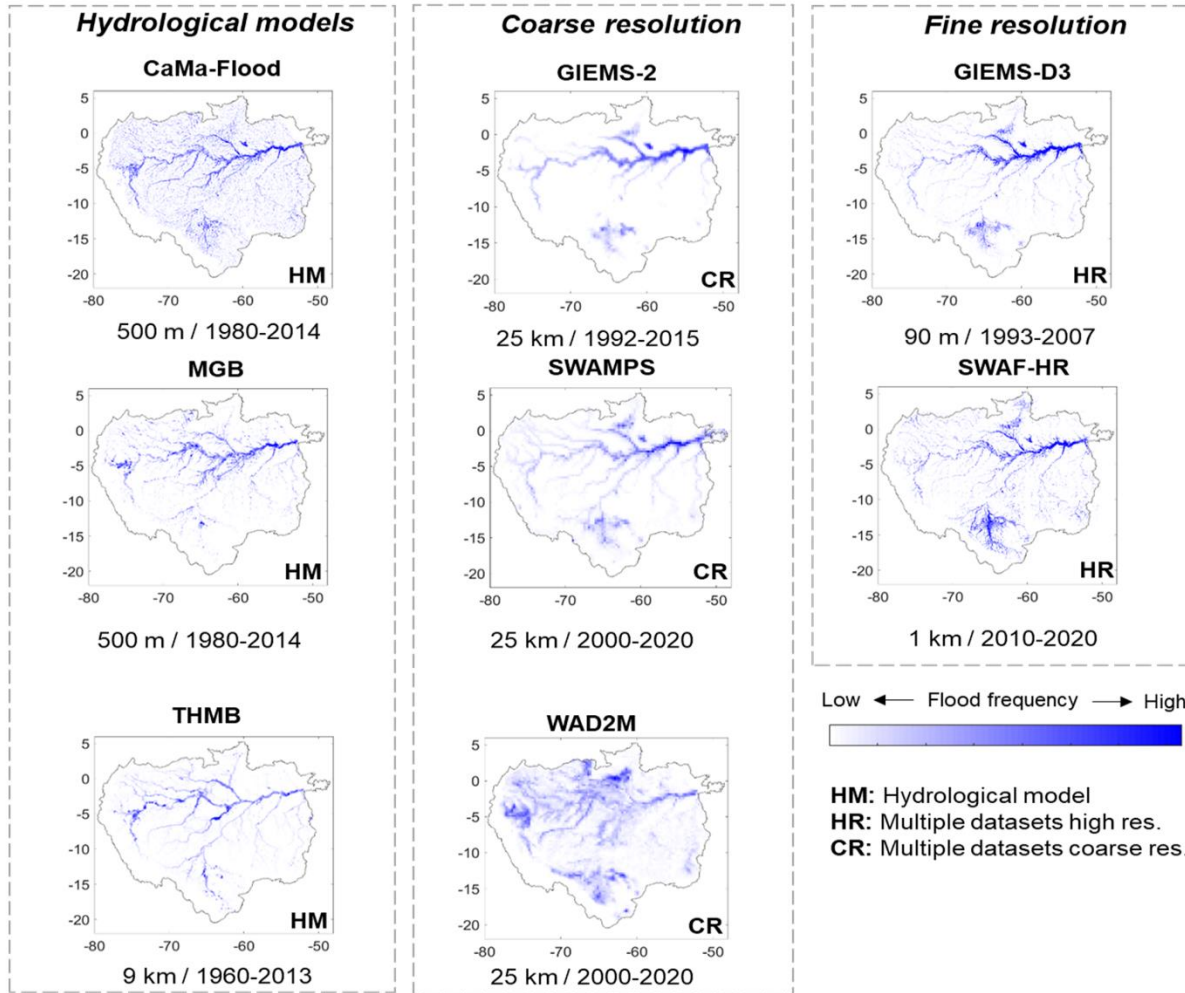
245 of interest for this study. Elevations lower than 500 m are shown in grey (based on SRTM DEM). The orange polygons
246 show the areas for which a subregional dataset was available for this study (Figure 4), and the green ones show wetland
247 areas of interest that do not have datasets specifically designed for these subregions. Photos depicting different wetland
248 complexes for (d) Mamirauá (courtesy of João Paulo Borges Pedro), (e) Llanos de Moxos (courtesy of Alex Ovando),
249 (f) Cabaliana floodplain lake close to Manacapuru (courtesy of Stephen Hamilton), and (g) Pacaya-Samiria (courtesy
250 of Katherine Jensen) regions, respectively.

251

252 **2.2 Datasets**

253 Twenty-nine inundation datasets covering areas ranging from the whole-basin scale to individual
254 wetland complexes, based on multiple data sources and spatiotemporal resolutions, were
255 assembled for our comparison (Table 1). Most of these datasets are recent, with 18 out of the 29
256 published since 2016, and 27 since 2011. They were chosen due to data availability and
257 representativeness; other datasets that were either unavailable or methodologically redundant to
258 those in our comparison were not used but are catalogued in Table S1. Overall, there are eight
259 dynamic (weekly to monthly; Figure 2) and 10 static (which include long-term maximum, annual
260 or dual-season categories; Figure 3) basin-scale datasets.

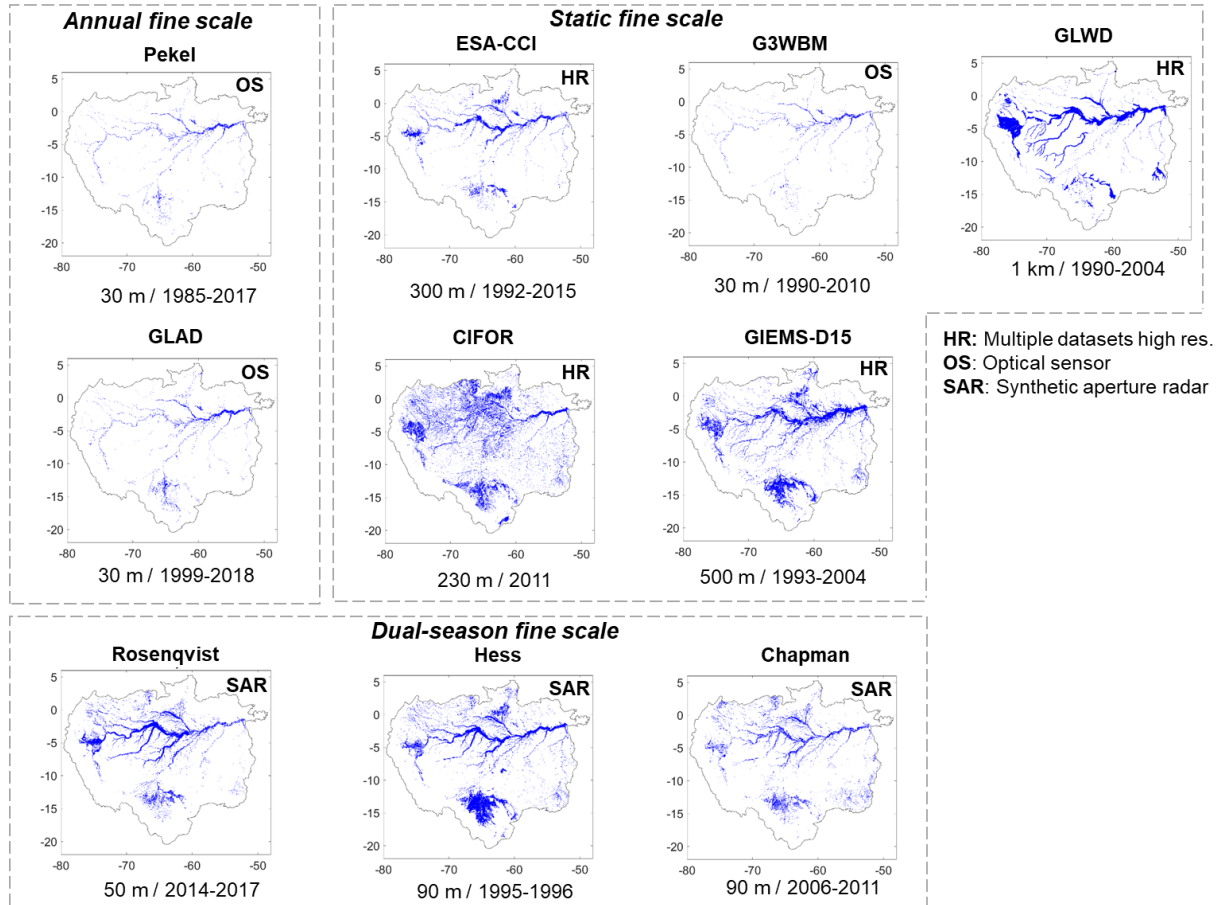
261



262

263 Figure 2. Basin-scale, dynamic inundation datasets used in this study, divided into three classes (hydrological models;
 264 merging of multiple datasets at high resolution; merging of multiple datasets at coarse resolution). Long-term flood
 265 frequency maps are provided for each dataset, calculated as the percentages of observations labelled as flooded
 266 throughout the entire time-series.

267



268

269 Figure 3. Basin-scale, static or dual-season inundation datasets used in this study, divided into three classes (merging
 270 of multiple datasets at high resolution; based on optical sensors; and based on SAR data). Flood frequency maps are
 271 not provided because the datasets are mainly static or annual-based.

272

273 Passive microwave (PM) data are the basis of SWAF-HR, GIEMS family (GIEMS-D15, GIEMS-
 274 D3, GIEMS-2), and SWAMPS, while ancillary data (i.e., optical imagery and microwave
 275 scatterometry) are used to complement the PM signal. SWAF-HR data result from the
 276 disaggregation of water surface fraction in a dataset at coarser spatial resolution (SWAF), based
 277 on L-band passive microwave observations from the Soil Moisture and Ocean Salinity (SMOS)
 278 satellite (Parrens et al. 2017). The disaggregation of SWAF relies on water occurrence maps from

279 GSWO and the Digital Elevation Model (DEM) Multi-Error-Removed-Improved-Terrain
280 (MERIT) (Parrens et al., 2019). A global implementation of SWAF based on multi-angular and
281 multi-polarization information has also been implemented (Al Bitar et al. 2020). GIEMS merges
282 multiple satellite passive and active microwave observations, along with the optically-derived
283 NDVI (Normalized Difference Vegetation Index), to detect the surface water and estimate the
284 vegetation attenuation, for a monthly quantification of the surface water extent at ~25 km spatial
285 resolution (Prigent et al., 2001, 2007, 2020; Papa et al., 2010). It is further disaggregated at 90-m
286 resolution (GIEMS-D3) using a topographical downscaling methodology (Aires et al. 2017).

287 Three basin-scale datasets are based mainly on SAR data from JERS-1 (Hess et al., 2003, 2015),
288 and its successor missions ALOS-PALSAR (Chapman et al., 2015) and ALOS-2 PALSAR-2
289 (Rosenqvist et al., 2020). These three datasets cover different decades of observation but are
290 methodologically similar.

291 Three of the optical-based datasets are based on Landsat data: GSWO (Pekel et al., 2016),
292 G3WBM (Yamazaki et al., 2015) and GLAD (Pickens et al., 2020). Although GSWO and GLAD
293 can provide monthly estimates for the Landsat archive (1984-today), given the inability of optical
294 data to estimate flooding under cloud cover or dense vegetation canopies, only annual maximum
295 and minimum values are used. For GLAD and GSWO, we consider a threshold of occurrence of
296 surface water of 95% to estimate the minimum inundation (i.e., for the permanently inundated
297 areas; Aires et al., 2018); otherwise, only a few isolated open water areas would be considered for
298 the minimum extent.

299 The European Space Agency Climate Change Initiative dataset (ESA-CCI) is based on surface
300 reflectance from MERIS, the Advanced Very High-Resolution Radiometer (AVHRR) and

301 PROBA-V data and Global Water Bodies from the Envisat Advanced Synthetic Aperture Radar
302 (ASAR) (Bontemps et al., 2013). Since the wetland pixels in ESA-CCI varied negligibly
303 throughout the years of observations, we use only the 2010 dataset as the ESA-CCI estimate for
304 maximum inundation.

305 Another set of data is based on the merging of multiple global datasets: GLWD, GIEMS-D15 and
306 WAD2M. GLWD is one of the first globally consistent databases of wetlands, which was based
307 on a collection of wetland estimates from diverse institutions worldwide (Lehner and Döll, 2004).
308 GIEMS-D15 combines GLWD, the Hydrosheds drainage network, and Global Land Cover 2000.
309 WAD2M is based on SWAMPS and CIFOR within its merging framework. WAD2M is the only
310 dataset to exclude open water areas (removal based on GSWO) due to its goal of estimating
311 wetland methane emissions. SWAF-HR (Parrens et al., 2019) and GIEMS-D3 (Aires et al., 2017)
312 use additional data and methodologies to downscale the original 25-km passive microwave-based
313 SWAF (Parrens et al., 2017) and GIEMS (Papa et al., 2010; Prigent et al., 2007) datasets to 1 km
314 and 90 m, respectively. While GIEMS-D3 has a different inundation magnitude than the original
315 GIEMS due to merging with ancillary data, SWAF-HR conserves the same inundation magnitude
316 across scales.

317 Among hydrological models, we selected representative datasets from each of the following broad
318 modeling types: 1) process-based hydrologic models that use flood routing to represent inundation
319 processes (i.e., from a simple kinematic wave model coupled to an inundation method to more
320 complex flow routing methods); or 2) hydraulic (or hydrodynamic) models that consider the
321 shallow water equations (or its simplifications) at any dimension (1D, 2D or 3D). For our analysis,
322 we adopted two basin-scale models – one hydrologic (THMB; Coe et al. (2008)) and one
323 hydrologic-hydrodynamic (MGB, Siqueira et al. (2018)), as well as a global-scale hydrodynamic

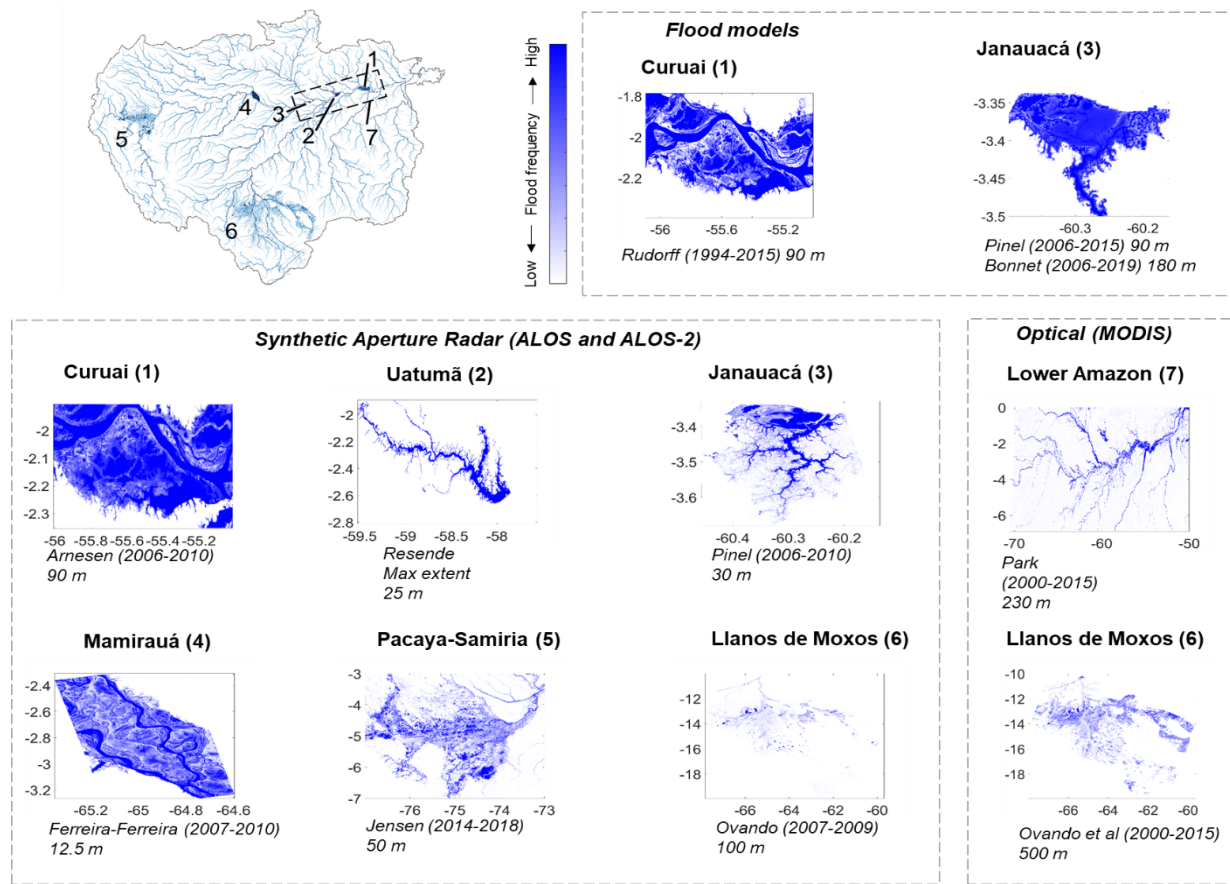
324 model (CaMa-Flood, Yamazaki et al. (2011)), in the Earth2Observe version available at
325 <http://www.earth2observe.eu/>). The inundated area estimation is largely affected by the DEMs.
326 The DEMs adopted in the model runs were: Bare-Earth (O’Loughlin et al., 2016) for MGB,
327 MERIT (Yamazaki et al., 2017) for CaMa-Flood, and SRTM (Farr et al., 2007) for THMB. The
328 rainfall/runoff input data are MSWEP v.1.1 daily precipitation (Beck et al., 2017) for MGB,
329 HTESSEL daily runoff (Balsamo et al., 2009) for CaMa-Flood, and CRU TS v.3.2.1 monthly
330 precipitation (Harris et al. 2014) for THMB. Although other hydrologic models have been applied
331 to the Amazon basin (Tables 1 and S1), the models chosen here were selected as representative of
332 global to local models, for having been well validated and applied over the Amazon basin, and for
333 representing state-of-the-art Amazon hydrologic modeling. All basin-scale models represent one-
334 dimensional (1D) flows only (i.e., floodplains are represented as storage units without active flow),
335 and thus do not represent 2D surface flows that occur in wetlands (Alsdorf et al., 2007;
336 Fleischmann et al., 2020). A detailed comparison of model capabilities and structural uncertainties
337 is beyond our current scope. Hydrologic models have different temporal resolution depending on
338 their numerical stability and forcing data. For instance, MGB and CaMa-Flood models run at an
339 adaptive time step (sub-minute timestep in the case of MGB), but are assessed at daily resolution
340 given their daily precipitation forcing. We aggregated the models’ estimates to monthly averages
341 to make them comparable to the remote sensing dynamic datasets.

342 The datasets available for individual wetland complexes are presented in Figure 4. ALOS-2
343 PALSAR-2 data were used for the Pacaya-Samiria region (Jensen et al., 2018), and the ScanSAR
344 mode of ALOS/PALSAR for the following datasets: Curuai floodplain lake (Arnesen et al., 2013),
345 Mamirauá Reserve (Ferreira-Ferreira et al., 2015), Uatumã river floodplain (Resende et al., 2019),
346 and Janauacá floodplain lake (Pinel et al., 2019). MODIS optical data were used for the Llanos de

347 Moxos savannas in the upper Madeira River basin (Ovando et al., 2016) and the lower Amazon
348 floodplain (Park and Latrubesse, 2019). Two local-scale 2D hydraulic models (LISFLOOD-FP for
349 Curuai lake, Rudorff et al. (2014), and TELEMAC-2D for Janauacá lake, Pinel et al. (2019)), and
350 one local-scale hydrologic model (for Janauacá lake; Bonnet et al. (2017)) were considered;
351 together, these are representative of the state-of-the-art of hydrological modeling in Amazon
352 wetlands.

353 The datasets were stored in various formats (i.e., raster and polygon shapefiles) and projections
354 (mainly projected UTM and geographic coordinate system with WGS84 datum), and were
355 converted to the WGS84 geographic coordinate system to compute areas. SWAMPS was provided
356 at the Equal-Area Scalable Earth (EASE) Grid, which was used to estimate its flooded areas.
357 Hydrologic model outputs were provided as either binary inundation maps or flood depth raster
358 files, which were then converted into binary maps by assuming depth > 0 m as inundated pixels.

359



360

361 Figure 4. Long-term flood frequency maps from subregional inundation datasets (i.e., for individual wetland
 362 complexes) used in this study. The Uatumã dataset (2) is static and is displayed as the maximum extent. Flood
 363 frequency maps are produced by computing the long-term average of all inundation maps available for each dataset.

364

365 2.3 Comparison framework

366 The comparison framework involved the following analyses, considering the entire basin and 11
 367 wetland complexes (seven areas with available subregional estimates, and four additional areas of
 368 interest without subregional estimates; Figure 1):

- 369 ● Annual maximum and minimum inundation estimates for each of the 18 basin-scale
370 datasets (section 3.1);
- 371 ● Basin-scale, long-term maximum and minimum inundation estimates for each of the 18
372 basin-scale datasets (section 3.1);
- 373 ● Long-term maximum and minimum inundation estimates for each of the 18 basin-scale
374 and 11 subregional datasets (section 3.2);
- 375 ● Comparison between basin-scale and subregional datasets with temporal (nRMSD and
376 Pearson correlation) and spatial (Fit metric) assessment (section 3.2);
- 377 ● Assessment of spatial agreement among the 18 basin-scale datasets at 1 km, for both long-
378 term maximum and minimum inundation maps (section 3.3);
- 379 ● Estimation of long-term maximum inundation for two classes of wetlands for the entire
380 basin: (i) medium to large river floodplains and (ii) interfluvial wetlands and small
381 floodplains (section 3.4).

382

383 The long-term maximum and minimum inundation extents were computed for each dataset as the
384 area of all pixels that were inundated at least once in the whole monthly time series, for the
385 maximum, and as those pixels that were always inundated, for the minimum. We stress that
386 analyzing long-term changes in inundation patterns is beyond the scope of this study, and thus we
387 assumed stationarity in our comparisons of long-term maximum and minimum inundation extents
388 from different time-periods.

389 The agreement of all basin-scale, high-resolution datasets (i.e., all basin-scale ones except for
390 THMB, GIEMS-2, SWAMPS and WAD2M, which have a coarse resolution between 9 and 25

391 km) was assessed for long-term maximum and minimum inundation at 1 km resolution, which is
392 the resolution of SWAF-HR, the coarsest resolution among the high-resolution datasets. For each
393 1 km pixel, the total number of datasets agreeing that it was inundated (either for maximum or
394 minimum extent) was computed, following Trigg et al. (2016). Given the size of the Amazon basin,
395 a 1 km resolution was considered adequate for the analysis. The analysis was done by aggregating
396 all datasets to 1 km, and considering that a 1 km pixel is flooded if more than 50% of its area is
397 flooded (following Hamilton et al., 2002). A sensitivity test was performed using a 25% threshold
398 and led to similar conclusions at the whole basin scale (Figure S1).

399 The basin-scale and four additional subregional datasets were compared to seven subregional ones,
400 which were used as independent validation datasets, and cover the following sites: Curuai
401 (Arnesen et al., 2013), Uatumã (Resende et al., 2019), Janauacá (Pinel et al., 2019), Mamirauá
402 (Ferreira-Ferreira et al., 2015), Pacaya-Samiria (Jensen et al., 2018), Llanos de Moxos MODIS
403 (Ovando et al., 2016) and lower Amazon River (Park and Latrubesse, 2019). Varying degrees of
404 validation exercises were performed for these validation datasets, with some being extensively
405 validated with airborne videography (Hess et al., 2003) or local surveys (Arnesen et al., 2013;
406 Ferreira-Ferreira et al., 2015; Jensen et al., 2018; Resende et al., 2019), while others were assessed
407 through comparisons with other datasets (Pinel et al., 2019), or visually inspected, as in the large
408 domains of the Llanos de Moxos (Ovando et al., 2016) and lower Amazon River (Park and
409 Latrubesse, 2019) subregional datasets. The four additional subregional datasets are: Curuai
410 LISFLOOD-FP model (Rudorff et al., 2014), Janauacá hydrological model (Bonnet et al., 2017),
411 Janauacá TELEMAC-2D model (Pinel et al., 2019), and Llanos de Moxos ALOS-PALSAR
412 (Ovando et al., 2016).

413 To use the subregional studies to assess the accuracy of the datasets covering broader areas, the
414 basin-scale and four additional subregional datasets were compared to the subregional validation
415 datasets at monthly temporal resolution, considering the total inundated area per wetland area (i.e.,
416 the whole Curuai Lake domain, the whole Uatumã floodplain, and so forth). The polygons of each
417 wetland area, which were used to extract the information from the basin-scale datasets, were
418 delineated as a 1-km buffer around the maximum inundated area, according to each subregional
419 dataset. For the four areas of interest without subregional datasets (Amazon mainstem and Purus
420 floodplains, and Roraima and Negro wetlands), the polygons were created considering the
421 maximum lateral extent in accordance with the MERIT DEM (Yamazaki et al., 2017) and ESA-
422 CCI land cover for savannas. The time series were compared with Pearson linear correlation (R)
423 and the normalized root mean square deviation (nRMSD), computed as the RMSD between a given
424 inundation map and the subregional validation map (i.e., the individual wetland complexes)
425 divided by the subregional long-term average inundation. The term ‘deviation’ was preferred over
426 ‘error’ to stress the uncertainties inherent to all datasets, for both basin and subregional scales,
427 although those derived for an individual wetland complex are considered as superior in accuracy
428 for having a more dedicated data processing for that particular area, and being validated with
429 ground surveys in some cases.

430 The ability of a particular dataset to estimate the local spatial patterns at maximum inundation was
431 assessed with the Fit metric (Bates and De Roo, 2000), which has been successfully applied to
432 compare inundation datasets (Bernhofen et al., 2018), and is computed as:

433
$$Fit = 100\% * \frac{A \cap B}{A \cup B} (1)$$

434 Where A and B are the subregional validation dataset estimates (e.g., the subregional map that
435 corresponds to maximum inundation) and the basin-scale maximum inundation maps.

436 To assess different wetland environments, we differentiate medium to large river floodplains from
437 interfluvial wetlands and small floodplains. An estimation of the total flooded area of large river
438 floodplains was computed, considering river reaches with upstream drainage area larger than 1,000
439 km², and a buffer mask around the river reaches (mask presented in Figure 1). The buffer was
440 defined based on the Hydrosheds drainage network (Lehner and Grill, 2013), segmented into 15
441 km-long reaches as in Siqueira et al. (2018). The buffer was proportional to the local reach drainage
442 area and further manually adjusted to include the maximum floodplain lateral extent, as estimated
443 from a visual inspection of the MERIT DEM (Yamazaki et al., 2017) and the three basin-scale
444 SAR-based datasets (Hess, Chapman and Rosenqvist datasets). Buffer values varied from 4 km in
445 upper reaches to 150 km on the Amazon mainstem close to the Mamirauá Reserve. Estimating
446 floodplain total inundated area is relevant to differentiate the Amazon riverine fringing floodplains
447 from non-floodplain wetlands (here referred to as interfluvial wetlands).

448 Finally, in order to assess the current capabilities of basin-scale mapping of inundation dynamics
449 at high spatial and temporal resolution, a further assessment of the four high-resolution dynamic
450 datasets (GIEMS-D3, CaMa-Flood, SWAF-HR and MGB) at their native resolutions was
451 performed by computing their long-term flood frequency for the entire basin.

452

453 **3. Results and Discussion**

454 **3.1 How much inundation is estimated to occur in the Amazon basin?**

455 *3.1.1 Overall assessment*

456 Comparisons among the various estimates of inundation area can begin with the maximum and
457 minimum inundated area across the entire Amazon basin. We found wide variation in the annual
458 maximum and minimum inundation estimates for the entire basin scale (Figure 5), as well as the
459 long-term maxima and minima (Figure 6 and Table 2). The annual maximum inundation area
460 represents the total area subject to inundation at some point over the year, whereas the annual
461 minimum inundation area represents the area that remained inundated all year. SAR estimates,
462 especially those based on L-band sensors and those having undergone validation (i.e., the Hess et
463 al. (2003) dataset), are assumed to be the most accurate given their high spatial resolution and
464 capability of mapping flooded areas under dense vegetation canopies and cloud cover. Given the
465 lack of ground validation for most basin-scale datasets, we assess their accuracy by comparing
466 them to subregional validation datasets in section 3.2.

467 By computing means and standard deviations of the long-term maximum area subject to inundation
468 by type of data (Table 2), we obtain the following values: $138,200 \pm 45,300 \text{ km}^2$ (mean \pm S.D.) for
469 optical, $533,500 \pm 217,800 \text{ km}^2$ for multiple datasets at high resolution, $579,100 \pm 108,900 \text{ km}^2$
470 for those at coarse resolution, $542,800 \pm 80,600 \text{ km}^2$ for hydrological models, and $599,700 \pm$
471 $81,800 \text{ km}^2$ for SAR. The mean area for optical-based datasets is thus around 23% of the SAR-
472 based estimate. If we assume that the ensemble of datasets could be a proxy of inundation
473 uncertainty in the Amazon basin, and neglecting the optical and land cover-based data (G3WBM,
474 GLAD, GSWO and ESA-CCI) and CIFOR datasets, given their lower capability to map inundation
475 as discussed below, 13 datasets are left, yielding an estimation for the long-term maximum
476 inundation of $559,300 \pm 81,100 \text{ km}^2$. This value is around $40,000 \text{ km}^2$ lower than the mean of the
477 maximum inundation area from the three SAR datasets. The mean of the maximum inundation

478 area considering all 18 datasets is $490,300 \pm 204,800$ km². Compared to the maximum inundation
479 area, the relative deviation among available estimates is higher for the long-term minimum area
480 inundated — $125,900 \pm 77,600$ km² (mean \pm S.D.), with a coefficient of variation of 0.62, for the
481 12 basin-scale datasets that provide minimum area, and $139,300 \pm 127,800$ km² for the three SAR-
482 based datasets, with a coefficient of variation of 0.92.

483 None of the datasets can map small, narrow floodplains or riparian zones, for which only simple
484 calculations are currently available (e.g., Junk et al., 1993), and whose total area can only be
485 estimated through statistical extrapolation of observable rivers. These small zones contribute to
486 the overall uncertainties of the inundation estimates. For instance, a wetland mask developed by
487 Hess et al. (2015) for SAR-based wetland classification yielded a basin-scale estimation of wetland
488 area including the smallest floodplains of 840,000 km². This estimate is much larger than the
489 largest long-term maximum inundated area obtained with SAR data (659,100 km² with
490 Rosenqvist’s dataset). In section 3.2, it will be shown that almost all datasets tend to underestimate
491 the maximum inundation, when compared to subregional ones. The two SAR-based datasets with
492 highest accuracy underestimate maximum inundation by 9% (Rosenqvist) and 13% (Hess), based
493 on the average difference between these and the subregional estimates for the seven locations with
494 available data. If this holds true for the whole basin, the basin-scale maximum inundation would
495 be around 10% higher.

496

497 *3.1.2 Estimates based on SAR datasets*

498 At the basin scale, SAR-based estimates of maximum annual inundation range from 424,600 km²
499 (Rosenqvist) to 633,500 km² (Hess), and minimum inundation from 53,900 km² (Rosenqvist) to

500 284,200 km² (Hess), as shown in Figure 5. By considering long-term maximum inundation (i.e.,
501 all pixels that were inundated at least once in the entire available time series), instead of annual
502 maxima, the SAR-based estimates range from 506,400 km² (Chapman) to 659,100 km²
503 (Rosenqvist) for the entire basin (Table 2). The minima vary from 42,400 km² (Rosenqvist) to
504 284,200 km² (Hess). This highlights the large differences that exist, especially for the minima,
505 usually referred to as the “low-water period.” Chapman’s dataset, based on the 2006-2011 ALOS-
506 PALSAR archive, has a smaller total maximum inundation area than the other two SAR datasets,
507 as well as a smaller estimate for minimum inundation in relation to Hess’ estimate, which in turn
508 was developed from SAR mosaics at two seasons spanning only one year (1995–1996).
509 Differences among the three datasets may originate from differences in acquisition dates,
510 interannual and seasonal inundation variability, algorithms, spatial resolutions, or inconsistencies
511 regarding the data processing. For example, Chapman estimates long-term maxima and minima
512 based on multiple years, while Hess and Rosenqvist provide annual values. The calibration
513 uncertainty was also higher for the JERS-1 data used in Hess’ mapping than in the subsequent
514 satellites (ALOS-PALSAR and ALOS-2 PALSAR-2) (Hess et al., 2003). For long-term minimum
515 inundation, the interannual variability seems to be a minor factor since the Hess dataset, which
516 estimated a larger figure than the other ones, was developed for a year with minimum water levels
517 higher than those during Chapman’s acquisition dates, but lower than those during Rosenqvist’s
518 ones (see Fig. 8 in Rosenqvist et al., 2020). Thus, the larger minimum inundation extent by Hess
519 et al. (2015) seems to be more related to algorithm differences (Figure S2). For the maximum
520 water levels, Hess’ period was associated with an average year, below the water levels in Chapman
521 and Rosenqvist, and this may explain the relatively higher long-term maximum inundation by
522 Rosenqvist, while Chapman’s smaller values are likely due to algorithm differences. For the

523 western basin, Hess' estimate is based on JERS-1 data mostly from June 1996 (Hess et al., 2015),
524 which likely missed some of the inundation in this region as in the Pacaya-Samiria region, and
525 may partly explain the larger value by Rosenqvist (see section 3.2.2). Spatial resolution is also an
526 important factor: Rosenqvist's resolution is 50 m, and it is capable of representing smaller
527 floodplains than the other two (Figure S3), as will be discussed in section 3.2.2.

528

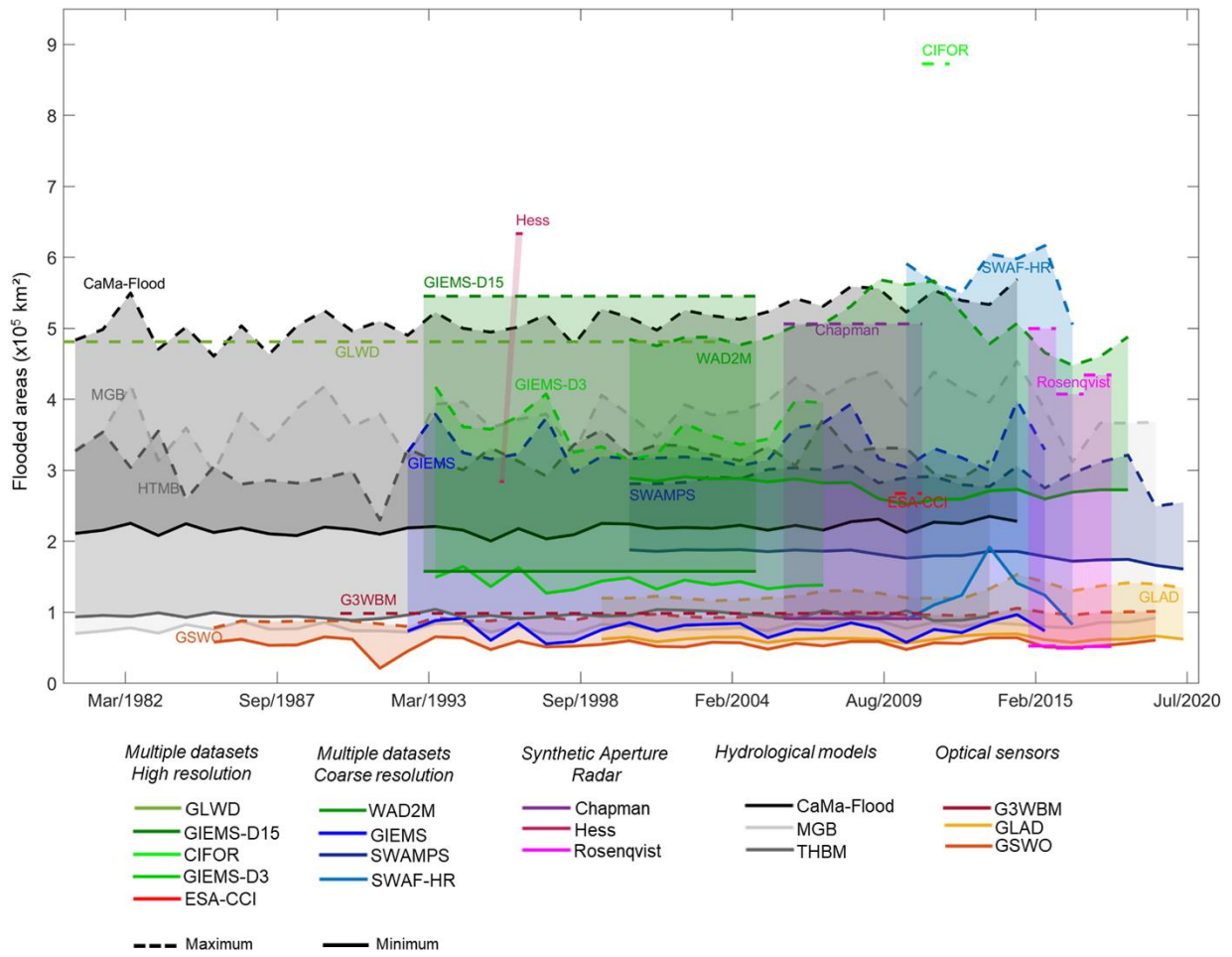
529 *3.1.3 Assessment of other datasets*

530 The coarse-resolution datasets and hydrologic models generally estimate smaller annual maximum
531 inundation areas in comparison to the SAR datasets, with the exception of SWAF-HR, WAD2M
532 and CaMa-Flood that yield similar annual maximum inundation. This results from the low
533 sensitivity of the passive microwave signal, which underlies most coarse-resolution datasets, to
534 detect small fractional flooded areas within the grid cells, flooding under particularly dense
535 vegetation, and flooding of short duration (i.e., less than one month of consecutive inundation)
536 (Hamilton et al., 2002). The higher sensitivity of the SWAF-HR may be associated with the use of
537 L-band passive microwave emission. Given the long-term data availability from dynamic, coarse-
538 resolution datasets, their long-term mean estimates are closer to the SAR ones, varying from
539 450,800 km² (THMB) to 630,900 km² (SWAF-HR), when compared to the annual scale analysis.
540 Therefore, no clear relationship between long-term minimum or maximum inundation and the
541 spatial resolution of the datasets is observed (Figure 6), which could be expected when analyzing
542 the annual values (Figure 5).

543 As expected, the optical-based datasets (GSWO, G3WBM, GLAD) cannot map inundation under
544 dense vegetation canopies and thus lead to much lower estimates of basin-wide inundation area

545 (Aires et al., 2018; Parrens et al. 2017). Similarly, ESA-CCI, which is based on land cover
546 classification of optical imagery with the addition of SAR inputs for delineation of wetland areas,
547 yields low basin-wide inundation areas, although relatively higher than the purely optical-based
548 estimates. In contrast, the multi-satellite-based CIFOR provides an unrealistically large estimate
549 of maximum inundation area (872,700 km²), which may be due to overestimation of soil moisture
550 by the topographic index used. This method is sensitive to rainfall overestimation, which may have
551 occurred in 2011, the year for which CIFOR was developed (Gumbricht et al., 2017). While the
552 dataset does represent well the spatial extent of peatlands across the Pacaya-Samiria region
553 (Gumbricht et al., 2017), its estimation of widespread inundation across the basin has limitations
554 to represent the large Amazon river floodplains, especially the forested ones, which are classified
555 as “swamps (including bogs)” by this dataset together with extensive interfluvial areas (Figure S4).

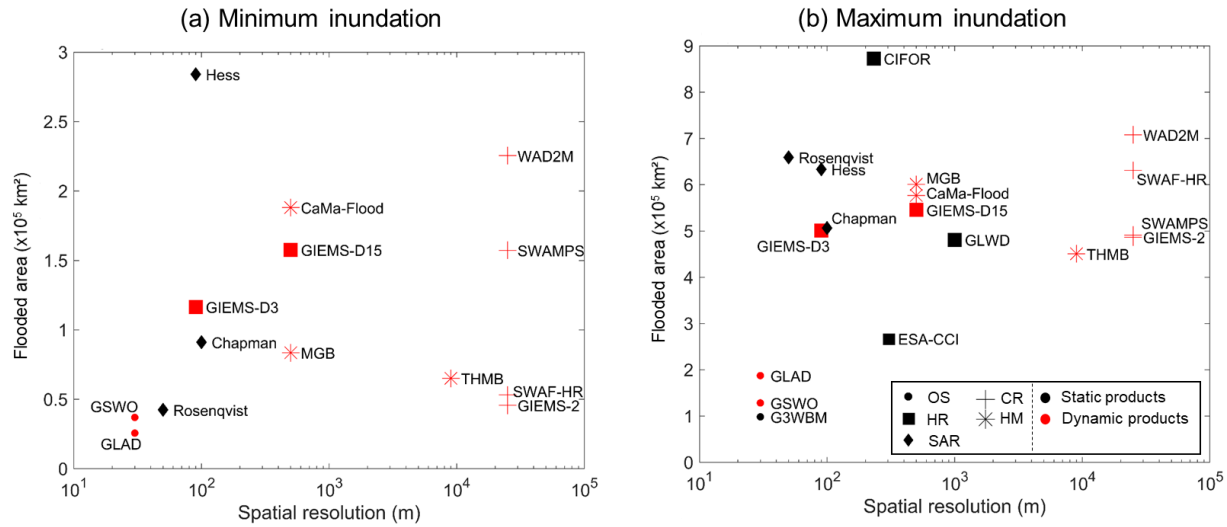
556



558 Figure 5. (a) Annual maximum and minimum flooded areas for the Amazon basin (< 500 m in elevation) for 18 basin-
 559 scale datasets over their respective observation time periods. Note that some datasets provide only average estimates
 560 based on multiple years of observation (e.g., GLWD, Chapman, G3WBM), and are marked as horizontal lines for the
 561 period of observation.

562

563



564

565 Figure 6. Summary of long-term (a) minimum and (b) maximum inundation for the 18 basin-scale datasets, which are
 566 categorized into five types (optical data; combination of datasets at high resolution; combination of datasets at low
 567 resolution; synthetic aperture radar; and hydrological models). Estimates by dynamic datasets are not directly
 568 comparable to the static ones; thus, each is colored differently: red (dynamic) and black (static). Legend for dataset
 569 types: OS: Optical Sensor; SAR: Synthetic Aperture Radar; HM: Hydrological Model; HR: multiple datasets at High
 570 Resolution; CR: multiple datasets at Coarse Resolution.

571 Table 2. Basin-scale, long-term minimum and maximum inundation estimates for 18 datasets.

	Dataset	Minimum (km ²)	Maximum (km ²)
Multiple datasets at coarse resolution	GIEMS-2	45,800	486,600
	SWAMPS	157,400	491,100
	WAD2M	225,500	707,900
Multiple datasets at high resolution	GIEMS-D3	116,600	500,700
	CIFOR	-	872,700
	ESA-CCI	-	267,400
	GIEMS-D15	157,700	545,400
	GLWD	-	481,200

	SWAF-HR	53,200	630,900
Hydrological model	THMB	65,200	450,800
	CaMa-Flood	188,100	576,700
	MGB	83,600	600,900
Optical sensor	G3WBM	-	98,500
	GLAD	25,700	187,600
	GSWO	37,000	128,500
Synthetic Aperture Radar	Hess	284,200	633,500
	Chapman	91,200	506,400
	Rosenqvist	42,400	659,100

572

573 **3.2 How much inundation is estimated to occur in individual wetland regions?**

574 **3.2.1 Overall assessment**

575 The 18 basin-scale inundation datasets were compared with the 11 subregional ones through
576 analysis of long-term means of annual maximum inundated areas (Table 3), long-term means of
577 annual minimum areas (Supplementary Table S3), and multiple comparison metrics
578 (Supplementary Table S4). The subregional datasets, covering individual wetland complexes, are
579 considered as independent validation datasets, given the ground validation performed for most of
580 them, as well as the use of a region-specific classification, and the often higher spatial resolution
581 (e.g., 12.5 m for some based on ALOS-PALSAR imagery).

582 The Amazon River floodplains (from Iquitos to Gurupá) and the Llanos de Moxos regions are the
583 largest Amazon wetland complexes: $106,800 \pm 25,800$ km² and $113,500 \pm 53,400$ km², respectively
584 when considering the three SAR-based datasets, and $94,100 \pm 32,500$ km² and $85,300 \pm 52,400$

585 km² when considering all 18 basin-scale datasets. Besides these two areas, the third largest
586 Amazon wetland region is Pacaya-Samiria, with $29,700 \pm 20,600$ km² (all datasets) and $40,000 \pm$
587 $4,200$ km² (SAR datasets).

588 The comparison of the long-term means of annual maximum and minimum observed inundation
589 over the available time periods indicates differences between basin-scale datasets and the
590 subregional validation datasets. Overall, the subregional datasets had a larger maximum inundation
591 extent than that estimated for the subregion from the basin-scale datasets. The underestimation by
592 the basin-scale ones varied from 49% for the Pacaya-Samiria region to 5% for the lower Amazon
593 River floodplain. Only three datasets overestimated the maximum extent of inundation: GIEMS-
594 D3, GIEMS-D15 and GLWD. The basin-scale, SAR-based ones (Hess, Chapman and Rosenqvist)
595 underestimated the maximum extent in the regions represented by all subregional datasets, except
596 Rosenqvist for Janauacá Lake, and Hess for the Llanos de Moxos region. This is likely related to
597 the higher resolution of many of the subregional datasets (e.g., 12.5 m original and 25 m final
598 resolution for the Uatumã ALOS-PALSAR classification by Resende et al., 2019), differences in
599 image acquisition period, and fine-tuning that may occur with dedicated processing for a particular
600 region.

601 To investigate the depiction seasonal patterns of inundation by the various datasets, we assessed
602 the correlation between the time series of absolute inundated areas from the dynamic ones and the
603 estimates for individual wetland complexes (Table S3). Overall, all datasets agreed well (average
604 Pearson correlation larger than 0.63 for the four wetland complexes with available time series),
605 showing a similar depiction of the inundation seasonality. However, their ability to monitor high-
606 resolution flood frequency is limited, as will be further discussed in section 4. A visual comparison
607 of the time series (Figure S6) shows agreement on seasonal timing of flooding and drainage, but

608 disagreement in the extent of inundation. In particular, two datasets have a small overall annual
609 amplitude (SWAMPS and WAD2M).

610 Overall, four datasets had the best overall representation of spatial patterns in inundation (Fit
611 metric; see Equation 1), as analyzed at 1 km pixel resolution, in comparison to the subregional
612 validation datasets: Hess, GLWD and the two hydrodynamic models (MGB and CaMa-Flood),
613 which were associated with average Fit metric between 0.64 and 0.67 (Table S3). While hydrologic
614 models such as MGB, CaMa-Flood and THMB have a satisfactory agreement basin wide, they are
615 unable to represent wetlands not primarily inundated by rivers (Fleischmann et al., 2020; Zhou et
616 al., 2021). For example, the Llanos de Moxos inundation is underestimated by both CaMa-Flood
617 and MGB with low Fit metric values (0.19-0.28; Table S3). This is expected for interfluvial
618 wetlands such as Llanos de Moxos and Roraima, where much of the flooding is caused by poor
619 drainage of local rainfall and tends to be shallower, as opposed to overflow of large rivers onto
620 adjacent floodplains. The four alternative subregional datasets assessed here - three hydrological
621 models (one for Curuai and two for Janauacá) and one classification of ALOS-PALSAR data for
622 the Llanos de Moxos area - were generally better or similar to some of the best-performing basin-
623 scale ones, as could be expected given their fine tuning for the specific areas, which often includes
624 local topographic surveys.

625 Some of the datasets merging multiple data sources overestimated the inundation area of individual
626 wetland complexes the most, especially GIEMS-D15, GIEMS-D3 and GLWD. Furthermore,
627 CIFOR was originally designed for peatland mapping in the tropics, and generally overestimates
628 inundation, suggesting a widespread distribution of wetlands along interfluvial terraces across the
629 whole basin that may include areas of poorly drained soils lacking surface water. For the individual
630 wetland complexes, however, CIFOR generally underestimated inundation and had a poor

631 representation of spatial patterns of inundation (low Fit metric). WAD2M underestimated the
632 maximum inundation the most, which is understandable given its removal of open water areas and
633 because its main inputs (CIFOR and SWAMPS) also underestimated inundated areas as indicated
634 by the subregional validation datasets.

635

636 *3.2.2 Individual inundation patterns based on SAR data*

637 Regarding the maximum inundation extent, the Janauacá case provides a representative
638 example to understand the differences among multiple L-band SAR datasets: these estimated total
639 inundated area as 209 km², 184 km² and 446 km² for Hess, Chapman and Rosenqvist, respectively,
640 in contrast to 404 km² with the subregional ALOS-PALSAR-based dataset (12.5 m resolution;
641 Pinel et al., 2019). Part of these differences occur because of interannual variability, but other
642 factors such as spatial resolution and algorithm differences seem relevant. Rosenqvist led to a more
643 consistent estimation of the spatial inundation extent in terms of maximum inundation (Table 3)
644 and inundation spatial patterns (Fit metric; Table S3), which can be a consequence of its higher
645 spatial resolution (50 m) in contrast to the other two (90 m; Figure S3). Overall, Rosenqvist
646 provided the largest inundation extent among SAR datasets across all areas along the Amazon
647 mainstem floodplain, except for the Curuai floodplain and the savanna wetlands, as well as the
648 closest agreement with subregional validation datasets ($-9\% \pm 13\%$; average \pm S.D.). Hess
649 estimated the largest inundation area in the savanna wetlands (Llanos de Moxos, Roraima and
650 Negro). However, Hess' estimate is 39% larger than the subregional validation dataset for Llanos
651 de Moxos, while the other two SAR estimates are lower (-26% and -41% for Chapman and
652 Rosenqvist, respectively).

653 One important question remains about the low-water period, as discussed in the previous section
654 for the basin-scale analysis. Hess suggests much more inundation for this period for the Amazon
655 mainstem floodplains (54,500 km²), mainly for the upstream forested reaches, and for the whole
656 basin in general (284,200 km²), than recent estimates with ALOS (28,500 and 91,200 km²) and
657 ALOS-2 data (19,500 and 42,400 km²). An assessment with the subregional datasets along the
658 Amazon floodplain suggests that Hess overestimates the minimum extent for Curuai, Mamirauá
659 and lower Amazon River, and is accurate for the Janauacá floodplain lake. Rosenqvist generally
660 underestimates the minimum inundation. For instance, for the Mamirauá dataset, the minimum
661 extent (i.e., permanently flooded areas) sums up to 715 km², which is increased to 1545 km² if
662 considering all pixels flooded for more than 295 days per year. For this area, the SAR estimates
663 are 1756 km² (Hess), 866 km² (Chapman) and 422 km² (Rosenqvist). Overall, this suggests that
664 the actual value of minimum inundation across the central Amazon floodplains is somewhere
665 between the Hess and Rosenqvist estimates.

666

667 *3.2.3 Challenges over floodable savannas*

668 Large discrepancies are observed for the Roraima and Negro floodable savannas. Roraima
669 wetlands are small river floodplains interspersed with open savannas subject to flooding, which
670 can be identified by optical data. In addition, the typical timing of high and low water in the
671 Roraima region coincides approximately with the JERS-1 dual-season mosaics that were designed
672 to reflect the seasonality of the central Amazon River floodplain (Hamilton et al. 2002). For these
673 reasons, the JERS-1-based dataset by Hess et al. (2015) seems to satisfactorily represent most of
674 the Roraima wetlands. However, it misses some small-scale riparian forests, given its 90 m spatial
675 resolution and snapshot coverage that likely missed flooding events on smaller, flashier rivers

676 (Figure S5). Thus, the maximum inundation is likely higher than the Hess estimate (8,900 km²),
677 which in turn is larger than the other ones based on SAR (1,900 - 4,100 km²). The only dataset to
678 estimate a higher value is the coarse SWAF-HR (18,100 km²), which is similar to the value
679 previously estimated by Hamilton et al. (2002) (16,500 km²), also with coarse data (SMMR passive
680 microwave), though a part of the discrepancy may be due to interannual variability. More studies
681 are necessary for this area to understand its actual inundation extent and dynamics. Similarly, the
682 inundation estimates in the Negro interfluvial savannas are subject to large uncertainty, with the
683 long-term maximum inundation varying between 95 (GLWD) and 20,700 km² (CIFOR),
684 considering all basin-scale datasets. SAR-based estimates were between 5,900 and 15,800 km². In
685 contrast, for the Pacaya-Samiria interfluvial area, which includes a large complex of forested
686 wetlands, peatlands and palm swamps, the discrepancies are smaller than for the savanna
687 interfluvial regions, although still considerable. The basin-scale SAR ranged between 24,000 km²
688 (Chapman) and 56,200 km² (Rosenqvist), with the subregional validation dataset yielding 57,900
689 km². The good agreement between Rosenqvist and the subregional dataset was already reported
690 by Rosenqvist et al. (2020).

691

692 Table 3. Long-term maximum inundation areas (km²) for the 11 wetland complexes (up to three subregional datasets
693 per complex) and the 18 basin-scale datasets. The subregional values refer to the following datasets, in this order
694 (comma-separated values relate to areas with more than one dataset available): Curuai - ALOS (Arnesen et al., 2013)
695 and LISFLOOD-FP model (Rudorff et al., 2014); Uatumã - ALOS (Resende et al., 2019); Janauacá - ALOS (Pinel et
696 al., 2019), hydrologic model (Bonnet et al., 2017) and TELEMAC-2D model (Pinel et al., 2019); Mamirauá - ALOS
697 (Ferreira-Ferreira et al., 2015); Pacaya-Samiria - ALOS-2 PALSAR-2 (Jensen et al., 2020); Llanos de Moxos -
698 MODIS (Ovando et al., 2016) and ALOS (Ovando et al., 2016); and Lower Amazon River - MODIS (Park et al.,

699 2019). Average, standard deviation (S.D.) and coefficient of variation (CV) are presented for each area in the last
700 rows.

	Dataset	Curuai	Uatumã	Janauacá	Mamirauá	Pacaya-Samiria	Llanos de Moxos	Lower Amazon	Amazon mainstem	Purus	Roraima savannas	Negro savannas
	Subregional	4162, 3720	1471	404, 336, 176	4476	57913	125422, 133470	56722	-	-	-	-
Multiple datasets at coarse resolution	GIEMS-2	3080	984	623	3344	23344	156176	79871	116379	7208	7173	12237
	SWAMPS	3359	722	280	1131	9929	88753	58626	72468	5618	4970	8819
	WAD2M	681	243	166	888	42635	102780	29276	49261	6698	3173	15450
Multiple datasets at high resolution	GIEMS-D3	4643	2732	505	3569	11562	150285	92908	127552	9045	12355	15123
	CIFOR	3796	994	177	1714	52590	116201	43509	86301	10844	3728	20712
	ESA-CCI	3236	855	260	3045	28727	39795	37475	84803	8883	510	12623
	GIEMS-D15	4635	2681	416	2444	44536	117979	86123	127150	11186	8129	14854
	GLWD	4275	2267	535	4259	79124	40661	67746	140921	14840	1048	95
	SWAF-HR	4439	2199	388	3205	16900	159712	69539	110468	10785	18146	15375
Hydrological model	THMB	2883	554	164	2840	27748	52693	39193	89658	19733	4307	3640
	CaMa-Flood	4246	1613	534	3208	34096	80725	63963	118577	20947	3454	6560
	MGB	4098	1549	474	3750	33344	21757	61997	115047	20394	240	3224
Optical sensors	G3WBM	2732	628	135	795	2694	9564	27451	37718	2351	352	1238
	GLAD	3479	832	204	1141	4196	38897	36930	53121	3903	3495	3885
	GSWO	3163	675	150	962	3637	19240	31191	44731	2982	1442	1880
Synthetic Aperture Radar	Chapman	2796	934	184	2694	24001	73710	39677	77632	12499	4077	5935
	Hess	3996	1045	209	3985	39741	174198	52156	115822	15155	8950	15758
	Rosenqvist	3055	1238	446	4362	56160	92693	55262	126806	20738	1867	9935
	Average	3477	1264	325	2630	29720	85323	54050	94134	11323	4856	9297
	S.D.	949	748	163	1226	20591	52387	19956	32503	6185	4666	6201
	CV	27%	59%	50%	47%	69%	61%	37%	35%	55%	96%	67%

701

702 3.3 How much do the datasets agree on the spatial distribution of inundation?

703 Agreement maps of the high resolution datasets (≤ 1 km spatial resolution) were developed for
704 both long-term maximum (14 datasets available) and minimum inundation areas (10 datasets),
705 based on the number of inundation datasets coinciding over a 1 km pixel (Figures 7 and 8 and their
706 categorization for specific regions in Figure 9). Overall, 31% of the Amazon lowlands area (i.e.,

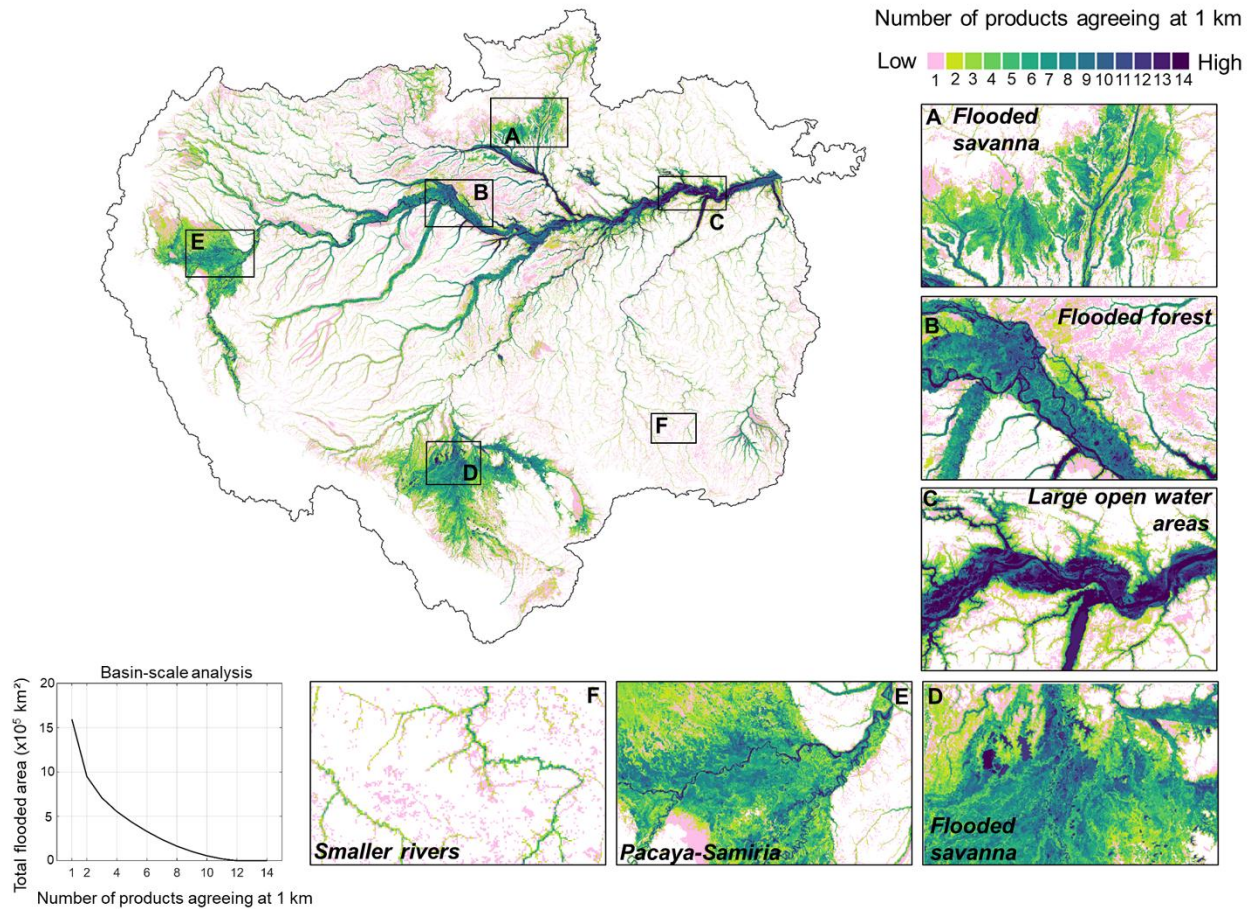
707 1.59 x 10⁶ km² out of 5.11 x 10⁶ km²) has been estimated as subject to inundation by at least one
708 dataset (bottom left panel, Figure 7). Based on the agreement between two datasets, this value
709 decreases to 948,300 km², which is larger than the value estimated when there is agreement among
710 four datasets (553,200 km²). This latter estimate is more similar to the average maximum
711 inundation as estimated by the ensemble of datasets (559,300 km²) and the three SAR-based ones
712 (599,700 km²). Furthermore, there is a lower agreement for the minimum inundation than for the
713 maximum inundation among individual regions (Figure 9).

714 For specific regions, a high degree of agreement for floodplains dominated by open water areas is
715 evident for the lower Amazon River reaches, followed by the forested floodplains fringing large
716 rivers, especially along the Amazon mainstem, Purus and Negro rivers. The generally higher
717 accuracies over central Amazon floodplains may also be related to the attention that dataset
718 developers have devoted to it, in contrast to other regions. Furthermore, the maximum floodplain
719 extent can be somewhat delineated with terrain elevation data (i.e., DEMs) using algorithms such
720 as HAND (Rennó et al., 2008), which helps to explain the relatively small disagreement for
721 floodplains fringing the largest rivers, and is particularly effective with vegetation bias-removed
722 DEMs (O'Loughlin et al., 2016; Yamazaki et al., 2017). The best agreement (for both maximum
723 and minimum inundation extent) occurred over the Curuai floodplain along the lower Amazon
724 mainstem, with 37% of its area being estimated as subject to inundation by all 14 datasets (Figure
725 9a). An agreement among all 14 datasets occurred, in part (i.e., more than 10% of the wetland
726 area), for the central Amazon floodplains (Curuai, Uatumã, Janauacá and lower Amazon River)
727 because of their relatively large fractions of open water areas.

728 In the interfluvial wetlands (Negro and Roraima savannas, Pacaya-Samiria and Llanos de Moxos),
729 the inundation patterns are less dependent on riverine overflow and more dependent on local

730 rainfall, making them less predictable (Hess et al., 2003). The disagreement for both maximum
731 and minimum inundation area is the largest across all regions, e.g., 65–78% of their flooded areas
732 were mapped by only one model for the minimum inundation (Figure 9b). The Llanos de Moxos
733 is conspicuous as a region of particular disagreement, perhaps because flooding is mainly shallow
734 and in vegetated areas (mainly savannas/grasslands), and is highly variable from year to year. In
735 general, the smaller the flooded patches the higher the challenge to map them, not only because of
736 resolution but also due to small-scale variation in topography. Similar disagreement occurred in
737 other interfluvial wetlands such as the Negro and Roraima savannas, and would be expected
738 elsewhere in savanna floodplains of South America (e.g., Pantanal, Llanos de Orinoco and Bananal
739 Island; Hamilton et al., 2002). The poor agreement over interfluvial areas, however, may also
740 partly reflect the longer history of study of Amazon mainstem floodplains, for which there are
741 river gage records that reflect floodplain water levels and inundation, while more remote areas
742 such as the Negro savannas and Pacaya-Samiria regions are more challenging to represent with a
743 few gages, and have received less attention. The challenges in estimating inundation over
744 interfluvial areas also affect the SAR-based datasets, which disagreed the most over these regions
745 (see section 3.5 and discussion in Rosenqvist et al., 2020).

746

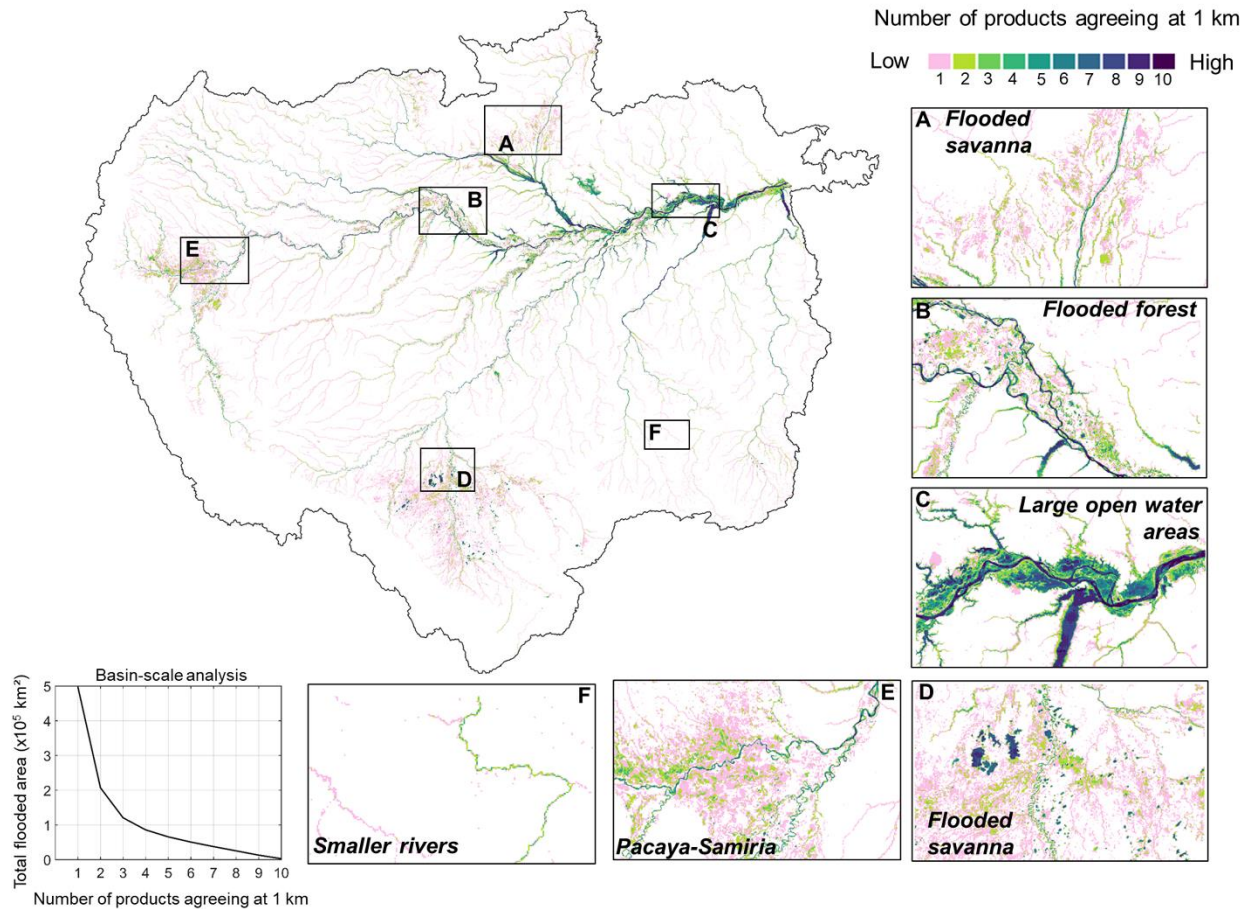


747

748 Figure 7. Agreement for maximum inundation area among 14 basin-scale datasets at high resolution (≤ 1 km spatial
 749 resolution): G3WBM, ESA-CCI, GLAD, GSWO, GLWD, CIFOR, GIEMS-D15, GIEMS-D3, Chapman, Hess,
 750 Rosenqvist, SWAF-HR, CaMa-Flood and MGB. A given pixel of a dataset with resolution higher than 1 km that had
 751 more than 50% of flooding at the maximum inundation extent is classified as inundated.

752

753



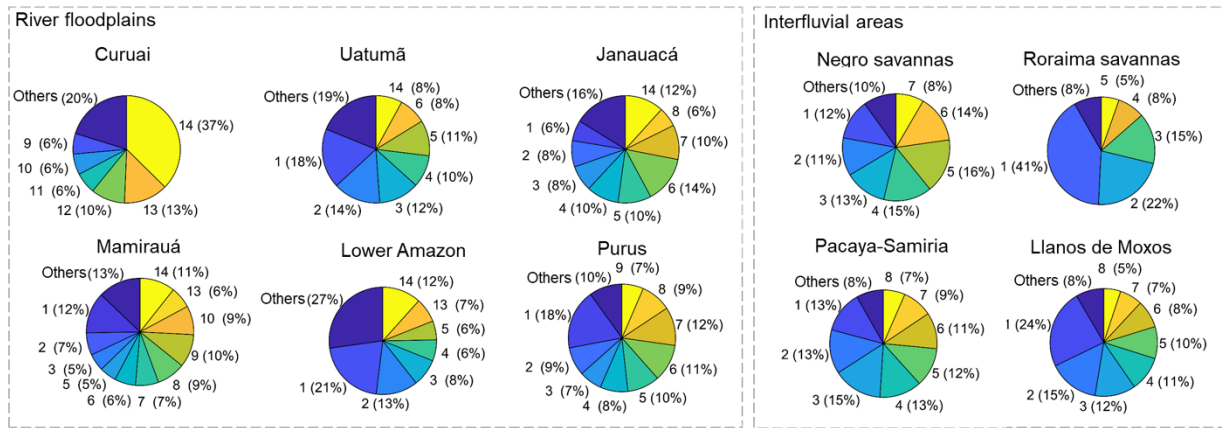
754

755 Figure 8. Agreement for minimum inundation area among 10 basin-scale datasets at high resolution (≤ 1 km spatial
 756 resolution): GIEMS-D15, Chapman, Hess, Rosenqvist, SWAF-HR, CaMa-Flood, MGB, GIEMS-D3, GSWO and
 757 GLAD. A given pixel of a dataset with resolution higher than 1 km that had more than 50% of flooding at the minimum
 758 inundation extent is classified as inundated.

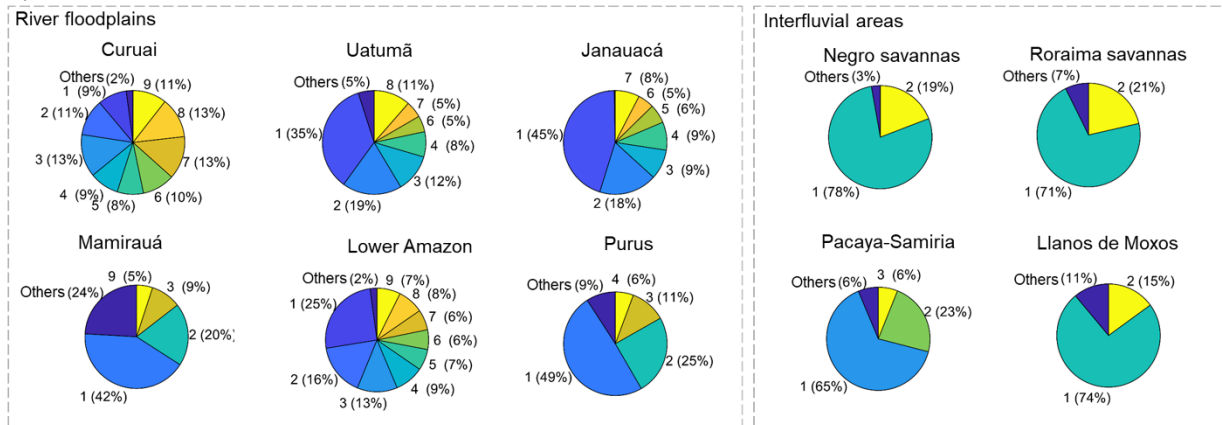
759

760

a) Maximum inundation



b) Minimum inundation



761

762 Figure 9. Degree of agreement for (a) maximum and (b) minimum inundation area for 10 individual wetland
 763 complexes, based on the 1 km agreement map (Figures 7 and 8). The percentage values indicate the fraction of each
 764 area where a given number of datasets agreed that it was flooded, e.g., 14 models agreed that 37% of the Curuai area
 765 was flooded in the maximum inundation extent. The class with number 1 indicates the fraction of the area that only
 766 one dataset estimated as being inundated. The class “others” refers to all classes that had less than 5% of pixels
 767 estimated as being inundated.

768

769 **3.4 Quantifying the inundation extent of different wetland types**

770 Amazon wetlands include a myriad of ecosystems varying in geomorphology, hydrology, and
771 vegetation cover. The classification system proposed by Junk et al. (2011) differentiated Amazon
772 wetlands according to amplitude and range of water level change. Wetland types ranged from the
773 forested swamps with stable water levels to river floodplains with oscillating water levels, and to
774 interfluvial areas with small seasonal water level amplitude due to the main contribution of local
775 rainfall and runoff (Fleischmann et al., 2020; Junk et al., 2011; Ovando et al., 2018).

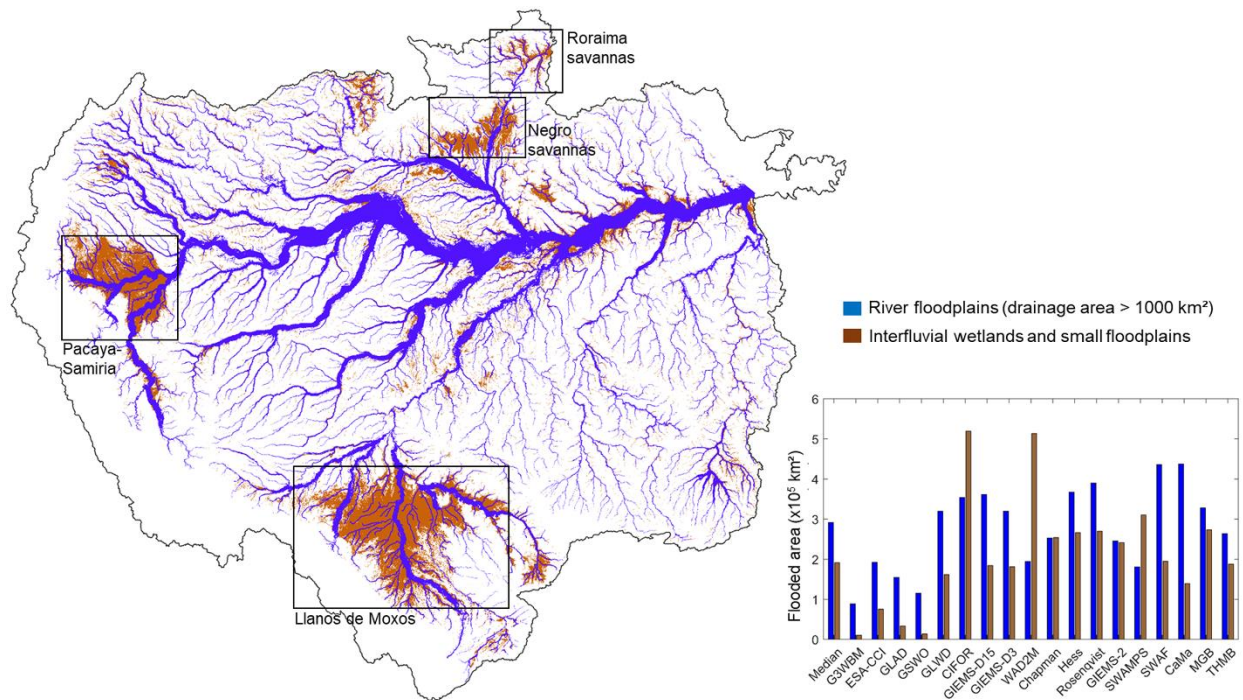
776 A simpler yet hydrologically meaningful classification is the categorization into river floodplains
777 and interfluvial wetlands adopted here, since the former typically have a greater hydrological
778 connection to the main river and thus are subject to a different control of inundation area by river
779 levels (Reis et al., 2019a). We performed a quantitative analysis of the inundation area in these
780 two main hydrological classes. All pixels considered flooded by at least two datasets, based on the
781 1 km agreement map for maximum inundation extent (Figure 7), are presented in Figure 10.
782 Overall, the medium to large river floodplains (upstream drainage area > 1000 km²) have a larger
783 inundation extent than the category with small floodplains and interfluvial areas. An average total
784 area subject to inundation of $317,800 \pm 84,400$ km² (average \pm S.D.; median equal to 323,700 km²)
785 was obtained for the medium to large floodplains, not including the optical and land cover datasets
786 (G3WBM, GLAD, GSWO and ESA-CCI). A greater area for large floodplains was estimated by
787 all except for CIFOR, SWAMPS and WAD2M. Two datasets estimated a similar value between
788 the two classes (Chapman and GIEMS-2), which may be related to an overestimation of basin-
789 scale isolated flooded patches.

790 Large floodplains fringing the main rivers, especially along the Amazon River, have been largely
791 addressed by previous studies (Table 1 and Table S1). However, large river floodplains are also
792 present in less studied reaches, e.g., in the upper Napo and Içá rivers in northwest Amazon basin,

793 and upper Xingu in the southeastern portion (see location in Figure 1). These upper reaches are
794 subject to more sporadic, flashy river hydrological regimes (Hamilton et al., 2007), which make
795 their inundation area difficult to map with current datasets of relatively low temporal resolution.
796 In our analysis, the non-floodplain areas include mainly the large interfluvial areas (black
797 rectangles in Figure 10), small river floodplains that are challenging to detect with currently
798 available datasets, and some reservoirs, such as Balbina reservoir on the Uatumã River.

799 Besides the central Amazon floodplains, which have been widely studied, other wetland
800 complexes require more attention, such as the Negro and Roraima savannas; the latter was only
801 assessed by a single study to our knowledge (Hamilton et al., 2002). The inundation mapping of
802 the Pacaya-Samiria region in the upper Amazon has received scientific attention recently (Jensen
803 et al., 2018; Rodriguez-Alvarez et al., 2019), partially because of the region's role as a carbon sink
804 via formation of peat (Draper et al., 2014; Lähteenoja et al., 2012). Regarding open water areas,
805 Melack (2016) reported values ranging from 64,800 km² (Melack and Hess, 2010) to 72,000 km²
806 (SRTM Water Body Data) and 92,000 km² (Hansen et al., 2013) for the Amazon basin (< 500 m
807 in elevation). The three Landsat-based datasets assessed here, which are mainly capable of
808 detecting open water areas, estimate 98,500 km² (G3WBM), 128,500 km² (GSWO) and 187,600
809 km² (GLAD).

810



811

812 Figure 10. Quantification of maximum inundated areas over river floodplains with drainage area larger than 1,000
 813 km², and interfluvial wetlands and small floodplains (area < 1,000 km²) within the Amazon basin. The maximum
 814 inundation map depicts all 1 km pixels with at least two datasets agreeing (i.e., a reclassification of Fig. 7), in order to
 815 avoid overestimation caused by pixels with only one dataset classifying them as subject to inundation. The four large
 816 areas of interfluvial wetlands are highlighted with black rectangles (Pacaya-Samiria, Llanos de Moxos, Negro and
 817 Roraima savannas).

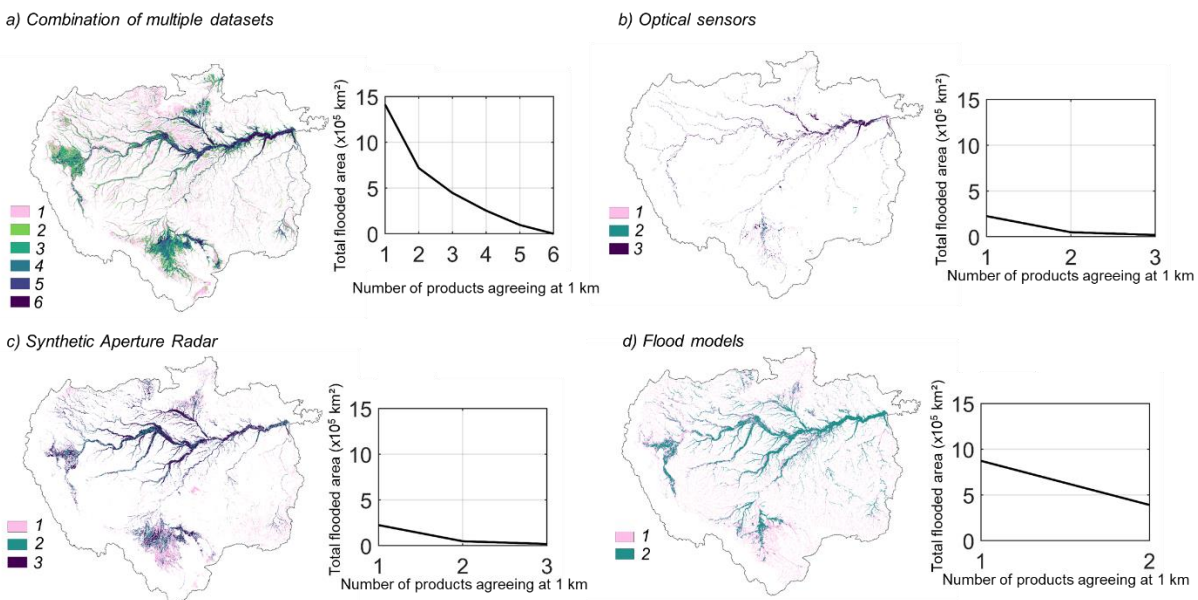
818

819 3.5 Limitations in comparing the inundation area datasets

820 Some of the differences in large-scale inundation mapping highlighted by our comparison occur
 821 because distinct datasets map temporal variation in inundation in different ways, varying for
 822 example in sensor type, post processing, and spatial resolution. Figure 11 shows the agreement
 823 maps for maximum inundation for four classes of datasets, considering the 14 basin-scale high-

824 resolution datasets. Those based on multiple datasets (GLWD, CIFOR, GIEMS-D3, GIEMS-D15,
 825 SWAF-HR) have the best agreement for the Llanos de Moxos area, and to a smaller degree, for
 826 Pacaya-Samiria, Negro and Roraima wetlands. The L-band SAR datasets have less overall
 827 agreement (Figure 11c), while the optical data are mainly applicable to open water areas in the
 828 Amazon mainstem floodplain (Figure 11b). The 1D hydrological models cannot represent
 829 interfluvial wetlands where flooding is not controlled by river level and discharge (Figure 11d).

830



831

832 Figure 11. Amazon basin (< 500 m elevation) agreement maps at 1 km resolution, for maximum inundation and for
 833 each type of dataset, considering only the high-resolution datasets ($\leq 1 \text{ km}$ spatial resolution): (a) six datasets based
 834 on merging of multiple datasets (GLWD, CIFOR, GIEMS-D3, GIEMS-D15, SWAF-HR, ESA-CCI), (b) three datasets
 835 based on optical sensors (G3WBM, GLAD, GSWO), (c) three datasets based on synthetic aperture radar (Hess,
 836 Chapman, Rosenqvist), and (d) two hydrological models (MGB and CaMa-Flood). The right column graphs present
 837 the total inundation area in the Amazon basin for a given number of datasets agreeing, e.g., the basin area where the
 838 two hydrological models (Fig. d) agree to be flooded is 390,900 km^2 .

839 The different methodologies used to produce each dataset complicate their direct comparison
840 (Rosenqvist et al., 2020), and some methodological differences produce systematic differences and
841 bias among the data sources included in our comparison. Here we used datasets covering long-
842 term dynamics (e.g., GIEMS or hydrologic models), short-term dual-season (e.g., Rosenqvist,
843 spanning four years), and a particular year (e.g., Hess). Some datasets use alternative approaches
844 to derive long-term maximum inundation area, such as GIEMS-D15, which generated estimates
845 by merging 3-year moving-window maximum values of GIEMS with the GLWD dataset.
846 Therefore, a comparison of all these datasets must be performed with consideration of their
847 methodology. For instance, the comparison of dual-season datasets against monthly datasets can
848 yield erroneous conclusions, although it has been a common practice to directly compare such
849 datasets. Some datasets also consider a “high-water assumption” (Ferreira-Ferreira et al., 2015;
850 Hess et al., 2003), whereby the high-water maps are forced to contain all flooded pixels from the
851 low-water map.

852 In addition to methodological differences, each dataset was developed for different periods (Table
853 1), and thus interannual and seasonal variability accounts for some of the differences among them.
854 To address this, we performed an annual analysis (Figure 5), which suggests that the long-term
855 inundation estimate is fairly stable for each dataset despite some interannual differences. In fact,
856 the temporal variability of each dataset is generally smaller than the differences in comparison
857 with the other estimates. However, the Amazon hydrological cycle has been shifting over decades
858 (Barichivich et al., 2018; Gloor et al., 2013), and a recent increase in maximum water levels in the
859 central Amazon suggests a new hydroclimatic state (Espinoza et al., 2019). Some wetlands have
860 also been subject to forest loss, and so the detectability of inundation by remote sensing may have
861 increased over time, e.g., major deforestation has occurred along the lower Amazon River

862 floodplain (Renó et al., 2011). Similarly, widespread burning might be converting black-water
863 floodplain forests into savanna vegetation (Flores and Holmgren, 2021). In addition, in some
864 regions, such as the southern Amazon, an increase in the dry-season length has been observed,
865 which is a major climatic constraint for forest sustainability (Fu et al. 2013; Staver et al., 2011).
866 However, analyzing long-term change in inundation patterns is beyond the scope of this study, and
867 thus we assumed stationarity in our comparison framework.

868 Another important challenge is to find a common definition of wetlands among datasets. Here we
869 focused on inundation extent, however some datasets (e.g., CIFOR) represent peatland locations
870 instead of inundated areas, although their areas of peat formation often include inundated areas.
871 Estimates based on SAR or passive microwave emission may also be sensitive to saturated soil
872 without standing water above it, and thus the observed inundation estimates can have some
873 ambiguity. Hydrologic models provide simulated surface water extent, and we mapped inundation
874 accounting for pixels with water depth greater than zero. While hydrologic models have
875 uncertainties related to model structure (e.g., inadequate representation of inundation processes),
876 input data (e.g., DEM and climate forcing) and parameterization (e.g., soil water capacity and river
877 channel width and depth; assumptions of level water surfaces between rivers and their floodplains),
878 remote sensing-based datasets have uncertainties related to spatial and temporal resolutions (e.g.,
879 coarse spatial resolution not capable of detecting small patches), and detection uncertainty (e.g.,
880 dense vegetation canopies can obscure passive microwave emission from underlying surfaces).
881 Thus, a comparative framework provides an opportunity to highlight and stress the uncertainties
882 and limitations of each dataset.

883 Hydrologic models currently available at the Amazon basin scale are one-dimensional, and thus
884 are capable of simulating flooding mainly along river floodplains, as corroborated by various

885 validation exercises in the Amazon that have relied on the Hess, GIEMS and SWAF-HR datasets
886 (Fleischmann et al., 2020; Luo et al., 2017; Paiva et al., 2013; Zhou et al., 2021). These models
887 are also largely dependent upon accurate DEMs, which are still challenging to obtain over tropical
888 forested floodplains. Furthermore, given that a 500 m elevation mask (Amazon lowlands) has been
889 used for some SAR datasets (Hess et al., 2015), and the difficulty of some radar and passive
890 microwave ones to detect inundation at high elevations due to slope and snow effects, for instance
891 (Parrens et al., 2017), we have adopted the same 500 m threshold in our lowland mask to improve
892 the comparability among datasets. However, even though higher elevation wetlands amount to
893 much less total area compared to lowland wetlands, understanding their flooding dynamics is
894 important for some parts of the Amazon basin. Although some datasets, especially the hydrological
895 models (MGB, CaMa-Flood and THMB), are capable of estimating inundation in higher elevation
896 parts of the basin, in this case uncertainties may also be large given errors in precipitation (low
897 density of in situ gauges and high rainfall spatial heterogeneity) and thus runoff fields over
898 mountainous areas, as well as the tendency for river flows to vary over short time scales (Espinoza
899 Villar et al., 2009; Zubieta et al., 2015). Furthermore, the availability of in situ river discharge
900 measurements for model calibration and validation is lower in the Andean Amazon (Feng et al.,
901 2020; Wongchuig et al., 2019; Zubieta et al., 2017).

902 Our analyses were performed at 1 km resolution and at regional scales, which avoids geolocation
903 problems that affect analyses at higher resolutions (e.g., 30 or 90 m). Small disagreements among
904 our estimates and the values presented in the original publications may also arise from the use of
905 the WGS84 datum with a geographical coordinate system for all datasets (except for SWAMPS
906 which was provided in the EASE-Grid format). Also, the coarse-resolution datasets, especially
907 GIEMS-2 and SWAMPS with 25 km spatial resolution, can be difficult to compare with estimates

908 for individual wetland complexes (e.g., Curuai and Janauacá), since only a few 25-km pixels may
909 be located within the wetland boundaries.

910 The quantification of inundation over larger river floodplains (Figure 10) is also subject to
911 uncertainties. The maximum floodplain lateral extent was estimated based on an automatic buffer
912 procedure around the Hydrosheds drainage network, further manually edited by considering the
913 three SAR-based, basin-scale datasets and the MERIT DEM-based topography. Although it
914 captures the basin-scale geomorphological differences along major floodplains, some uncertainties
915 remain regarding the true lateral extent for areas where rain-fed savanna floodplains are present
916 (e.g., Llanos de Moxos, Roraima), and where flooding extend far from the main rivers (e.g.,
917 Pacaya-Samiria). For these areas in particular, we assumed buffer values similar to adjacent
918 upstream and downstream floodplains (e.g., the Amazon River downstream of Pacaya-Samiria),
919 which is reasonable but should undergo future scrutiny, including local ground-based surveys.

920

921 **4. Perspectives and recommendations**

922 Considerable advances have been achieved in recent decades in the mapping of inundation extent
923 across the Amazon basin. Here, we have presented an analysis of 29 inundation datasets for the
924 basin, covering multiple scales, spatial and temporal resolutions, and data sources. We showed
925 that large discrepancies persist, and this is especially true at local scales. Below we present some
926 perspectives and recommendations for future development of inundation mapping in the world's
927 largest river basin.

928

929 *4.1 Which are the most reliable data sources for inundation mapping in the Amazon River*
930 *basin?*

931 At basin scale, the Rosenqvist ALOS-2 PALSAR-2 dataset is available at 50 m, and shows a good
932 overall agreement with the 90 m Hess one over the large river floodplains, while the latter seems
933 more accurate for interfluvial savanna floodplains (e.g., Negro and Roraima). The high agreement
934 is observed mainly for the maximum inundation estimates, while for the minimum inundation area,
935 important disagreements persist and more studies should be performed to understand them.
936 Overall, the Hess' dataset has been the Amazon inundation benchmark for many years, and still
937 provides satisfactory estimates. Detection of inundation by L-band SAR has a sound theoretical
938 and empirical basis that has been validated for the Amazon (Rosenqvist et al., 2002; Hess et al.,
939 2003). Optical datasets with resolution higher than 30 m are available, but detection of inundation
940 is restricted to non-vegetated wetlands and clear-sky periods, and is most applicable in the lower
941 Amazon River floodplains. ALOS-PALSAR at 12.5 m resolution and Sentinel SAR at 10 m
942 resolution (with C-band and limited vegetation penetration) can be applied to specific regions.
943 Time series of these datasets can estimate seasonal variations in inundation, but are limited by the
944 length of the acquisitions. Weekly to monthly, spatially coarser data (25 km) are available from
945 passive microwave-based datasets such as GIEMS, SWAF and SWAMPS. Downscaling
946 techniques have improved their spatial resolution to 90 m (GIEMS-D3) and 1 km (SWAF-HR).
947 Hydrological models (e.g., CaMa-Flood and MGB) are capable of accurately estimating
948 inundation over river floodplains, and at high temporal resolution depending on the input rainfall
949 data (e.g., hourly to daily). However, they are still limited over interfluvial wetlands with less

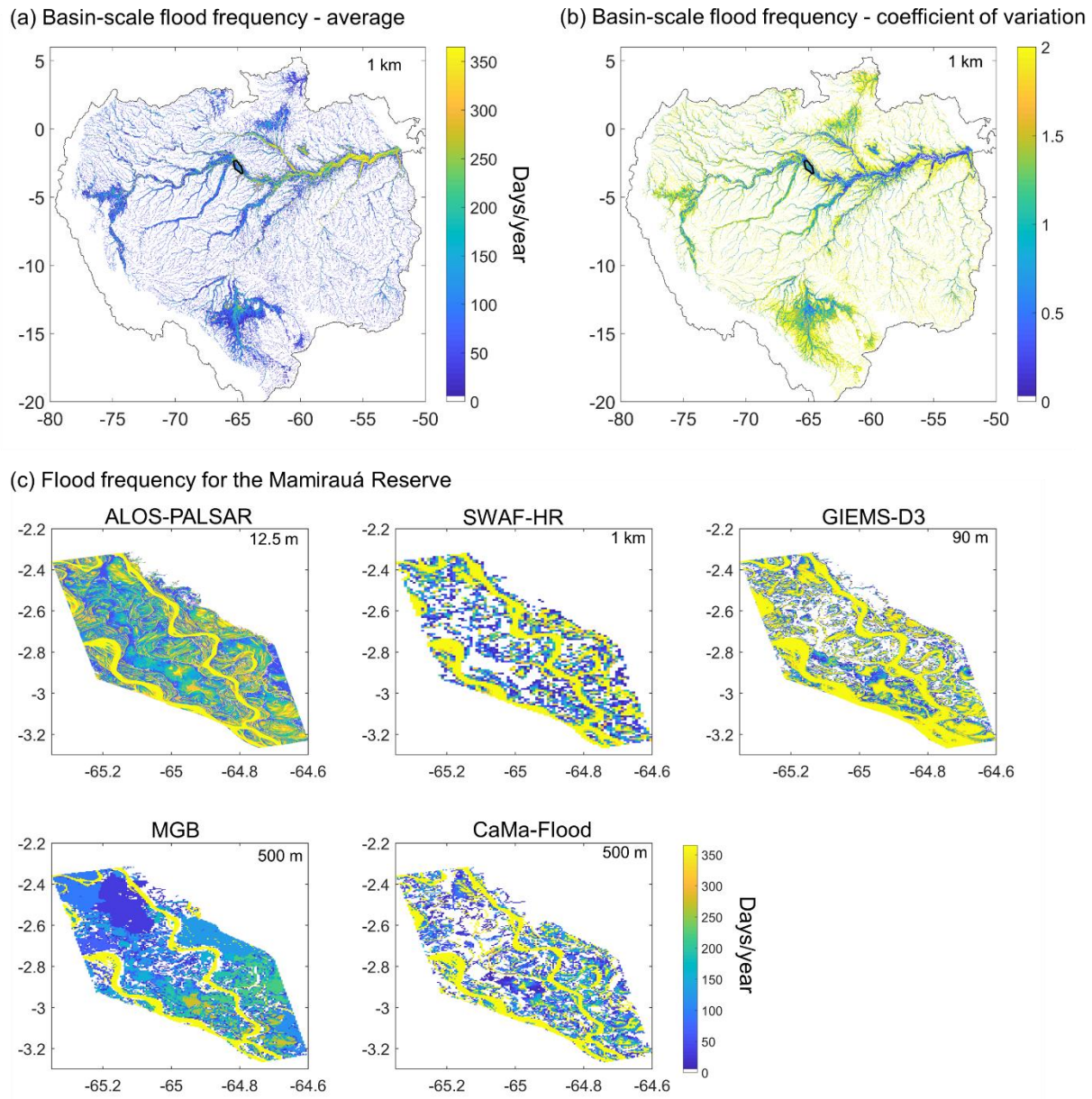
950 connection with rivers, unless they are upgraded for simulating 2D inundation processes and
951 complex floodplain flow paths (Fleischmann et al., 2020; Yamazaki et al., 2014).

952

953 ***4.2 What are the current capabilities of flood frequency mapping?***

954 At the basin scale, high-resolution, long-term average flood frequency can be estimated by four of
955 the datasets analyzed here (GIEMS-D3, SWAF-HR, MGB and CaMa-Flood), with spatial
956 resolutions ranging from 90 m to 1 km. Although multiple SAR data are currently available (e.g.,
957 Sentinel-1, ALOS-PALSAR and ALOS-2 PALSAR-2), they have a limited temporal resolution,
958 and we still do not have a flood frequency dataset of higher spatial resolution (i.e., better than 90
959 m) for the whole basin based on SAR. The discrepancies among the available datasets are notable
960 (Figure 12). The average of the basin-scale flood frequency shows a higher agreement for areas
961 with high flood frequency along the lower Amazon River (Figure 12a). These are associated with
962 a high proportion of open water areas, and have lower uncertainty (Figure 12b). Generally, there
963 is a smaller variation along floodplains bordering the major rivers (except for their fringes) than in
964 interfluvial areas, especially in the Negro and Roraima wetlands (Figure 12b). Detailed inundation
965 mapping for the Mamirauá Sustainable Development Reserve in the Amazon mainstem floodplain
966 (Figure 12c) reinforces the challenges for mapping local spatio-temporal inundation dynamics.
967 The northern part of the Mamirauá reserve has a shorter flood frequency in all datasets, while three
968 of them (SWAF-HR, GIEMS-D3, CaMa-Flood) estimate that large portions are never flooded. For
969 the southern part, there is some convergence for areas that are frequently flooded.

970



971

972 Figure 12. Analysis of flood frequency for (a) basin-scale average and (b) coefficient of variation of the long-term
 973 flood frequency estimated from four high-resolution dynamic datasets (GIEMS-D3, SWAF-HR, CaMa-Flood and
 974 MGB). (c) The four basin-scale datasets are compared to a subregional validation dataset (i.e., the ALOS-PALSAR-
 975 based classification by Ferreira-Ferreira et al. (2015), displayed in the top left panel) for the Mamirauá Sustainable
 976 Development Reserve along the central Amazon River mainstem (location shown by black outline in figure a).

977

978 ***4.3 Implications for biogeochemistry, ecology and flood management***

979 The divergent estimates of Amazon inundation extent have major implications for the
980 quantification of the role of wetlands in global biogeochemical cycles, ecosystem processes and
981 natural disaster management.

982 First, different datasets have been used to quantify the role of Amazon wetlands in the carbon cycle
983 (Guilhen et al., 2020; Melack et al., 2004; Richey et al., 2002; Saunois et al., 2020). An
984 intercomparison assessment of global models forced with different inundation datasets for the
985 Amazon could provide insights into their sensitivity to the estimated inundation. This would be
986 particularly important for modeled estimates of methane flux, given the region's significant
987 contribution to global methane emissions from natural wetlands (Basso et al., 2021). Furthermore,
988 for a proper estimation of methane and carbon dioxide fluxes, dynamic inundation estimates are
989 necessary; this study shows that most coarse-resolution dynamic datasets capture relatively well
990 the seasonality (i.e., the timing of high and low water periods) of annual flooding at a large scale
991 (but not at the local scales), but the magnitude of inundation area over time is still associated with
992 significant errors (Fig. S6).

993 The understanding of the ecology of Amazon freshwaters has benefited from advances in remote
994 sensing-based mapping of inundation. Hydrological variables of interest in relation to wildlife
995 (Alvarenga et al., 2018; Bodmer et al., 2018) and vegetation distribution (Hess et al., 2015, 2003)
996 include hydroperiod, floodplain water depth (Arantes et al., 2013; Fassoni-Andrade et al., 2020),
997 and (lateral) surface water connectivity (Castello, 2008; Duponchelle et al., 2021; Reis et al.,
998 2019a, 2019b), and should be better estimated by future datasets. In addition, many wetland
999 ecosystem studies are performed at the tree stand level (e.g., floristic inventories) and require high

1000 spatial resolution inundation estimates to perform meaningful spatial analyses accounting for
1001 spatial heterogeneity of wetland vegetation. Furthermore, besides a simple interfluvial/floodplain
1002 categorization of wetlands as performed here (section 3.4), which is reasonable from a hydrologic
1003 perspective, improving our understanding of the ecology of Amazon freshwater systems requires
1004 accurate mapping of habitats and their diverse vegetation types (e.g., grasslands, particular
1005 monodominant tree species, herbaceous plants). For instance, floodplain forest cover has been
1006 positively correlated to fishery yields (Arantes et al., 2018) and fish abundance (Lobón-Cerviá et
1007 al., 2015). While this wetland habitat mapping has already been done by some initiatives at the
1008 basin (Hess et al., 2015, 2003) and subregional scales (Ferreira-Ferreira et al., 2015; Silva et al.,
1009 2013), there is still a need for higher resolution and dynamic datasets.

1010 Regarding flood monitoring in the context of natural hazard management, the flood warning
1011 systems of regional water authorities in the basin provide information based on river discharge and
1012 water level at monitoring stations (e.g., Brazil's Geological Survey SACE system;
1013 <<http://sace.cprm.gov.br/amazonas/#>>). In addition, there are other available monitoring and
1014 forecasting services that have been developed for the global scale, such as the Global Flood
1015 Detection System (<https://www.gdacs.org/flooddetection/>), based on remote sensing, and the
1016 Global Flood Monitoring System (<http://flood.umd.edu/>) and the Global Flood Awareness System
1017 (<https://www.globalfloods.eu/>), based on hydrological modeling. The currently available, basin-
1018 scale inundation datasets are unable to map flood hazard at the detailed resolution required for
1019 flood management applications, especially concerning urban areas (Almeida et al., 2018). High-
1020 resolution flood mapping has been achieved using hydraulic modeling based on local surveys of
1021 river bathymetry and floodplain LiDAR DTM, but only for a few specific sites such as the lower
1022 Madeira River (Fleischmann et al., 2021).

1023

1024 *4.4 Future opportunities and recommendations*

1025 Future satellite missions will provide opportunities for improved inundation mapping in the
1026 Amazon, especially the polarimetric and interferometric L-band SAR data from the upcoming
1027 NASA/ISRO mission (NISAR), the P-Band BIOMASS mission from ESA, and the Ka-band Radar
1028 Interferometer (KaRIn) swath observations from the forthcoming SWOT mission (Biancamaria et
1029 al., 2016). New inundation detection technology under development with Global Navigation
1030 Satellite System-Reflectometry (GNSS-R), such as the Cyclone GNSS (CYGNSS) constellation
1031 of GNSS-R satellites, holds promise to provide higher frequency observations of water level
1032 changes (Jensen et al., 2018; Ruf et al., 2018; Rodriguez-Alvarez et al., 2019). Further studies with
1033 the ALOS-2 PALSAR-2 data also are promising, in order to achieve new dynamic inundation
1034 detection, as well as ongoing assessments of the accuracy of the newly available high temporal
1035 resolution inundation datasets (e.g., SWAF-HR with 3-day availability). Consistent and updated
1036 validation products of Amazon inundation are required, which could be derived from airborne,
1037 satellite, or UAV-based LiDAR surveys along multiple wetlands, in particular for overlooked
1038 wetlands such as the Negro and Roraima floodable savannas where measured water levels in rivers
1039 may not adequately predict inundation area. This is especially important for the minimum
1040 inundation extent, which showed large uncertainties among the multiple datasets.

1041 Comprehensive comparisons among multiple inundation datasets are scarce in the literature, yet
1042 are valuable ways to understand benefits and limitations of each of them. A few examples include
1043 a continental-scale assessment of flood model hazard maps in Africa (Trigg et al., 2016) and
1044 regional assessment of inundation in floodplains of Nigeria and Mozambique (Bernhofen et al.,
1045 2018), both based on global hydrological models. Similar initiatives for other areas worldwide

1046 would be welcome, especially for those that lack consistent flood mapping, such as the Congo and
1047 other large wetland systems in Africa (Papa et al., 2022). Furthermore, the combination and
1048 integration of multiple inundation datasets present a promising and effective approach (Gumbrecht
1049 et al., 2017; Hu et al., 2017). We recommend that future developments include optimal data
1050 merging approaches, e.g., by integrating inundation extent into models accounting for water cycle
1051 components with multiple constraints (Meyer Oliveira et al, 2020; Pellet et al., 2021), and by
1052 considering new types of datasets (e.g., GNSS-R; Jensen et al., 2018). Bias of different datasets
1053 could be corrected based on intercomparisons such as those we present here. For instance, recent
1054 studies have performed inundation bias correction using the Hess dataset (Aires et al., 2013;
1055 Sorribas et al., 2016). However, merging of different datasets must be performed with caution, in
1056 a consistent way, avoiding double counting of surfaces, as well as missing others: its success
1057 critically depends upon a good understanding of the limitations and assets of each individual
1058 dataset. The optimal combination of hydrological-hydraulic models with satellite flood maps using
1059 techniques such as data assimilation is also a promising alternative at the basin scale (Wongchuig
1060 et al., 2020).

1061 There is a need for the development of more large-scale 2D hydrological model applications,
1062 especially for large wetland complexes such as the Llanos de Moxos and Pacaya-Samiria, to better
1063 represent inundation dynamics (Fleischmann et al., 2020). 2D models have been applied mainly
1064 to some local-scale areas in the Amazon mainstem floodplain (Pinel et al., 2019; Rudorff et al.,
1065 2014; Trigg et al., 2009; Wilson et al., 2007). Furthermore, inundation anomalies are still poorly
1066 understood owing to the lack of ground-based inundation observations during extreme floods and
1067 droughts. Therefore, validation of estimates for extreme years has usually been performed with
1068 river water level data (in situ or from satellite altimetry) (Silva et al., 2018; Wongchuig et al.,

1069 2019). Future works should address which datasets and methodologies are the most suitable for
1070 mapping extreme events. Furthermore, besides inundation extent, flood storage (Frappart et al.,
1071 2005; Papa et al., 2008; Schumann et al., 2016; Papa and Frappart, 2021) and water velocity (Pinel
1072 et al., 2019) are necessary hydraulic variables to properly address multiple environmental studies
1073 (e.g., flood monitoring, flood attenuation by floodplains, fish floodplain habitats), but to date have
1074 not been well studied in the Amazon.

1075 Finally, there is a need for better-informed usage of the currently available inundation datasets by
1076 multiple local and regional stakeholders (e.g., local water authorities, national water agencies), as
1077 well as research communities not close to remote sensing groups. This will only be achieved
1078 through a two-way interaction with these actors and development of easy-to-access visualization
1079 platforms (i.e., investment in hydroinformatics), as well as training of regional/local user
1080 communities. To this end, we have developed a WebGIS platform ([https://amazon-
1081 inundation.herokuapp.com/](https://amazon-inundation.herokuapp.com/)) to display and provide data acquisition links for the inundation
1082 datasets assessed here, which will be continuously updated once new datasets are made available.
1083 The interaction with local users would bring important feedback on the large-scale datasets as well,
1084 for instance through citizen science initiatives that are ongoing in the Amazon
1085 (<https://www.amazoniacienciaciudadana.org/>).

1086

1087

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1105

1106 **References**

1107 Abril, G., Martinez, J.M., Artigas, L.F., Moreira-Turcq, P., Benedetti, M.F., Vidal, L., Meziane,
1108 T., Kim, J.-H., Bernardes, M.C., Savoye, N., Deborde, J., Souza, E.L., Albéric, P., Landim
1109 de Souza, M.F., Roland, F., 2014. Amazon River carbon dioxide outgassing fuelled by
1110 wetlands. *Nature* 505, 395–398. <https://doi.org/10.1038/nature12797>

1111 Aires, F., Miolane, L., Prigent, C., Pham, B., Fluet-Chouinard, E., Lehner, B., Papa, F., 2017. A
1112 Global Dynamic Long-Term Inundation Extent Dataset at High Spatial Resolution Derived
1113 through Downscaling of Satellite Observations. *J. Hydrometeorol.* 18, 1305–1325.

- 1114 <https://doi.org/10.1175/JHM-D-16-0155.1>
- 1115 Aires, F., Papa, F., Prigent, C., 2013. A Long-Term, High-Resolution Wetland Dataset over the
1116 Amazon Basin, Downscaled from a Multiwavelength Retrieval Using SAR Data. *J.*
1117 *Hydrometeorol.* 14, 594–607. <https://doi.org/10.1175/JHM-D-12-093.1>
- 1118 Aires, F., Prigent, C., Fluet-Chouinard, E., Yamazaki, D., Papa, F., Lehner, B., 2018.
1119 Comparison of visible and multi-satellite global inundation datasets at high-spatial
1120 resolution. *Remote Sens. Environ.* 216, 427–441. <https://doi.org/10.1016/j.rse.2018.06.015>
- 1121 Al Bitar, A., Parrens, M., Fatras, C., Luque, S. P., 2020. Global Weekly Inland Surface Water
1122 Dynamics from L-Band Microwave. In *IGARSS 2020-2020 IEEE International Geoscience*
1123 *and Remote Sensing Symposium*, 5089-5092.
- 1124 Alsdorf, D., Bates, P., Melack, J., Wilson, M., Dunne, T., 2007. Spatial and temporal complexity
1125 of the Amazon flood measured from space. *Geophys. Res. Lett.* 34.
1126 <https://doi.org/10.1029/2007GL029447>
- 1127 Alvarenga, G.C., Ramalho, E.E., Baccaro, F.B., da Rocha, D.G., Ferreira-Ferreira, J., Dineli
1128 Bobrowiec, P.E., 2018. Spatial patterns of medium and large size mammal assemblages in
1129 várzea and terra firme forests, Central Amazonia, Brazil. *PLoS One* 13, 1–19.
1130 <https://doi.org/10.1371/journal.pone.0198120>
- 1131 Andrade, M.M.N. de, Bandeira, I.C.N., Fonseca, D.D.F., Bezerra, P.E.S., Andrade, Á. de S.,
1132 Oliveira, R.S. de, 2017. Flood Risk Mapping in the Amazon, in: *Flood Risk Management.*
1133 *InTech*, p. 13. <https://doi.org/10.5772/intechopen.68912>

- 1134 Arantes, C.C., Castello, L., Cetra, M., Schilling, A., 2013. Environmental influences on the
1135 distribution of arapaima in Amazon floodplains. *Environ. Biol. Fishes* 96, 1257–1267.
1136 <https://doi.org/10.1007/s10641-011-9917-9>
- 1137 Arantes, C.C., Winemiller, K.O., Petrere, M., Castello, L., Hess, L.L., Freitas, C.E.C., 2018.
1138 Relationships between forest cover and fish diversity in the Amazon River floodplain. *J.*
1139 *Appl. Ecol.* 55, 386–395. <https://doi.org/10.1111/1365-2664.12967>
- 1140 Armijos, E., Crave, A., Espinoza, J.C., Filizola, N., Espinoza-Villar, R., Ayes, Fonseca, P.,
1141 Fraizy, P., Gutierrez, O., Vauchel, P., Camenen, B., Martinez, J.M., Dos Santos, A.,
1142 Santini, W., Cochonneau, G., Guyot, J.L., 2020. Rainfall control on Amazon sediment flux:
1143 synthesis from 20 years of monitoring. *Environ. Res. Commun.* 2, 051008.
1144 <https://doi.org/10.1088/2515-7620/ab9003>
- 1145 Arnesen, A.S., Silva, T.S.F., Hess, L.L., Novo, E.M.L.M., Rudorff, C.M., Chapman, B.D.,
1146 McDonald, K.C., 2013. Monitoring flood extent in the lower Amazon River floodplain
1147 using ALOS/PALSAR ScanSAR images. *Remote Sens. Environ.* 130, 51–61.
1148 <https://doi.org/10.1016/j.rse.2012.10.035>
- 1149 Asner, G.P., 2001. Cloud cover in Landsat observations of the Brazilian Amazon. *Int. J. Remote*
1150 *Sens.* 22, 3855–3862. <https://doi.org/10.1080/01431160010006926>
- 1151 Balsamo, G., Beljaars, A., Scipal, K., Viterbo, P., van den Hurk, B., Hirschi, M., and Betts, A.
1152 K.: A Revised Hydrology for the ECMWF Model: Verification from Field Site to
1153 Terrestrial Water Storage and Impact in the Integrated Forecast System, *J. Hydrometeorol.*,
1154 10, 623–643, <https://doi.org/10.1175/2008jhm1068.1>, 2009.

- 1155 Barichivich, J., Gloor, E., Peylin, P., Brienen, R.J.W., Schöngart, J., Espinoza, J.C., Pattnayak,
1156 K.C., 2018. Recent intensification of Amazon flooding extremes driven by strengthened
1157 Walker circulation. *Sci. Adv.* 4. <https://doi.org/10.1126/sciadv.aat8785>
- 1158 Basso, L.S., Marani, L., Gatti, L. V., Miller, J.B., Gloor, M., Melack, J., Cassol, H.L.G., Tejada,
1159 G., Domingues, L.G., Arai, E., Sanchez, A.H., Corrêa, S.M., Anderson, L., Aragão,
1160 L.E.O.C., Correia, C.S.C., Crispim, S.P., Neves, R.A.L., 2021. Amazon methane budget
1161 derived from multi-year airborne observations highlights regional variations in emissions.
1162 *Commun. Earth Environ.* 2, 1–14. <https://doi.org/10.1038/s43247-021-00314-4>
- 1163 Batalha, M.A., Cianciaruso, M. V., Silva, I.A., Delitti, W.B.C., 2005. Hyperseasonal cerrado, a
1164 new brazilian vegetation form. *Brazilian J. Biol.* 65, 735–738.
1165 <https://doi.org/10.1590/S1519-69842005000400021>
- 1166 Bates, P.D., De Roo, A.P.J., 2000. A simple raster-based model for flood inundation simulation.
1167 *J. Hydrol.* [https://doi.org/10.1016/S0022-1694\(00\)00278-X](https://doi.org/10.1016/S0022-1694(00)00278-X)
- 1168 Beck, H.E., Van Dijk, A.I.J.M., Levizzani, V., Schellekens, J., Miralles, D.G., Martens, B., De
1169 Roo, A., 2017. MSWEP: 3-hourly 0.25° global gridded precipitation (1979-2015) by
1170 merging gauge, satellite, and reanalysis data. *Hydrol. Earth Syst. Sci.*
1171 <https://doi.org/10.5194/hess-21-589-2017>
- 1172 Beighley, R.E., Eggert, K.G., Dunne, T., He, Y., Gummadi, V., Verdin, K.L., 2009. Simulating
1173 hydrologic and hydraulic processes throughout the Amazon River Basin. *Hydrol. Process.*
1174 23, 1221–1235. <https://doi.org/10.1002/hyp.7252>
- 1175 Belger, L., Forsberg, B. R. Melack, J. M., 2011. Carbon dioxide and methane emissions from

1176 interfluvial wetlands in the upper Negro River basin, Brazil. *Biogeochemistry* 105, 171–
1177 183. <https://doi.org/10.1007/s10533-010-9536-0>

1178 Bernhofen, M. V, Whyman, C., Trigg, M.A., Sleigh, P.A., Smith, A.M., Sampson, C.C.,
1179 Yamazaki, D., Ward, P.J., Rudari, R., Pappenberger, F., Dottori, F., Salamon, P.,
1180 Winsemius, H.C., 2018. A first collective validation of global fluvial flood models for
1181 major floods in Nigeria and Mozambique. *Environ. Res. Lett.* 13, 104007.
1182 <https://doi.org/10.1088/1748-9326/aae014>

1183 Biancamaria, S., Lettenmaier, D. P., Pavelsky, T. M, 2016. The SWOT Mission and Its
1184 Capabilities for Land Hydrology. *Surv. Geophys.* 37, 307–337.
1185 <https://doi.org/10.1007/s10712-015-9346-y>

1186 Blatrix, R., Roux, B., Béarez, P., Prestes-Carneiro, G., Amaya, M., Aramayo, J.L., Rodrigues, L.,
1187 Lombardo, U., Iriarte, J., De Souza, J.G., Robinson, M., Bernard, C., Pouilly, M., Durécu,
1188 M., Huchzermeyer, C.F., Kalebe, M., Ovando, A., McKey, D., 2018. The unique
1189 functioning of a pre-Columbian Amazonian floodplain fishery. *Sci. Rep.* 8.
1190 <https://doi.org/10.1038/s41598-018-24454-4>

1191 Bodmer, R., Mayor, P., Antunez, M., Chota, K., Fang, T., Puertas, P., Pittet, M., Kirkland, M.,
1192 Walkey, M., Rios, C., Perez-Peña, P., Henderson, P., Bodmer, W., Bicerra, A., Zegarra, J.,
1193 Docherty, E., 2018. Major shifts in Amazon wildlife populations from recent intensification
1194 of floods and drought. *Conserv. Biol.* 32, 333–344. <https://doi.org/10.1111/cobi.12993>

1195 Bonnet, M.P., Barroux, G., Martinez, J.M., Seyler, F., Moreira-Turcq, P., Cochonneau, G.,
1196 Melack, J.M., Boaventura, G., Maurice-Bourgoin, L., León, J.G., Roux, E., Calmant, S.,

1197 Kosuth, P., Guyot, J.L., Seyler, P., 2008. Floodplain hydrology in an Amazon floodplain
1198 lake (Lago Grande de Curuaí). *J. Hydrol.* 349, 18–30.
1199 <https://doi.org/10.1016/j.jhydrol.2007.10.055>

1200 Bonnet, M.P., Pinel, S., Garnier, J., Bois, J., Resende Boaventura, G., Seyler, P., Motta Marques,
1201 D., 2017. Amazonian floodplain water balance based on modelling and analyses of
1202 hydrologic and electrical conductivity data. *Hydrol. Process.* 31, 1702–1718.
1203 <https://doi.org/10.1002/hyp.11138>

1204 Bontemps, S., Defourny, P., Radoux, J., Van Bogaert, E. Lamarche, C., Achard, F., Mayaux, P.,
1205 Boettcher, M., Brockmann, C., Kirches, G., 2013. Consistent global land cover maps for
1206 climate modelling communities: current achievements of the ESA’s land cover CCI, in:
1207 Proceedings of the ESA Living Planet Symposium. Edinburgh.

1208 Bourgoin, L.M., Bonnet, M.P., Martinez, J.M., Kosuth, P., Cochonneau, G., Moreira-Turcq, P.,
1209 Guyot, J.L., Vauchel, P., Filizola, N., Seyler, P., 2007. Temporal dynamics of water and
1210 sediment exchanges between the Curuaí floodplain and the Amazon River, Brazil. *J.*
1211 *Hydrol.* 335, 140–156. <https://doi.org/10.1016/j.jhydrol.2006.11.023>

1212 Canisius, F., Brisco, B., Murnaghan, K., Van Der Kooij, M., Keizer, E., 2019. SAR backscatter
1213 and InSAR coherence for monitoring wetland extent, flood pulse and vegetation: A study of
1214 the Amazon lowland. *Remote Sens.* 11, 1–18. <https://doi.org/10.3390/RS11060720>

1215 Castello, L., 2008. Lateral migration of *Arapaima gigas* in floodplains of the Amazon. *Ecol.*
1216 *Freshw. Fish* 17, 38–46. <https://doi.org/10.1111/j.1600-0633.2007.00255.x>

1217 Chapman, B., McDonald, K., Shimada, M., Rosenqvist, A., Schroeder, R., Hess, L., 2015.

- 1218 Mapping Regional Inundation with Spaceborne L-Band SAR. *Remote Sens.* 7, 5440–5470.
1219 <https://doi.org/10.3390/rs70505440>
- 1220 Coe, M.T., Costa, M.H., Howard, E.A., 2008. Simulating the surface waters of the Amazon
1221 River basin: impacts of new river geomorphic and flow parameterizations. *Hydrol. Process.*
1222 22, 2542–2553. <https://doi.org/10.1002/hyp.6850>
- 1223 Coomes, O.T., Lapointe, M., Templeton, M., List, G., 2016. Amazon river flow regime and flood
1224 recession agriculture: Flood stage reversals and risk of annual crop loss. *J. Hydrol.* 539,
1225 214–222. <https://doi.org/10.1016/j.jhydrol.2016.05.027>
- 1226 Coomes, O.T., Takasaki, Y., Abizaid, C., Barham, B.L., 2010. Floodplain fisheries as natural
1227 insurance for the rural poor in tropical forest environments: Evidence from Amazonia. *Fish.*
1228 *Manag. Ecol.* 17, 513–521. <https://doi.org/10.1111/j.1365-2400.2010.00750.x>
- 1229 Dalmagro, H.J., de A. Lobo, F., Vourlitis, G.L., Dalmolin, Â.C., Antunes, M.Z., Ortíz, C.E.R.,
1230 de S. Nogueira, J., 2016. Photosynthetic response of a wetland- and an upland-adapted tree
1231 species to seasonal variations in hydrology in the Brazilian Cerrado and Pantanal. *Acta*
1232 *Physiol. Plant.* 38, 107. <https://doi.org/10.1007/s11738-016-2125-7>
- 1233 de Almeida, G.A.M., Bates, P., Ozdemir, H., 2018. Modelling urban floods at submetre
1234 resolution: challenges or opportunities for flood risk management? *J. Flood Risk Manag.*
1235 11, S855–S865. <https://doi.org/10.1111/jfr3.12276>
- 1236 Denevan, W.M., 1996. A Bluff Model of Riverine Settlement in Prehistoric Amazonia. *Ann.*
1237 *Assoc. Am. Geogr.* 86, 654–681. <https://doi.org/10.1111/j.1467-8306.1996.tb01771.x>

- 1238 Draper, F.C., Roucoux, K.H., Lawson, I.T., Mitchard, E.T.A., Honorio Coronado, E.N.,
1239 Lahteenoja, O., Montenegro, L.T., Sandoval, E.V., Zarate, R., Baker, T.R., 2014. The
1240 distribution and amount of carbon in the largest peatland complex in Amazonia. *Environ.*
1241 *Res. Lett.* 9. <https://doi.org/10.1088/1748-9326/9/12/124017>
- 1242 Dunne, T., Mertes, L.A.K., Meade, R.H., Richey, J.E., Forsberg, B.R., 1998. Exchanges of
1243 sediment between the flood plain and channel of the Amazon River in Brazil. *Bull. Geol.*
1244 *Soc. Am.* 110, 450–467. [https://doi.org/10.1130/0016-](https://doi.org/10.1130/0016-7606(1998)110<0450:EOSBTF>2.3.CO;2)
1245 [7606\(1998\)110<0450:EOSBTF>2.3.CO;2](https://doi.org/10.1130/0016-7606(1998)110<0450:EOSBTF>2.3.CO;2)
- 1246 Duponchelle, F., Isaac, V.J., Doria, C., Van Damme, P.A., Herrera-R, G.A., Anderson, E.P.,
1247 Cruz, R.E.A., Hauser, M., Hermann, T.W., Agudelo, E., Bonilla-Castillo, C., Barthem, R.,
1248 Freitas, C.E.C., Garcıa-Davila, C., Garcıa-Vasquez, A., Renno, J., Castello, L., 2021.
1249 Conservation of migratory fishes in the Amazon basin. *Aquat. Conserv. Mar. Freshw.*
1250 *Ecosyst. aqc.3550*. <https://doi.org/10.1002/aqc.3550>
- 1251 Espinoza Villar, J.C., Ronchail, J., Guyot, J.L., Cochonneau, G., Naziano, F., Lavado, W., De
1252 Oliveira, E., Pombosa, R., Vauchel, P., 2009. Spatio-temporal rainfall variability in the
1253 Amazon basin countries (Brazil, Peru, Bolivia, Colombia, and Ecuador). *Int. J. Climatol.*
1254 29, 1574–1594. <https://doi.org/10.1002/joc.1791>
- 1255 Espinoza Villar, J.C., Ronchail, J., Marengo, J.A., Segura, H., 2019. Contrasting North–South
1256 changes in Amazon wet-day and dry-day frequency and related atmospheric features (1981–
1257 2017). *Clim. Dyn.* 52, 5413–5430. <https://doi.org/10.1007/s00382-018-4462-2>

1258 Farr, T.G., Rosen, P.A., Caro, E., Crippen, R., Duren, R., Hensley, S., Kobrick, M., Paller, M.,
1259 Rodriguez, E., Roth, L., Seal, D., Shaffer, S., Shimada, J., Umland, J., Werner, M., Oskin,
1260 M., Burbank, D., Alsdorf, D.E., 2007. The shuttle radar topography mission. *Rev. Geophys.*
1261 45, 1–25. <https://doi.org/10.1029/2005RG000183>

1262 Fassoni-Andrade, 2020. PhD thesis. Mapeamento e caracterização do sistema rio-planície da
1263 Amazônia central via sensoriamento remoto e modelagem hidráulica. Federal University of
1264 Rio Grande do Sul. Available at <<https://lume.ufrgs.br/handle/10183/211269>>.

1265 Fassoni-Andrade, A.C., Fleischmann, A.S., Papa, F., Paiva, R.C.D. de, Wongchuig, S., Melack,
1266 J.M., Moreira, A.A., Paris, A., Ruhoff, A., Barbosa, C., Maciel, D.A., Novo, E., Durand, F.,
1267 Frappart, F., Aires, F., Abrahão, G.M., Ferreira-Ferreira, J., Espinoza, J.C., Laipelt, L.,
1268 Costa, M.H., Espinoza-Villar, R., Calmant, S., Pellet, V., 2021. Amazon Hydrology From
1269 Space: Scientific Advances and Future Challenges. *Rev. Geophys.* 59, 1–97.
1270 <https://doi.org/10.1029/2020RG000728>

1271 Fassoni-Andrade, A.C., Paiva, R.C.D. de, 2019. Mapping spatial-temporal sediment dynamics of
1272 river-floodplains in the Amazon. *Remote Sens. Environ.*
1273 <https://doi.org/10.1016/j.rse.2018.10.038>

1274 Fassoni-Andrade, A.C., Paiva, R.C.D. de, Rudorff, C. de M., Barbosa, C.C.F., Novo, E.M.L. de
1275 M., 2020. High-resolution mapping of floodplain topography from space: A case study in
1276 the Amazon. *Remote Sens. Environ.* 251, 112065. <https://doi.org/10.1016/j.rse.2020.112065>

1277 Feng, D., Raoufi, R., Beighley, E., Melack, J.M., Goulding, M., Barthem, R.B., Venticinque, E.,
1278 Cañas, C., Forsberg, B., Sorribas, M.V., 2020. Future climate impacts on the hydrology of

1279 headwater streams in the Amazon River Basin: Implications for migratory goliath catfishes.
1280 Hydrol. Process. hyp.13952. <https://doi.org/10.1002/hyp.13952>

1281 Ferreira-Ferreira, J., Silva, T.S.F., Streher, A.S., Affonso, A.G., de Almeida Furtado, L.F.,
1282 Forsberg, B.R., Valsecchi, J., Queiroz, H.L., de Moraes Novo, E.M.L., 2015. Combining
1283 ALOS/PALSAR derived vegetation structure and inundation patterns to characterize major
1284 vegetation types in the Mamirauá Sustainable Development Reserve, Central Amazon
1285 floodplain, Brazil. *Wetl. Ecol. Manag.* 23, 41–59. [https://doi.org/10.1007/s11273-014-9359-](https://doi.org/10.1007/s11273-014-9359-1)
1286 1

1287 Fleischmann, A.S., Fialho Brêda, J.P., Rudorff, C., Dias de Paiva, R.C., Collischonn, W., Papa,
1288 F., Ravello, M.M., 2021. River Flood Modeling and Remote Sensing Across Scales:
1289 Lessons from Brazil, in: Schumann, G.J.P. (Ed.), *Earth Observation for Flood Applications*.
1290 Elsevier, pp. 61–103. <https://doi.org/10.1016/B978-0-12-819412-6.00004-3>

1291 Fleischmann, A.S., Paiva, R.C.D., Collischonn, W., Siqueira, V.A., Paris, A., Moreira, D.M.,
1292 Papa, F., Bitar, A.A., Parrens, M., Aires, F., Garambois, P.A., 2020. Trade-Offs Between 1-
1293 D and 2-D Regional River Hydrodynamic Models. *Water Resour. Res.* 56.
1294 <https://doi.org/10.1029/2019WR026812>

1295 Flores, B.M., Holmgren, M., 2021. White-Sand Savannas Expand at the Core of the Amazon
1296 After Forest Wildfires. *Ecosystems* 24, 1624–1637. [https://doi.org/10.1007/s10021-021-](https://doi.org/10.1007/s10021-021-00607-x)
1297 00607-x

1298 Fluet-Chouinard, E., Lehner, B., Rebelo, L.M., Papa, F., Hamilton, S.K., 2015. Development of
1299 a global inundation map at high spatial resolution from topographic downscaling of coarse-

1300 scale remote sensing data. *Remote Sens. Environ.* 158, 348–361.
1301 <https://doi.org/10.1016/j.rse.2014.10.015>

1302 Frappart, F., Seyler, F., Martinez, J., León, J.G., Cazenave, A., 2005. Floodplain water storage in
1303 the Negro River basin estimated from microwave remote sensing of inundation area and
1304 water levels. *Remote Sens. Environ.* 99, 387–399. <https://doi.org/10.1016/j.rse.2005.08.016>

1305 Fu, R., Yin, L., Li, W., Arias, P.A., Dickinson, R.E., Huang, L., Chakraborty, S., Fernandes, K.,
1306 Liebmann, B., Fisher, R., Myneni, R.B., 2013. Increased dry-season length over southern
1307 Amazonia in recent decades and its implication for future climate projection. *Proc. Natl.*
1308 *Acad. Sci. U. S. A.* 110, 18110–18115. <https://doi.org/10.1073/pnas.1302584110>

1309 Getirana, A., Boone, A., Yamazaki, D., Decharme, B., Papa, F., Mognard, N., 2012. The
1310 Hydrological Modeling and Analysis Platform (HyMAP): Evaluation in the Amazon Basin.
1311 *J. Hydrometeorol.* 13, 1641–1665. <https://doi.org/10.1175/JHM-D-12-021.1>

1312 Getirana, A., Peters-Lidard, C., Rodell, M., Bates, P.D., 2017. Trade-off between cost and
1313 accuracy in large-scale surface water dynamic modeling. *Water Resour. Res.* 53, 4942–
1314 4955. <https://doi.org/10.1002/2017WR020519>

1315 Gloor, M., Brienen, R.J.W., Galbraith, D., Feldpausch, T.R., Schöngart, J., Guyot, J.L.,
1316 Espinoza, J.C., Lloyd, J., Phillips, O.L., 2013. Intensification of the Amazon hydrological
1317 cycle over the last two decades. *Geophys. Res. Lett.* 40, 1729–1733.
1318 <https://doi.org/10.1002/grl.50377>

1319 Guilhen, J., Al Bitar, A., Sauvage, S., Parrens, M., Martinez, J., Abril, G., Moreira-Turcq, P.,
1320 Sánchez-Pérez, J.-M., 2020. Denitrification and associated nitrous oxide and carbon dioxide

1321 emissions from the Amazonian wetlands. *Biogeosciences* 17, 4297–4311.
1322 <https://doi.org/10.5194/bg-17-4297-2020>

1323 Guimberteau, M., Drapeau, G., Ronchail, J., Sultan, B., Polcher, J., Martinez, J.-M., Prigent, C.,
1324 Guyot, J.-L., Cochonneau, G., Espinoza, J.C., Filizola, N., Fraizy, P., Lavado, W., De
1325 Oliveira, E., Pombosa, R., Noriega, L., Vauchel, P., 2012. Discharge simulation in the sub-
1326 basins of the Amazon using ORCHIDEE forced by new datasets. *Hydrol. Earth Syst. Sci.*
1327 16, 911–935. <https://doi.org/10.5194/hess-16-911-2012>

1328 Gumbrecht, T., Roman-Cuesta, R.M., Verchot, L., Herold, M., Wittmann, F., Householder, E.,
1329 Herold, N., Murdiyarso, D., 2017. An expert system model for mapping tropical wetlands
1330 and peatlands reveals South America as the largest contributor. *Glob. Chang. Biol.* 23,
1331 3581–3599. <https://doi.org/10.1111/gcb.13689>

1332 Hamilton, S.K., Kellndorfer, J., Lehner, B., Tobler, M., 2007. Remote sensing of floodplain
1333 geomorphology as a surrogate for biodiversity in a tropical river system (Madre de Dios,
1334 Peru). *Geomorphology* 89, 23–38. <https://doi.org/10.1016/j.geomorph.2006.07.024>

1335 Hamilton, S.K., Sippel, S.J., Melack, J.M., 2002. Comparison of inundation patterns among
1336 major South American floodplains. *Journal of Geophysical Research* 107 (D20): Art. No.
1337 8038, 1–14. <https://doi.org/10.1029/2000JD000306>

1338 Hamilton, S.K., Sippel, S.J., Melack, J.M., 2004. Seasonal inundation patterns in two large
1339 savanna floodplains of South America: the Llanos de Moxos(Bolivia) and the Llanos del
1340 Orinoco(Venezuela and Colombia). *Hydrol. Process.* 18, 2103–2116.
1341 <https://doi.org/10.1002/hyp.5559>

1342 Hansen, M.C., Potapov, P. V., Moore, R., Hancher, M., Turubanova, S.A., Tyukavina, A., Thau,
1343 D., Stehman, S. V., Goetz, S.J., Loveland, T.R., Kommareddy, A., Egorov, A., Chini, L.,
1344 Justice, C.O., Townshend, J.R.G., 2013. High-Resolution Global Maps of 21st-Century
1345 Forest Cover Change. *Science* (80-.). 342, 850–853.
1346 <https://doi.org/10.1126/science.1244693>

1347 Harris, I., Jones, P.D., Osborn, T.J., Lister, D.H., 2014. Updated high-resolution grids of monthly
1348 climatic observations - the CRU TS3.10 Dataset. *Int. J. Climatol.* 34, 623–642.
1349 <https://doi.org/10.1002/joc.3711>

1350 Hawes, J.E., Peres, C.A., Riley, L.B., Hess, L.L., 2012. Landscape-scale variation in structure
1351 and biomass of Amazonian seasonally flooded and unflooded forests. *For. Ecol. Manage.*
1352 281, 163–176. <https://doi.org/10.1016/j.foreco.2012.06.023>

1353 Hess, L.L., Melack, J.M., Affonso, A.G., Barbosa, C., Gastil-Buhl, M., Novo, E.M.L.M., 2015.
1354 Wetlands of the Lowland Amazon Basin: Extent, Vegetative Cover, and Dual-season
1355 Inundated Area as Mapped with JERS-1 Synthetic Aperture Radar. *Wetlands* 35, 745–756.
1356 <https://doi.org/10.1007/s13157-015-0666-y>

1357 Hess, L.L., Melack, J.M., Novo, E.M.L.M., Barbosa, C.C.F., Gastil, M., 2003. Dual-season
1358 mapping of wetland inundation and vegetation for the central Amazon basin. *Remote Sens.*
1359 *Environ.* 87, 404–428. <https://doi.org/10.1016/j.rse.2003.04.001>

1360 Hoch, J.M., Haag, A. V., Van Dam, A., Winsemius, H.C., Van Beek, L.P.H., Bierkens, M.F.P.,
1361 2017. Assessing the impact of hydrodynamics on large-scale flood wave propagation: A
1362 case study for the Amazon Basin. *Hydrol. Earth Syst. Sci.* 21, 117–132.

- 1363 <https://doi.org/10.5194/hess-21-117-2017>
- 1364 Hu, S., Niu, Z., Chen, Y., 2017. Global Wetland Datasets: a Review. *Wetlands* 37, 807–817.
- 1365 <https://doi.org/10.1007/s13157-017-0927-z>
- 1366 Jardim, C.M., Nardoto, G.B., de Lima, A.C.B., de Jesus Silva, R., Schor, T., de Oliveira, J.A.,
1367 Martinelli, L.A., 2020. The influence of seasonal river flooding in food consumption of
1368 riverine dwellers in the central Amazon region: an isotopic approach. *Archaeol. Anthropol.*
1369 *Sci.* 12. <https://doi.org/10.1007/s12520-020-01172-5>
- 1370 Jensen, K., McDonald, K., 2019. Surface Water Microwave Product Series Version 3: A Near-
1371 Real Time and 25-Year Historical Global Inundated Area Fraction Time Series From Active
1372 and Passive Microwave Remote Sensing. *IEEE Geosci. Remote Sens. Lett.* 16, 1402–1406.
1373 <https://doi.org/10.1109/lgrs.2019.2898779>
- 1374 Jensen, K., McDonald, K., Podest, E., Rodriguez-Alvarez, N., Horna, V., Steiner, N., 2018.
1375 Assessing L-Band GNSS-Reflectometry and Imaging Radar for Detecting Sub-Canopy
1376 Inundation Dynamics in a Tropical Wetlands Complex. *Remote Sens.* 10, 1431.
1377 <https://doi.org/10.3390/rs10091431>
- 1378 Ji, X., Lesack, L.F.W., Melack, J.M., Wang, S., Riley, W.J., Shen, C., 2019. Seasonal and
1379 Interannual Patterns and Controls of Hydrological Fluxes in an Amazon Floodplain Lake
1380 With a Surface-Subsurface Process Model. *Water Resour. Res.* 55, 3056–3075.
1381 <https://doi.org/10.1029/2018WR023897>
- 1382 Junk, W.J., Bayley, P.B., Sparks, R.E, 1989. The flood pulse concept in river-floodplain systems.
1383 *Canadian special publication of fisheries and aquatic sciences*, 106.1, 110-127.

- 1384 Junk, W.J., Furch, K., Limnologie, M., Tropenökologie, A., Plon, W.-, 1993. A general review
1385 of tropical South American floodplains. *Wetl. Ecol. Manag.* 2, 231–238.
- 1386 Junk, W.J., Piedade, M.T.F., Schöngart, J., Cohn-Haft, M., Adeney, J.M., Wittmann, F., 2011. A
1387 classification of major naturally-occurring amazonian lowland wetlands. *Wetlands* 31, 623–
1388 640. <https://doi.org/10.1007/s13157-011-0190-7>
- 1389 Lähteenoja, O., Reátegui, Y.R., Räsänen, M., Torres, D.D.C., Oinonen, M., Page, S., 2012. The
1390 large Amazonian peatland carbon sink in the subsiding Pastaza-Marañón foreland basin,
1391 Peru. *Glob. Chang. Biol.* 18, 164–178. <https://doi.org/10.1111/j.1365-2486.2011.02504.x>
- 1392 Langerwisch, F., Rost, S., Gerten, D., Poulter, B., Rammig, A., Cramer, W., 2013. Potential
1393 effects of climate change on inundation patterns in the Amazon Basin. *Hydrol. Earth Syst.*
1394 *Sci.* 17, 2247–2262. <https://doi.org/10.5194/hess-17-2247-2013>
- 1395 Langill, J.C., Abizaid, C., 2020. What is a bad flood? Local perspectives of extreme floods in the
1396 Peruvian Amazon. *Ambio* 49, 1423–1436. <https://doi.org/10.1007/s13280-019-01278-8>
- 1397 Latrubesse, E.M., 2012. Amazon lakes, in: *Encyclopedia of Earth Sciences Series*.
1398 https://doi.org/10.1007/978-1-4020-4410-6_36
- 1399 Lauerwald, R., Regnier, P., Camino-Serrano, M., Guenet, B., Guimberteau, M., Ducharne, A.,
1400 Polcher, J., Ciais, P., 2017. ORCHILEAK (revision 3875): a new model branch to simulate
1401 carbon transfers along the terrestrial–aquatic continuum of the Amazon basin. *Geosci.*
1402 *Model Dev.* 10, 3821–3859. <https://doi.org/10.5194/gmd-10-3821-2017>
- 1403 Lehner, B., Döll, P., 2004. Development and validation of a global database of lakes, reservoirs

1404 and wetlands. *J. Hydrol.* 296, 1–22. <https://doi.org/10.1016/j.jhydrol.2004.03.028>

1405 Lehner, B., Grill, G., 2013. Global river hydrography and network routing: Baseline data and
1406 new approaches to study the world’s large river systems. *Hydrol. Process.*
1407 <https://doi.org/10.1002/hyp.9740>

1408 Lesack, L.F.W., Melack, J.M., 1995. Flooding Hydrology and Mixture Dynamics of Lake Water
1409 Derived from Multiple Sources in an Amazon Floodplain Lake. *Water Resour. Res.* 31,
1410 329–345. <https://doi.org/10.1029/94WR02271>

1411 Li, D., Lu, D., Moran, E., da Silva, R.F.B., 2020. Examining water area changes accompanying
1412 dam construction in the Madeira River in the Brazilian Amazon. *Water (Switzerland)* 12.
1413 <https://doi.org/10.3390/w12071921>

1414 Lobón-Cerviá, J., Hess, L.L., Melack, J.M., Araujo-Lima, C.A.R.M., 2015. The importance of
1415 forest cover for fish richness and abundance on the Amazon floodplain. *Hydrobiologia* 750,
1416 245–255. <https://doi.org/10.1007/s10750-014-2040-0>

1417 Luize, B.G., Silva, T.S.F., Wittmann, F., Assis, R.L., Venticinque, E.M., 2015. Effects of the
1418 Flooding Gradient on Tree Community Diversity in Várzea Forests of the Purus River,
1419 Central Amazon, Brazil. *Biotropica* 47, 137–142. <https://doi.org/10.1111/btp.12203>

1420 Luo, X., Li, H.-Y., Leung, L.R., Tesfa, T.K., Getirana, A., Papa, F., Hess, L.L., 2017. Modeling
1421 surface water dynamics in the Amazon Basin using MOSART-Inundation v1.0: impacts of
1422 geomorphological parameters and river flow representation. *Geosci. Model Dev.* 10, 1233–
1423 1259. <https://doi.org/10.5194/gmd-10-1233-2017>

- 1424 Mansur, A. V., Brondízio, E.S., Roy, S., Hetrick, S., Vogt, N.D., Newton, A., 2016. An
1425 assessment of urban vulnerability in the Amazon Delta and Estuary: a multi-criterion index
1426 of flood exposure, socio-economic conditions and infrastructure. *Sustain. Sci.* 11, 625–643.
1427 <https://doi.org/10.1007/s11625-016-0355-7>
- 1428 Martinez, J.M., Le Toan, T., 2007. Mapping of flood dynamics and spatial distribution of
1429 vegetation in the Amazon floodplain using multitemporal SAR data. *Remote Sens. Environ.*
1430 108, 209–223. <https://doi.org/10.1016/j.rse.2006.11.012>
- 1431 Martínez-Espinosa, C., Sauvage, S., Al Bitar, A., Green, P. A., Vörösmarty, C. J., Sánchez-
1432 Pérez, J. M. (2020). Denitrification in wetlands: A review towards a quantification at global
1433 scale. *Science of the total environment* 754, 142398.
1434 <https://doi.org/10.1016/j.scitotenv.2020.142398>
- 1435 Matthews, E., Fung, I., 1987. Methane emission from natural wetlands: Global distribution, area,
1436 and environmental characteristics of sources. *Global Biogeochem. Cycles* 1, 61–86.
1437 <https://doi.org/10.1029/GB001i001p00061>
- 1438 Melack, J.M., 2016. Aquatic Ecosystems, in: *Ecological Studies*. pp. 119–148.
1439 https://doi.org/10.1007/978-3-662-49902-3_7
- 1440 Melack, J.M., Coe, M.T., 2021. Amazon floodplain hydrology and implications for aquatic
1441 conservation. *Aquat. Conserv. Mar. Freshw. Ecosyst.* 1029–1040.
1442 <https://doi.org/10.1002/aqc.3558>
- 1443 Melack, J.M., Forsberg, B.R., 2001. Biogeochemistry of Amazon Floodplain, in: McClain, M.E.,
1444 Victoria, R., Richey, J.E. (Eds.), *The Biogeochemistry of the Amazon Basin*. Oxford

1445 University Press, New York, USA.

1446 Melack, J.M., Hess, L.L., 2010. Remote Sensing of the Distribution and Extent of Wetlands in
1447 the Amazon Basin, in: Amazonian Floodplain Forests. pp. 43–59.
1448 https://doi.org/10.1007/978-90-481-8725-6_3

1449 Melack, J.M., Hess, L.L., Gastil, M., Forsberg, B.R., Hamilton, S.K., Lima, I.B.T., Novo,
1450 E.M.L.M., 2004. Regionalization of methane emissions in the Amazon Basin with
1451 microwave remote sensing. *Glob. Chang. Biol.* 10, 530–544. [https://doi.org/10.1111/j.1365-
1452 2486.2004.00763.x](https://doi.org/10.1111/j.1365-2486.2004.00763.x)

1453 Meyer Oliveira, A., Fleischmann, A., Paiva, R., 2020. On the contribution of remote sensing-
1454 based calibration to model multiple hydrological variables. *J. Hydrol.*
1455 <https://doi.org/10.1016/j.jhydrol.2021.126184>

1456 Miguez-Macho, G., Fan, Y., 2012. The role of groundwater in the Amazon water cycle: 1.
1457 Influence on seasonal streamflow, flooding and wetlands. *J. Geophys. Res. Atmos.* 117, 1–
1458 30. <https://doi.org/10.1029/2012JD017539>

1459 Nardi, F., Annis, A., Di Baldassarre, G., Vivoni, E.R., Grimaldi, S., 2019. GFPLAIN250m, a
1460 global high-resolution dataset of Earth’s floodplains. *Sci. Data* 6, 180309.
1461 <https://doi.org/10.1038/sdata.2018.309>

1462 O’Loughlin, F.E., Paiva, R.C.D., Durand, M., Alsdorf, D.E., Bates, P.D., 2016. A multi-sensor
1463 approach towards a global vegetation corrected SRTM DEM product. *Remote Sens.*
1464 *Environ.* 182, 49–59. <https://doi.org/10.1016/j.rse.2016.04.018>

- 1465 Ovando, A., Martinez, J.M., Tomasella, J., Rodriguez, D.A., von Randow, C., 2018. Multi-
1466 temporal flood mapping and satellite altimetry used to evaluate the flood dynamics of the
1467 Bolivian Amazon wetlands. *Int. J. Appl. Earth Obs. Geoinf.* 69, 27–40.
1468 <https://doi.org/10.1016/j.jag.2018.02.013>
- 1469 Ovando, A., Tomasella, J., Rodriguez, D.A., Martinez, J.M., Siqueira-Junior, J.L., Pinto, G.L.N.,
1470 Passy, P., Vauchel, P., Noriega, L., von Randow, C., 2016. Extreme flood events in the
1471 Bolivian Amazon wetlands. *J. Hydrol. Reg. Stud.* 5, 293–308.
1472 <https://doi.org/10.1016/j.ejrh.2015.11.004>
- 1473 Paca, V.H. da M., Espinoza-Dávalos, G.E., Hessels, T.M., Moreira, D.M., Comair, G.F.,
1474 Bastiaanssen, W.G.M., 2019. The spatial variability of actual evapotranspiration across the
1475 Amazon River Basin based on remote sensing products validated with flux towers. *Ecol.*
1476 *Process.* 8, 6. <https://doi.org/10.1186/s13717-019-0158-8>
- 1477 Paiva, R., Buarque, D.C., Collischonn, W., Bonnet, M.P., Frappart, F., Calmant, S., Bulhões
1478 Mendes, C.A., 2013. Large-scale hydrologic and hydrodynamic modeling of the Amazon
1479 River basin. *Water Resour. Res.* 49, 1226–1243. <https://doi.org/10.1002/wrcr.20067>
- 1480 Pangala, S.R., Enrich-Prast, A., Basso, L.S., Peixoto, R.B., Bastviken, D., Hornibrook, E.R.C.,
1481 Gatti, L. V., Marotta, H., Calazans, L.S.B., Sakuragui, C.M., Bastos, W.R., Malm, O.,
1482 Gloor, E., Miller, J.B., Gauci, V., 2017. Large emissions from floodplain trees close the
1483 Amazon methane budget. *Nature* 552, 230–234. <https://doi.org/10.1038/nature24639>
- 1484 Papa, F., J.-F. Crétaux, M. Grippa, E. Robert, M. Trigg, R. Tshimanga, B. Kitambo, A. Paris, A.
1485 Carr, A.S. Fleischmann, M. de Fleury, P.G. Gbetkom, B. Calmettes, Calmant, S., 2022.

1486 Water Resources in Africa under Global Change: Monitoring Surface Waters from Space.
1487 Surveys in Geophysics. <https://doi.org/10.1007/s10712-022-09700-9>

1488 Papa, F., and F. Frappart (2021), Surface Water Storage in Rivers and Wetlands Derived from
1489 Satellite Observations: A Review of Current Advances and Future Opportunities for
1490 Hydrological Sciences, *Remote Sensing*, 13(20), 4162, doi:10.3390/rs13204162

1491 Papa, F., Güntner, A., Frappart, F., Prigent, C., Rossow, W.B., 2008. Variations of surface water
1492 extent and water storage in large river basins: A comparison of different global data
1493 sources. *Geophys. Res. Lett.* 35, L11401. <https://doi.org/10.1029/2008GL033857>

1494 Papa, F., Prigent, C., Aires, F., Jimenez, C., Rossow, W.B., Matthews, E., 2010. Interannual
1495 variability of surface water extent at the global scale, 1993–2004. *J. Geophys. Res.* 115,
1496 D12111. <https://doi.org/10.1029/2009JD012674>

1497 Park, E., Latrubesse, E.M., 2017. The hydro-geomorphologic complexity of the lower Amazon
1498 River floodplain and hydrological connectivity assessed by remote sensing and field
1499 control. *Remote Sens. Environ.* 198, 321–332. <https://doi.org/10.1016/j.rse.2017.06.021>

1500 Park, E., Latrubesse, E.M., 2019. A geomorphological assessment of wash-load sediment fluxes
1501 and floodplain sediment sinks along the lower Amazon River. *Geology* 47, 403–406.
1502 <https://doi.org/10.1130/G45769.1>

1503 Parrens, M., Al Bitar, A., Frappart, F., Papa, F., Calmant, S., Crétaux, J.-F., Wigneron, J.-P.,
1504 Kerr, Y., 2017. Mapping Dynamic Water Fraction under the Tropical Rain Forests of the
1505 Amazonian Basin from SMOS Brightness Temperatures. *Water* 9, 350.
1506 <https://doi.org/10.3390/w9050350>

1507 Parrens, M., Bitar, A. Al, Frappart, F., Paiva, R., Wongchuig, S., Papa, F., Yamasaki, D., Kerr,
1508 Y., 2019. High resolution mapping of inundation area in the Amazon basin from a
1509 combination of L-band passive microwave, optical and radar datasets. *Int. J. Appl. Earth*
1510 *Obs. Geoinf.* 81, 58–71. <https://doi.org/10.1016/j.jag.2019.04.011>

1511 Pekel, J., Cottam, A., Gorelick, N., Belward, A.S., 2016. High-resolution mapping of global
1512 surface water and its long-term changes. *Nature* 540, 418–422.
1513 <https://doi.org/10.1038/nature20584>

1514 Pellet, V., Aires, F., Yamazaki, D., Papa, F., 2021. Coherent Satellite Monitoring of the Water
1515 Cycle Over the Amazon. Part 1: Methodology and Initial Evaluation. *Water Resour. Res.*
1516 57, 1–21. <https://doi.org/10.1029/2020wr028647>

1517 Pham-Duc, B., Prigent, C., Aires, F., Papa, F., 2017. Comparisons of global terrestrial surface
1518 water datasets over 15 years. *J. Hydrometeorol.* 18, 993–1007.
1519 <https://doi.org/10.1175/JHM-D-16-0206.1>

1520 Pickens, A.H., Hansen, M.C., Hancher, M., Stehman, S. V., Tyukavina, A., Potapov, P.,
1521 Marroquin, B., Sherani, Z., 2020. Mapping and sampling to characterize global inland water
1522 dynamics from 1999 to 2018 with full Landsat time-series. *Remote Sens. Environ.* 243,
1523 111792. <https://doi.org/10.1016/j.rse.2020.111792>

1524 Pinel, S., Bonnet, M., S. Da Silva, J., Sampaio, T.C., Garnier, J., Catry, T., Calmant, S., Fragoso,
1525 C.R., Moreira, D., Motta Marques, D., Seyler, F., 2019. Flooding dynamics within an
1526 Amazonian floodplain: water circulation patterns and inundation duration. *Water Resour.*
1527 *Res.* 2019WR026081. <https://doi.org/10.1029/2019WR026081>

1528 Prigent, C., Jimenez, C., Bousquet, P., 2020. Satellite-Derived Global Surface Water Extent and
1529 Dynamics Over the Last 25 Years (GIEMS-2). *J. Geophys. Res. Atmos.* 125, 1–18.
1530 <https://doi.org/10.1029/2019JD030711>

1531 Prigent, C., Matthews, E., Aires, F., Rossow, W.B., 2001. Remote sensing of global wetland
1532 dynamics with multiple satellite data sets. *Geophys. Res. Lett.* 28, 4631–4634.
1533 <https://doi.org/10.1029/2001GL013263>

1534 Prigent, C., Papa, F., Aires, F., Rossow, W.B., Matthews, E., 2007. Global inundation dynamics
1535 inferred from multiple satellite observations, 1993-2000. *J. Geophys. Res. Atmos.* 112,
1536 1993–2000. <https://doi.org/10.1029/2006JD007847>

1537 Prigent, C., Rochetin, N., Aires, F., Defer, E., Grandpeix, J.-Y., Jimenez, C., Papa, F., 2011.
1538 Impact of the inundation occurrence on the deep convection at continental scale from
1539 satellite observations and modeling experiments. *J. Geophys. Res. Atmos.* 116, n/a-n/a.
1540 <https://doi.org/10.1029/2011JD016311>

1541 Reis, V., Hermoso, V., Hamilton, S.K., Bunn, S.E., Fluet-Chouinard, E., Venables, B., Linke, S.,
1542 2019. Characterizing seasonal dynamics of Amazonian wetlands for conservation and
1543 decision making. *Aquat. Conserv. Mar. Freshw. Ecosyst.* 29, 1073–1082.
1544 <https://doi.org/10.1002/aqc.3051>

1545 Reis, V., Hermoso, V., Hamilton, S.K., Bunn, S.E., Linke, S., 2019b. Conservation planning for
1546 river-wetland mosaics: A flexible spatial approach to integrate floodplain and upstream
1547 catchment connectivity. *Biol. Conserv.* 236, 356–365.
1548 <https://doi.org/10.1016/j.biocon.2019.05.042>

1549 Rennó, C.D., Nobre, A.D., Cuartas, L.A., Soares, J.V., Hodnett, M.G., Tomasella, J., Waterloo,
1550 M.J., 2008. HAND, a new terrain descriptor using SRTM-DEM: Mapping terra-firme
1551 rainforest environments in Amazonia. *Remote Sens. Environ.* 112, 3469–3481.
1552 <https://doi.org/10.1016/j.rse.2008.03.018>

1553 Renó, V.F., Novo, E.M.L.M., Suemitsu, C., Rennó, C.D., Silva, T.S.F., 2011. Assessment of
1554 deforestation in the Lower Amazon floodplain using historical Landsat MSS/TM imagery.
1555 *Remote Sens. Environ.* 115, 3446–3456. <https://doi.org/10.1016/j.rse.2011.08.008>

1556 Resende, A.F. de, Schöngart, J., Streher, A.S., Ferreira-Ferreira, J., Piedade, M.T.F., Silva,
1557 T.S.F., 2019. Massive tree mortality from flood pulse disturbances in Amazonian floodplain
1558 forests: The collateral effects of hydropower production. *Sci. Total Environ.* 659, 587–598.
1559 <https://doi.org/10.1016/j.scitotenv.2018.12.208>

1560 Restrepo A, J.D., Kettner, A.J., Robert Brakenridge, G., 2020. Monitoring water discharge and
1561 floodplain connectivity for the northern Andes utilizing satellite data: A tool for river
1562 planning and science-based decision-making. *J. Hydrol.* 586, 124887.
1563 <https://doi.org/10.1016/j.jhydrol.2020.124887>

1564 Richey, J.E., Hedges, J.I., Devol, A.H., Quay, P.D., Victoria, R., Martinelli, L., Forsberg, B.R.,
1565 1990. Biogeochemistry of carbon in the Amazon River. *Limnol. Oceanogr.* 35, 352–371.
1566 <https://doi.org/10.4319/lo.1990.35.2.0352>

1567 Richey, J.E., Melack, J.M., Aufdenkampe, A.K., Ballester, V.M., Hess, L.L., 2002. Outgassing
1568 from Amazonian rivers and wetlands as a large tropical source of atmospheric CO₂. *Nature*
1569 416, 617–620. <https://doi.org/10.1038/416617a>

1570 Ringeval, B., Decharme, B., Piao, S.L., Ciais, P., Papa, F., de Noblet-Ducoudré, N., Prigent, C.,
1571 Friedlingstein, P., Gouttevin, I., Koven, C., Ducharne, A., 2012. Modelling sub-grid
1572 wetland in the ORCHIDEE global land surface model: evaluation against river discharges
1573 and remotely sensed data. *Geosci. Model Dev.* 5, 941–962. [https://doi.org/10.5194/gmd-5-](https://doi.org/10.5194/gmd-5-941-2012)
1574 941-2012

1575 Ringeval, B., Houweling, S., van Bodegom, P.M., Spahni, R., van Beek, R., Joos, F., Röckmann,
1576 T., 2014. Methane emissions from floodplains in the Amazon Basin: challenges in
1577 developing a process-based model for global applications. *Biogeosciences* 11, 1519–1558.
1578 <https://doi.org/10.5194/bg-11-1519-2014>

1579 Rodriguez-Alvarez, N., Podest, E., Jensen, K., McDonald, K.C., 2019. Classifying Inundation in
1580 a Tropical Wetlands Complex with GNSS-R. *Remote Sens.* 11, 1053.
1581 <https://doi.org/10.3390/rs11091053>

1582 Rosenqvist, A., Forsberg, B.R., Pimentel, T., Rauste, Y.A., Richey, J.E., 2002. The use of
1583 spaceborne radar data to model inundation patterns and trace gas emissions in the central
1584 Amazon floodplain. *Int. J. Remote Sens.* 23, 1303–1328.
1585 <https://doi.org/10.1080/01431160110092911>

1586 Rosenqvist, J., Rosenqvist, A., Jensen, K., McDonald, K., 2020. Mapping of Maximum and
1587 Minimum Inundation Extents in the Amazon Basin 2014–2017 with ALOS-2 PALSAR-2
1588 ScanSAR Time-Series Data. *Remote Sens.* 12, 1326. <https://doi.org/10.3390/rs12081326>

1589 Rosinger, A.Y., 2018. Household water insecurity after a historic flood: Diarrhea and
1590 dehydration in the Bolivian Amazon. *Soc. Sci. Med.* 197, 192–202.

1591 <https://doi.org/10.1016/j.socscimed.2017.12.016>

1592 Rossetti, D.F., Gribel, R., Rennó, C.D., Cohen, M.C.L., Moulatlet, G.M., Cordeiro, C.L. de O.,
1593 Rodrigues, E. do S.F., 2017a. Late Holocene tectonic influence on hydrology and vegetation
1594 patterns in a northern Amazonian megafan. *Catena* 158, 121–130.
1595 <https://doi.org/10.1016/j.catena.2017.06.022>

1596 Rossetti, D.F., Valeriano, M.M., Gribel, R., Cohen, M.C.L., Tatumi, S.H., Yee, M., 2017b. The
1597 imprint of Late Holocene tectonic reactivation on a megafan landscape in the northern
1598 Amazonian wetlands. *Geomorphology* 295, 406–418.
1599 <https://doi.org/10.1016/j.geomorph.2017.07.026>

1600 Rudorff, C.M., Melack, J.M., Bates, P.D., 2014. Flooding dynamics on the lower Amazon
1601 floodplain: 1. Hydraulic controls on water elevation, inundation extent, and river-floodplain
1602 discharge. *Water Resour. Res.* 50, 619–634. <https://doi.org/10.1002/2013WR014091>

1603 Ruf, C.S., Chew, C., Lang, T. et al. A New Paradigm in Earth Environmental Monitoring with
1604 the CYGNSS Small Satellite Constellation. *Sci Rep* 8, 8782 (2018).
1605 <https://doi.org/10.1038/s41598-018-27127-4>

1606 Ruiz Agudelo, C.A., Mazzeo, N., Díaz, I., Barral, M.P., Piñeiro, G., Gadino, I., Roche, I., Acuña-
1607 Posada, R.J., 2020. Land use planning in the amazon basin: Challenges from resilience
1608 thinking. *Ecol. Soc.* 25. <https://doi.org/10.5751/ES-11352-250108>

1609 Santos, J.O.S., Nelson, B.W., Giovannini, C.A., 1993. Corpos de areia sob leitos abandonados de
1610 grandes rios. *Ciência Hoje* 16, 22–25.

1611 Santos, L.B.L., Carvalho, T., Anderson, L.O., Rudorff, C.M., Marchezini, V., Londe, L.R., Saito,
1612 S.M., 2017. An RS-GIS-Based Comprehensive Impact Assessment of Floods—A Case
1613 Study in Madeira River, Western Brazilian Amazon. *IEEE Geosci. Remote Sens. Lett.* 14,
1614 1614–1617. <https://doi.org/10.1109/LGRS.2017.2726524>

1615 Santos, M.J., Medvigy, D., Silva Dias, M.A.F., Freitas, E.D., Kim, H., 2019. Seasonal Flooding
1616 Causes Intensification of the River Breeze in the Central Amazon. *J. Geophys. Res. Atmos.*
1617 124, 5178–5197. <https://doi.org/10.1029/2018JD029439>

1618 Saunio, M., Stavert, A.R., Poulter, B., Bousquet, P., Canadell, J.G., Jackson, R.B., Raymond,
1619 P.A., Dlugokencky, E.J., Houweling, S., Patra, P.K., Ciais, P., Arora, V.K., Bastviken, D.,
1620 Bergamaschi, P., Blake, D.R., Brailsford, G., Bruhwiler, L., Carlson, K.M., Carrol, M.,
1621 Castaldi, S., Chandra, N., Crevoisier, C., Crill, P.M., Covey, K., Curry, C.L., Etiope, G.,
1622 Frankenberg, C., Gedney, N., Hegglin, M.I., Höglund-Isaksson, L., Hugelius, G., Ishizawa,
1623 M., Ito, A., Janssens-Maenhout, G., Jensen, K.M., Joos, F., Kleinen, T., Krummel, P.B.,
1624 Langenfelds, R.L., Laruelle, G.G., Liu, L., Machida, T., Maksyutov, S., McDonald, K.C.,
1625 McNorton, J., Miller, P.A., Melton, J.R., Morino, I., Müller, J., Murguia-Flores, F., Naik,
1626 V., Niwa, Y., Noce, S., O’Doherty, S., Parker, R.J., Peng, C., Peng, S., Peters, G.P., Prigent,
1627 C., Prinn, R., Ramonet, M., Regnier, P., Riley, W.J., Rosentreter, J.A., Segers, A., Simpson,
1628 I.J., Shi, H., Smith, S.J., Steele, L.P., Thornton, B.F., Tian, H., Tohjima, Y., Tubiello, F.N.,
1629 Tsuruta, A., Viovy, N., Voulgarakis, A., Weber, T.S., van Weele, M., van der Werf, G.R.,
1630 Weiss, R.F., Worthy, D., Wunch, D., Yin, Y., Yoshida, Y., Zhang, W., Zhang, Z., Zhao, Y.,
1631 Zheng, B., Zhu, Qing, Zhu, Qian, Zhuang, Q., 2020. The Global Methane Budget 2000–
1632 2017. *Earth Syst. Sci. Data* 12, 1561–1623. <https://doi.org/10.5194/essd-12-1561-2020>

1633 Schöngart, J., Wittmann, F., Faria de Resende, A., Assahira, C., Sousa Lobo, G., Rocha Duarte
1634 Neves, J., Rocha, M., Biem Mori, G., Costa Quaresma, A., Oreste Demarchi, L., Weiss
1635 Albuquerque, B., Oliveira Feitosa, Y., Silva Costa, G., Vieira Feitoza, G., Machado
1636 Durgante, F., Lopes, A., Trumbore, S.E., Sanna Freire Silva, T., Steege, H., Val, A.L., Junk,
1637 W.J., Piedade, M.T.F., 2021. The shadow of the Balbina dam: A synthesis of over 35 years
1638 of downstream impacts on floodplain forests in Central Amazonia. *Aquat. Conserv. Mar.*
1639 *Freshw. Ecosyst.* 31, 1117–1135. <https://doi.org/10.1002/aqc.3526>

1640 Schroeder, R., McDonald, K., Chapman, B., Jensen, K., Podest, E., Tessler, Z., Bohn, T.,
1641 Zimmermann, R., 2015. Development and Evaluation of a Multi-Year Fractional Surface
1642 Water Data Set Derived from Active/Passive Microwave Remote Sensing Data. *Remote*
1643 *Sens.* 7, 16688–16732. <https://doi.org/10.3390/rs71215843>

1644 Schumann, G.J.P., Stampoulis, D., Smith, A.M., Sampson, C.C., Andreadis, K.M., Neal, J.C.,
1645 Bates, P.D., 2016. Rethinking flood hazard at the global scale. *Geophys. Res. Lett.* 43,
1646 10,249-10,256. <https://doi.org/10.1002/2016GL070260>

1647 Silva, M.V., Paris, A., Calmant, S., Cândido, L.A., Silva, J.S. da, 2018. Relationships between
1648 Pacific and Atlantic ocean sea surface temperatures and water levels from satellite altimetry
1649 data in the Amazon rivers. *Brazilian Journal of Water Resources* 23.
1650 <https://doi.org/10.1590/2318-0331.231820170148>

1651 Silva, T.S.F., Melack, J.M., Novo, E.M.L.M., 2013. Responses of aquatic macrophyte cover and
1652 productivity to flooding variability on the Amazon floodplain. *Glob. Chang. Biol.* 19, 3379-
1653 3389. <https://doi.org/10.1111/gcb.12308>

- 1654 Sippel, S.J., Hamilton, S.K., Melack, J.M., 1992. Inundation area and morphometry of lakes on
1655 the Amazon River floodplain, Brazil. *Arch. Hydrobiol* 123, 385–400.
- 1656 Sippel, S.J., Hamilton, S.K., Melack, J.M., Novo, E.M.M., 1998. Passive microwave
1657 observations of inundation area and the area/stage relation in the amazon river floodplain.
1658 *Int. J. Remote Sens.* 19, 3055–3074. <https://doi.org/10.1080/014311698214181>
- 1659 Siqueira, V.A., Paiva, R.C.D., Fleischmann, A.S., Fan, F.M., Ruhoff, A.L., Pontes, P.R.M.,
1660 Paris, A., Calmant, S., Collischonn, W., 2018. Toward continental hydrologic–
1661 hydrodynamic modeling in South America. *Hydrol. Earth Syst. Sci.* 22, 4815–4842.
1662 <https://doi.org/10.5194/hess-22-4815-2018>
- 1663 Sorribas, M.V., Paiva, R.C.D., Melack, J.M., Bravo, J.M., Jones, C., Carvalho, L., Beighley, E.,
1664 Forsberg, B., Costa, M.H., 2016. Projections of climate change effects on discharge and
1665 inundation in the Amazon basin. *Clim. Change* 136, 555–570.
1666 <https://doi.org/10.1007/s10584-016-1640-2>
- 1667 Souza, C.M., Kirchoff, F.T., Oliveira, B.C., Ribeiro, J.G., Sales, M.H., 2019. Long-term annual
1668 surface water change in the Brazilian Amazon Biome: Potential links with deforestation,
1669 infrastructure development and climate change. *Water (Switzerland)* 11, 566.
1670 <https://doi.org/10.3390/w11030566>
- 1671 Staver, A.C., Archibald, S., Levin, S.A., 2011. The Global Extent and Determinants of Savanna
1672 and Forest as Alternative Biome States. *Science (80-.)*. 334, 230–232.
1673 <https://doi.org/10.1126/science.1210465>
- 1674 Taylor, C.M., Prigent, C., Dadson, S.J., 2018. Mesoscale rainfall patterns observed around

1675 wetlands in sub-Saharan Africa. *Q. J. R. Meteorol. Soc.* 144, 2118–2132.
1676 <https://doi.org/10.1002/qj.3311>

1677 Towner, J., Cloke, H.L., Zsoter, E., Flamig, Z., Hoch, J.M., Bazo, J., Coughlan de Perez, E.,
1678 Stephens, E.M., 2019. Assessing the performance of global hydrological models for
1679 capturing peak river flows in the Amazon basin. *Hydrol. Earth Syst. Sci.* 23, 3057–3080.
1680 <https://doi.org/10.5194/hess-23-3057-2019>

1681 Trigg, M.A., Birch, C.E., Neal, J.C., Bates, P.D., Smith, A., Sampson, C.C., Yamazaki, D.,
1682 Hirabayashi, Y., Pappenberger, F., Dutra, E., Ward, P.J., Winsemius, H.C., Salamon, P.,
1683 Dottori, F., Rudari, R., Kappes, M.S., Simpson, A.L., Hadzilacos, G., Fewtrell, T.J., 2016.
1684 The credibility challenge for global fluvial flood risk analysis. *Environ. Res. Lett.* 11,
1685 094014. <https://doi.org/10.1088/1748-9326/11/9/094014>

1686 Trigg, M.A., Wilson, M.D., Bates, P.D., Horritt, M.S., Alsdorf, D.E., Forsberg, B.R., Vega,
1687 M.C., 2009. Amazon flood wave hydraulics. *J. Hydrol.* 374, 92–105.
1688 <https://doi.org/10.1016/j.jhydrol.2009.06.004>

1689 Ward, N.D., Bianchi, T.S., Medeiros, P.M., Seidel, M., Richey, J.E., Keil, R.G., Sawakuchi,
1690 H.O., 2017. Where Carbon Goes When Water Flows: Carbon Cycling across the Aquatic
1691 Continuum. *Front. Mar. Sci.* 4, 1–27. <https://doi.org/10.3389/fmars.2017.00007>

1692 Wilson, M.D., Bates, P., Alsdorf, D., Forsberg, B., Horritt, M., Melack, J., Frappart, F.,
1693 Famiglietti, J., 2007. Modeling large-scale inundation of Amazonian seasonally flooded
1694 wetlands. *Geophys. Res. Lett.* 34, 4–9. <https://doi.org/10.1029/2007GL030156>

1695 Wongchuig, S.C., Paiva, R.C.D. de, Espinoza, J.C., Collischonn, W., 2017. Multi-decadal

1696 Hydrological Retrospective: Case study of Amazon floods and droughts. *J. Hydrol.* 549,
1697 667–684. <https://doi.org/10.1016/j.jhydrol.2017.04.019>

1698 Wongchuig, S.C., de Paiva, R.C.D., Siqueira, V., Collischonn, W., 2019. Hydrological reanalysis
1699 across the 20th century: A case study of the Amazon Basin. *J. Hydrol.* 570, 755–773.
1700 <https://doi.org/10.1016/j.jhydrol.2019.01.025>

1701 Wongchuig, S.C., Paiva, R.C.D., Biancamaria, S., Collischonn, W., 2020. Assimilation of future
1702 SWOT-based river elevations, surface extent observations and discharge estimations into
1703 uncertain global hydrological models. *J. Hydrol.* 590, 125473.
1704 <https://doi.org/10.1016/j.jhydrol.2020.125473>

1705 Wu, J., Lakshmi, V., Wang, D., Lin, P., Pan, M., Cai, X., Wood, E.F., Zeng, Z., 2020. The
1706 Reliability of Global Remote Sensing Evapotranspiration Products over Amazon. *Remote*
1707 *Sens.* 12, 2211. <https://doi.org/10.3390/rs12142211>

1708 Yamazaki, D., Ikeshima, D., Tawatari, R., Yamaguchi, T., O’Loughlin, F., Neal, J.C., Sampson,
1709 C.C., Kanae, S., Bates, P.D., 2017. A high-accuracy map of global terrain elevations.
1710 *Geophys. Res. Lett.* <https://doi.org/10.1002/2017GL072874>

1711 Yamazaki, D., Kanae, S., Kim, H., Oki, T., 2011. A physically based description of floodplain
1712 inundation dynamics in a global river routing model. *Water Resour. Res.* 47, 1–21.
1713 <https://doi.org/10.1029/2010WR009726>

1714 Yamazaki, D., Sato, T., Kanae, S., Hirabayashi, Y., Bates, P.D., 2014. Regional flood dynamics
1715 in a bifurcating mega delta simulated in a global river model. *Geophys. Res. Lett.* 41, 3127–
1716 3135. <https://doi.org/10.1002/2014GL059744>

1717 Yamazaki, D., Trigg, M.A., Ikeshima, D., 2015. Development of a global ~90m water body map
1718 using multi-temporal Landsat images. *Remote Sens. Environ.* 171, 337–351.
1719 <https://doi.org/10.1016/j.rse.2015.10.014>

1720 Zhang, Z., Poulter, B., Fluet-Chouinard, E., Jensen, K., McDonald, K., Hugelius, G., Gumbrecht,
1721 T., Carroll, M., Prigent, C., Bartsch, A., 2020. Development and evaluation of the global
1722 Wetland Area and Dynamics for Methane Modeling dataset (WAD2M). *Earth Syst. Sci.*
1723 Data in review, 1–50. <https://doi.org/10.5194/essd-2020-262>

1724 Zhou, X., Prigent, C., Yamazaki, D., 2021. Toward improved comparisons between land-
1725 surface-water-area estimates from a global river model and satellite observations. *Water*
1726 *Resour. Res.* 57, e2020WR029256. <https://doi.org/10.1029/2020WR029256>

1727 Zubieta, R., Getirana, A., Espinoza, J.C., Lavado, W., 2015. Impacts of satellite-based
1728 precipitation datasets on rainfall-runoff modeling of the Western Amazon basin of Peru and
1729 Ecuador. *J. Hydrol.* 528, 599–612. <https://doi.org/10.1016/j.jhydrol.2015.06.064>

1730 Zubieta, R., Getirana, A., Espinoza, J.C., Lavado-Casimiro, W., Aragon, L., 2017. Hydrological
1731 modeling of the Peruvian-Ecuadorian Amazon Basin using GPM-IMERG satellite-based
1732 precipitation dataset. *Hydrol. Earth Syst. Sci.* 21, 3543–3555. [https://doi.org/10.5194/hess-](https://doi.org/10.5194/hess-21-3543-2017)
1733 [21-3543-2017](https://doi.org/10.5194/hess-21-3543-2017)

1734 Zubieta, R., Saavedra, M., Espinoza, J.C., Ronchail, J., Sulca, J., Drapeau, G., Martin-Vide, J.,
1735 2019. Assessing precipitation concentration in the Amazon basin from different satellite-
1736 based data sets. *Int. J. Climatol.* 39, 3171–3187. <https://doi.org/10.1002/joc.6009>

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1742 **Supplementary Material**

1743

1744 Table S1. List of additional studies that mapped inundation in the Amazon, which were not
1745 included in the article analysis because of redundancy with the used datasets, or data unavailability.

	Reference	Dataset name / Type	Spatial. resolution	Temporal resolution	Time period	Region	Type of inundation captured
1	Aires et al. (2013)	GIEMS + downscaling with SAR	500 m	Monthly	1993-2007	Central Amazon	All
2	Belger et al. (2011)	Radarsat-1 / C-band SAR	25 m	Irregular	2004-2005	Cuini and Itu (Negro basin)	All
3	Bonnet et al. (2008)	Hydrological model		Daily	1997-2003	Curuai	All
4	Canisius et al., 2019)	Radarsat-2 / C-band SAR	2.5-2.6 m	Irregular	2014-2016	Lower Amazon river	All
5	Fleischman et al. (2020)	MGB / Hydrological-hydraulic model	4 km	Daily	1999-2015	Negro River basin	All
6	Frappart et al. (2005)	JERS-1 / L-band SAR	90 m	Static (high and low water)	1995-1996	Negro River basin	All

7	Getirana et al. (2012)	HYMAP / Hydrological model		Daily	1986–2006	Negro River basin	All
8	Guimberteau et al. (2012)	ORCHIDEE / Hydrological model	0.5 degrees	Daily	1980–2000	Basin	All
9	Hawes et al. (2012)	ALOS-PALSAR / L-band SAR	100 m	Irregular	2006–2009	Juruá floodplain	All
10	Hoch et al. (2017)	PCR-GLOBWB / Hydraulic model	30 arcmin	Daily	1985–1990	Central Amazon	All
11	Langerwisch et al. (2013)	LPJmL / Hydrological model	0.5 degrees	Monthly	1961–1990	Basin	All
12	Lauerwald et al. (2017)	ORCHIDEE-ORCHILEAK / Land surface model	0.5 degrees	Daily	1980–2000	Basin	All
13	Lesack and Melack (1995)	In situ data	-	-	-	Lake Calado	All
14	Li et al. (2020)	Landsat (Mapbiomas)	30 m	Annual	1985–2019	Madeira river close to Santo Antônio and Jirau dams	All
15	Luo et al. (2017)	MOSART / Hydraulic model	-	-	-	Basin	All
16	Martinez and Le Toan (2007)	JERS-1 / SAR	25 m	Irregular (21 images)	1993–1997	Curuai	All
17	Miguez-Macho and Fan (2012)	LEAF-Hydro-Flood / Hydrological-hydraulic model	~2 km	Daily	2000–2010	Basin	All

18	Meyer Oliveira et al. (2020)	ALOS-PALSAR / L-band SAR	100 m	Irregular	2006-2010	Purus River basin	All
19	Nardi et al. (2019)	GFPLAIN250m / geomorphic approach	250 m	Static	2002 (SRTM mission)	Basin	Floodplains
20	Paiva et al. (2013)	MGB / Hydrological-hydraulic model	500 m	Daily	1998-2010	Basin	All
21	Ringeval et al. (2012)	TOPMODEL - LSM / Hydrological model	1 degree	Monthly	1993-2004	Basin	All
22	Ringeval et al. (2014)	PCR-GLOBWB / Hydrological model	0.5 degrees	Daily	1979 - 2009	Basin	All
23	Rodriguez-Alvarez et al. (2019)	CYGNSS / GNSS-R	500 m - 7 km	Daily-14 days	2017	Pacaya-Samiria	All
24	Rosenqvist et al. (2002)	JERS-1 / L-band SAR	100 m	Irregular	1996-1997	Jauí river basin	All
25	Silva et al. (2013)	Radarsat-1 / C-band SAR	25 m	Irregular	2003 - 2005	Amazon river (Juruti - Monte alegre)	All
26	Sippel et al. (1992)	RADAMBRASIL / Side-looking Airborne Radar	0.25 degrees	Monthly	1979-1987	Amazon river in Brazil	All
27	Souza et al. (2019)	Landsat	30 m	Annual	1985-2017	Brazilian Amazon	Open water
28	Trigg et al. (2009)	LISFLOOD-FP and HEC-RAS / Hydraulic models	180 m / irregular	Daily	1995-1997	Solimões River (Itapeua - Manaus)	All
29	Wilson et al. (2007)	LISFLOOD-FP / Hydraulic model	270 m	Daily	1995-1997	Solimões River (Itapeua - Manaus)	All
30	Fassoni-Andrade et al., 2019	MODIS	250 m	8-Days	2003-2017	Central Amazon	Open water

1747 Table S2. Main characteristics of the assessed wetlands.

	Name	Location	Characteristics
1	Curuai floodplain	Lower Amazon River	Shallow lakes with high suspended sediment concentrations
2	Janauacá floodplain	Middle Amazon River	Ria lake and “várzea” environments (white-water floodplains)
3	Uatumã floodplain	300-km reach between Balbina dam and the confluence with the Amazon River	Black-water floodplain
4	Mamirauá Reserve	Confluence between Solimões and Japurá rivers	Mosaic of chavascal, herbaceous, and low and high várzea vegetation
5	Purus floodplain	Purus River	Large floodplain to river width ratio
6	Pacaya-Samiria wetlands	Upper Solimões River	Flooded forests, palm swamps and peatlands
7	Llanos de Moxos floodable savannas	Upper Madeira River basin	Interfluvial areas among Beni, Mamoré and Madre de Dios rivers
8	Negro savannas	Negro-Branco interfluvial area	Regional neotectonic depressions
9	Roraima savannas	Smaller river floodplains interspersed with areas subject to flooding by local rainfall in the upper Branco River basin	Poorly drained interfluvial savannas

1748

1749 Table S3. Comparison metrics - Pearson correlation (R) and normalized root mean square error (nRMSD) for time

1750 series, and Fit metric for the spatial analysis of maximum observed inundation area for all datasets against the

1751 subregional estimates for individual wetland complexes: Curuai (Arnesen et al., 2013), Uatumã (Resende et al., 2019),
 1752 Janauacá (Pinel et al., 2019), Mamirauá (Ferreira-Ferreira et al., 2015), Pacaya-Samiria (Jensen et al., 2020), Llanos
 1753 de Moxos (Ovando et al., 2016) and Lower Amazon (Park et al., 2019). Four additional subregional datasets were
 1754 compared to the local ones mentioned above: Curuai LISFLOOD-FP model (Rudorff et al., 2014), Janauacá
 1755 hydrological model (Bonnet et al., 2017), Janauacá TELEMAC-2D model (Pinel et al., 2019), and Llanos de Moxos
 1756 ALOS-PALSAR (Ovando et al., 2016). The Fit metric was applied by converting all maps to 1 km, considering a pixel
 1757 with inundation fraction higher than 50% as inundated.

	Dataset	-	Curuai				Uatu mã	Janauacá			Mamira uá	Pacaya-Samiria			Llanos de Moxos			Lower Amaz on
	-	Period	2006-2010				2006- 2011	2007-2011			2007- 2010	2014-2018			2001-2014			2000- 2020
			R	nRM SD	Fit	Fit	R	nR M SD	Fit	Fit	R	nRM SD	Fit	R	nRM SD	Fit	Fit	
Other subregio nal datasets	Curuai- Model	1994- 2015	0.8 2	12%	0.8 6	-	-	-	-	-	-	-	-	-	-	-	-	
	Janauacá -Bonnet	2006- 2019	-	-	-	-	0.7 5	25 %	0.49	-	-	-	-	-	-	-	-	
	Janauacá -Pinel	2006- 2015	-	-	-	-	0.5 7	17 %	0.82	-	-	-	-	-	-	-	-	
	Llanos de Moxos - ALOS	2006- 2010	-	-	-	-	-	-	-	-	-	-	-	0.5 2	99%	0.3 3	-	
Multiple datasets at coarse resolutio n	GIEMS- 2	1992- 2015	0.9 6	21%	-	-	0.7 8	15 7%	-	-	0.8 8	68%	-	0.9 1	85%	-	-	
	SWAMP S	2000- 2020	0.9 1	2%	-	-	0.8	38 %	-	-	0.5 2	74%	-	0.9 2	171%	-	-	
	WAD2 M	2000- 2018	0.9	82%	-	-	0.7 9	63 %	-	-	0.4 6	2%	-	0.9	123%	-	-	
Multiple datasets at high	GIEMS- D3	1993- 2007	-	-	0.9 2	0.61	-	-	0.80	0.81	-	-	0.1 4	-	-	0.4 4	0.45	
	CIFOR	2011	-	-	0.9 1	0.39	-	-	0.24	0.33	-	-	0.5 5	-	-	0.3 0	0.69	

resolutio n	ESA- CCI	1992- 2015	-	-	0.7 6	0.40	-	-	0.40	0.70	-	-	0.3 6	-	-	0.1 4	0.69
	GIEMS- D15	1993- 2004	-	-	0.9 2	0.58	-	-	0.68	0.59	-	-	0.5 1	-	-	0.3 8	0.46
	GLWD	1992- 2004	-	-	0.8 8	0.45	-	-	0.79	0.93	-	-	0.6 3	-	-	0.0 8	0.51
	SWAF- HR	2010- 2019	-	-	0.9 5	0.64	-	-	0.63	0.71	0.6 6	73%	0.2 2	0.7 5	213%	0.3 9	0.57
Hydro- logical models	THMB	1961- 2013	0.7 2	62%	-	-	0.7 3	73 %	-	-	-	-	-	0.5 4	7%	-	-
	CaMa- Flood	1980- 2014	0.8 0	11%	0.9 7	0.73	0.6 8	11 1%	0.88	0.83	-	-	0.4 9	0.8 2	218%	0.2 8	0.58
	MGB	1980- 2014	0.8 3	7%	0.9 6	0.58	0.6 4	29 3%	0.82	0.93	-	-	0.5 2	0.9 1	26%	0.1 9	0.52
Optical sensors	G3WB M	1990- 2010	-	-	0.6 4	0.29	-	-	0.19	0.14	-	-	0.0 3	-	-	0.0 4	0.59
	GLAD	1999- 2018	-	-	0.8 4	0.39	-	-	0.30	0.20	-	-	0.0 4	-	-	0.1 6	0.78
	GSWO	1984- 2019	-	-	0.7 5	0.31	-	-	0.21	0.17	-	-	0.0 4	-	-	0.0 9	0.68
SAR	Hess	1995- 1996	-	-	0.9 6	0.47	-	-	0.28	0.98	-	-	0.4 8	-	-	0.4 7	0.69
	Chapma n	2006- 2011	-	-	0.6 5	0.27	-	-	0.22	0.68	-	-	0.2 8	-	-	0.2 4	0.50
	Rosenqv ist	2014- 2018	-	-	0.5 9	0.34	-	-	0.59	0.98	-	-	0.6 4	-	-	0.1 9	0.48

1758

1759 Table S4. Long-term minimum inundation areas (km²) for 11 wetland complexes (up to three datasets per complex)
1760 and the 18 basin-scale datasets. The local-scale values refer to the following datasets, in this order (comma-separated
1761 values relate to areas with more than one dataset available) : Curuai - ALOS (Arnesen et al., 2013) and LISFLOOD-
1762 FP model (Rudorff et al., 2014); Uatumã - ALOS (Resende et al., 2019); Janauacá - ALOS (Pinel et al., 2019),
1763 hydrologic model (Bonnet et al., 2017) and TELEMATAC-2D model (Pinel et al., 2019); Mamirauá - ALOS (Ferreira-
1764 Ferreira et al., 2015); Pacaya-Samiria - ALOS-2 PALSAR-2 (Jensen et al., 2020); Llanos de Moxos - MODIS (Ovando
1765 et al., 2016) and ALOS (Ovando et al., 2016); and lower Amazon - MODIS (Park et al., 2019). Average, standard
1766 deviation (S.D.) and coefficient of variation (CV) are presented for each area in the last row.

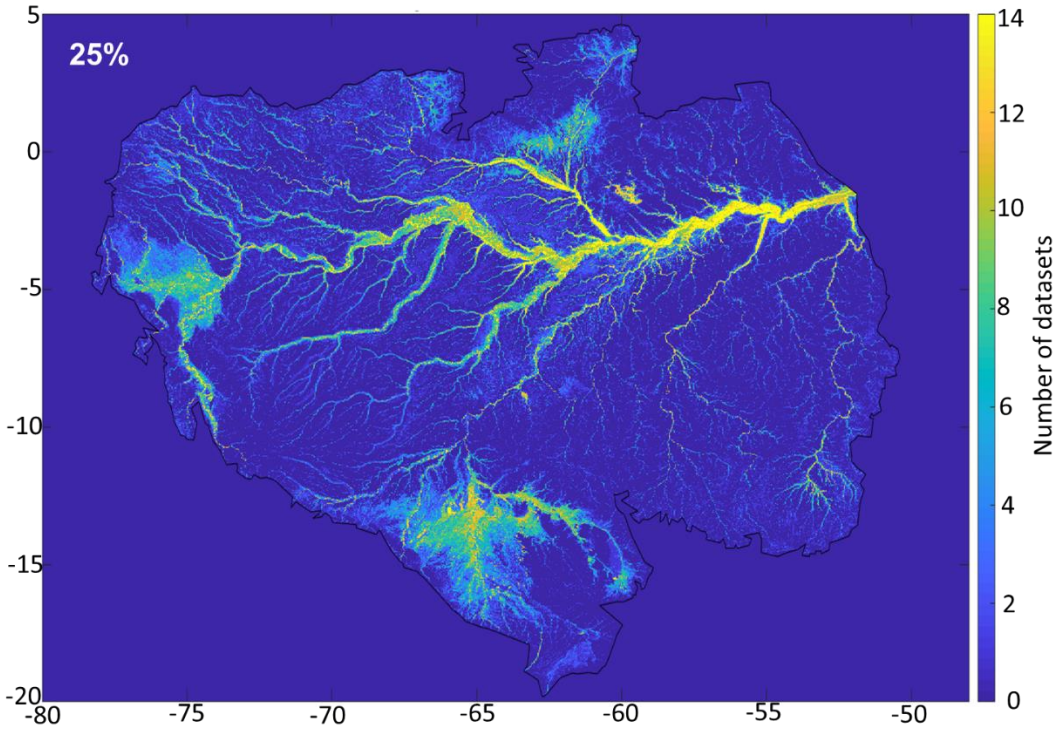
	Datas et	Curuai	Uatumã	Janauacá	Mamirauá	Pacaya-Samiria	Llanos de Moxos	Lower Amazon	Amazon mainstem	Purus	Roraima savannas	Negro savannas
	Local	1690, 1278	-	108, 38, 18	715	3824	1014, 3962	17797				
Multiple datasets at coarse resolution	GIEM S-2	995	263	183	1117	1578	500	19717	26807	349	0	0
	SWA MPS	2840	479	197	790	4433	24622	38345	53256	3492	309	6375
	WAD 2M	403	97	97	633	20421	31713	14728	29932	4240	258	10443
Multiple datasets at high resolution	GIEM S-D3	2712	861	151	1115	2731	8375	33253	44853	2696	383	146
	CIFOR	-	-	-	-	-	-	-	-	-	-	-
	ESA-CCI	-	-	-	-	-	-	-	-	-	-	-
	GIEM S-D15	3942	1265	116	1077	3409	15074	44277	59066	3401	2966	2622
	GLWD	-	-	-	-	-	-	-	-	-	-	-
	SWA F-HR	1502	544	69	469	215	8304	20944	30242	784	0	3
Hydrological model	THMB	487	38	1	266	5349	7172	6708	18099	5596	383	195
	CaMa - Flood	2741	861	184	1135	8269	17776	31569	45848	4128	1001	672
	MGB	3005	212	0	587	6101	4508	21333	32073	1769	226	35
Optical sensors	G3W BM	-	-	-	-	-	-	-	-	-	-	-
	GLAD	474	77	8	288	514	1513	6243	9857	335	13	20
	GSWO	736	345	10	314	401	2934	11908	16428	735	117	2
Synthetic Aperture Radar	Hess	2770	584	106	1756	32107	56337	28981	54493	7061	1217	6084
	Chapman	1894	385	68	866	6775	10090	18413	28539	2951	1025	2843
	Rosenqvist	1514	313	49	422	1077	4566	13413	19512	575	60	5
	Average	1858	452	89	774	6670	13820	22131	33500	2722	568	2103
	S.D.	1148	350	71	430	8978	15190	11637	15551	2094	801	3285
	CV	0.62	0.77	0.80	0.56	1.35	1.10	0.53	0.46	0.77	1.41	1.56

1768

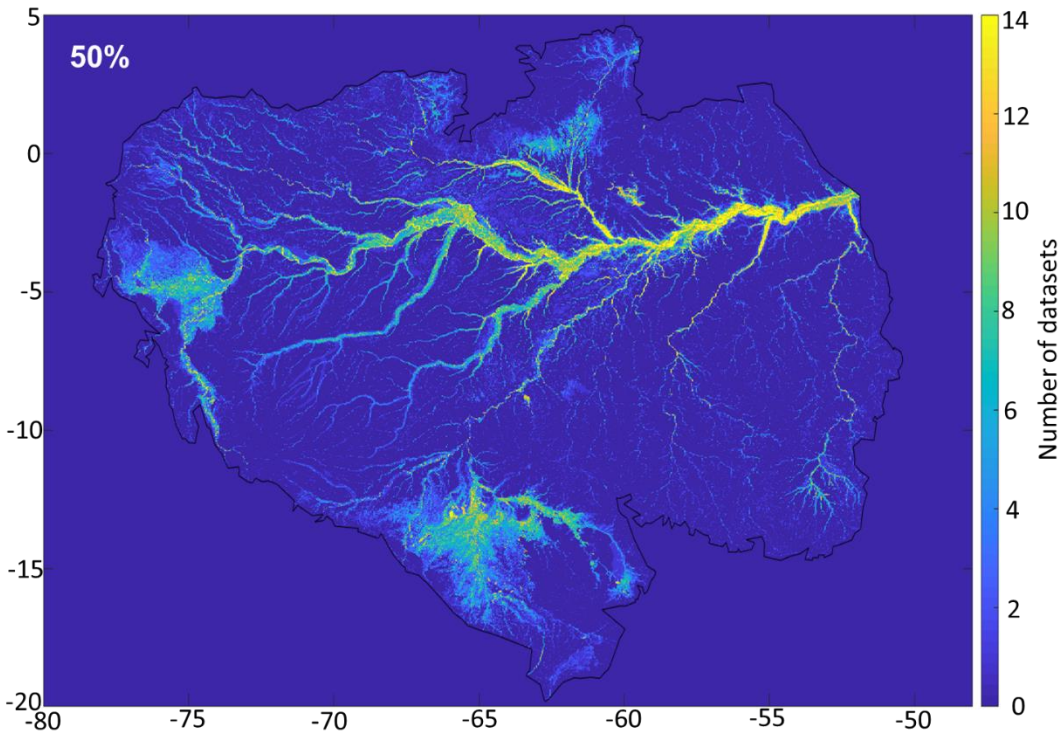
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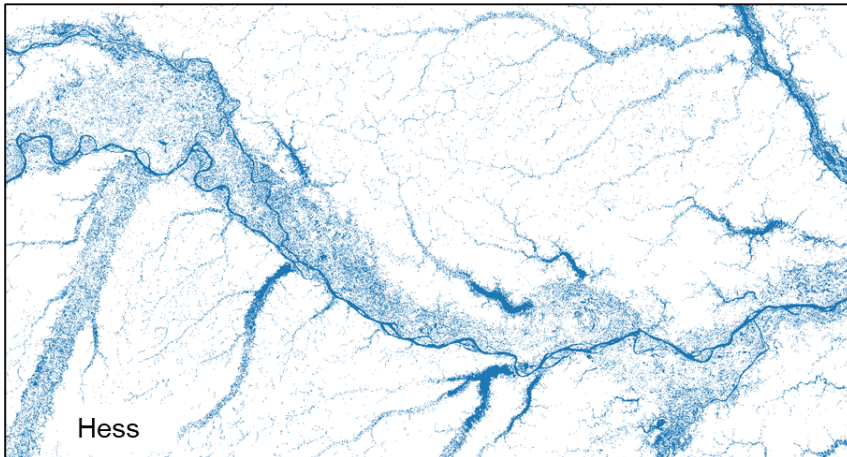
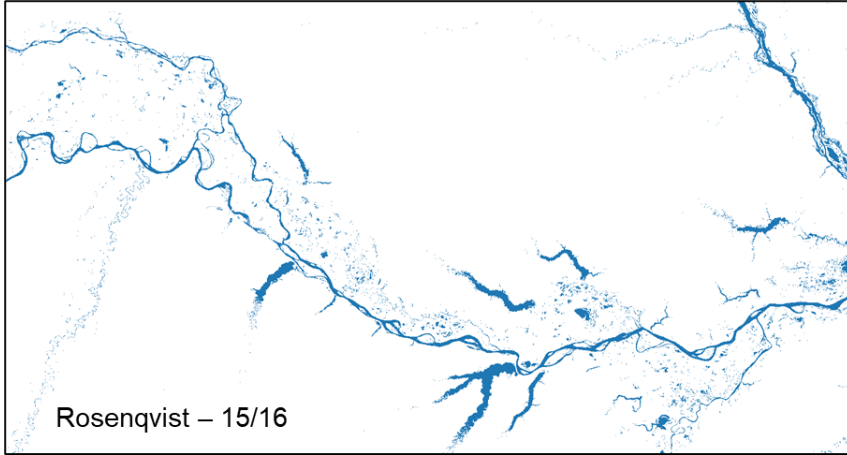


1773

1774 Figure S1. Sensitivity of the fraction used to define a flooded 1km pixel (25% and 50%).

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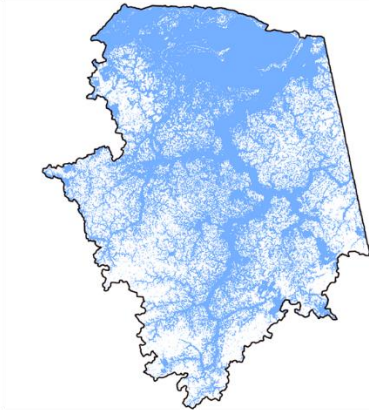
1777

1778 Figure S2. Minimum inundation extent for the central Amazon River, as estimated by the

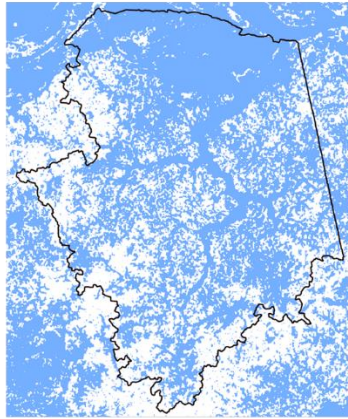
1779 Rosenqvist (years 2015-2016) and Hess (1995) datasets.

1780

(A) Janauacá



Pinel et al. (2019) – 30 m

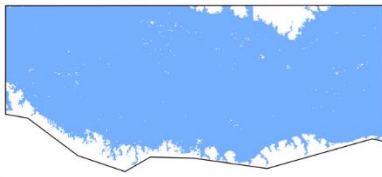


Rosenqvist et al. (2020) – 50 m

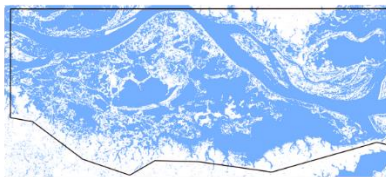


Hess et al. (2015) – 90 m

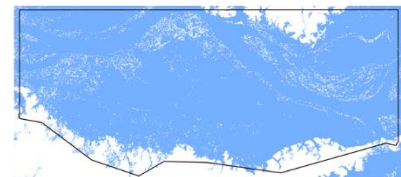
(B) Curuai



Arnesen et al. (2013) – 12.5 m



Rosenqvist et al. (2020) – 50 m



Hess et al. (2015) – 90 m

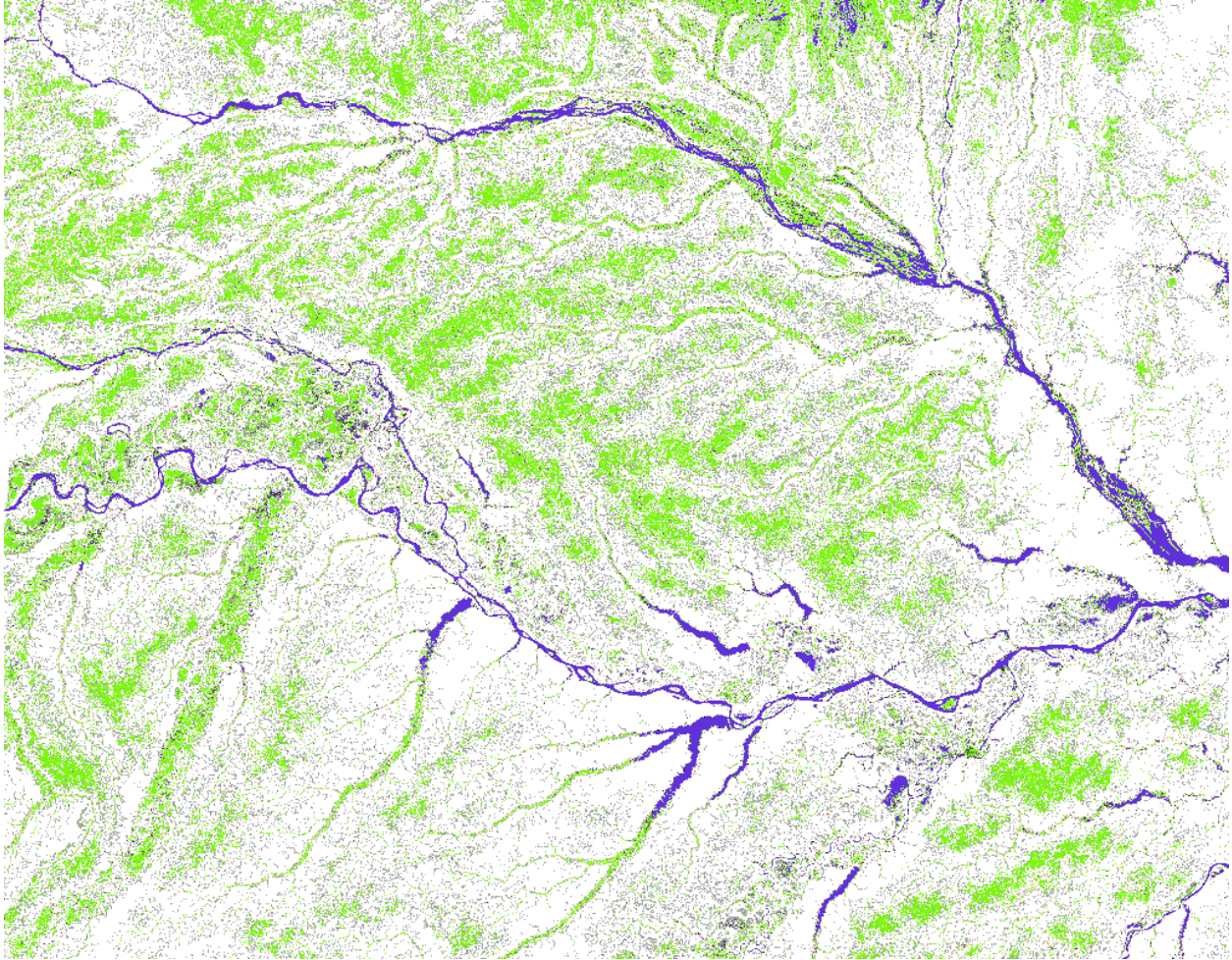
1781

1782 Figure S3. Comparison between the long-term maximum inundation for subregional validation

1783 locations (Pinel and Arnesen datasets) as well as the Rosenqvist and Hess datasets for the (a)

1784 Janauacá and (b) Curuai areas. The polygons refer to the area used to extract the values presented

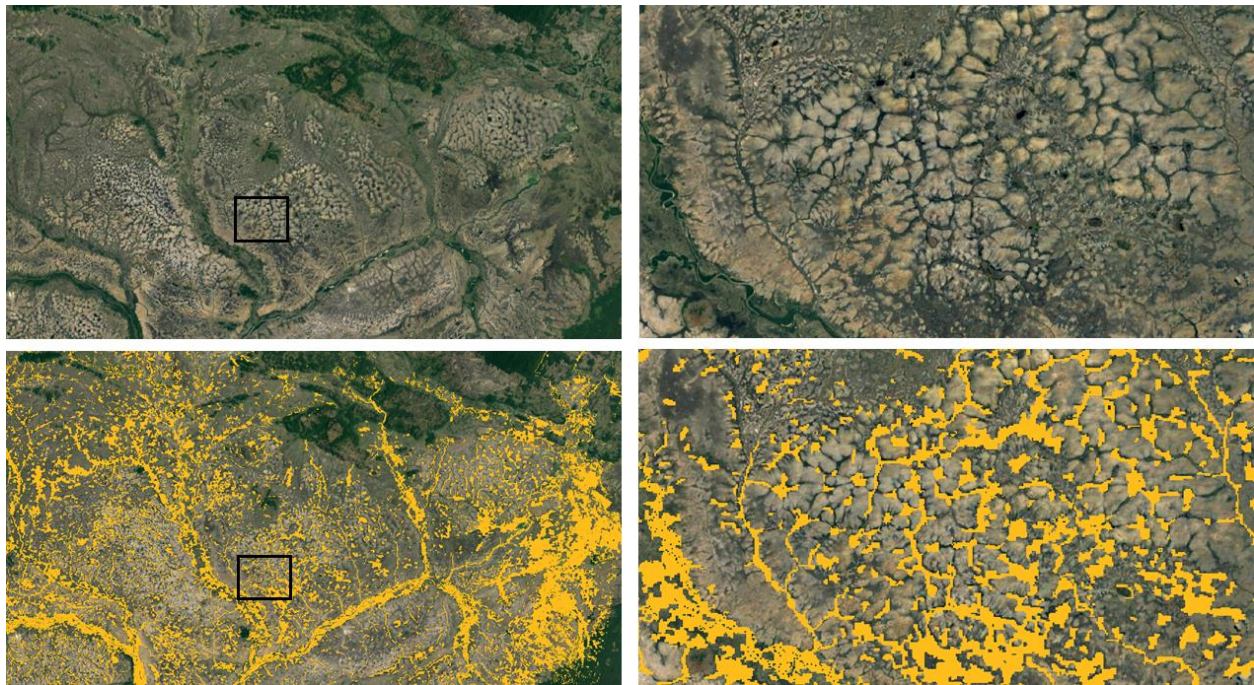
1785 in Tables 3, S3 and S4. The spatial resolution of each dataset is noted.



1786

1787 Figure S4. Estimation of wetland areas by Gumbrecht et al. (2017) across the central Amazon River
1788 basin. Green pixels relate to the “swamps (incl. bogs)” category, which is defined as “Wet all year
1789 around, but not necessarily inundated.”

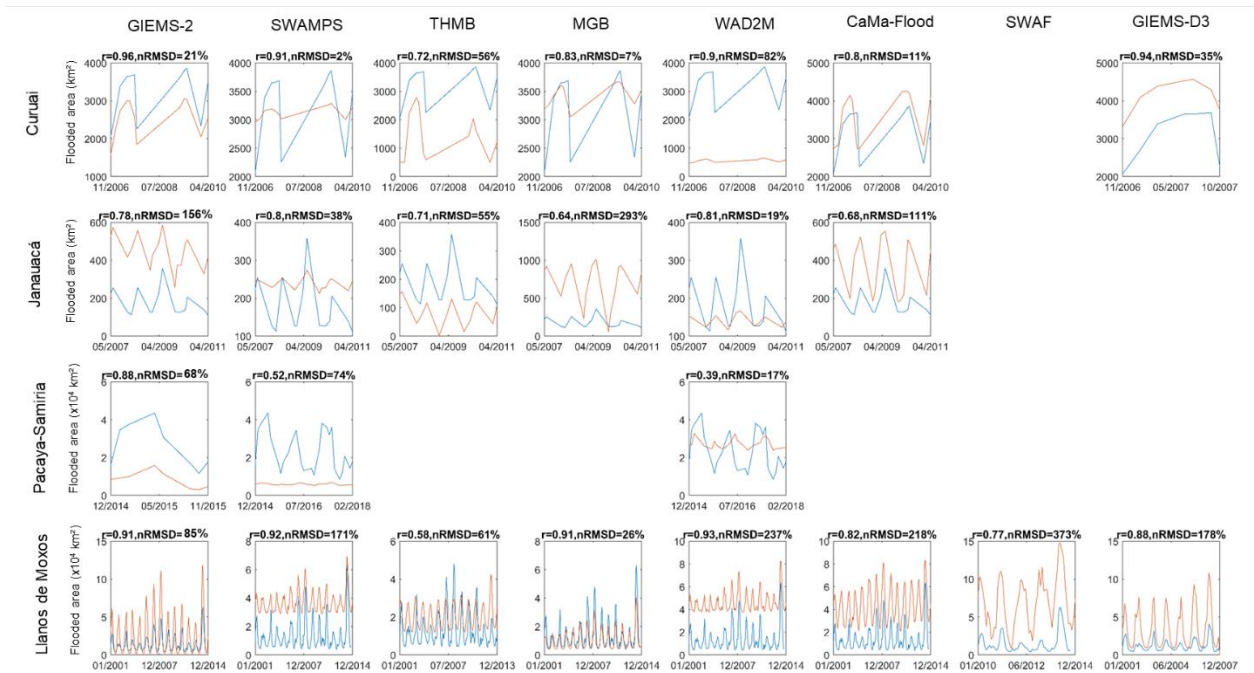
1790



1791

1792 Fig S5. Roraima wetlands. Above: Google Earth imagery. Below: Hess SAR classification of
 1793 floodable areas (at large scale in the left, and detailed scale in the right), displayed as orange areas.

1794



1795

1796 Fig S6. Inundation time series for the four wetlands with available datasets, and for the eight basin-
1797 scale dynamic datasets (GIEMS-2, SWAMPS, THMB, MGB, WAD2M, CaMa-Flood, SWAF-HR
1798 and GIEMS-D3). The subplots that are empty refer to areas where the basin-scale dataset time
1799 spans did not overlap with the subregional dataset ones. The subregional dataset is displayed in
1800 blue, and each of the basin-scale datasets in red.