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1 **Faeces traits as unifying predictors of detritivore effects on organic matter turnover**

2

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20

## 21 **Abstract**

22 In the last decade, our understanding of plant litter decomposition and soil organic matter  
23 formation substantially improved but critical blind spots remain. Particularly, the role of  
24 detritivores, i.e. soil animals that feed on litter and soil, is poorly understood and notoriously  
25 missing from biogeochemical models. This major gap results from methodological difficulties to  
26 isolate their effect and from the astonishing diversity of detritivorous organisms with few  
27 common features, thereby hampering the identification of general patterns. In this viewpoint, we  
28 propose that the characteristics of their faeces can predict detritivore effects on soil processes  
29 related to organic matter turnover across the large detritivore diversity. Indeed, faeces are  
30 common to all detritivores, and a large part of organic matter is transformed into faeces in many  
31 ecosystems. Two recent studies presented here showed that faeces characteristics are powerful  
32 predictors of the fate and turnover of this transformed organic matter. We suggest that faeces  
33 characteristics, such as water-holding capacity, size and spatial organisation of the faecal pellets  
34 and of their constituting particles, particulate organic matter connectivity, as well as the  
35 characteristics of dissolved organic matter in faecal pellets, are promising ‘effect traits’. By  
36 focusing on similar features rather than differences, this approach has the potential to break  
37 down barriers of this highly fragmented soil animal group, in particular between earthworms that  
38 are often studied as ecosystem engineers and classical litter transformers such as millipedes,  
39 woodlice, or snails. We discuss ways of tackling the complexity of using such traits, particularly  
40 regarding the composite determinism of faeces characteristics that are driven both by the  
41 detritivore identity and the ingested organic matter. Rigorous and hypothesis-based use of faeces  
42 characteristics as effect traits, including clear identification of studied processes, could allow  
43 integrating detritivores in our current understanding of organic matter turnover.

44 **Key-words**

45 Macroarthropods ; Soil functioning ; Soil invertebrates ; Soil processes ; Trait-based approaches



## 46 **1. Introduction**

47           Plant litter decomposition and the subsequent formation of soil organic matter (SOM) are  
48 key ecosystem processes that control biogeochemical cycling and the ability of soils to store  
49 large amounts of carbon (Lehmann and Kleber, 2015). In the last decade, a new understanding of  
50 litter decomposition and SOM formation has emerged (Basile-Doelsch et al., 2020; Dignac et al.,  
51 2017; Schmidt et al., 2011), through (i) a renewed characterization of the chemical nature and  
52 protection mechanisms of SOM (Kögel-Knabner and Rumpel, 2018; Lehman and Kleber, 2015),  
53 (ii) a growing recognition that interactions between plant litter, microbial communities and  
54 minerals rather than litter recalcitrance control SOM formation (Cotrufo et al., 2013, 2015;  
55 Dynarski et al., 2020), and (iii) the recognition of the important role roots play in SOM  
56 formation (Adamczyk et al., 2019, Clemmensen et al., 2013; Rasse et al., 2005; Sokol et al.,  
57 2019a). In contrast, while the importance of soil invertebrates in soil processes is often  
58 acknowledged (Briones, 2018; Griffith et al., 2021), our understanding of their roles in SOM  
59 dynamics is still poor (Filser et al., 2016; Prescott and Vesterdal, 2021). Detritivores in  
60 particular, i.e., soil invertebrates that feed on dead organic matter, importantly contribute to  
61 organic matter turnover (Prescott and Vesterdal, 2021). They do so by ingesting large amounts of  
62 organic matter, assimilating a part of it and rejecting the main part as faeces (David, 2014). This  
63 processing greatly affects the organic matter physicochemical characteristics (e.g., Coulis et al.,  
64 2009, 2016; Hedde et al., 2005; Joly et al., 2018; Le Mer et al., 2020; Vidal et al., 2016) and its  
65 contribution to SOM formation (Angst et al., 2019, Vidal et al., 2019). Despite clear evidence  
66 that in many ecosystems detritivores process large amounts of organic matter, we lack a general  
67 understanding of their role in its turnover.

68           One of the main obstacles to understanding the detritivores' influence on organic matter  
69 turnover is the difficulty to isolate these effects experimentally. Traditionally, the role of soil  
70 invertebrates in decomposition processes has been studied using litterbags of different mesh sizes  
71 (0.1 mm, 2 mm, 4 to 8 mm), sequentially excluding soil invertebrates based on their body width  
72 (e.g., Handa et al., 2014; Wall et al., 2008). A meta-analysis of such studies reported that micro-  
73 and mesofauna (body width < 2 mm) presence increased litter mass loss by 37% on average  
74 across biomes (Garcia-Palacios et al., 2013). This figure emphasises the importance of soil  
75 invertebrates in decomposition, but has several limitations. The focus on body width means that  
76 the measured effect includes not only the effect of soil invertebrates feeding on plant litter  
77 (detritivores), but also of other functional groups with potential top-down effects such as  
78 microbivores and predators (Koltz et al., 2018; Lenoir et al., 2007). Moreover, the large mesh  
79 sizes used for treatments allowing faunal access entail that the litter consumed by detritivores but  
80 returned to soil as faeces is not retrieved in litterbags and considered as lost mass. The  
81 decomposition of these faeces and their contribution to SOM formation is a major unknown  
82 (Prescott, 2010). Studies on decomposition in reconstructed detritivore communities in  
83 microcosms (e.g., Hattenschwiler and Gasser, 2005; Joly et al., 2021; Vidal et al., 2019), or on  
84 the detritivore faeces fate (e.g., Coulis et al. 2016; Decaëns, 2000; Joly et al., 2020) contributed  
85 to overcoming the limitations of the litterbag technique. Yet, the complexity of such studies  
86 limited the number of detritivore species considered and thus the identification of general  
87 patterns across the diversity of detritivores.

88           The extreme diversity of detritivores is the other dominant obstacle towards identifying  
89 general principles of detritivore effects on organic matter turnover. Detritivores include  
90 millipedes, woodlice, earthworms, snails, and insect larvae, which greatly differ in their

91 morphologies, behaviours, and the ways they process organic matter. Historically, the role of  
92 these animals on soil processes was studied considering separate broad functional groups and  
93 subgroups. Specialists of millipedes (e.g., David and Gillon, 2002), woodlice (Zimmer, 2002), or  
94 snails (Astor et al., 2015) often studied the role of a few species on litter decomposition  
95 separately. In addition, since earthworms also modify their environment through their burrowing  
96 activities, most earthworm studies focussed on their global role as ‘ecosystem engineers’  
97 (Lavelle and Spain, 2001; Wardle, 2002) rather than ‘detritivores’. Clearly, the separate study of  
98 the various groups of detritivores have limited the identification of general principles of  
99 detritivore effects on soil processes. Trait-based approaches allow moving beyond broad  
100 categorical characterisation of organisms based on their assumed differences in ecosystem  
101 function, to more precise continuous characterisation based on characteristics that relate to their  
102 differences in ecosystem function, known as *effect traits* (Garnier et al., 2016; Violle et al.,  
103 2007). In plant studies, effect traits have proved very useful for upscaling from organisms to  
104 ecosystems, whether aboveground traits (Lavorel and Garnier, 2002; Violle et al. 2007) or  
105 belowground ones (Freschet et al., 2021), particularly as predictors of litter decomposition  
106 (Cornwell et al., 2008; Rosenfield et al., 2020). For example, plant litter with high specific leaf  
107 area and leaf nitrogen and phosphorus concentrations generally decomposes rapidly, while high  
108 dry matter content and tannin concentrations of leaves are associated with slow decomposition  
109 (Cortez et al., 2007; de la Riva et al., 2019; Kazakou et al., 2006). To better integrate detritivores  
110 into the current framework of SOM dynamics, we need to identify effect traits that link the  
111 activity of these animals to processes controlling SOM dynamics, but this challenge lags behind.  
112 In the European invertebrate trait database BETSI (<https://portail.betsi.cnrs.fr/>, Pey et al., 2014),  
113 out of 76 traits recorded, only 11 can be considered as effect traits according to Brousseau et al.

114 (2018), and only two of these have direct links to ecosystem function (i.e. burrowing strategy and  
115 feeding traits). This calls for a common effort to identify effect traits of detritivores relevant to  
116 organic matter turnover that enable meaningful comparisons amongst taxa. Such traits should (i)  
117 be measurable on all kinds of detritivores and (ii) have a demonstrated link to the studied  
118 function. This point is crucial as current applications of trait-based approaches often lack such a  
119 clear link (Brousseau et al., 2018; Shipley et al., 2016).

120         In this viewpoint paper, we argue that detritivore faeces are a promising yet overlooked  
121 part of their phenotype, which characteristics, measurable on all soil fauna, can predict their  
122 effect on key soil processes related to organic matter turnover. First, we show that detritivore  
123 faeces represent important by-products of detritivore activity and that their characteristics are  
124 directly related to organic matter turnover. Then, with two selected recent case studies, focusing  
125 on litter-feeding and soil-feeding detritivores respectively, we show that characteristics of faeces  
126 can predict their fate, and thus predict the effect of these detritivores species on litter  
127 decomposition and SOM formation. We thus advocate for the consideration of faeces  
128 characteristics as detritivore effect traits. Such traits could be powerful unifying traits across the  
129 large diversity of detritivores that otherwise share few common features with little link to  
130 ecosystem function.

131

## 132 **2. Faeces as key by-products of detritivory**

133         Detritivores are soil animals that feed on dead organic matter, either on leaf litter  
134 (arthropods, snails, epigeic earthworms), on soil and root litter (endogeic earthworms), or both  
135 (anecic earthworms). Since these food sources are rather nutrient-depleted and hard to digest  
136 (Sturner and Elser, 2002), detritivores typically have low assimilation efficiencies and high

137 consumption rates (Crossley et al., 1971; Curry and Schmidt 2007; David, 2014). Thus, they  
138 ingest a lot of dead organic matter, assimilate a small part of it, and egest most of it to soils as  
139 faeces (Fig. 1). Studies from temperate (Schaefer et al., 1990), Mediterranean (David and Gillon,  
140 2002), arid (Sagi et al., 2019), and tropical ecosystems (Dangerfield and Milner, 1996) estimated  
141 that in these ecosystems, 40-50% of the annual litterfall is consumed by detritivores and returned  
142 to soils as faeces. In ecosystems where detritivores are abundant, these faeces thus represent a  
143 substantial part of the soil profile, e.g., in temperate (Zanella, 2018) or tropical ecosystems  
144 (Bottinelli et al., 2021). Undeniably, in many ecosystems, large quantities of organic matter  
145 originating from litter are decomposed and stabilised only after conversion into faeces.  
146 Determining the physicochemical characteristics of these faeces and how they affect their fate is  
147 thus critical to understand detritivore effects on organic matter turnover (Prescott and Vesterdal,  
148 2021).

149         The conversion of plant litter and/or soil into detritivore faeces leads to profound  
150 physicochemical changes that can affect the fate of organic matter in soils. For detritivores that  
151 preferentially feed on plant litter (e.g. millipedes, woodlice, snails), faeces have higher  
152 concentrations of dissolved organic carbon and nitrogen, water-holding capacity and surface area  
153 available for microbial colonisation, and lower C:N ratio and tannins content compared to the  
154 plant litter from which they are derived (Coulis et al., 2009, 2016; Ganault et al., 2022; Joly et  
155 al., 2018, 2020). Because these physicochemical characteristics are known to influence  
156 decomposition rates (Makkonen et al., 2012) and the contribution of litter to SOM formation  
157 (Cotrufo et al., 2013), their changes during gut passage are likely to drive the fate of the egested  
158 organic matter. In fact, faeces of detritivores such as millipedes typically decompose faster than  
159 the litter from which they are derived - an acceleration previously linked to the higher lability of

160 the faeces compared to the ingested litter (Coulis et al., 2016; Joly et al., 2018). Similar to the  
161 physicochemical characteristics of leaf litter, those of faeces from distinct detritivore species  
162 could thus predict their fate.

163 For detritivores such as endogeic earthworms that feed on mineral soil, ingestion of soil  
164 and its incorporation into earthworm faeces (known as ‘casts’) also lead to major changes in soil  
165 physicochemical characteristics. Compared to bulk soil, the faeces are richer in organic carbon,  
166 total and mineral nitrogen, total and available phosphorus, and exhibit higher cation-exchange  
167 capacity, base saturation and pH (van Groenigen et al., 2019). Similar to litter, these  
168 characteristics are known to relate to further microbial degradation and organic matter  
169 decomposition (Jouquet et al., 2008), so their changes following gut passage can affect the  
170 formation and stabilisation of SOM (Clause et al., 2014). For instance, increased soil compaction  
171 and reduced pore size distribution that allow air and water circulation can limit the accessibility  
172 of microbial communities to organic matter, and thus physically protect SOM (Angst et al.,  
173 2017). Recently, Barthod et al. (2020, 2021) reported that faeces produced by *Eisenia* sp. fed  
174 with different clay minerals have a contrasting composition, which in turn differently affected  
175 the microbial decomposition of organic matter occluded in these faeces incubated in the soil.  
176 This demonstrates a clear link between earthworm faeces characteristics and their fate.

177 Generally, there is thus growing evidence that detritivore faeces are important  
178 decomposition by-products and that their characteristics can be linked to their fate in soils. This  
179 suggests that faeces characteristics of different detritivore species could predict the species-  
180 specific effect on organic matter turnover. Recently, two studies, each focusing on multiple  
181 detritivore species, used this approach to predict their effects on organic matter turnover.

182

### 183 **3. Case studies using faeces traits to predict detritivore effects on organic** 184 **matter turnover**

#### 185 *3.1 Case study 1: Detritivore faeces traits as predictors of organic matter turnover*

186         The potential of faeces traits as predictors of organic matter turnover was recently  
187 illustrated in a study on the role of detritivores on litter decomposition (Joly et al., 2020). The  
188 authors explored how detritivores affect litter decomposition, by converting litter into faeces, and  
189 how this effect varies across six phylogenetically-diverse invertebrates species. To do so, they  
190 collected faeces from six detritivore species (three millipede, two woodlouse and one snail  
191 species) feeding on litter of six tree species, separately, resulting in 36 faeces types (Fig. 2).  
192 Then, they measured physicochemical characteristics on the 36 faeces types and on the six intact  
193 litter types as controls. They then placed all substrates to decompose on top of soil to study the  
194 detritivore effect on organic matter turnover. Faeces varied in colour depending on the nature of  
195 ingested litter, and in shape depending on detritivore identity (Fig. 2), whereas their  
196 physicochemical characteristics (e.g., elemental composition, surface area, water-holding  
197 capacity) were driven both by the nature of the ingested litter and the animal identity.  
198 Importantly, these faeces traits were tightly correlated with faeces decomposition. Indeed, faeces  
199 C and N losses correlated with faeces concentration in dissolved organic carbon and total  
200 dissolved nitrogen, respectively. This shows that faeces traits may be predictors of organic  
201 matter turnover across detritivore species as different as millipedes and snails, suggesting that  
202 extending trait measures to detritivore faeces may allow predicting their effects on soil processes.

203         Another major finding of this study was that the detritivore effect – that is, the difference  
204 in organic matter quality or element cycling rate between faeces and intact litter – depended on  
205 the ingested litter species, with larger positive effects for low-quality and slow-cycling litter, and

206 small or negative effects for high-quality and fast cycling litter (Fig. 3). This general pattern was  
207 consistent across detritivore species, suggesting that diverse detritivores play a similar role in  
208 organic matter turnover. Yet, the magnitude of the effect, and its relationship with the intact litter  
209 characteristics were detritivore species-specific. The parameters of the relationship between litter  
210 quality/cycling and the change in quality/cycling following litter conversion into faeces, could  
211 thus be used as powerful effect traits. The intercept describes the extent to which a given  
212 detritivore species increases organic matter quality/cycling. The slope, in turn, describes the  
213 extent to which the effect of this detritivore species varies depending on the initial  
214 quality/cycling rate of the ingested litter.

### 215 *3.2 Case study 2: Microstructural organisation of earthworm faeces as predictor of earthworm* 216 *effect on organic matter turnover*

217 The potential use of earthworm faeces properties as predictors of organic matter turnover  
218 was also recently investigated for six earthworm species (Le Mer et al., 2022). In this study, the  
219 authors explored how earthworms affect SOM stability by occluding fresh organic matter within  
220 their faeces, and how this effect varies between different earthworm species. To do so, they  
221 collected six earthworm species, from three ecological categories (epigeic, anecic, and  
222 endogeic), fed the earthworms with the same organic matter and soil and collected the resulting  
223 six faeces types. They then incubated each faeces type individually under optimal conditions and  
224 measured CO<sub>2</sub> respiration rates after 7, 42 and 140 days of incubation as indicators of SOM  
225 stability. Finally, they measured the characteristics and physical organisation of the six faeces  
226 types and the control soil without earthworm activity. To characterise the SOM occluded by  
227 earthworms in their faeces, the authors measured several faeces traits such as organic C content  
228 and organic matter stability by Rock-Eval 6 analysis. Moreover, thanks to x-ray



229 microtomography and image analyses, the spatial organisation between pore and POM structures  
230 at micro-scale (9.5  $\mu\text{m}$ ) was also characterised. For each faeces sample, the authors computed the  
231 (i) pore and (ii) POM volumes, as well as the (iii) pore subvolumes directly connected to the air  
232 outside the faeces and the (iv) POM subvolumes connected, directly or indirectly (through the  
233 connected pores), to the outside of the faeces. The contribution to the faeces volume (%) and the  
234 mean volume of individualised pores and organic matter fragments ( $\text{mm}^3$ ) was computed for  
235 each of these faeces pores and POM compartments (total, connected and unconnected ones).

236 Despite deriving from the same soil and same plant litter, the physicochemical  
237 characteristics of fresh faeces, such as elemental content and physical cast organisation (content  
238 of particulate organic matter and pores) varied amongst earthworm species (Fig. 4a and 4b).  
239 SOM stability in faeces depended on the identity of earthworms that produced the faeces, and at  
240 least half of the variation in respiration rates amongst faeces of different earthworm species  
241 could be explained by species-specific variations of the microstructural traits of faeces (Fig. 4).

242 One of the major findings of this study was that, regardless of the earthworm species or  
243 the stage of faeces decomposition considered, a substantial part of the variability in faeces  
244 mineralisation rates observed could be explained by the physical organisation of these faeces.  
245 These included volume contribution of POM, and especially its connection with the  
246 microporosity, which possibly favoured the accessibility of SOM to microorganisms (Fig. 4c).  
247 This study therefore suggests that earthworm faeces traits can ultimately contribute to  
248 determining the effect of different earthworm species on SOM dynamics.

249

## 250 **4. Discussion, challenges and perspectives**

251           The two case studies highlight the pertinence of using faeces characteristics to predict the  
252 effects of the myriad of detritivore species on soil processes. Despite major differences in  
253 morphology, and feeding and behavioural habits, diverse detritivore species share faeces as a  
254 common feature. We believe that measuring characteristics on detritivore faeces is a promising  
255 research avenue to unify research areas so far compartmentalised into subgroups of soil fauna  
256 (Fig. 5). In the following sections, we discuss key aspects related to the use of faeces  
257 characteristics as predictors of soil processes, including potential difficulties and precautions,  
258 research directions and integration within current frameworks.

259

### 260 *4.1. Which faeces traits for which soil processes?*

261           A pertinent use of effect traits requires (i) a clear identification of the process of interest  
262 and (ii) the formulation of clear hypotheses on the link between the measured traits and the  
263 process of interest. In a literature review, Brousseau et al. (2018) identified a detrimental lack of  
264 such clarification in 39% of the reviewed studies on arthropod effect traits. This is especially  
265 important for organic matter turnover, which results from multiple processes including leaching  
266 of water-soluble compounds, enzymatic degradation by microorganisms, physical and physico-  
267 chemical protection, that can all contribute to the stabilisation and destabilisation of SOM (Fig.  
268 5, right panel). Because of the strong control of physicochemical characteristics on these  
269 processes, detritivore faeces traits may be linked to organic matter turnover through their effect  
270 on specific processes, but to varying degrees depending on the process considered and the  
271 temporal scale. In the two aforementioned case studies, such links between processes and faeces  
272 traits were hypothesised. For example, Joly et al. (2020) hypothesised that the link between

273 concentrations of DOC in faeces and faeces C loss over time was due to an increased leaching of  
274 water-soluble compounds following litter conversion into detritivore faeces, which would  
275 facilitate decomposition and increase the amount of organic matter transferred to the underlying  
276 soil. Similarly, Le Mer et al. (2022) hypothesised that increasing volume of POM in earthworm  
277 faeces, connected with the pore space presenting an uninterrupted path to the edge of the cast,  
278 facilitates microbial activity and thus SOM mineralisation. In the long term, however, it remains  
279 unknown if faeces traits related to organic matter C loss (Joly et al., 2020) or CO<sub>2</sub> emissions (Le  
280 Mer et al., 2022) translate into changes in the persistence of SOM. Because both leaching and  
281 microbial activity can favour the production of microbial biomass and thus necromass, and  
282 ultimately affect the formation of mineral-associated organic matter (Sokol et al. 2019b), faeces  
283 DOC concentrations or POM connectivity may predict the contribution of faeces to SOM  
284 formation. However, once the easily degradable compounds are leached or used by  
285 microorganisms, the remaining fragments that compose the faeces may contribute to the  
286 formation of a partly decomposed POM pool, which is not necessarily subject to stabilisation  
287 processes. The formation of this kind of POM may be linked to different faeces traits, such as the  
288 average faeces particle size (Joly et al., 2020) or its location within the pore structure of the  
289 faeces (Le Mer et al., 2022). The feeding of detritivores on faeces (known as coprophagy), either  
290 on their own (e.g., Kautz et al., 2002), or that of other species (e.g., Bonkowski et al., 1998),  
291 could further affect the fate of organic matter. Faeces characteristics determining their  
292 palatability to detritivores may thus also be considered as faeces traits.

293         Future use of detritivore faeces traits should thus carefully consider the mechanistic links  
294 between the traits and soil processes/ parameters considered, and the timescale at which the traits  
295 are relevant as predictor of soil processes. The study of faeces characteristics is still in its infancy

296 and characteristics not yet considered may prove useful in the future. As starting points, we  
297 recommend that future studies should consider physical traits such as water-holding capacity,  
298 faecal pellets specific area and density, faeces particle size, pore structure, and POM  
299 connectivity, as well as chemical characteristics such as elemental composition and DOC or  
300 TDN concentrations, as predictor of faeces decomposition, mineralisation, and contribution to  
301 SOM formation. These faeces characteristics could also possibly explain detritivore-species  
302 specific effects on aggregate size distribution and stability. Such characteristics are easily  
303 measurable and on relatively small amounts of faeces (see. Joly et al., 2018, 2020, Le Mer et al.,  
304 2022). We encourage future studies to explore relationships between faeces traits and soil  
305 processes in order to build a conceptual framework linking detritivores and organic matter  
306 turnover.

307

#### 308 *4.2. Bridging research between litter- and soil-feeding detritivores*

309 While the two case studies presented here focused on different groups of detritivores with  
310 different food sources (feeding on leaf litter in Joly et al. (2020), and feeding on soil and litter in  
311 Le Mer et al. (2022)) and considered different soil processes (organic matter C and N loss in Joly  
312 et al., 2020; soil C mineralisation in Le Mer et al., 2022), we argue that their respective  
313 approaches could be combined by considering similar faeces characteristics and processes  
314 (leaching, microbial degradation, stabilisation) across a diversity of organisms feeding on plant  
315 litter and mineral soil (Fig. 5, left panel). Notably, earthworms do not solely feed on mineral soil  
316 but also, depending on species, ingest varying quantities of litter at various stages of  
317 decomposition. Simultaneously, litter-feeding detritivores also integrate substantial amounts of  
318 soil as part of their diet (David, 2014). Yet, most earthworm studies compared earthworm faeces

319 to the bulk soil often ignoring the ingestion and fate of litter, and in turn the faeces of litter-  
320 feeding detritivores was mostly compared to the intact litter ignoring the ingestion of soil and its  
321 fate. Because both groups ingest and mix soil and litter to some extent, they may affect similar  
322 soil processes to varying degrees. We thus suggest that these groups be placed along gradients of  
323 litter-soil ingestion, and that their faeces be compared to the average characteristics and fate of  
324 their food source (soil and litter).

325         The depth at which produced faeces are returned to the soil may also be an important  
326 faeces trait to predict detritivore effects of organic matter turnover and combine the roles of litter  
327 and soil feeding detritivores. Although many detritivore species live and feed in the litter layer,  
328 some live deeper in the soil and most at least move through the soil, as recently illustrated with  
329 3D image analyses of soil burrows in mesocosms occupied by earthworms and millipedes (Mele  
330 et al., 2021). The creation of biopores by millipedes, well-known by soil zoologists, has been  
331 rarely considered by ecologists. A direct consequence of this is that faeces may also be deposited  
332 deeper than the ingested food in the soil, thus possibly changing decomposition rate. Indeed,  
333 Coulis et al. (2016) showed that faeces decomposition was faster than intact litter at soil surface,  
334 and that this decomposition was even faster when faeces were buried. Instead, an isopod species  
335 in the Negev desert that lives in deep burrows deposits its faeces at the soil surface (Sagi et al.,  
336 2019; Yair and Rutin, 1981). The average depth at which a given detritivore species typically  
337 deposits its faeces, and the proportion of buried faeces compared to surface ones, may thus be  
338 important factors for the fate of the faeces, and could place detritivore species along a continuous  
339 axis rather than categorise detritivore into soil-dwelling and litter-dwelling groups.

340

#### 341 4.3. *The composite determinism of detritivore faeces traits*

342 A main difficulty in the use of faeces traits as predictors of organic matter turnover is that  
343 these traits have a composite determinism, originating both from the identity of the detritivore  
344 and from the quality of its resources (Fig. 5). We argue that this feature does not contradict the  
345 consideration that faeces characteristics are relevant effect traits. The composite determinism of  
346 faeces traits does not prevent identifying which traits are powerful predictors of organic matter  
347 turnover. For example, in the case studies presented above, the authors identified faeces DOC  
348 concentrations as a good predictor of faeces C loss (Joly et al., 2020), and Le Mer et al. (2022)  
349 similarly demonstrated that microstructural traits predicted CO<sub>2</sub> emissions from earthworm  
350 faeces. The composite determinism of faeces traits, however, clearly makes it more challenging  
351 to use species-specific trait values to upscale to the community and ecosystem levels. For  
352 example, the use of community-weighted means is based on measurements of the local  
353 community structure and on taxon-specific trait values, averaged from local measurements or  
354 from databases. While some traits are mainly determined by the detritivore species (e.g., size,  
355 shape and location of faeces), for traits related to chemical characteristics, the attribution of a  
356 trait value to a detritivore species is not straightforward. Indeed, their value depends on the  
357 ingested resource and its interaction with the detritivore species. For such traits, the approach  
358 presented in the case study 1 (Joly et al., 2020) might be a promising solution: the relevant trait is  
359 not the faeces trait per se, but the change in trait value between the food and the produced faeces.  
360 Building relationships between the quality of the ingested organic matter and relevant faeces  
361 traits, for major groups of detritivores or even for individual species as proposed in Fig. 3,  
362 appears as a relevant way to overcome the difficulties arising from the composite determinism of

363 faeces traits. With this framework, the knowledge of litter quality and of the local community of  
364 detritivores could allow a reasonable prediction of the effect of litter transformation into faeces.

365

#### 366 4.4. Integration into current frameworks of trait-based ecology

367 While studying detritivore faeces characteristics appears as a promising way to better  
368 understand and integrate the role of detritivore in organic matter dynamics, we must also ask  
369 whether they can be considered as *traits*. Traits are defined as “any morphological,  
370 physiological, or phenological heritable feature measurable at the individual level, from the cell  
371 to the whole organism, without reference to the environment or any other level of organization”  
372 (Pey et al., 2014; Violle et al., 2007). If applied rigorously, faeces characteristics do not fit to this  
373 definition, since faeces are not part of the individual, strictly speaking. Yet, because they are  
374 largely shaped by the identity of the detritivore, faeces characteristics can to a large extent be  
375 conceptualised and analysed as traits. Similar extensions of the use of traits beyond the living  
376 organisms is commonly applied, for example for plant litter traits as an extension of plant traits  
377 (e.g., Fujii et al., 2020; Garcia-Palacios et al., 2016; Makkonen et al., 2012) or enzymatic  
378 production as microbial trait (Piton et al., 2020; Weimann, 2016). Thus, we argue that including  
379 faeces characteristics as traits of the detritivores that produced them is a reasonable and fruitful  
380 option.

381 We then must answer: can faeces traits be considered as *functional* traits? Defining what  
382 makes a trait functional is far from trivial, because several definitions of functions have been  
383 used in ecology (Malaterre et al., 2019). From a selectionist approach, the functions of a trait of  
384 biological entities are “the effect for which those entities were favoured under past natural  
385 selection” (Malaterre et al., 2019). This definition bears similarity with the functional trait

386 definition proposed by Violle et al. (2007) or Garnier et al. (2016), as traits “indirectly  
387 influencing the fitness of an individual via its effects on growth, reproduction, or survival”.

388 Response traits, which vary in response to changes in environmental conditions, fit well with  
389 these selectionist approaches. The question to answer to determine if faeces characteristics fit  
390 this selectionist definition is therefore: do the characteristics of faeces feed back to the fitness of  
391 the organisms producing the faeces? This question was explored for soil engineers by Jouquet et  
392 al. (2006) who differentiated, following Jones et al. (1994, 1997), between ‘extended phenotype  
393 engineers’ as organisms creating biogenic structures that directly influence the fitness of the  
394 organism producing it, and ‘accidental engineers’ for which no such positive effect is recorded. It  
395 was recently shown that earthworm activity in European forests could increase soil pH, thereby  
396 making soil conditions more favourable for themselves and reinforcing earthworm abundance  
397 (Desie et al., 2020). This suggests that the feeding activity of soil fauna and transformation of  
398 organic matter can alter soil properties in a way that affects soil fauna fitness. For other  
399 detritivores, we are not aware of studies demonstrating that faeces properties modify  
400 environmental conditions in a way that benefits fitness, and the answer might depend on the  
401 studied species. When the term *functional* is used in a selectionist meaning, faeces traits are thus  
402 not unequivocally functional. However, other authors proposed non-selectionist, alternative  
403 definitions of function, and therefore of functional traits (Dussault, 2018, Malaterre et al. 2019).

404 In this approach, traits are functional when they enable the organism to achieve particular  
405 contribution to ecosystem processes (Dussault, 2018). Following this alternative definition of  
406 function, they can also be considered functional traits. Regardless of the definition of function  
407 and functional traits, faeces traits are unambiguously *effect trait*, which influences ecosystem  
408 properties (Garnier and Navas, 2012).



## 409 **5. Conclusions**

410 In conceptual and mechanistic biogeochemical models, soil fauna are the ‘*elephant in the*  
411 *room*’ (Briones, 2018; Filser et al., 2016; Griffiths et al., 2021, Prescott and Vesterdal, 2021),  
412 likely because of the difficulty of studying and synthesising such a diverse group of organisms,  
413 which roles are difficult to isolate. As a first step towards bridging this gap, our viewpoint  
414 proposes a way to integrate detritivorous soil animals by focusing on their faeces, which is a  
415 common feature amongst detritivores and represents a key decomposition by-product in  
416 detritivore-rich ecosystems. Faeces characteristics of distinct detritivore species were recently  
417 shown to predict relatively well processes involved in organic matter turnover, and we thus  
418 formalised faeces characteristics as *effect traits*. This appears as a promising way to deal with the  
419 astonishing diversity of detritivores in soils, which may in particular unify historical soil fauna  
420 groups such as soil engineers and litter transformers. This approach could overall contribute to  
421 the inclusion of detritivores in biogeochemical models, thereby improving our understanding and  
422 modelling of carbon cycling.

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427

428 **References**

- 429 Adamczyk, B., Sietiö, O.M., Straková, P., Prommer, J., Wild, B., Hagner, M., Pihlatie, M.,  
430 Fritze, H., Richter, A., Heinonsalo, J., 2019. Plant roots increase both decomposition and stable  
431 organic matter formation in boreal forest soil. *Nat Commun* 10, 3982.  
432 <https://doi.org/10.1038/s41467-019-11993-1>.
- 433 Angst, Š., Mueller, C.W., Cajthaml, T., Angst, G., Lhotáková, Z., Bartuška, M., Špaldoňová, A.,  
434 Frouz, J., 2017. Stabilization of soil organic matter by earthworms is connected with physical  
435 protection rather than with chemical changes of organic matter. *Geoderma*, 289, 29-35.  
436 <https://doi.org/10.1016/j.geoderma.2016.11.017>.
- 437 Angst, G., Mueller, C. W., Prater, I., Angst, Š., Frouz, J., Jílková, V., Peterse, F., Nierop, K.G.,  
438 2019. Earthworms act as biochemical reactors to convert labile plant compounds into stabilized  
439 soil microbial necromass. *Commun Biol*, 2, 441. <https://doi.org/10.1038/s42003-019-0684-z>.
- 440 Astor, T., Lenoir, L., Berg, M.P., 2015. Measuring feeding traits of a range of litter-consuming  
441 terrestrial snails: leaf litter consumption, faeces production and scaling with body size.  
442 *Oecologia*, 178, 833-845. <https://doi.org/10.1007/s00442-015-3257-y>.
- 443 Barthod, J., Dignac, M.F., Le Mer, G., Bottinelli, N., Watteau, F., Kögel-Knabner, I., Rumpel,  
444 C., 2020. How do earthworms affect organic matter decomposition in the presence of clay-sized  
445 minerals? *Soil Biol Biochem*, 143, 107730. <https://doi.org/10.1016/j.soilbio.2020.107730>.
- 446 Barthod, J., Dignac, M.F., Rumpel, C., 2021. Effect of decomposition products produced in the  
447 presence or absence of epigeic earthworms and minerals on soil carbon stabilization. *Soil Biol*  
448 *Biochem*, 108308. <https://doi.org/10.1016/j.soilbio.2021.108308>.

449 Basile-Doelsch, I., Balesdent, J., Pellerin, S., 2020. Reviews and syntheses: The mechanisms  
450 underlying carbon storage in soil, *Biogeosciences*, 17, 5223-5242. [https://doi.org/10.5194/bg-17-](https://doi.org/10.5194/bg-17-5223-2020)  
451 5223-2020.

452 BETSI database, 2014. Biological and Ecological Traits for Soil Invertebrates.  
453 <https://portail.betsi.cnrs.fr/> (Accessed 08 December 2021)

454 Bonkowski, M., Scheu, S., Schaefer, M. (1998). Interactions of earthworms (*Octolasion*  
455 *lacteum*), millipedes (*Glomeris marginata*) and plants (*Hordelymus europaeus*) in a beechwood  
456 on a basalt hill: implications for litter decomposition and soil formation. *Appl Soil Ecol*, 9(1-3),  
457 161-166. [https://doi.org/10.1016/S0929-1393\(98\)00070-5](https://doi.org/10.1016/S0929-1393(98)00070-5).

458 Bottinelli, N., Maeght, J.L., Pham, R.D., Valentin, C., Rumpel, C., Pham, Q.V., Nguyen, T.T.,  
459 Lam, D.H., Nguyen, A.D., Tran, T.M., Zaiss, R., Jouquet, P. 2021. Anecic earthworms generate  
460 more topsoil than they contribute to erosion—Evidence at catchment scale in northern Vietnam.  
461 *Catena*, 201, 105-186. <https://doi.org/10.1016/j.catena.2021.105186>.

462 Briones, M.J., 2018. The serendipitous value of soil fauna in ecosystem functioning: the  
463 unexplained explained. *Front Environ Sci*, 6, 149. <https://doi.org/10.3389/fenvs.2018.00149>.

464 Brousseau, P.M., Gravel, D., Handa, I.T., 2018. On the development of a predictive functional  
465 trait approach for studying terrestrial arthropods. *J Anim Ecol*, 87(5), 1209-1220.  
466 <https://doi.org/10.1111/1365-2656.12834>.

467 Clause, J., Barot, S., Richard, B., Decaens, T., Forey, E., 2014. The interactions between soil  
468 type and earthworm species determine the properties of earthworm casts. *Appl Soil Ecol*, 83,  
469 149-158. <https://doi.org/10.1016/j.apsoil.2013.12.006>.

470 Clemmensen, K.E., Bahr, A., Ovaskainen, O., Dahlberg, A., Ekblad, A., Wallander, H., Stenlid,  
471 J., Finlay, R.D., Wardle, D.A., Lindahl, B.D., 2013. Roots and associated fungi drive long-term  
472 carbon sequestration in boreal forest. *Science*, 339(6127), 1615-1618.  
473 <https://doi.org/10.1126/science.1231923>.

474 Cornwell, W. K., Cornelissen, J. H., Amatangelo, K., Dorrepaal, E., Eviner, V. T., Godoy, O.,  
475 Hobbie, S.E., Hoorens, B., Kurokawa, H., Pérez-Harguindeguy, N., Quested, H.M., Santiago,  
476 L.S., Wardle, D.A., Wright, I.J., Aerts, R., Allison, S.D., Van Bodegom, P., Brovkin, V.,  
477 Chatain, A., Callaghan, T.V., Díaz, S., Garnier, E., Gurvich, D.E., Kazakou, E., Klein, J.A.,  
478 Read, J., Reich, P.B., Soudzilovskaia, N.A., Vaieretti, M.V., Westoby, M. 2008. Plant species  
479 traits are the predominant control on litter decomposition rates within biomes worldwide. *Ecol*  
480 *Lett*, 11(10), 1065-1071. <https://doi.org/10.1111/j.1461-0248.2008.01219.x>.

481 Cortez, J., Garnier, E., Pérez-Harguindeguy, N., Debussche, M., Gillon, D., 2007. Plant traits,  
482 litter quality and decomposition in a Mediterranean old-field succession. *Plant Soil*, 296(1), 19-  
483 34. <https://doi.org/10.1007/s11104-007-9285-6>.

484 Cotrufo, M.F., Soong, J., Horton, A., Campbell, E.E., Haddix, M.L., Wall, D.H., Parton, W.J.,  
485 2015. Formation of soil organic matter via biochemical and physical pathways of litter mass loss.  
486 *Nat Geosci* 8, 776-779. <https://doi.org/10.1038/ngeo2520>.

487 Cotrufo, M.F., Wallenstein, M.D., Boot, C.M., Deneff, K., Paul, E., 2013. The Microbial  
488 Efficiency-Matrix Stabilization (MEMS) framework integrates plant litter decomposition with  
489 soil organic matter stabilization: do labile plant inputs form stable soil organic matter?. *Glob*  
490 *Chang Biol*, 19: 988-995. <https://doi.org/10.1111/gcb.12113>.

491 Coulis, M., Hättenschwiler, S., Coq, S., David, J.F., 2016. Leaf litter consumption by  
492 macroarthropods and burial of their faeces enhance decomposition in a Mediterranean  
493 ecosystem. *Ecosystems*, 19(6), 1104-1115. <https://doi.org/10.1007/s10021-016-9990-1>.

494 Coulis, M., Hättenschwiler, S., Rapior, S., Coq, S., 2009. The fate of condensed tannins during  
495 litter consumption by soil animals. *Soil Biol Biochem*, 41(12), 2573-2578.  
496 <https://doi.org/10.1016/j.soilbio.2009.09.022>.

497 Crossley, D.A., Reichle, D.E., Edwards, C.A., 1971. Intake and turnover of radioactive cesium  
498 by earthworms (*Lumbricidae*). *Pedobiologia*, 11, 71-76.

499 Curry, J. P., Schmidt, O., 2007. The feeding ecology of earthworms—a review. *Pedobiologia*,  
500 50(6), 463-477. <https://doi.org/10.1016/j.pedobi.2006.09.001>

501 David, J.F., Gillon, D., 2002. Annual feeding rate of the millipede *Glomeris marginata* on holm  
502 oak (*Quercus ilex*) leaf litter under Mediterranean conditions. *Pedobiologia*, 46, 42-52.  
503 <https://doi.org/10.1078/0031-4056-00112>.

504 David, J.F., 2014. The role of litter-feeding macroarthropods in decomposition processes: a  
505 reappraisal of common views. *Soil Biol Biochem*, 76, 109-118.  
506 <https://doi.org/10.1016/j.soilbio.2014.05.009>.

507 Dangerfield, J.M., Milner, A.E., 1996. Millipede fecal pellet production in selected natural and  
508 managed habitats of southern Africa: implications for litter dynamics. *Biotropica*, 113-120.  
509 <https://doi.org/10.2307/2388776>.

510 de la Riva, E.G., Prieto, I., Villar, R., 2019. The leaf economic spectrum drives leaf litter  
511 decomposition in Mediterranean forests. *Plant Soil*, 435, 353-366.  
512 <https://doi.org/10.1007/s11104-018-3883-3>.

513 Decaëns, T., 2000. Degradation dynamics of surface earthworm casts in grasslands of the eastern  
514 plains of Colombia. *Biol Fertil Soils*, 32, 149156. <https://doi.org/10.1007/s003740000229>.

515 Desie, E., Van Meerbeek, K., De Wandeler, H., Bruelheide, H., Domisch, T., Jaroszewicz, B.,  
516 Joly, F.X., Vancampenhout, K., Vesterdal, L., Muys, B., 2020. Positive feedback loop between  
517 earthworms, humus form and soil pH reinforces earthworm abundance in European forests.  
518 *Funct Ecol*, 34(12), 2598-2610. <https://doi.org/10.1111/1365-2435.13668>.

519 Dignac, M.F., Derrien D., Barré P., Barot S., Cécillon L., Chenu, C., Chevallier, C., Freschet,  
520 G.T., Garnier, P., Guenet, B., Hedde, M., Klumpp, K., Lashermes, G., Maron, P.A., Nunan, N.,  
521 Roumet, C., Basile-Doelsch, I., 2017. Increasing soil carbon storage: mechanisms, effects of  
522 agricultural practices and proxies. A review. *Agron Sustain Dev* 37, 14.  
523 <https://doi.org/10.1007/s13593-017-0421-2>.

524 Dungait, J.A., Hopkins, D.W., Gregory, A.S., Whitmore, A.P., 2012. Soil organic matter  
525 turnover is governed by accessibility not recalcitrance. *Glob Chang Biol*, 18(6), 1781-1796.  
526 <https://doi.org/10.1111/j.1365-2486.2012.02665.x>.

527 Dussault, A.C., 2018. Functional ecology's non-selectionist understanding of function. *Stud Hist*  
528 *Philos Biol Biomed Sci C*, 70, 1-9. <https://doi.org/10.1016/j.shpsc.2018.05.001>.

529 Dynarski, K.A., Bossio D.A., Scow K.M., 2020. Dynamic stability of soil C: Reassessing the  
530 "permanence" of carbon sequestration. *Front Environ Sci*, 8, 218.  
531 <https://doi.org/10.3389/fenvs.2020.514701>.

532 Filser, J., Faber, J.H., Tiunov, A.V., Brussaard, L., Frouz, J., De Deyn, G., Uvarov, A.V., Berg,  
533 M.P., Lavelle, P., Loreau, M., Wall, D.H., Querner, P., Eijsackers, H., Jiménez, J.J., 2016. Soil  
534 fauna: key to new carbon models, *Soil*, 2, 565-582, <https://doi.org/10.5194/soil-2-565-2016>.

535 Freschet, G.T., Roumet, C., Comas, L.H., Weemstra, M., Bengough, A.G., Rewald, B., Bardgett,  
536 R.D., De Deyn, G.B., Johnson, D., Klimešová, J., Lukac, M., McCormack, M.L., Meier, I.C.,  
537 Pagès, L., Poorter, H., Prieto, I., Wurzbürger, N., Zadworny, M., Bagniewska-Zadworna, A.,  
538 Blancaflor, E.B., Brunner, I., Gessler, A., Hobbie, S.E., Iversen, C.M., Mommer, L., Picon-  
539 Cochard, C., Postma, J.A., Rose, L., Ryser, P., Scherer-Lorenzen, M., Soudzilovskaia, N.A.,  
540 Sun, T., Valverde-Barrantes, O.J., Weigelt, A., York, L.M., Stokes, A., 2021. Root traits as  
541 drivers of plant and ecosystem functioning: current understanding, pitfalls and future research  
542 needs. *New Phytol*, 232, 1123-1158. <https://doi.org/10.1111/nph.17072>.

543 Freschet, G.T., Cornwell, W.K., Wardle, D.A., Elumeeva, T.G., Liu, W., Jackson, B.G.,  
544 Onipchenko, V.G., Soudzilovskaia, N.A., Tao, J., Cornelissen, J.H.C., 2013. Linking litter  
545 decomposition of above- and below-ground organs to plant-soil feedbacks worldwide. *J Ecol*,  
546 101(4), 943-952. <http://www.jstor.org/stable/42580325>.

547 Fujii, S., Berg, M.P., Cornelissen, J.H., 2020. Living litter: Dynamic trait spectra predict fauna  
548 composition. *Trends Ecol Evol*, 35(10), 886-896. <https://doi.org/10.1016/j.tree.2020.05.007>.

549 Ganault, P., Barantal, S., Coq, S., Hättenschwiler, S., Lucas, S., Decaëns, T., Nahmani, J., 2022.  
550 Leaf litter morphological traits, invertebrate body mass and phylogenetic affiliation explain the  
551 feeding and feces properties of saprophagous macroarthropods. *Eur J Soil Biol*, 109, 103383

552 García- Palacios, P., Maestre, F.T., Kattge, J., Wall, D.H., 2013. Climate and litter quality  
553 differently modulate the effects of soil fauna on litter decomposition across biomes. *Ecol lett*,  
554 16(8), 1045-1053. <https://doi.org/10.1111/ele.12137>.

555 García- Palacios, P., McKie, B.G., Handa, I.T., Frainer, A., Hättenschwiler, S., 2016. The  
556 importance of litter traits and decomposers for litter decomposition: a comparison of aquatic and  
557 terrestrial ecosystems within and across biomes. *Funct Ecol*, 30(5), 819-829.  
558 <https://doi.org/10.1111/1365-2435.12589>.

559 Garnier, E., Navas, M.L. 2012. A trait-based approach to comparative functional plant ecology:  
560 concepts, methods and applications for agroecology. A review. *Agron Sustain Dev*, 32(2), 365-  
561 399. <https://doi.org/10.1007/s13593-011-0036-y>.

562 Garnier, E., Navas, M.L., Grigulis, K., 2016. *Plant functional diversity: organism traits,*  
563 *community structure, and ecosystem properties.* Oxford University Press.

564 Griffiths, H.M., Ashton, L.A., Parr, C.L., Eggleton, P., 2021. The impact of invertebrate  
565 decomposers on plants and soil. *New Phytol*, 231, 2142-2149.  
566 <https://doi.org/10.1111/nph.17553>.

567 Handa, I., Aerts, R., Berendse, F., Berg, M.P., Bruder, A., Butenschoen, O., Chauvet, E.,  
568 Gessner, M.O., Jabiol, J., Makkonen, M., McKie, B.G., Malmqvist, B., Peeters, E.T.H.M.,  
569 Scheu, S., Schmid, B., van Ruijven, J., Vos, V.C.A., Hättenschwiler, S., 2014. Consequences of



570 biodiversity loss for litter decomposition across biomes. *Nature*, 509, 218-221.  
571 <https://doi.org/10.1038/nature13247>.

572 Hättenschwiler, S., Gasser, P., 2005. Soil animals alter plant litter diversity effects on  
573 decomposition. *Proc Natl Acad Sci*, 102(5), 1519-1524.  
574 <https://doi.org/10.1073/pnas.0404977102>.

575 Hedde, M., Lavelle, P., Joffre, R., Jiménez, J.J., Decaëns, T., 2005. Specific functional signature  
576 in soil macro- invertebrate biostructures. *Funct Ecol*, 19 (5), 785-793.  
577 <https://doi.org/10.1111/j.1365-2435.2005.01026.x>

578 Joly, F.X., Coq, S., Coulis, M., David, J.F., Hättenschwiler, S., Mueller, C.W., Proater, I., Subke,  
579 J.A., 2020. Detritivore conversion of litter into faeces accelerates organic matter turnover.  
580 *Commun Biol*, 3(1), 1-9. <https://doi.org/10.1038/s42003-020-01392-4>.

581 Joly, F.X., Coq, S., Coulis, M., Nahmani, J., Hättenschwiler, S., 2018. Litter conversion into  
582 detritivore faeces reshuffles the quality control over C and N dynamics during decomposition.  
583 *Funct Ecol*, 32(11), 2605-2614. <https://doi.org/10.1111/1365-2435.13178>.

584 Joly, F.X., McAvoy, E., Subke, J.A., 2021. Synergistic interactions between detritivores  
585 disappear under reduced rainfall. *Ecology*, 102(4), e03299. <https://doi.org/10.1002/ecy.3299>.

586 Jones, C.G., Lawton, J.H., Shachak, M., 1994. Organisms as ecosystem engineers. In *Ecosystem*  
587 *management*. Springer, New York, NY. pp. 130-147.

588 Jones, C.G., Lawton, J.H., Shachak, M., 1997. Positive and negative effects of organisms as  
589 physical ecosystem engineers. *Ecology*, 78(7), 1946-1957.

590 Jouquet, P., Dauber, J., Lagerlöf, J., Lavelle, P., Lepage, M., 2006. Soil invertebrates as  
591 ecosystem engineers: intended and accidental effects on soil and feedback loops. *Appl Soil Ecol*,  
592 32(2), 153-164. <https://doi.org/10.1016/j.apsoil.2005.07.004>.

593 Jouquet, P., Bottinelli, N., Podwojewski, P., Hallaire, V., Duc, T.T., 2008. Chemical and  
594 physical properties of earthworm casts as compared to bulk soil under a range of different land-  
595 use systems in Vietnam. *Geoderma*, 146, 231-238.  
596 <https://doi.org/10.1016/j.geoderma.2008.05.030>.

597 Kautz G., Zimmer M., Topp W. 2002. Does *Porcellio scaber* (Isopoda: Oniscidea) gain from  
598 coprophagy? *Soil Biol Biochem* 34(9), 1253-1259. <https://doi.org/10.1016/S0038->  
599 [0717\(02\)00065-2](https://doi.org/10.1016/S0038-0717(02)00065-2).

600 Kazakou, E., Vile, D., Shipley, B., Gallet, C., Garnier, E., 2006. Co- variations in litter  
601 decomposition, leaf traits and plant growth in species from a Mediterranean old- field  
602 succession. *Funct Ecol*, 20(1), 21-30. <https://doi.org/10.1111/j.1365-2435.2006.01080.x>.

603 Kögel-Knabner, I., Rumpel, C., 2018. Chapter One - Advances in Molecular Approaches for  
604 Understanding Soil Organic Matter Composition, Origin, and Turnover: A Historical Overview.  
605 *Advances in Agronomy*, Academic Press, Volume 149, pp. 1-48, Ed Sparks D.  
606 <https://doi.org/10.1016/bs.agron.2018.01.003>.

607 Koltz, A.M., Classen, A.T., Wright, J.P., 2018. Warming reverses top-down effects of predators  
608 on belowground ecosystem function in Arctic tundra. *Proc Natl Acad Sci*, 115(32), E7541-  
609 E7549.

610 Lavelle, P., Martin, A., 1992. Small-scale and large-scale effects of endogeic earthworms on soil  
611 organic matter dynamics in soils of the humid tropics. *Soil Biol Biochem*, 24(12), 1491-1498.  
612 [https://doi.org/10.1016/0038-0717\(92\)90138-N](https://doi.org/10.1016/0038-0717(92)90138-N).

613 Lavelle, P., Spain, A.V., 2001. *Soil Ecology*. Kluwer Academic Publishers, Dordrecht.

614 Lavorel, S., Garnier, E., 2002. Predicting changes in community composition and ecosystem  
615 functioning from plant traits: revisiting the Holy Grail. *Funct Ecol*, 16(5), 545-556.  
616 <https://doi.org/10.1046/j.1365-2435.2002.00664.x>.

617 Le Mer, G, Bottinelli, N., Dignac, M.F., Jouquet, P., Mazurier, A., Capowiez, Y., Rumpel, C.,  
618 2022. Exploring the control of earthworm faeces micro- and macro-scale features on soil organic  
619 matter dynamics across species and ecological categories. *Geoderma*, submitted.

620 Le Mer, G., Barthod, J., Dignac, M.F., Barré, P., Baudin, F., Rumpel, C., 2020. Inferring the  
621 impact of earthworms on the stability of organo-mineral associations, by Rock-Eval thermal  
622 analysis and <sup>13</sup>C NMR spectroscopy. *Org Geochem*, 144, 104016.  
623 <https://doi.org/10.1016/j.orggeochem.2020.104016>.

624 Lehmann, J., Kleber, M. 2015. The contentious nature of soil organic matter. *Nature*, 528, 60-68.  
625 <https://doi.org/10.1038/nature16069>.

626 Lenoir, L., Persson, T., Bengtsson, J., Wallander, H., Wiren, A., 2007. Bottom-up or top-down  
627 control in forest soil microcosms? Effects of soil fauna on fungal biomass and C/N  
628 mineralisation. *Biol Fertil Soils*, 43(3), 281-294. <https://doi.org/10.1007/s00374-006-0103-8>.

629 Makkonen, M., Berg, M. P., Handa, I. T., Hättenschwiler, S., van Ruijven, J., van Bodegom, P.  
630 M., Aerts, R., 2012. Highly consistent effects of plant litter identity and functional traits on  
631 decomposition across a latitudinal gradient. *Ecol Lett*, 15(9), 1033-1041.  
632 <https://doi.org/10.1111/j.1461-0248.2012.01826.x>.

633 Malaterre, C., Dussault, A.C., Rousseau-Mermans, S., Barker, G., Beisner, B.E., Bouchard, F.,  
634 Desjardins, E., Handa, I.T., Kembel, S.W., Lajoie, G., Maris, V., Munson, A.D., Odenbaugh, J.,  
635 Poisot, T., Shapiro, B.J., Suttle, C.A., 2019. Functional diversity: An epistemic roadmap.  
636 *BioScience*, 69(10), 800-811. <https://doi.org/10.1093/biosci/biz089>.

637 Mele, G., Buscemi, G., Gargiulo, L., Terribile, F., 2021. Soil burrow characterization by 3D  
638 image analysis: Prediction of macroinvertebrate groups from biopore size distribution  
639 parameters. *Geoderma*, 404, 115292. <https://doi.org/10.1016/j.geoderma.2021.115292>.

640 Pey, B., Nahmani, J., Auclerc, A., Capowiez, Y., Cluzeau, D., Cortet, J., Decaëns, T.,  
641 Deharveng, L. Dubs, F., Joimel, S., Briard, C., Grumiaux, F., Laporte, M.A., Pasquet, A., Pelosi,  
642 C., Pernin, C., Ponge, J.F., Salmon, S., Santorufo, L., Hedde, M., 2014. Current use of and future  
643 needs for soil invertebrate functional traits in community ecology. *Basic Appl Ecol*, 15(3), 194-  
644 206. <https://doi.org/10.1016/j.baae.2014.03.007>.

645 Piton, G., Foulquier, A., Martínez-García, L.B., Legay, N., Hedlund, K., da Silva, P.M.,  
646 Nascimento, E., Reis, F., Sousa, J.P., De Deyn, G.B., Clement, J.C., 2020. Disentangling drivers  
647 of soil microbial potential enzyme activity across rain regimes: An approach based on the  
648 functional trait framework. *Soil Biol Biochem*, 148, 107881.  
649 <https://doi.org/10.1016/j.soilbio.2020.107881>.

650 Prescott, C.E., 2010. Litter decomposition: what controls it and how can we alter it to sequester  
651 more carbon in forest soils? *Biogeochemistry*, 101, 133-149. [https://doi.org/10.1007/s10533-](https://doi.org/10.1007/s10533-010-9439-0)  
652 010-9439-0.

653 Prescott, C.E., Vesterdal, L., 2021. Decomposition and transformations along the continuum  
654 from litter to soil organic matter in forest soils. *For Ecol Manage*, 498, 119522.  
655 <https://doi.org/10.1016/j.foreco.2021.119522>.

656 Rasse, D.P., Rumpel, C., Dignac, M.F., 2005. Is soil carbon mostly root carbon? Mechanisms for  
657 a specific stabilisation. *Plant Soil* 269, 341–356. <https://doi.org/10.1007/s11104-004-0907-y>.

658 Rosenfield, M.V., Keller, J.K., Clausen, C., Cyphers, K., Funk, J.L., 2020. Leaf traits can be  
659 used to predict rates of litter decomposition. *Oikos*, 129(10), 1589-1596.  
660 <https://doi.org/10.1111/oik.06470>.

661 Sagi, N., Grünzweig, J.M., Hawlena, D., 2019. Burrowing detritivores regulate nutrient cycling  
662 in a desert ecosystem. *Proc R Soc B*, 286(1914), 20191647.  
663 <https://doi.org/10.1098/rspb.2019.1647>

664 Satchell, J.E., 1983. Earthworm ecology in forest soils. In: Satchell J.E. (eds) *Earthworm  
665 Ecology*. Springer, Dordrecht. [https://doi.org/10.1007/978-94-009-5965-1\\_13](https://doi.org/10.1007/978-94-009-5965-1_13).

666 Schaefer, M., 1990. The soil fauna of a beech forest on limestone: trophic structure and energy  
667 budget. *Oecologia*, 82(1), 128-136. <https://doi.org/10.1007/BF00318544>.

668 Schmidt, M., Torn, M., Abiven, S., Dittmar, T., Guggenberger, G., Janssens, I., Kleber, M.,  
669 Kögel-Knabner, I., Lehmann, J., Manning, D., Nannipieri, P., Rasse, D., Weiner, S., Trumbore,

670 E., 2011. Persistence of soil organic matter as an ecosystem property. *Nature* 478, 49-56.  
671 <https://doi.org/10.1038/nature10386>.

672 Shipley, B., De Bello, F., Cornelissen, J.H.C., Laliberté, E., Laughlin, D.C., Reich, P.B., 2016.  
673 Reinforcing loose foundation stones in trait-based plant ecology. *Oecologia*, 180(4), 923-931.  
674 <https://doi.org/10.1007/s00442-016-3549-x>.

675 Sokol, N.W., Kuebbing, S.E., Karlsen- Ayala, E., Bradford, M.A., 2019a. Evidence for the  
676 primacy of living root inputs, not root or shoot litter, in forming soil organic carbon. *New Phytol*,  
677 221(1), 233-246. <https://doi.org/10.1111/nph.15361>.

678 Sokol, N.W., Sanderman J., Bradford, M.A., 2019b. Pathways of mineral-associated soil organic  
679 matter formation: Integrating the role of plant carbon source, chemistry, and point of entry. *Glob*  
680 *Chang Biol*, 25(1): 12-24. <https://doi.org/10.1111/gcb.14482>

681 Sterner, R.W., Elser, J.J., 2002. *Ecological stoichiometry: the biology of elements from*  
682 *molecules to the biosphere*. Princeton University Press, Princeton.

683 Van Groenigen, J.W., Van Groenigen, K.J., Koopmans, G.F., Stokkermans, L., Vos, H.M.,  
684 Lubbers, I.M., 2019. How fertile are earthworm casts? A meta-analysis. *Geoderma*, 338, 525-  
685 535. <https://doi.org/10.1016/j.geoderma.2018.11.001>.

686 Vidal, A., Watteau, F., Remusat, L., Mueller, C.W., Nguyen Tu, T.T., Buegger, F., Derenne, S.,  
687 Quenea K., 2019. Earthworm cast formation and development: a shift from plant litter to mineral  
688 associated organic matter. *Front Environ Sci* 7-55. <https://doi.org/10.3389/fenvs.2019.00055>.

689 Vidal, A., Quenea, K., Alexis, M., Derenne, S., 2016. Molecular fate of root and shoot litter on  
690 incorporation and decomposition in earthworm casts. *Org Geochem*, 101, 1-10.  
691 <https://doi.org/10.1016/j.orggeochem.2016.08.003>.

692 Violle, C., Navas, M.L., Vile, D., Kazakou, E., Fortunel, C., Hummel, I., Garnier, E., 2007. Let  
693 the concept of trait be functional! *Oikos*, 116(5), 882-892. <https://doi.org/10.1111/j.0030->  
694 [1299.2007.15559.x](https://doi.org/10.1111/j.0030-1299.2007.15559.x).

695 Wall, D.H., Bradford, M.A., St. John, M.G., Trofymow, J. A., Behan- Pelletier, V., Bignell,  
696 D.E., Dangerfield, J.M., Parton, W.J., Rusek, J., Voigt, W., Wolters V., Gardel, H.Z., Ayuke,  
697 F.O., Bashford, R., Beljakova, O.I., Bohlen, P.J., Brauman, A., Flemming, S., Henschel, J.R.,  
698 Johnson, D.L., Jones, T.H., Kovarova, M., Kranabetter, J.M., Kutny, L., Lin, K.C., Maryati, M.,  
699 Masse, D., Pokarzhevskii, A., Rahman, H., Sabara, M.G., Salamon, J.A., Swift, M.J., Varela, A.,  
700 Vasconcelos, H.L., White, D., Zou, X. 2008. Global decomposition experiment shows soil  
701 animal impacts on decomposition are climate- dependent. *Glob Chang Biol*, 14(11), 2661-2677.  
702 <https://doi.org/10.1111/j.1365-2486.2008.01672.x>.

703 Wardle, D.A., 2002. *Communities and Ecosystems Linking the Aboveground and Belowground*  
704 *Components*. Princeton University Press, Princeton.

705 Weimann, A., Mooren, K., Frank, J., Pope, P. B., Bremges, A., McHardy, A.C., 2016. From  
706 genomes to phenotypes: Traitair, the microbial trait analyzer. *mSystems*, 1(6), e00101-16.  
707 <https://doi.org/10.1128/mSystems.00101-16>.

708 Yair, A., Ruting, J., 1981. Some aspects of the regional variation in the amount of available  
709 sediment produced by isopods and porcupines, northern Negev, Israel. *Earth Surf Process Landf*,  
710 6(3-4), 221-234.

711 Zanella, A., Ponge, J.F., Briones, M.J.L., 2018. *Appl Soil Ecol. Humusica* 1, article 8: Terrestrial  
712 humus systems and forms – Biological activity and soil aggregates, space-time dynamics  
713 <https://doi.org/10.1016/j.apsoil.2017.07.020>.

714 Zimmer, M., 2002. Nutrition in terrestrial isopods (Isopoda: Oniscidea): an evolutionary-  
715 ecological approach. *Biol Rev*, 77(4), 455-493. <https://doi.org/10.1017/S1464793102005912>.

## 716 **Figure captions**

717 **Figure 1:** A sample of the diversity of detritivores and their faeces.

718 **Figure 2:** Diversity of detritivore faeces resulting from leaf litter of six tree species eaten by six  
719 detritivore species. From Joly et al., 2020.

720 **Figure 3:** Schematic representation of the relationships between the detritivore effect (i.e.  
721 changes in litter characteristics following detritivore conversion of litter into faeces) and intact  
722 litter characteristics, as observed in Joly et al., 2020. Changes in the magnitude of the detritivore  
723 effect following litter conversion into faeces are described by the intercept (e.g. **m** for species 1  
724 and **n** species 2). Changes in the interaction between the detritivore effect and the intact litter  
725 characteristics are described by the slope (e.g. **a** for species 2 and **b** for species 3). The intercept  
726 and slope value for each species can then be used to determine the change in organic matter  
727 characteristics following conversion into detritivore faeces.

728 **Figure 4:** Microstructural traits as predictors of carbon mineralization rates in faeces. a)

729 Multidimensional representation of earthworm faeces microstructures during decomposition. For



730 all species, porosity increases through time, while POM and fresh organic matter decreases. b)  
731 Mineralization rates of faeces produced by six earthworm species belonging to three ecological  
732 categories, measured after 7, 42 and 140 days of incubation. Mineralisation rates depend on  
733 earthworm species and faeces age. c) Respective importance of faeces traits as predictors of C  
734 mineralisation. Altogether, these microstructural traits explain more than 50% of the variability  
735 in faeces CO<sub>2</sub> emissions.

736 **Figure 5:** Conceptual framework formalising faeces traits as unifying predictors of detritivore  
737 effects on organic matter turnover. Amongst the diversity of detritivore, each individual can be  
738 placed along a gradient of litter and soil ingestion. Detritivores produce faeces whose traits are  
739 governed by the composite determinism of the identity of the detritivore, the characteristics of its  
740 resource and the interactions between both factors. These faeces traits are related to the several  
741 processes that contribute to organic matter turnover in soils.

742

743

Snails



Millipedes



Woodlice



Earthworms



1 mm



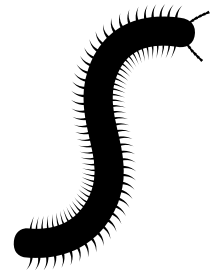
1 mm



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Millipedes

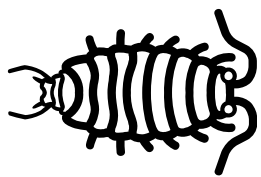
*Glomeris marginata*



*Ommatoiulus sabulosus*



*Tachypodoiulus niger*



Woodlice

*Armadillidium vulgare*



*Porcellio scaber*



Snail

*Cepaea nemoralis*



*Fagus sylvatica*



*Quercus robur*



*Acer pseudo-platanus*



*Aesculus hippocastanum*



*Corylus avellana*

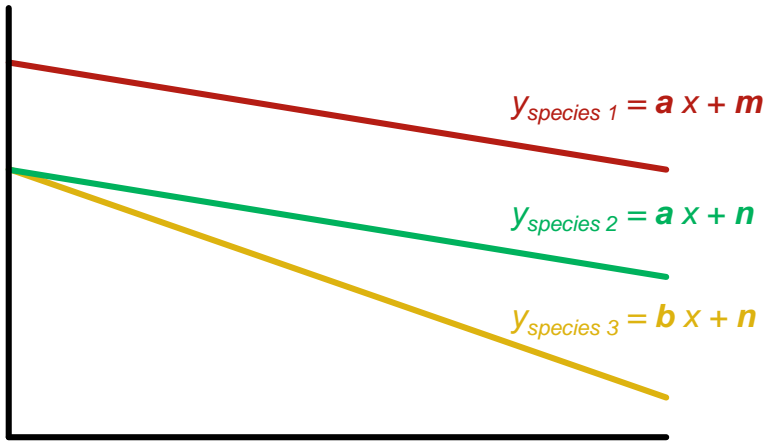


*Tilia platyphyllos*



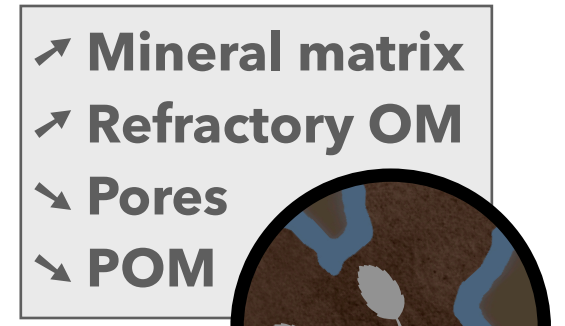
1mm

Differences in characteristics  
between faeces and intact litter



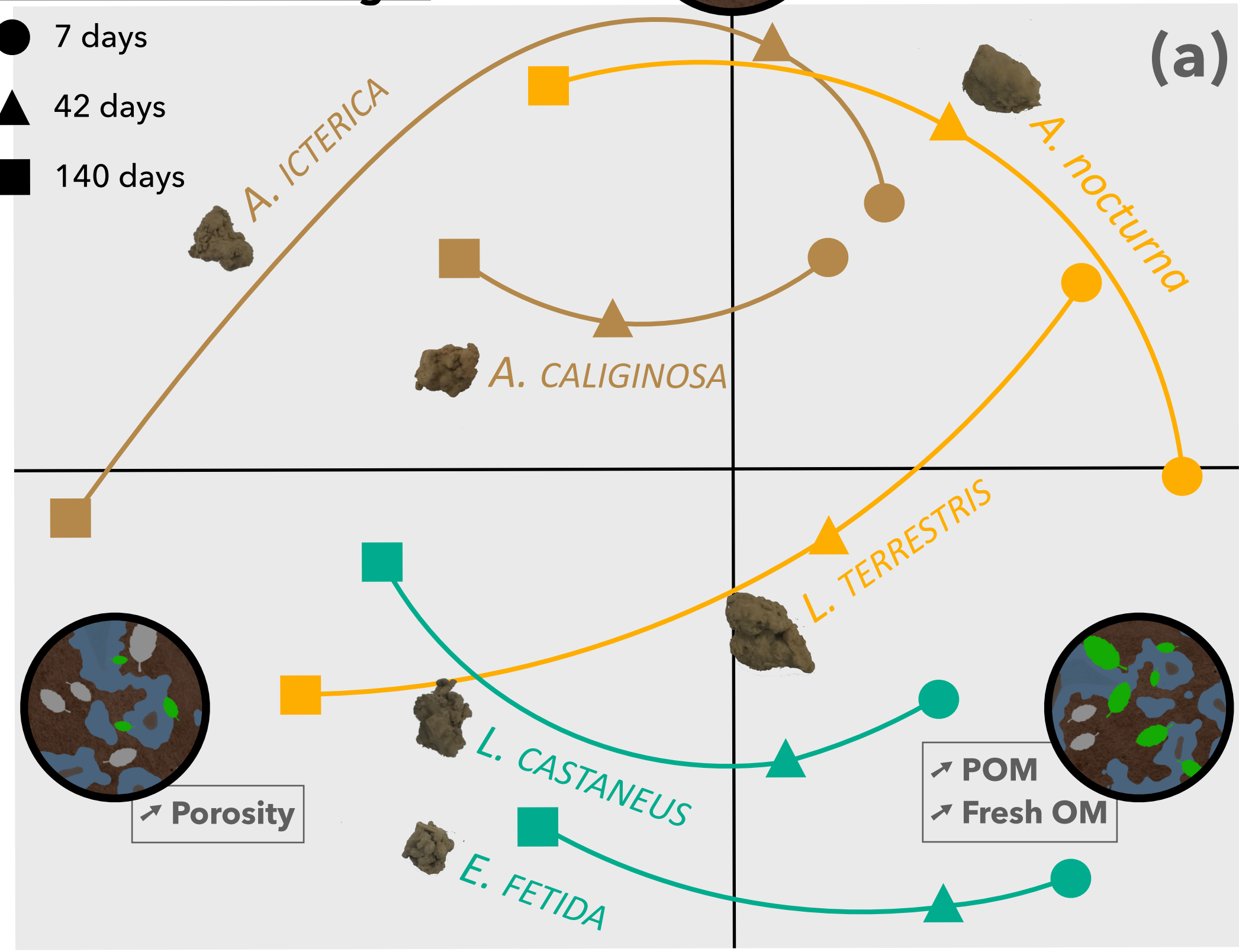
Intact litter characteristics





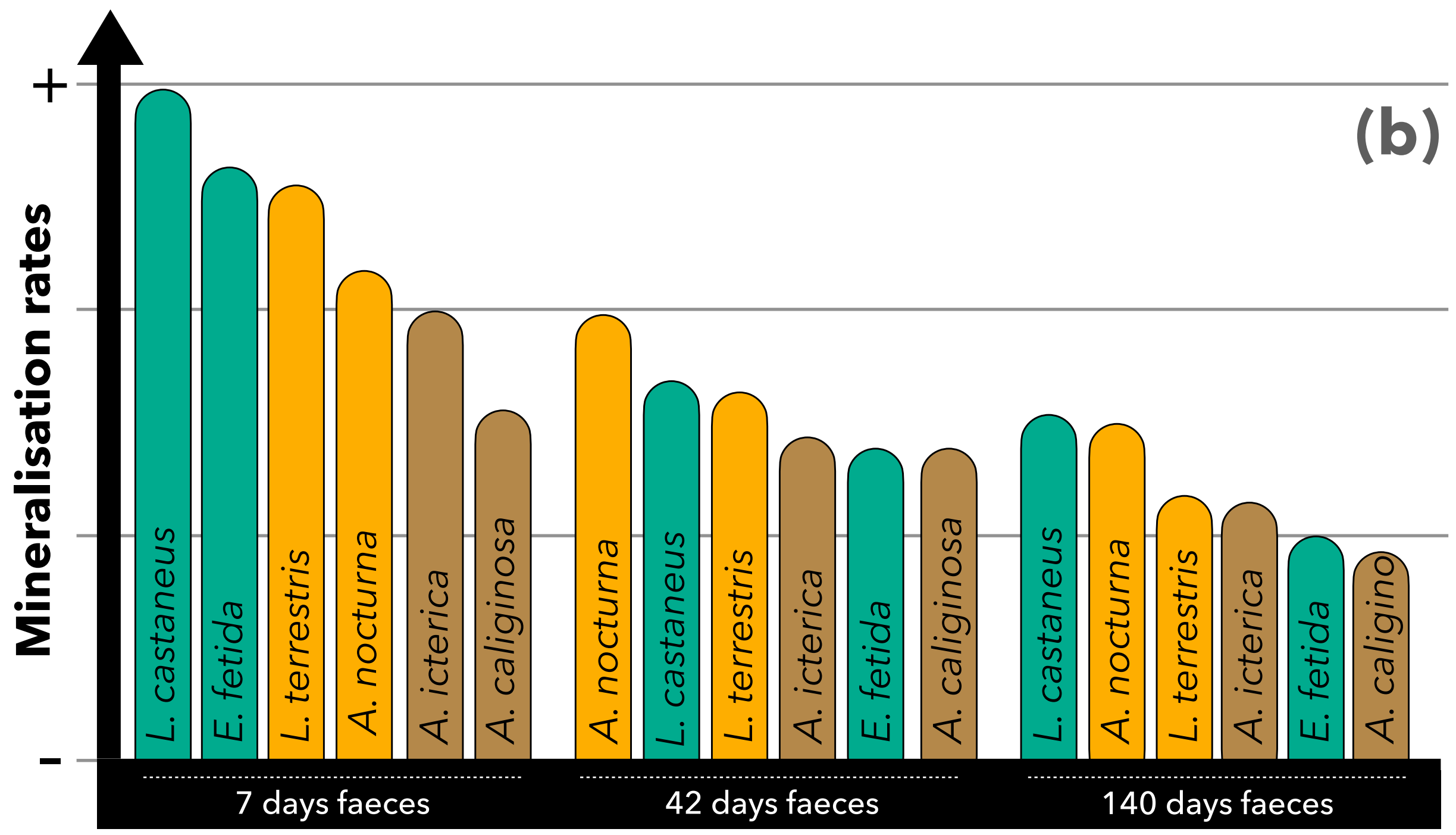
**Earthworm faeces age :**

- 7 days
- ▲ 42 days
- 140 days

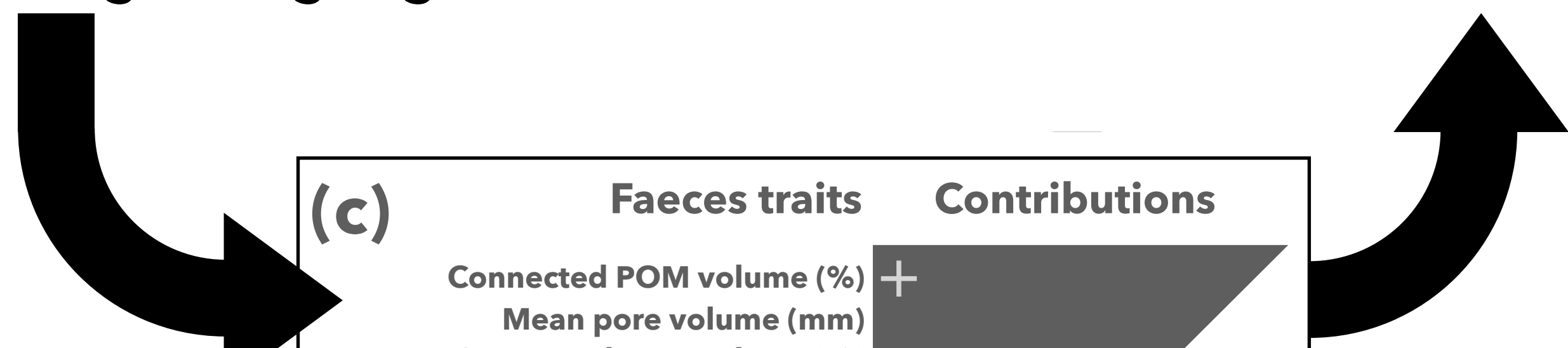


**Variability of earthworm faeces microstructures throughout ageing**

**Earthworm ecological group :**



**Variability of earthworm faeces mineralisation rates throughout ageing**



**(c)**

Faeces traits	Contributions
Connected POM volume (%)	+
Mean pore volume (mm)	
Connected pore volume (%)	
Cast specific area (mm <sup>2</sup> )	
Unconnected pore volume (%)	
Cast thickness (mm)	
Mean connected POM volume (mm)	
Total POM volume (%)	
Mean connected pore volume (mm)	

# Detritivore effects on organic matter turnover?

