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Hélène Tribouillois, Julie Constantin, Clément Murgue, Jean Villerd, Olivier Therond. Integrated modeling of crop and water management at the watershed scale: Optimizing irrigation and modifying crop succession. *European Journal of Agronomy*, 2022, 140, 10.1016/j.eja.2022.126592 . hal-03758473

HAL Id: hal-03758473

<https://hal.inrae.fr/hal-03758473>

Submitted on 28 Aug 2023

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1 Integrated modeling of crop and water management at the watershed 2 scale: optimizing irrigation and modifying crop succession

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8

9 **Keywords**

10 Water; modeling; watershed; withdrawals; irrigation; scenarios

11

12 **1 Introduction**

13 Due to population growth and the expansion of the agricultural, energy, and manufacturing sectors,
14 the demand for water has increased. Water scarcity has become a major concern for the future of
15 sustainable food production and the conservation of important ecosystem functions (Rosa et al.,
16 2020). Climate change is further exacerbating water scarcity and is expected to increase the frequency
17 of extremely dry periods in Europe (Lelieveld et al., 2012). In southern Europe, precipitation levels
18 declined 20% during the 20th century, additionally contributing to water supply issues (Caballero et
19 al., 2007).

20 Hydrological droughts occur when there are “inadequate surface and subsurface water resources for
21 the established water uses of a given water resources management system” (Mishra and Singh, 2010).

22 This definition underscores that drought occurrence, like other natural events requiring management,
23 is not only affected by resource levels, but also by usage patterns (i.e., needs and practices) and

24 management approaches, including operational water management and water use regulations.
25 Worldwide, over the past 50 years, cultivated irrigated surfaces have almost doubled in area, and 70%
26 of freshwater withdrawals are devoted to irrigation (Foley et al., 2011). This situation has placed
27 additional stress on water bodies and aquifers. The proliferation of irrigated, water-intensive crops,
28 such as maize, has increased soil drying and water scarcity in many watersheds with large areas
29 dedicated to irrigated crops (Liu et al., 2015).

30 Given climbing drought risks, it is crucial to consider how water management in cropping systems can
31 be improved in irrigated watersheds. In the scientific literature, agronomic studies at watershed scale
32 and on water management are mainly based on coarse farm typologies and focus on withdrawals
33 volume for irrigation and crop needs (e.g. Clavel et al., 2012; Neupane and Guo, 2019). These studies
34 generally do not include hydrological issues, such as rivers flows. Conversely, there is plentiful studies
35 on the hydrology of watersheds, but the functioning of crops and farming systems are often
36 represented very roughly (e.g. Francesconi et al., 2016; Van Emmerik et al., 2014). Furthermore, most
37 studies focus only on one issue in particular, such as drainage (e.g. Ale et al., 2012) or irrigation (e.g.
38 Clavel et al., 2012).

39 Another type of approach is to implement integrated assessment and modelling which accounts for
40 the complex interactions among cropping systems, farming systems, water resources, and water
41 management within watersheds both within and across years (Mazzega et al., 2014; Therond et al.,
42 2014). Integrated modelling is a useful method for functionally exploring the impacts of most of these
43 factors (Alcamo et al., 2003; Jakeman et al., 2006; Letcher et al., 2006; Therond et al., 2014; Zhang and
44 Guo, 2016). It can be used to evaluate and compare alternative agricultural and/or water management
45 systems at the watershed scale (March et al., 2012). It is essential to explicitly model anthropogenic
46 management strategies and their consequences to properly simulate the hydrological dynamics of
47 highly entropized basins (Martin et al., 2016). In irrigated watersheds, it is crucial to represent the
48 multiple ways in which different cropping systems (i.e., rainfed vs. irrigated) can affect water resource

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49 dynamics and, if relevant, reservoir releases and water use regulations (Martin et al., 2016; Mazzega
50 et al., 2014).

51 MAELIA is an integrated assessment and modeling platform representing fine-scale spatiotemporal
52 interactions among agricultural practices (e.g., crop succession and management strategies), the
53 hydrology of different water resources, and water resource management approaches (e.g., reservoir
54 releases, water use restrictions) within watersheds (Gaudou et al., 2014; Therond et al., 2014). To
55 properly perform such agent-based modeling, it is necessary to integrate scientific and practical
56 knowledge, local expertise, and large-scale census data (see Murgue et al., 2016, for example). This
57 integrated modeling platform could be very useful to test and evaluate the impact of agricultural
58 scenarios on water dynamics at larger scale.

59 The objectives of this study was to assess the impacts of three contrasting, realistic scenarios for
60 managing water use in irrigated maize agricultural systems on water dynamics at the field and
61 watershed scales, using MAELIA. We aimed to evaluate how much irrigation withdrawal volumes could
62 be lowered via strategies to reduce water deficits, which are traditionally recommended in France
63 (Mayor et al., 2012).

64

65 2 Materials and Methods

66 2.1 General approach

67 We first evaluated MAELIA predictions on withdrawal volumes and river flow rates at the watershed
68 scale over several years for the current situation of our studied system. We then designed and analyzed
69 the effect of three scenarios to enhance water management in irrigated maize systems on
70 environmental, water management, and agricultural indicators. These three alternative scenarios
71 were compared with the benchmark scenario based on actual irrigation management practiced in the
72 watershed. In the first, maize irrigation efficiency was optimized based on crop's water requirement.

73 In the second, maize monocultures were replaced by a wheat/maize succession. In the third, the two
74 previous scenarios were combined.

75 2.2 Study system

76 For this study, we choose the downstream Aveyron watershed, located in southwestern France
77 because it experiences strong and recurrent water deficits (Martin et al. 2016). This watershed is
78 around 800 km² large and is located in the Adour-Garonne basin, which is the basin with the greatest
79 structural water deficit in France. It contains numerous irrigated farming systems that account for up
80 to 80% of total anthropogenic water consumption during the area's low-flow period, from June to
81 September (Mazzega et al., 2014). The Aveyron watershed is among the most irrigated watersheds in
82 France and has the greatest annual water deficit: 7 hm³ on average. It is a prominent example of a
83 water-imbalance situation, in which low water inputs co-occur with high agricultural demands. Two
84 main rivers cross the watershed: the Aveyron and the Lère. The flow rate of the Aveyron River
85 frequently drops below the target regulatory threshold, which was established to guarantee aquatic
86 ecosystem health. This low-flow threshold is 4 m³.s⁻¹ and 0.1 m³.s⁻¹ for the Aveyron and Lère Rivers,
87 respectively. Furthermore, the flow of the smaller rivers connected to the Aveyron is also regularly
88 disrupted (Mazzega et al., 2014). In this area, daily river discharge dynamics are strongly determined
89 by irrigation demands and the resulting withdrawals. Flow deficits are managed daily by releasing
90 water from the reservoirs associated with the two main rivers and by the implementation of
91 withdrawal restrictions by regulatory authorities.

92 Our study focuses on the downstream portion of the Aveyron watershed, an area characterized by the
93 Aveyron terraces and their surrounding clay-limestone slopes. The watershed contains an irrigated
94 agricultural landscape which represents 16% of the watershed's surface area but accounts for 80% of
95 its irrigation withdrawals (18 hm³). In total, around 40,000 ha are under cultivation by 1,150 farms, and
96 approximately 40% of this land may be irrigated (according to the French Land Parcel Identification
97 System, 2014). The area mainly hosts large-scale production systems: maize, grains (wheat and some
98 barley), oilseeds (sunflower and a little bit of soybean), and a few protein crops (peas). The hilly slopes

99 are mainly dotted with fields of rainfed grains and sunflower, while the alluvial plain is covered by
100 irrigated, largely monocropped grain and silage maize, which require the most water. There are also
101 fields of fruit crops and seed maize; the latter has a high added agricultural value but requires a
102 consistent water supply.

103 **2.3** Climate data

104 We obtained daily climate data using SAFRAN (8 × 8 km grid) (Vidal et al., 2010). Over the 10-year study
105 period, 2007–2016, the daily mean minimum and maximum temperatures were 9.0°C and 17.2°C,
106 respectively. Mean cumulative annual precipitation and potential evapotranspiration (Penman-
107 Montheith formula) were 757 mm and 849 mm, respectively, leading to a mean annual water deficit
108 of 91 mm. Over the study period, the years 2011 and 2015 were the driest (547 and 673 mm of
109 precipitation, respectively) and were characterized by an annual water deficit of 300 mm or more
110 (Figure 1). Looking exclusively at the low-flow period, from June to September, the years 2009, 2012,
111 2015 and 2016 were the driest—low precipitation resulted in a summer water deficit of 300 mm or
112 more.

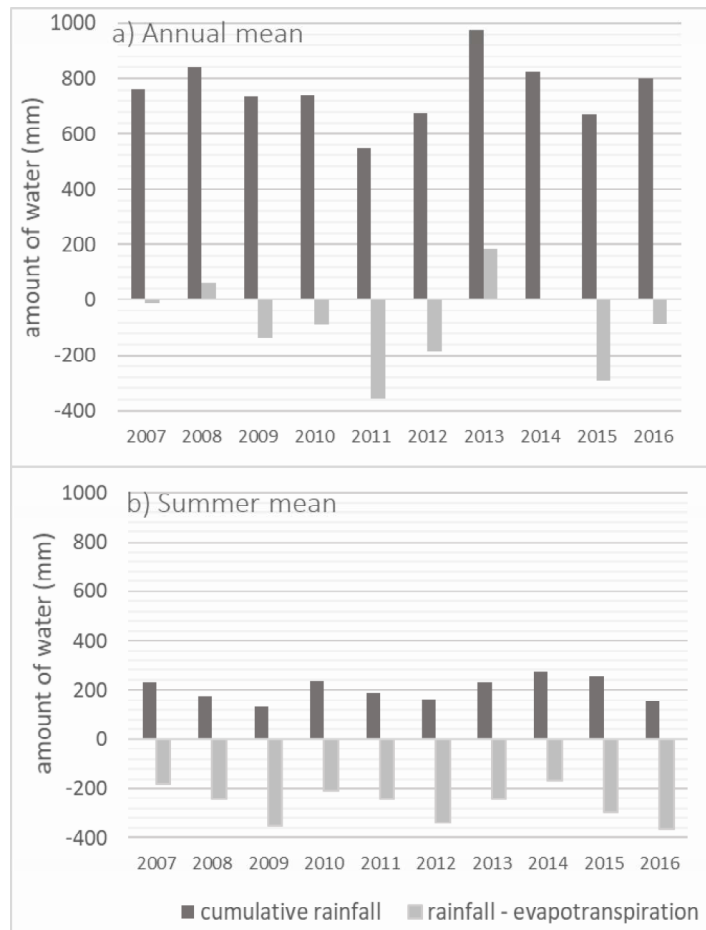


Figure 1. Cumulative rainfall and water deficit levels (rainfall – evapotranspiration) from 2007 to 2016: a) on a yearly basis and b) during the low-flow period (June to September). Depicted are the means for all the SAFRAN data obtained for the downstream Aveyron watershed.

2.4 *The MAELIA model*

2.4.1 *Overview*

MAELIA, a multiagent integrated modeling and assessment platform, was developed to study the environmental and socioeconomic impacts of different combinations of water management and land use strategies (Martin et al., 2016; Murgue et al., 2015, 2014). It employs GAMA® (GIS & agent-based modeling architecture), an open-source generic modeling and simulation environment for building rich, spatially explicit agent-based simulations (Grignard et al., 2013). MAELIA simulates the interactions between four modules:

- 126 (i) the agriculture module, which includes a soil-crop model for simulating plant growth,
127 irrigation, and water dynamics at the field level and a farmer-as-agent model for simulating
128 crop management dynamics
- 129 (ii) the hydrology module, in which different water resources (i.e., rivers, small reservoirs,
130 large reservoirs, and groundwater) are simulated via the soil and water assessment tool
131 (SWAT) (Arnold et al., 2010)
- 132 (iii) the other-water-uses module, in which domestic and industrial water use is simulated via
133 econometric equations (Therond et al., 2014)
- 134 (iv) the water-management module, in which withdrawal restrictions and water-release
135 decisions are simulated

136 The modules interact at different spatial levels at daily time steps; feedback is also incorporated. Field-
137 level water requirements affect farmers' decisions and, consequently, determine irrigation
138 withdrawals from different water resources. There are subsequently impacts on river flow rates, which
139 influence decisions around reservoir releases and/or withdrawal restrictions. The resulting
140 hydrological conditions determine how much water is available for irrigation, which then affects
141 farmers' irrigation levels and crop growth. MAELIA has already been used to explore issues in the
142 downstream Aveyron watershed (Martin et al., 2014; Murgue et al., 2015, 2014). Based on local
143 databases, expert knowledge, and geographic information system (GIS) analysis, each irrigated field in
144 this area is connected to one or more water sources (e.g., river, small reservoir, and/or groundwater).
145 More information about MAELIA can be found in Martin et al. (2016b) and Therond et al. (2014).

146 2.4.2 *Agriculture module and data*

147 Within the agriculture module, the soil-crop model AqYield simulates crop growth, soil water
148 dynamics, and crop yield at the field level. The equations it employs have been previously described
149 (Constantin et al., 2015). Crop development is simulated using the accumulated mean daily effective
150 temperature. AqYield uses a FAO-56 approach to simulate evapotranspiration through a crop
151 coefficient depending on crop development (Allen, 2000). Evaporation and transpiration are daily

152 simulated separately. They are both dependent on a potential, calculated with potential
153 evapotranspiration and crop coefficient, and soil water availability. Soil water dynamics is modelled
154 using a capacity approach. Soil water content levels range from wilting point to 100% field capacity
155 and are calculated for each day by taking the balance of inputs (irrigation, rainfall) to outputs
156 (transpiration, evaporation, drainage, and runoff). The crops represented in the module include the 10
157 species that, when taken together, cover 100% of the irrigated areas: maize, sunflower, wheat, peas,
158 rapeseed, soybean, temporary and permanent grasslands, orchards, and vineyards. Different
159 parameters are used for grain maize and silage maize; there are six earliness categories for grain maize:
160 from very early to very late. AqYield funnels soil and plant indicators to the farmer-agent and
161 hydrological information (field drainage and runoff) to the hydrology module.

162 AqYield simulates events in each field for each crop in the crop successions identified via the French
163 Land Parcel Identification System (Murgue et al., 2016; Martin et al., 2016).

164 In the farmer-as-agent model, crop management strategies are modeled using a set of IF-THEN-ELSE
165 decision rules that represent conditions related to daily soil water content, crop development stage or
166 stress, climate, and/or water resource levels that necessitate technical operations (also simulated),
167 including tillage, seeding, irrigation, and harvesting. They are simulated for each crop based on field
168 (pedoclimatic) and farm characteristics. Farmer-agents carry out technical operations on a daily basis
169 in accordance with these rules, the time available for conducting the work, and the spatial distribution
170 of fields on their farm (i.e., with a view to minimizing distance when changing fields). Each farm and
171 field are individually simulated, taking into consideration their biophysical characteristics (soil, climate)
172 and agricultural characteristics (crop succession, crop management and workload). More details are
173 available elsewhere (Murgue et al., 2014 and Rizzo et al., 2019).

174 Regarding our case study, farmer's crop management strategies, including irrigation strategies, were
175 modeled according to information collected via a dedicated farm survey (see details in Murgue et al.,
176 2016).

177 **2.4.3 Irrigation management**

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2 178 Irrigated fields are organized into islets using the same irrigation materials, resulting in a minimum
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4 179 irrigation rate and volume. This model component makes it possible to simulate how irrigation
5
6 180 materials constrain irrigation dynamics. The fields are watered sequentially based on material
7
8 181 characteristics. Irrigation decision rules determine the amounts and dates of the water supplied for
9
10 182 each irrigated field of each irrigated farm.
11

14 183 **2.4.4 Water-management module**

16 184 The water-management module exploits MAELIA's basic framework, via decision rules, to simulate
17
18 185 regional policies regarding reservoir releases and withdrawal restrictions (Mazzega et al., 2014; Mayor
19
20 186 et al., 2012). These decision rules were determined by obtaining information from regional authorities
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22 187 about their current practices. This module estimates water releases on a daily basis after taking into
23
24 188 consideration the hydrological network and filling levels. In the downstream Aveyron watershed, there
25
26 189 are four dams (Les Falquettes, Saint-Géraud, Thuriès, and Pareloup) whose reservoirs are available to
27
28 190 boost river flow rates. Similarly, the module can simulate withdrawal restrictions using decision rules
29
30 191 concerning flow rates at hydrological nodes and water availability in the reservoirs (ibid.).
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36 192 **2.4.5 Calibration and evaluation**

38 193 Calibration and evaluation of MAELIA models were implemented by previous studies. The AqYield
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40 194 model was calibrated and evaluated on several water balance components through time, for various
41
42 195 crops at the field scale and over a 7-year succession period (Constantin et al., 2015; Tribouillois et al.,
43
44 196 2018). The results highlighted the effectiveness of the model in simulating the different water-related
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46 197 variables, such as soil water content, evapotranspiration, and drainage in irrigated and rainfed
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48 198 systems, with annual crops such as maize, sorghum, sunflower, rapeseed and wheat and on bare soil.
49
50 199 The model was evaluated over crop rotation on daily evapotranspiration measured on Eddy covariance
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52 200 and on drainage measured on lysimeters for several years and with or without crop. The results
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54 201 confirmed that the model was able to simulate evaporation during bare soil period and transpiration
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56 202 during crop development. AqYield predicted monthly ET on calibration and validation datasets with a
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203 model efficiency of 0.84 and 0.69, respectively. It also predicted soil available water content in the
204 various soil and climate situations with an efficiency of 0.83 (Tribouillois et al., 2018).

205 Parameters of irrigation decision rules were calibrated to reproduce observed annual irrigation
206 withdrawal volumes provided by the Adour-Garonne Water Agency for the period from 2003 to 2010
207 (Martin et al., 2016).

208 The hydrology module was calibrated using an approach similar to that of Lardy et al. (2014), in which
209 simulated and observed flow rates were compared for the two main rivers using a daily time step;
210 irrigation withdrawal volumes were also compared. Water release patterns and water restrictions
211 were calibrated using iterative decision rules, drawing upon regional government records from 2008
212 to 2010; this information was used to consistently simulate the volumes released annually.

213 Initial evaluations of model performance concluded that simulated withdrawals at the watershed scale
214 were very satisfactory (Martin et al., 2016; Murgue et al., 2016); they were based on data provided by
215 the Adour-Garonne Water Agency for the period from 2003 to 2010. However, these first evaluations
216 conducted in the previous studies did not provide any statistical performance figures except for an
217 underestimation of 10% on average of the withdrawal. The predictive quality of the model was then
218 entrusted to the appreciation of the field experts in relation to their expertise and knowledge of the
219 expected values and dynamics. They noted a good ability of the model to reproduce annual water
220 withdrawal volumes and interannual differences, particularly in comparison to the observed data,
221 which they also considered not very accurate, with a possible error of 10%. Local experts also validated
222 the results of the simulations of crop management practices (e.g., sowing dates, first and last dates of
223 irrigation). Additionally, they confirmed that, for the same time period, the model accurately
224 reproduced observed flow dynamics during the low-flow season.

225 To further assess the model, we compared simulated and observed values of withdrawal volumes for
226 irrigation and river flow rates over a 10-year period (2007–2016) in the two main rivers (Aveyron and
227 Lère). The observed data were provided by the water agency. Three statistics were calculated:

228 (1) the root mean square error (RMSE), which quantifies the difference between predicted and
229 observed values.

230 (2) the Nash–Sutcliffe efficiency (NSE) coefficient, which was described by Nash and Sutcliffe (1970)
231 and is defined as one minus the sum of the absolute squared differences between the predicted and
232 observed values divided by the variance of the observed values. NSE ranges between $-\infty$ and 1, where
233 the latter is indicative of a perfect fit. When the coefficient value is less than zero, the observed mean
234 is a better predictor than the predicted mean.

235 (3) the mean bias error (MBE), which reflects the average bias in predicted values.

236 These three statistics were calculated as follows:

$$237 \quad RMSE = \sqrt{\frac{\sum_{i=1}^n (E_i - O_i)^2}{n}} \quad (1)$$

$$238 \quad NSE = 1 - \frac{\sum_{i=1}^n (E_i - O_i)^2}{\sum_{i=1}^n (O_i - \bar{O})^2} \quad (2)$$

$$239 \quad MBE = \sum_{i=1}^n \frac{E_i - O_i}{n} \quad (3)$$

240 where E_i = the estimated value, O_i = the observed value, \bar{O} = the mean of the observed values, and n =
241 the number of values.

242

243 2.5 Simulations of the current scenario and three alternative scenarios

244 Maize is the main irrigated crop found in the downstream Aveyron watershed. Consequently, we used
245 MAELIA modeling to compare the performance of the current maize management scenario with that
246 of three alternative scenarios to reduce demand for irrigation water and, thus, the overall water deficit
247 in the watershed:

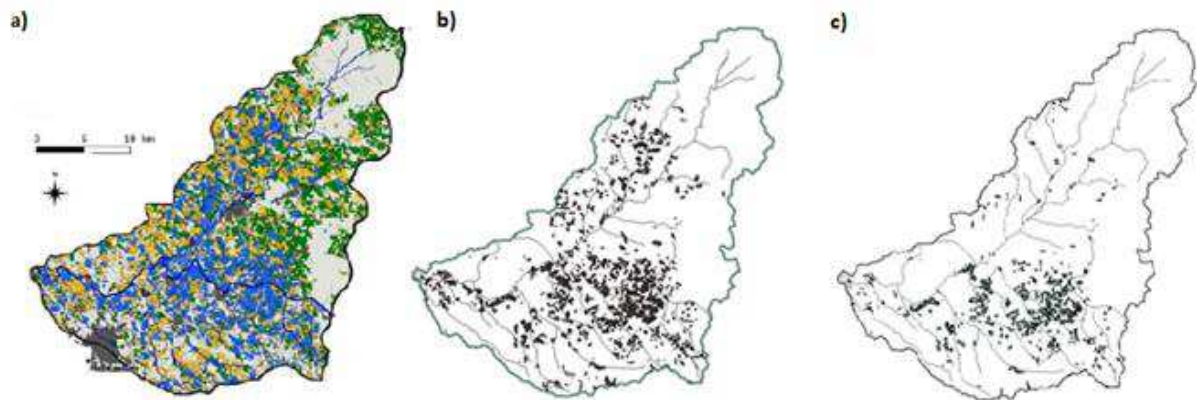
248 i) The **benchmark scenario** represented current management practices

249 ii) The **irrigation efficiency scenario** provided water to maize on an as-need basis

- 250 iii) The **modified crop succession scenario** replaced maize monocultures with a wheat/maize
 251 succession
- 252 iv) The **combined scenario** brought together the irrigation efficiency and succession
 253 modification scenarios

254 In the three alternative scenarios, seed maize production was left unchanged because of the crop’s
 255 economic value. Its importance within the watershed’s economy also means it is guaranteed a
 256 consistent water supply.

257 In these four scenarios, the model simulated withdrawal volumes, drainage, river flow rates, the
 258 number of days where river flow rates fell below the target threshold, and reservoir releases. The
 259 simulations took place over a 10-year period (2007–2016) at the scale of the watershed, in which 1,142
 260 farms with 15,224 crop fields cover a total of 38,880 ha.



261
 262 **Figure 2. Maps of the fields affected by the different scenarios: a) benchmark scenario, b) irrigation**
 263 **efficiency scenario, and c) modified crop succession scenario. For the benchmark scenario, irrigated**
 264 **fields are in blue, rainfed fields are in yellow, and grasslands are in green. For the irrigation**
 265 **efficiency and modified succession scenarios, the surfaces impacted are in dark gray.**

267 2.5.1 *Irrigation efficiency scenario*

268 In this scenario, only the irrigation strategy was changed. All other variables were the same as in the
 269 benchmark scenario, including farm constraints, crop successions, and irrigation islets as well as tillage,
 270 sowing, and harvesting strategies. Instead of irrigation rules based on precipitation and soil moisture
 271 used in the benchmark, irrigation was based on plant water requirements, using the water satisfaction

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272 index (i.e. ratio between actual and potential transpiration) predicted by AqYield. The threshold value
273 for this ratio to trigger irrigation was fixed between 0.8 and 0.9, depending on crop development
274 (Supplementary Material A). This scenario sought to quantify the gain with optimized irrigation utilizing
275 irrigation decision-making tools. This scenario affect approximately 16% of the watershed’s agricultural
276 lands: 2,045 crop fields with at least one maize in crop succession covering 6,365 ha, including 2,820
277 ha dedicated to maize monocultures and 1,670 ha dedicated to wheat/maize successions (Figure 2b).

278

279 *2.5.2 Modified crop succession scenario*

280 The modified crop succession scenario was designed based on recommendations made by local
281 stakeholders (Murgue et al., 2015). Maize monocultures and crop successions containing more than
282 75% maize were replaced by a wheat/maize succession. The hypothesis was that replacing irrigated
283 maize by irrigated wheat would strongly reduce water withdrawals (from about 10 water supplies for
284 maize to 2-3 supplies for wheat) and position them around the wheat flowering that is most often
285 before the intense low-flow season.

286 We randomly assigned the initial crop to have half of the fields starting with maize and half with winter
287 wheat. The wheat was irrigated if necessary since the farmers already have access to irrigation in those
288 fields. To ensure crop compatibility, the maize was the most common variety used in the watershed
289 (mid-late maize). This choice of maturity also ensures that maize could reach physiological maturity to
290 be correctly harvest before wheat sowing. This scenario affected 7% of agricultural lands within the
291 watershed: 706 crop fields and 2,820 ha of maize monoculture successions (Figure 2c). It led to a 25%
292 decrease in the area covered by maize and a 15% increase in the area covered by wheat.

293 *2.5.3 Combined scenario*

294 In the third alternative scenario, the practices from the two previous scenarios were applied in tandem.
295 As in the irrigation efficiency scenario, approximately 16% of the watershed’s agricultural lands (2,045
296 fields covering 6,365 ha) were affected; this figure included 2,820 ha dedicated to maize monocultures

297 that were converted to the wheat/maize succession and 1,670 ha already subject to a wheat/maize
298 succession (Figure 2).

299 *2.5.4 Data analysis*

300 To assess the irrigation efficiency and modified crop succession scenarios at a small scale, we first
301 selected an example field to explore the main processes influencing water management and balance
302 (i.e., evapotranspiration, irrigation, amount of water in the soil, and drainage). This 2.6-ha field hosted
303 an irrigated maize monoculture (semi-late maize) and was located on deep, clay-limestone soil.

304 In a second step, irrigation withdrawals, drainage (i.e., percolation across the rhizosphere) and river
305 flow rate were analyzed at the watershed scale over the 10-year period. They were also analyzed on a
306 yearly basis and/or for each low-flow period (i.e. time of the year when natural river flows are lowest
307 and irrigation occurs in June, July, August, and September), depending on the variable.

308 For withdrawal volumes, we also conducted an analysis on two farms (spanning 24 and 110 ha,
309 respectively) largely dedicated to growing irrigated maize monocultures. We expected that they would
310 show the greatest response to the changes implemented in the three alternative scenarios. The two
311 farms are located on deep soils (clay-limestone and silty soils, respectively).

312 For drainage, we analyzed it at the watershed scale by taking the mean of all the fields as well as for
313 20 target islets (blocks of fields), which contained the largest surface areas dedicated to maize
314 monocultures (mean: 92%; range: 71–100%). Their maize monocultures spanned a mean area of 24 ha
315 (range: 16–41 ha). Like the above farms, the target islets were expected to show the greatest response
316 to the three alternative scenarios. This approach allowed us to analyze the maximum effect of the
317 scenarios on annual drainage. The drainage calculation takes into account the soil depth of each field
318 to account for the soil heterogeneity of the watershed. Depending on the fields, this soil depth varies
319 between 50 and 120 cm with an average of 84 cm.

320 Finally, we evaluated the scenarios' effects on the number of days that river flow rates fell below the
321 low-flow target threshold and, thus, on reservoir releases. The low-flow target threshold is a regulatory
322 standard that helps ensure the coexistence of all uses, the proper functioning of the aquatic

environment, and good-quality water conditions. It is defined by the watershed's water management policies. When the river's flow falls below this threshold, restrictions can be applied to irrigation withdrawals.

326

3 Results

3.1 Evaluation of MAELIA on the actual watershed hydrology

For the study period, the water agency had observed that the mean withdrawal volume was 13.1 hm³ (range: 9.4–15.9 hm³). MAELIA performed well in simulating volumes and interannual variation in total annual withdrawal volumes (RMSE = 1.64 hm³, MBE = -0.78 hm³ and NSE = 0.43). Predictions were poorest for 2011, which was characterized by atypical climatic conditions (Figure 1), namely a dry spring and a rainy July (88 mm). As a result, in real life, significant volumes were withdrawn early on, and no adjustments were made to the wetter summer. Since this withdrawal pattern was not modeled, poorer predictions resulted.

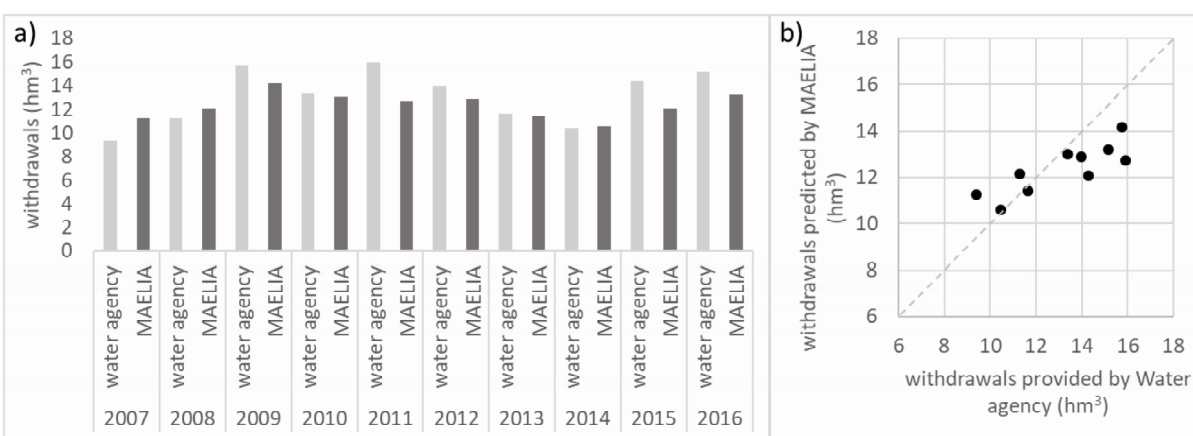
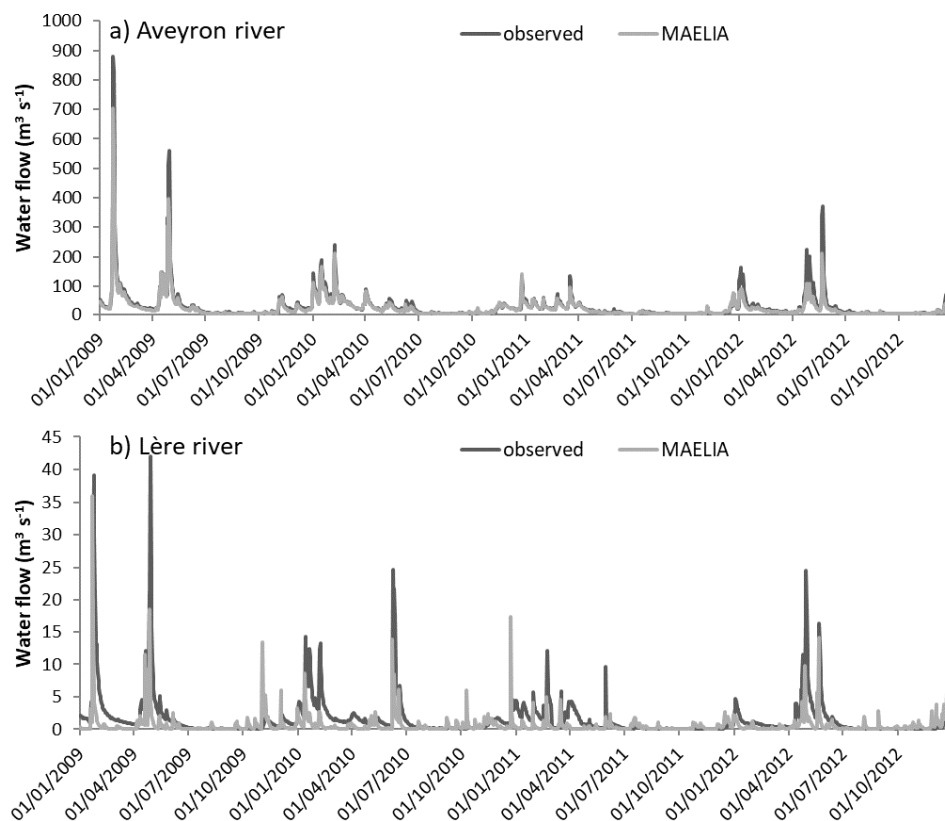


Figure 3. Annual withdrawal volumes observed by the water agency and predicted by the MAELIA model's benchmark scenario

During the study period, the mean monthly flow rate during the low-flow period was 7.4 m³.s⁻¹ for the Aveyron (range: 3.1–19.6 m³.s⁻¹) and 0.7 m³.s⁻¹ for the Lère (range: 0.1–5.3 m³.s⁻¹). The model's results showed that observed flow dynamics were well recreated (Figure 4), especially those during the low-

343 flow period, months of great importance for water management. Indeed, the predicted values highly
 1
 2 344 resembled the observed values, with a mean flow rate of $6.2 \text{ m}^3 \cdot \text{s}^{-1}$ for the Aveyron (range: $2.6\text{--}17.1$
 3
 4 345 $\text{m}^3 \cdot \text{s}^{-1}$) and $0.4 \text{ m}^3 \cdot \text{s}^{-1}$ for the Lère (range: 0.0 to $1.8 \text{ m}^3 \cdot \text{s}^{-1}$). The calculated NSE for the monthly flow
 5
 6 346 rate during the low-flow period for the entirety of the study period was 0.85 for the Aveyron and 0.75
 7
 8 347 for the Lère. This result confirms that the model could effectively represent these flow dynamics, even
 9
 10 348 if high flow peaks appeared to be underestimated and were flatter in the simulations.
 11
 12
 13

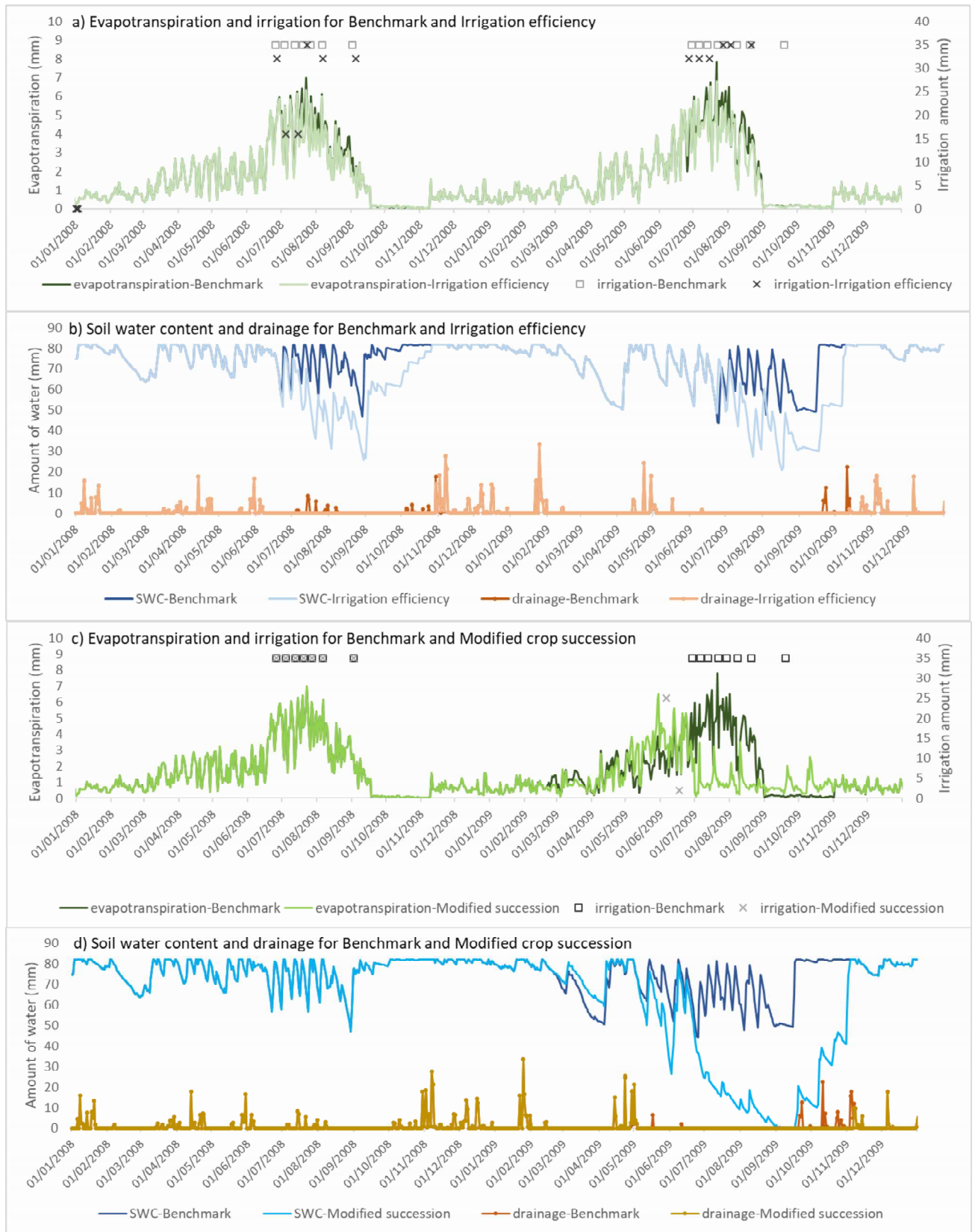


349
 350 **Figure 4. Example of observed and predicted daily flow rates (from 2009 to 2012) for the a)**
 351 **Aveyron River and b) Lère River. Please note that the scale for the y-axis is not the same for the**
 352 **two rivers because of their very different flow volumes.**
 353

354 3.2 Effects of the scenarios on field water balance

355
 356 At the field scale, the irrigation efficiency scenario resulted in evapotranspiration levels similar to those
 357 in the benchmark scenario, which confirms that the customized irrigation strategy did not negatively
 358 impact maize development despite the reduced irrigation levels during the summer (Figure 5a). In the

359 example field, irrigation levels dropped by 80 mm in 2008 and 2009, which corresponds to
 360 approximately two summer irrigation events.



361

362 **Figure 5. Daily evapotranspiration, irrigation, soil water content and drainage dynamics for an**
1 363 **example field with maize grown in 2008 and 2009: a) and b) the benchmark scenario and the**
2 364 **irrigation scenario; c) and d) the benchmark scenario and the modified scenario. "SWC" is soil**
3 365 **water content.**

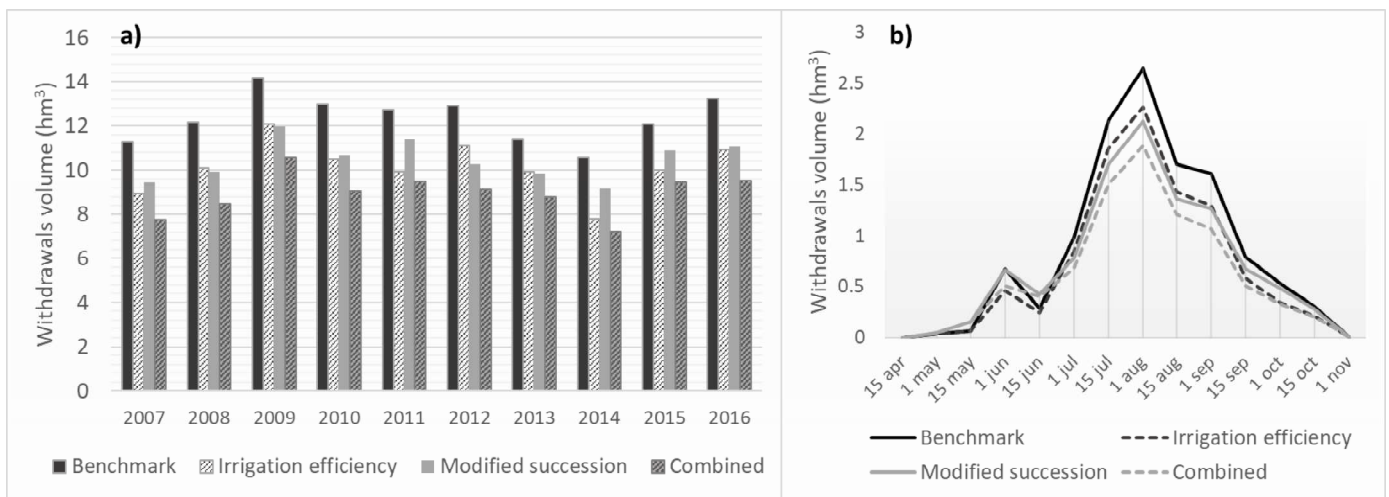
5 366 However, during the low-flow period, soil moisture levels dropped by a mean of 15 mm because
6
7 367 irrigation water was being efficiently taken by the maize (Figure 5b). This fact explains why drainage
8
9 368 declined in the late summer and fall in the irrigation efficiency scenario (by 18 mm in 2008 and by 53
10
11 369 mm in 2009; Figure 5b). In the modified crop succession scenario, evapotranspiration dynamics were
12
13 370 different due to differences between wheat and maize crops, as can be seen in 2009 (Figure 5c). The
14
15 371 earlier growth of wheat results in earlier evapotranspiration and irrigation events (i.e., in May);
16
17 372 however, because this period is wetter and cooler, overall levels are lower than in the summer (i.e.,
18
19 373 August). As a consequence of the wheat crop, spring rainfall almost entirely infiltrates the soil and
20
21 374 departs via evapotranspiration in June. Conversely, maize growth and evapotranspiration peaks in July
22
23 375 and August, resulting in a strong demand for irrigation water. In maize fields, the soil thus remains
24
25 376 moist, and rainfall may enter aquifers or drain into nearby water bodies. In wheat fields in August and
26
27 377 September, soil moisture levels are quite low due to the harvest and high evaporation pressure.
28
29 378 Rainfall is needed to renew soil water content. As a result, drainage declined by 77 mm in fall 2009
30
31 379 (Figure 5d).

380 3.3 Effects of the scenarios on watershed water balance

381 *3.3.1 Irrigation withdrawals*

382 Over the study period, the mean withdrawal volumes were 12.4 hm³ (range: 10.6–14.2 hm³), 10.1 hm³
383 (range: 7.8–12.1 hm³), 10.5 hm³ (range: 9.2–12.0 hm³), and 8.9 hm³ (range: 7.2–10.6 hm³) for the
384 benchmark, irrigation efficiency, modified crop succession, and combined scenarios, respectively
385 (Figure 6a). Compared to the benchmark scenario, the irrigation efficiency and modified crop
386 succession scenarios resulted in significantly lower withdrawal volumes, with an average reduction of
387 18% (annual range: 10–31%) and 15% (annual range: 13–27%), respectively. The combined scenario
388 led to an even greater reduction: 28% (annual range: 22–32%). In absolute terms, there was a mean
389 decrease of 2.2 hm³ in the irrigation efficiency scenario, 1.9 hm³ in the modified crop succession

390 scenario, and 3.4 hm³ in the combined scenario. The three alternative scenarios thus reduced the mean
 391 annual water deficit of the Aveyron watershed (7 hm³) by an average of 32%, 27%, and 49%,
 392 respectively. For the two farms used to explore the maximum potential effects of these scenarios,
 393 withdrawal volumes decreased by an average of 42% for the irrigation efficiency scenario, 35% for the
 394 modified crop succession scenario, and 60% for the combined scenario, relative to the benchmark
 395 scenario; the level of reduction depending on the year (data not shown). As early as May, current
 396 practices resulted in overirrigation relative to crop needs, as reflected in the irrigation efficiency
 397 scenario, resulting in larger withdrawal volumes (Figure 6b). In the modified crop succession scenario,
 398 withdrawal volumes were slightly greater than they were in the irrigation efficiency scenario in the
 399 spring, showing that optimizing irrigation was more efficient than replacing half of the maize area by
 400 irrigated wheat at the watershed scale to reduce withdrawals in spring. By July, this pattern was
 401 reversed because wheat has lower water requirements than maize with need-based irrigation. Finally,
 402 in mid and late September, withdrawal volumes were reduced more in the irrigation efficiency scenario
 403 than in the modified crop succession scenario. The combined scenario allowed for a synergy in these
 404 reductions, particularly during the summer, starting in July.



406
 407 **Figure 6. Irrigation withdrawal volumes in the four model scenarios for the study period from 2007**
 408 **to 2016: a) withdrawal volumes on an annual basis and b) biweekly withdrawal volumes over the**
 409 **irrigation period averaged for the different years**

410 3.3.2 *Drainage from agricultural fields*

1
2 411 Over the study period, mean annual drainage from the agricultural fields was 344 mm (range: 155–556
3
4 412 mm) in the benchmark scenario (Table 1). At the watershed scale, there were minimal differences in
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6
7 413 drainage among scenarios since only 7–16% of the agricultural surfaces were affected by the
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9 414 implementation of alternative practices. We found that annual drainage was reduced, on average, by
10
11 415 0.5% (-1.6 mm.yr⁻¹) in the irrigation efficiency scenario, by 0.2% (-0.7 mm.yr⁻¹) in the modified crop
12
13 416 succession scenario, and by 0.6% (-1.7 mm.yr⁻¹) in the combined scenario. At the watershed scale,
14
15
16 417 these figures translate into mean annual withdrawal reductions of 0.62 hm³ in the irrigation efficiency
17
18 418 scenario, 0.26 hm³ in the modified crop succession scenario, and 0.69 hm³ in the combined scenario.
19
20
21 419 In the irrigation efficiency and combined scenarios, a constant level of reduction occurred every year
22
23 420 that depended less on climatic conditions and more on optimizing maize irrigation. In contrast, there
24
25 421 was more interannual variability in the modified crop succession scenario, given that there was an
26
27 422 increase or no change in drainage in some years (2007, 2014), even if there was a decline overall across
28
29
30 423 the 10-year study period (Table 1).
31
32

33 424 In the 20 target islets used to explore the maximum potential effects of the scenarios, the mean
34
35 425 reduction in drainage was 27 mm.yr⁻¹ (range: 22–35 mm.yr⁻¹) in the irrigation efficiency scenario,
36
37 426 depending on the year (Table 1). Because irrigation was tailored to plant needs, no excess water ended
38
39 427 up in the soil, limiting drainage below the soil surface. In the modified crop succession scenario,
40
41 428 drainage dropped by 7 mm on average, although this amount varied dramatically among years. In 2007
42
43 429 and 2014, drainage actually climbed by 17 mm while, in other years, it was reduced as much as in the
44
45 430 irrigation efficiency scenario (2010: -25 mm; Table 1). This pattern stems from the fact that wheat
46
47 431 requires less irrigation and, therefore, less excess water ends up draining off. Interestingly, 2007 and
48
49 432 2014 had the driest low-flow periods, resulting in the wheat crops being irrigated earlier and at higher
50
51 433 levels at the beginning of the summer. These conditions led to wetter soils, resulting in larger drainage
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53 434 events when significant amounts of precipitation subsequently fell. The combined scenario reduced
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435 drainage by an average of 22 mm.yr⁻¹ (range: 0–43 mm.yr⁻¹), depending on the year. This effect was
 436 intermediate to those of the irrigation efficiency and modified crop succession scenarios.

437 **Table 1. Mean annual drainage (in mm) for the benchmark scenario compared to the three**
 438 **alternative scenarios for the 20 target islets with extensive maize monocultures and for the entire**
 439 **watershed. The values shown for the alternative scenarios are the differences from the benchmark**
 440 **scenario.**

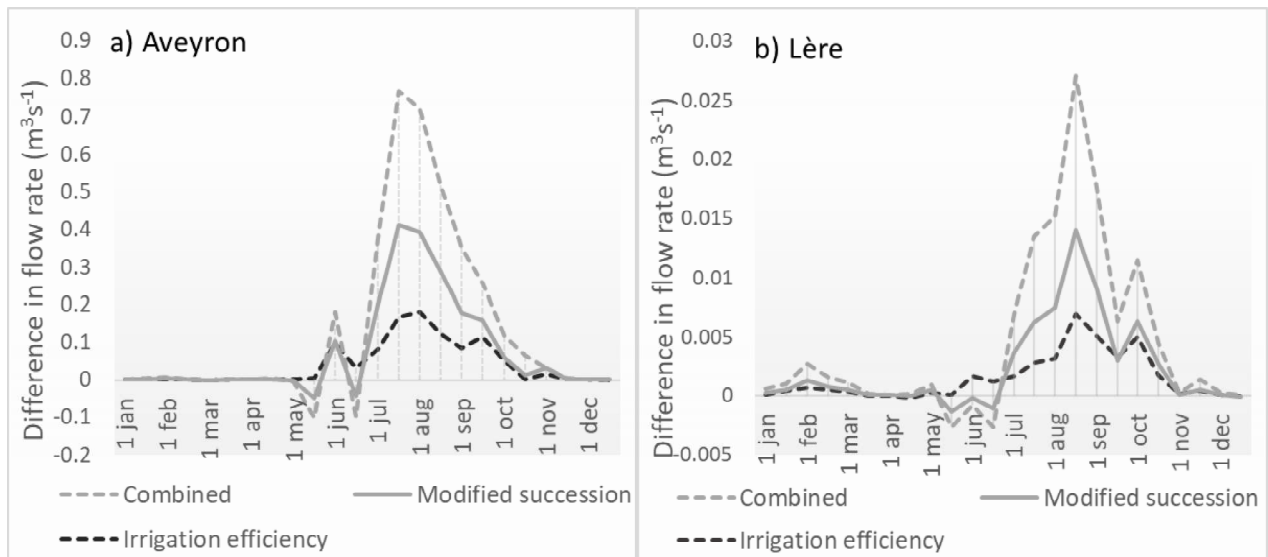
Year	Benchmark	20 target islets			Watershed			
		Irrigation efficiency	Modified crop succession	Combined	Benchmark	Irrigation efficiency	Modified crop succession	Combined
2007	262	-25	12	-1	342	-1.7	0.6	-0.6
2008	308	-26	-5	-18	408	-1.7	-0.5	-1.8
2009	307	-26	-16	-31	350	-1.5	-1.1	-2.1
2010	295	-31	-26	-43	330	-1.8	-1.2	-2.5
2011	81	-22	-3	-19	155	-1.4	-0.6	-1.6
2012	212	-35	-22	-39	288	-1.6	-0.9	-1.9
2013	476	-24	-12	-25	556	-1.1	-1.0	-1.8
2014	295	-31	17	0	391	-2.0	0.0	-1.5
2015	200	-26	-9	-22	260	-1.5	-1.2	-2.3
2016	260	-26	-5	-21	357	-1.7	-0.7	-1.9
Mean	270	-27	-7	-22	344	-1.6	-0.7	-2.0

442 3.3.3 River flow rates

443 In the benchmark situation, the mean daily flow rate in the summer was 6.2 m³.s⁻¹ for the Aveyron and
 444 0.29 m³.s⁻¹ for the Lère. The irrigation efficiency and modified crop succession scenarios had quite
 445 similar effects on river flow rates. Over the entire study period, they tended to increase flow rates
 446 slightly during the low-flow season. For the Aveyron, these mean increases amounted to 0.12 m³.s⁻¹
 447 (+2.2%) and 0.13 (+2.4%) for the two scenarios, respectively (Figure 7). For the Lère River, the increase
 448 was 0.004 m³.s⁻¹ (+1.9%) for both scenarios. In the combined scenario, river flow rates increased even
 449 more: by 0.21 m³.s⁻¹ (+3.9%) for the Aveyron and by +0.008 m³.s⁻¹ (+3.9%) for the Lère.

450 For the entire study period, in the summer, the maximum percent increase in flow rate for both rivers
 451 could be as high as 9% for the irrigation efficiency and modified crop succession scenarios, depending
 452 on the month. For the combined scenario, the maximum percent increase was 10% for the Aveyron
 453 and 12% for the Lère. Even if the irrigation efficiency and modified crop succession scenarios had
 454 similar average effects on summer flow rates, their flow rate distribution patterns differed across the

455 summer months (Figure 6). In the modified crop succession scenario, the flow rate was lower toward
 1
 2 456 the beginning of the irrigation period (May and June) because of increased withdrawals (Figure 6b) for
 3
 4 457 wheat, which starts earlier than maize. The effect on flow rates was clearly stronger in July and August
 5
 6 458 because there is less maize to be irrigated and therefore less water withdrawn from the rivers (Figure
 7
 8 459 6). In the irrigation efficiency scenario, flow rates appeared to be more constant during the summer,
 9
 10 460 probably because irrigation levels were constantly adapted to meet the maize's needs, resulting in less
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 12 461 drainage stemming from overirrigation and less water withdrawn from the rivers in comparison to the
 13
 14 462 benchmark scenario. In the combined scenario, spring flow dynamics resembled those in the modified
 15
 16 463 crop succession scenario, since more land was covered by wheat. However, flow rates experienced a
 17
 18 464 greater boost in July, August, and September because significantly less water was withdrawn since less
 19
 20 465 land was dedicated to maize and the remaining maize was irrigated in an optimized manner. Thus,
 21
 22 466 combining the two approaches significantly increased flow rates more than either individual approach
 23
 24 467 alone (Figure 7).



468
 469 **Figure 6. Differences in the biweekly flow rates between the benchmark scenario and each**
 470 **alternative scenario for the a) Aveyron River and b) Lère River.**

472 **3.3.4 Low-flow target and water restrictions**

473 For almost every year of the study period, the three alternative scenarios slightly reduced the number
 474 of days with flow rates below the low-flow target (Table 2), the threshold that can trigger withdrawal
 475 restrictions. The irrigation efficiency and combined scenarios reduced this metric slightly more than
 476 did the modified crop succession scenario. For the Aveyron, the number of below-target days declined
 477 by 7%, 6%, and 11% for the irrigation efficiency, modified crop succession, and combined scenarios,
 478 respectively. For the Lère, the percent decrease was 5%, 2%, and 3% in the same respective order. The
 479 effect was more pronounced for the Aveyron than for the Lère, likely because the fields most affected
 480 by the scenarios are more commonly found in the former's basin.
 481 Finally, in all three alternative scenarios, the volume of reservoir releases from various locations within
 482 the watershed (see Supplementary Materials B) dropped since demand for irrigation water was
 483 reduced overall and flow rates during the low-water period were improved. Taking the mean for the
 484 four main reservoirs over the entire study period, the percent reduction in release volume was 15%
 485 (range: 0–35%), 17% (range: 1–36%), and 22% (range: 0–45%) for the irrigation efficiency, modified
 486 crop succession, and combined scenarios, respectively (see Supplementary Materials B). There was
 487 marked variability in release volumes because of geographical and administrative differences among
 488 the reservoirs.

489 **Table 2. Number of days between June 15 and September 15 with flow rates below the low-flow**
 490 **target for the benchmark scenario compared to the three alternative scenarios. The values shown**
 491 **for the alternative scenarios are the differences from the benchmark scenario.**

Year	Benchmark	Aveyron			Lère			
		Irrigation efficiency	Modified crop succession	Combined	Benchmark	Irrigation efficiency	Modified crop succession	Combined
2007	1	-1	-1	-1	42	-3	0	-1
2008	20	-2	-1	-3	32	-2	-2	-2
2009	56	-3	-1	-5	67	-3	-2	-3
2010	56	0	-1	-1	52	-2	-2	-3
2011	69	-8	-2	-9	46	-1	-1	-1
2012	58	1	-4	-3	54	-1	-1	-3
2013	13	-3	-4	-4	36	-1	1	1
2014	1	0	0	0	16	0	1	-1
2015	46	-6	-4	-8	33	-1	1	1
2016	28	-3	-2	-4	54	-2	-2	-2
Overall	348	-25	-20	-38	432	-16	-7	-14

492 **4 Discussion**

493 **4.1 All three alternative scenarios improved watershed water management**

494 The three alternative scenarios that we modeled improved water management by reducing the overall
495 water deficit in the watershed. The deficit decreased by around 30% in both the irrigation efficiency
496 and the modified crop succession scenarios and by up to 50% in the combined scenario. There were
497 also slight improvements to river flow rates, an key contributor to efforts to preserve aquatic
498 biodiversity (Bunn and Arthington, 2002). It is striking that implementing just one change to farming
499 practices (i.e., either irrigating maize as needed or shifting to wheat/maize successions) could help
500 eliminate nearly one-third of the current water deficit. Combining the two changes reduced the water
501 deficit by half and seems to be a better overall approach to optimizing water usage in the watershed.
502 The effects are not necessarily additive since it is not possible to systematically apply both practices to
503 a given field (e.g., in a wheat/maize succession, the wheat cannot be optimally irrigated).

504 The alternative scenarios reduced withdrawals, by 18%, 15%, and 28% for the irrigation efficiency,
505 modified crop succession, and combined scenarios, respectively, and slightly boosted river flow rates
506 during the low-flow period. Thus, our simulations showed that the reduced drainage levels attributable
507 to the alternative management strategies did not have negative effects on flow rates. This reduction
508 in drainage can be explained at the field level: the soil was drier in the late summer because the wheat
509 fields were irrigated earlier in the season and the maize fields received optimized irrigation. It is
510 interesting to note that the two individual scenarios yielded relatively similar results, even though
511 there was a two-fold difference in the surface areas affected: 6,400 ha in the case of the irrigation
512 efficiency scenario and 2,800 ha in the case of the modified crop succession scenario. Our findings
513 indicate that irrigating grain and silage maize on an as-need basis with the help of a decision-support
514 tool could help limit withdrawals and ensure higher river flow rates in the summer, while having very
515 little impact on yield; in some cases, there may even be a slight improvement (see Supplementary
516 Materials C). Indeed, new techniques are increasingly used by advisors to optimize the efficiency of
517 allocated irrigation water, so as to give the highest crop production with the least water use. The use

1
2 519 of a tensiometer probe to evaluate the level of soil water deficit combined with a decision support
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4 520 tool, such as MODERATO for example, improves irrigation schedules according to maize crop needs
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7 521 (Bergez et al., 2002). These results are consistent with those of García-Vila and Fereres (2012), which
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9 522 showed that using irrigation levels slightly below full irrigation requirements (i.e., -16%) allow yield to
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11 523 be maintained. However, it is important to note that this effect of improving irrigation efficiency is
12
13 524 dependent on current irrigation efficiency in our case-study and on the capacity of MAELIA to simulate
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15 525 irrigation withdrawals. As we showed that MAELIA well reproduces levels and interannual variation of
16
17 526 observed annual withdrawals, we can conclude that, in our case study, there is an over-irrigation and
18
19 527 so increasing irrigation efficiency is a relevant option to reduce the water deficit in this watershed.

20
21 528 The results for the modified crop succession scenario demonstrated that replacing maize
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23 529 monocultures by a wheat/maize succession had a relatively large effect on irrigation withdrawal
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25 530 volumes even if a relatively small surface area was affected by this change.

26
27 531 The findings we obtained with the irrigation efficiency and modified crop succession scenarios are
28
29 532 consistent with those of a study on a water-deficient watershed in China (Yan et al., 2015). The
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31 533 researchers evaluated a scenario in which regulated deficit irrigation was used to grow maize (a
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33 534 scenario resembling that in our study); this approach reduced maize water demands by about 20%.
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35 535 The researchers also implemented a scenario in which a specific crop succession was adopted to
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37 536 reduce amounts of maize (like in our modified crop succession scenario); the scenario's wheat-maize-
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39 537 maize succession required 12% less water (water needs declined even further if maize was completely
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41 538 eliminated from the crop succession). The conclusion was that these scenarios can improve water
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43 539 management, eliminating up to two-thirds of the water deficit without hurting agricultural production
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1 544 crops with lower water needs in some fields if it ensured that farmers had access to water to properly
2 545 irrigate more profitable crops (García-Vila and Fereres, 2012).
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4
5 546 Despite these challenges, our multifarious results obtained at different scales can guide regional
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7 547 stakeholders and water resource managers who wish to develop integrated land management
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9 548 strategies adapted to local conditions.
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13 14 15 550 **4.2 Farm-scale economic evaluations are needed**

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17 551 An important future direction for this work is to evaluate the economic impacts of the three alternative
18
19 552 scenarios. For the irrigation efficiency scenario, there were negligible impacts or even slight
20
21 553 improvements to maize yields. Furthermore, water management decision-support tools are
22
23 554 inexpensive, supporting the economical acceptability of this scenario (Allain et al., 2018). Several
24
25 555 studies have found that irrigation optimization has little impact on yield and profitability. For example,
26
27 556 Playán and Mateos (2006) concluded that properly implementing irrigation efficiency strategies can
28
29 557 do more to boost economic performance than improvements to irrigation structures.
30
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33
34 558 In contrast, replacing maize monocultures with a wheat/maize succession can decrease farm income
35
36 559 by 9% (ibid.). To compensate for economic losses, it could be useful to deploy agroecological practices
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38 560 such as cover crops, which can provide a suite of ecosystem services with multiple agricultural and
39
40 561 societal benefits (Duru et al., 2015; Wezel et al., 2014; Zhang et al., 2007). Economic losses could also
41
42 562 be circumvented by introducing other crops, especially those with a high added value, into the crop
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44 563 successions. To subsidize the incorporation of more agroecological practices, farmers could receive
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46 564 remuneration for the ecosystem services their farms deliver to society (Redford and Adams, 2009).
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51 565 It is also important to underscore that the three alternative scenarios led to an approximately 20%
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53 566 decrease in reservoir release volumes. The latter represent a significant cost for regional governments
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55 567 (Debril and Therond, 2012). Indeed, Thuriès and Pareloup are hydroelectric dams operated by the main
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57 568 French electricity company (EDF). Thus, any water released from the reservoirs to maintain river flow
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1 569 rates must be purchased by public authorities at a rate of €0.049/m³ (i.e., cost in 2016). Consequently,
2 570 the three alternative scenarios resulted in an overall reduction in reservoir release volumes of 2.4 to
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4 571 3.6M m³, depending on the specific scenario (see Supplementary Materials B), which amounts to
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7 572 €120,000–180,000 in savings.
8

9 573 **4.3** MAELIA modeling delivered satisfactory results

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11 574 MAELIA modeling has advantages, as well as certain limitations. For example, we observed that our
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14 575 simulations tended to slightly underestimate withdrawal volumes and flow rates, mainly outside the
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17 576 low-flow period. However, overall, the model was successful in recreating observed results, which
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19 577 supports its usefulness for comparing the performance of current and potential agricultural practices.
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21 578 Our evaluation of MAELIA modeling performance was carried out using a new dataset that runs
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23
24 579 through 2016. Our findings strengthen and support the results of earlier studies (Martin et al., 2016;
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26 580 Murgue et al., 2016).

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28
29 581 A key advantage of MAELIA models is that they can be developed for different, contrasting French
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31 582 watersheds. With some adjustments, the type of model we employed can be adapted to represent
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33
34 583 most of the constraints found in watersheds, namely those related to soils, climatic conditions, water
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36 584 management strategies, cropping systems, and water use regulations. MAELIA is thus a very powerful
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38 585 tool that should be deployed to explore how this study's scenarios might play out in other regions of
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41 586 France.

42 43 587 **4.4** Broader implications for other watersheds, practices, and environmental issues

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45 588 Our findings are exclusively applicable to the downstream Aveyron watershed, which has specific
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48 589 climatic conditions (e.g., levels of precipitation and evapotranspiration), soil types, cropping systems
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51 590 (e.g., crop succession types and crop management practices), and hydrology. It would be interesting
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53 591 to see the results produced by our three alternative scenarios in other watersheds, to assess how
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55 592 general our conclusions are. Furthermore, our study watershed already experiences large water
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57
58 593 deficits, where particularly dry years can lead to numerous irrigation restrictions. Under climate
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60 594 change, such years are predicted to increase in frequency, and water management problems will

595 become more pronounced (see the introduction). Consequently, we must identify solutions, such as
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2 596 these three alternative scenarios, in order to reduce withdrawals, an essential task if we are to
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4 597 preserve water resources and agricultural systems. Since the three alternative scenarios reduced water
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7 598 percolation, they could have a negative influence on deep water infiltration and groundwater
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9 599 recharge. However, aquifer water quality must also be considered. For example, if the water in the
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11 600 system has lower levels of nitrates and plant protection products, then it might be environmentally
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14 601 beneficial to trade off groundwater recharge quantity for water quality. Moreover, the positive
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16 602 impacts of these scenarios should be enhanced via complementary agroecological practices that
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19 603 provide environmental benefits, such as cover crops or legume-based crop diversification (Constantin
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21 604 et al., 2012; Gregorich et al., 2001; Poeplau and Don, 2015; Tonitto et al., 2006; Tribouillois et al.,
22
23 605 2018a). The next step would be to incorporate nitrogen levels into the model, which could be
24
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26 606 accomplished via the AqYield-N module (Tribouillois et al., 2020), for example. This addition would
27
28 607 allow us to analyze how the three alternative scenarios affect nitrate leaching and thus water quality
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31 608 in the watershed. Finally, incorporating carbon storage and greenhouse gas indicators would also make
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33 609 it possible to comprehensively assess the scenarios' environmental impacts.
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37 38 39 611 5 Conclusion

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41 612 In this study, we demonstrated the usefulness of a multiagent integrated modeling platform that can
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44 613 conduct simulations at fine levels of spatial and temporal resolution. We used this tool to explore how
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46 614 shifts in agricultural practices could affect agrohydrological conditions in the downstream Aveyron
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49 615 watershed, which often experiences large water deficits. By either optimizing maize irrigation
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51 616 efficiency or replacing maize monocultures with a wheat/maize crop succession, it was possible to
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53 617 reduce annual irrigation withdrawal volumes and eliminate around 30% of the watershed's current
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56 618 water deficit; up to 50% of the deficit could be eliminated by combining the strategies. Furthermore,
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58 619 in the combination scenario, there were additive effects: withdrawal volumes declined by 28%, and
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2 621 river flow rates increased by 4%. Given the concurrent challenges of water scarcity and climate change
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4 622 that we are facing at present, it behooves us to adopt solutions that combine several water
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6 623 management practices. If farms are to implement need-based irrigation, they must also equip
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8 624 themselves with a decision-support tool or with probes to estimate the maize's water stress in real
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10 625 time. Therefore, it is more a question of improving the current system's efficiency, which entails few
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12 626 changes in maize management and thus constrains farmers less. Replacing maize monocultures with
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14 627 wheat/maize successions would affect a smaller agricultural surface area, but it would require more
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16 628 profound changes in cropping systems, which may have economic impacts for farmers. However, this
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18 629 scenario represents a first step toward crop diversification and puts farmers on a more agroecological
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20 630 path. It might be possible to compensate farmers for any economic losses with payments for
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22 631 ecosystem services. Expanding on this line of thought, it would be interesting to explore the
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24 632 implementation of more ambitious agroecological practices, like the incorporation of cover crops
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26 633 and/or greater succession diversification, in addition to looking at the scenario's impacts on nitrate
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28 634 leaching, GHG emissions, and/or carbon storage.
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33 635 6 Acknowledgments

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35
36 636 This research was performed as part of the BAG'AGES project and was funded by the Adour-Garonne
37
38 637 Water Agency. The authors gratefully acknowledge Laurène Casal and Renaud Misslin, who helped
39
40 638 develop the modified crop succession scenario. We thank Jessica Pearce-Duvet, who polished the
41
42 639 manuscript via English-language editing.
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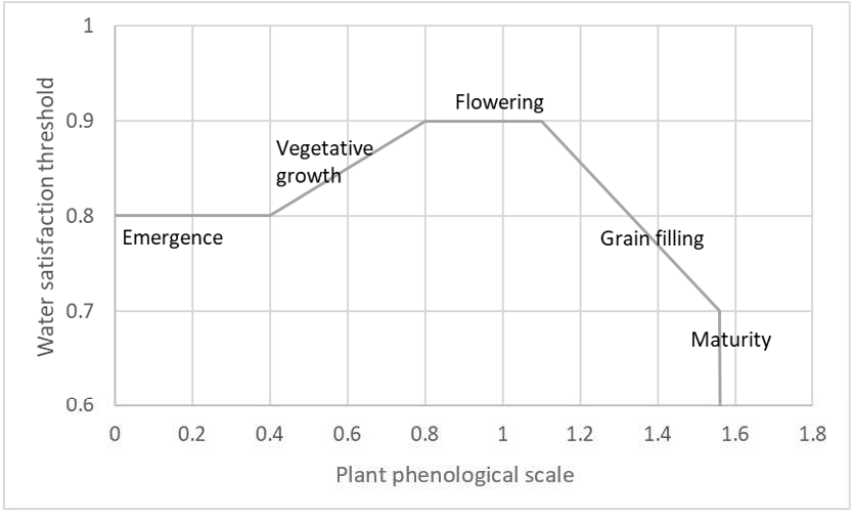
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781 Supplementary Materials A



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783 The water satisfaction threshold as a function of plant phenological stage

Supplementary Materials B

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786 **Reservoir release volumes (m³) for the watershed's four dams in the benchmark scenario versus the**
787 **three alternative scenarios (% difference); the abbreviations are as follows: BM = benchmark**
788 **scenario, IE = irrigation efficiency scenario, MCS = modified crop succession scenario, and CB =**
789 **combined scenario**

Year	Pareloup				Thuriès				Saint-Géraud				Les Falquettes			
	BM	IE	MCS	CB	BM	IE	MCS	CB	BM	IE	MCS	CB	BM	IE	MCS	CB
2007	0	-	-	-	15,849	100%	100%	-100%	0	-	-	-	800,000	0%	0%	0%
2008	0	-	-	-	770,000	0%	0%	0%	2,688,748	-24%	-19%	-41%	698,484	1%	-3%	0%
2009	468,089	100%	100%	-100%	770,000	0%	0%	0%	7,500,000	-4%	-8%	-15%	800,000	0%	0%	0%
2010	1,893,538	-35%	-42%	-66%	770,000	0%	0%	0%	7,500,000	0%	0%	0%	788,948	-1%	0%	-1%
2011	3,500,000	0%	0%	0%	1,100,000	-21%	0%	-30%	8,313,731	-10%	-7%	-10%	800,000	0%	0%	0%
2012	3,241,460	-40%	-37%	-58%	770,000	0%	0%	0%	7,500,000	0%	0%	0%	800,000	0%	0%	0%
2013	0	-	-	-	770,000	0%	-2%	-9%	374,287	-44%	-77%	-100%	800,000	0%	0%	0%
2014	0	-	-	-	0	-	-	-	0	-	-	-	659,004	-5%	-3%	-3%
2015	0	-	-	-	770,000	0%	0%	0%	2,691,686	-19%	-27%	-37%	800,000	0%	0%	0%
2016	1,070,564	0%	0.2%	0.2%	770,000	0%	0%	0%	5,451,314	-3%	-9%	-11%	800,000	0%	0%	0%
Mean	1,017,365	-35%	-36%	-45%	650,585	-13%	-11%	-15%	4,201,977	-13%	-19%	-27%	7,746,436	0%	-1%	0%

Supplementary Materials C

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793

794 **Effects of the benchmark (BM) and irrigation efficiency (IE) scenarios on the yields of different maize**

795 **varieties; the percent difference is for the IE scenario compared to the BM scenario; * = the area-**

796 **weighted effect across varieties**

	Very early maize	Early maize	Mid-early maize	Mid-late maize	Late maize	Very late maize	Mean
BM yield (t.ha ⁻¹)	6.1	4.8	5.6	8.5	10.9	10.7	7.8
IE yield (t.ha ⁻¹)	6.1	4.8	5.6	8.4	10.7	10.4	7.7
Difference (%)	-0.6%	-0.6%	-0.9%	1.4%	2.3%	2.5%	+0.7%
Area covered (ha)	49	168	288	1352	1035	323	*+1.5%

797