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1 **Exposure to polyethylene microplastics alters immature gut microbiome in**  
2 **an infant *in vitro* gut model**

3  
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AhR: aryl hydrocarbon receptor; ASV: amplicon sequence variant; DSC: differential scanning calorimetry; GAPDH: glyceraldehyde-3-phosphate dehydrogenase; GC-MS: gas chromatography-mass spectrometry; HRP: horseradish peroxidase; HPLC: high-performance liquid chromatography; LDPE: low density polyethylene; LY: lucifer yellow; MP: microplastic; MUC2: mucin 2 ; MUC5AC: mucin 5AC; M-ARCOL: mucosal artificial colon; OCLN: occludin; Papp: apparent permeability coefficient; PE: polyethylene; PET: polyethylene terephthalate; PS: polystyrene; qPCR: quantitative polymerase chain reaction; SCFA: short chain fatty acid; SPME: solid-phase microextraction; TEER: trans-epithelial electrical resistance; TGA: thermogravimetric analysis; TJ: tight junction; Tm-ARCOL: toddler mucosal artificial colon; VOC: volatile organic compound; ZO-1: zonula occludens-1

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22

23

24 **Abstract**

25 Infants are characterized by an immaturity of the gut ecosystem and a high exposure to  
26 microplastics (MPs) through diet, dust and suckling. However, the bidirectional interactions  
27 between MPs and immature infant intestinal microbiota remain unknown. Our study aims to  
28 investigate the impact of chronic exposure to polyethylene (PE) MPs on the gut microbiota and  
29 intestinal barrier of infants, using the new Toddler mucosal Artificial Colon coupled with a co-  
30 culture of epithelial and mucus-secreting cells. Gut microbiota composition was determined by  
31 16S metabarcoding and microbial activities were evaluated by gas, short chain fatty acid and  
32 volatolomics analyses. Gut barrier integrity was assessed *via* evaluation of intestinal  
33 permeability, inflammation and mucus synthesis. Exposure to PE MPs induced gut microbial  
34 shifts increasing  $\alpha$ -diversity and abundance of potentially harmful pathobionts, such as  
35 *Dethiosulfovibrionaceae* and *Enterobacteriaceae*. Those changes were associated to butyrate  
36 production decrease and major changes in volatile organic compounds profiles. In contrast, no  
37 significant impact of PE MPs on the gut barrier, as mediated by microbial metabolites, was  
38 reported. For the first time, this study indicates that ingestion of PE MPs can induce  
39 perturbations in the gut microbiome of infants. Next step would be to further investigate the  
40 potential vector effect of MPs.

41

42 **Environmental implication**

43 Plastic production has exploded the last half-century, resulting in environmental pollution and  
44 microplastics (MPs) contamination throughout the food chain. This raised concerns for human  
45 health, particularly in the highly exposed and at-risk infant population. We investigated for the  
46 first time the effects of a chronic exposure to polyethylene MPs on immature infant gut  
47 microbiome and intestinal barrier, using a novel approach combining a newly developed infant  
48 colon model with co-cultures of enterocytes and goblet cells. This work provides useful data

49 on MP interactions with infant microbiota and mucus, considering inter-individualities,  
50 increasing our knowledge on the health impact of MP pollution.

51

52 **Keywords**

53 Microplastics, infant, gut microbiota, *in vitro* gut models, cellular models

54

## 55 1. Introduction

56 Plastic is a multifunctional material, easy to process, cheap and resistant. These valuable  
57 properties have led to massive plastic production, which began in the 1950's and reaching 367  
58 million tons in 2020 [1]. Various types of plastic are produced around the world, but the most  
59 widely manufactured polymer is polyethylene (PE) [1]. Plastic materials are gradually  
60 accumulating in all ecological environments, which leads to major environmental threats [2–  
61 4]. Subject to weathering and (bio)degradation in the environment, plastic waste releases large  
62 amounts of smaller particles that could be sorted by size. Frias and Nash (2019) defined  
63 microplastics (MPs) as “any synthetic solid particles or polymeric matrices, with regular or  
64 irregular shape and with size ranging from 1  $\mu\text{m}$  to 5 mm, of either primary -voluntarily  
65 manufactured- or secondary -released from plastic debris weathering- manufacturing origin,  
66 which are insoluble in water” [5].

67 The ubiquity of MPs throughout the food chain has raised concerns about their potential health  
68 effects following ingestion in humans [6–9]. MPs have been detected in air, water and many  
69 food and beverage [10], as well as in human blood [11], colon [12] and stools [13–15]. Already  
70 in the mother's womb, the fetus seems to be exposed to MPs as they have detected in the  
71 placenta and meconium of newborns [16,17]. After birth, infants are highly exposed through  
72 breathing and ingesting dust during roaming [18], but also during breastfeeding [19] or bottle  
73 feeding, since milk [20,21] and baby bottles [22,23] are contaminated with MPs. In addition,  
74 the suckling of silicone rubber represents a source of exposure as steam sterilization leads to  
75 the release of MPs [24].

76 Despite this critical exposure, toxicological studies investigating the potential impact of MPs  
77 on infant health remain rare [25,26]. While the embryotoxic effect of polystyrene (PS) nano- or  
78 micro-sized plastics seems to vary depending on the model used [27–30], a high dose of PS  
79 MPs injected intraperitoneally into pregnant mice resulted in immune disturbance and

80 spontaneous abortion [31]. Oral exposure to PS MPs also induced fecal microbiota dysbiosis  
81 and gut barrier dysfunction in dams while offspring showed a higher risk of metabolic disorders  
82 [32,33]. Importantly, no studies to date have investigated the bidirectional interactions between  
83 ingested MPs and the immature and developing infant microbiome, despite its key role in gut  
84 homeostasis. Indeed, during the first 1000 days of life, while the digestive physiology is still  
85 under development, the infant intestinal microbiota (which mainly colonizes the colonic  
86 compartment) simultaneously undergoes significant maturation in terms of composition and  
87 metabolic activity [34–36]. The gut microbiome acts primarily in the degradation of undigested  
88 food, resulting in the production of secondary metabolites such as gases, short-chain fatty acids  
89 (SCFAs) [37,38], volatile organic compounds (VOCs) [39] and aryl hydrocarbon receptor  
90 (AhR) ligands [40]. SCFAs (i.e., butyrate, propionate and acetate) are byproducts of the  
91 fermentation of nondigestible dietary fibers under the action of intestinal bacteria, whereas AhR  
92 ligands are primarily derived from the metabolism of tryptophan. Both groups of metabolites  
93 are involved in maintaining the integrity of intercellular tight junction in the intestinal  
94 epithelium, thus linking gut microbial activity to the maintenance of effective intestinal barrier  
95 function. In addition, the gut microbiota plays an important role in the formation of host  
96 immunity, in the regulation of gut endocrine and neurological functions, and together with the  
97 intestinal mucus and epithelium, it forms a protective barrier against exogenous threats such as  
98 pathogens and foreign particles [41–43].

99 In this context, as an extension of a previous study by us on adult microbiota [44], we  
100 investigated for the first time the impact of oral exposure to PE MPs on the infant microbiome  
101 and the intestinal epithelium/mucus barrier. Due to major ethical constraints linked to clinical  
102 studies on young populations, but also the striking differences between human and rodent  
103 digestive physiology -including microbiota- [45], these experiments were carried out using an  
104 innovative *in vitro* approach, combining the Toddler mucosal Artificial Colon (Tm-ARCOL,

105 [46]) simulating the colonic ecosystem specific to infants (6 months to 3 years) and a co-culture  
106 of intestinal epithelial and mucus-secreting goblet cells (Caco-2 and HT29-MTX cells,  
107 respectively). After a 2-week daily exposure to PE MPs, the structure and metabolic activities  
108 of the infant microbiome were determined by 16S metabarcoding and measurements of  
109 gas/SCFAs/VOCs/AhR activity, respectively. Fecal water was then applied to intestinal cell co-  
110 cultures to assess the impact of gut microbe metabolites on cell-related inflammatory pathways,  
111 mucin synthesis and barrier function.

## 112 **2. Materials and methods**

### 113 **2.1 Microplastics characterization**

114 PE MPs (CPMS-0.96, 1-10  $\mu\text{m}$  – 0.2 g; non-fluorescent particles) were purchased from  
115 Cospheric (CA, USA). The initial powder was suspended (2.625 mg/mL) in a solution of sterile  
116 deionized water with 0.01% (w/v) Tween 80 before use. Particle size was determined using the  
117 smileview software from scanning electron microscopy images captured with the JSM 6060LV  
118 microscope (Jeol, Japan). In order to determine the polymeric composition, the powder was  
119 analyzed by differential scanning calorimetry (DSC) and by thermogravimetric analysis (TGA).  
120 DSC measurements were performed on a Mettler DSC3<sup>+</sup> device (N.V. Mettler-Toledo S.A.,  
121 Belgium). The equipment is calibrated with Indium. A few mg of material were placed in an  
122 aluminum pan and then subjected to a temperature cycle ranging from 25°C to 200°C under a  
123 flow of nitrogen gas. A rate of 10°C/min was applied to heat and cool the material during the  
124 experiment. TGA measurements were performed on a Mettler TGA2 device (N.V. Mettler-  
125 Toledo S.A., Belgium). The analysis was performed at a heating rate of 10°C/min under a flow  
126 of a nitrogen gas up to 800°C.

### 127 **2.2 Fecal sample collection and treatment**

128 Fecal samples were collected from four healthy infants (two boys aged 30 and 20 months -  
129 Toddlers 1 and 3, and two girls aged 25 and 22 months -Toddlers 2 and 4-) were collected. All

130 donors had no history of antibiotic treatment or probiotic consumption during the month prior  
131 sample collection. Immediately after defecation, the fecal samples (or diapers) were transferred  
132 to a sterile container placed in an airtight anaerobic box (GENbag anaer gas pack systems,  
133 Biomerieux, France), transported at ambient temperature and processed in the laboratory within  
134 6 h. In an anaerobic chamber (COY laboratories, USA), 11 g of each fresh stool sample were  
135 suspended into 110 mL of 30 mM sterile sodium phosphate buffer and filtered (500  $\mu$ m stainless  
136 steel sieve).

### 137 **2.3 Description and set-up of the colonic Tm-ARCOL model**

138 The Tm-ARCOL is a one-stage fermentation system (MiniBio, Applikon, The Netherlands),  
139 operated under continuous conditions, which simulates both physicochemical (pH, retention  
140 time, ileal effluents supply and anaerobiosis) and microbial conditions (lumen and mucus  
141 associated microbiota) encountered in the colon of young infant [46]. This model is composed  
142 of a main bioreactor (luminal compartment) connected to an external glass compartment  
143 (mucosal compartment) containing mucin-alginate beads (**Fig. 1a**). These beads were prepared  
144 with type II porcine stomach mucin (Sigma-Aldrich, USA) and sodium alginate (Sigma-  
145 Aldrich, USA), as previously described [47]. The mucin-alginate beads were stored at 4°C  
146 before introduction into the hermetically sealed glass compartment (total beads area of 556 cm<sup>2</sup>)  
147 connected to the bioreactor. Tm-ARCOL was set-up, based on *in vivo* data, to reproduce the  
148 average conditions found in the colon of a healthy toddler, as previously described [46]. Briefly,  
149 the composition of the ileal effluents was adjusted to reflect the specificities of the infant diet  
150 (**Fig. 1a**). The temperature of the main bioreactor and mucin-alginate bead compartment was  
151 maintained at 37°C, the average retention time was set at 19 h and the pH was maintained at a  
152 constant value of 6.1 by addition of 2M NaOH. After an initial N<sub>2</sub> flush upon introduction of  
153 the fecal inoculum, anaerobiosis was maintained inside the bioreactor by the activity of the  
154 resident microbiota alone, in order to faithfully reproduce the *in vivo* situation. The mucin-

155 alginate beads were renewed every two days while maintaining the external compartment in  
156 anaerobiosis thanks to a constant flow of CO<sub>2</sub>.

#### 157 **2.4 Experimental design and sampling during *in vitro* fermentations**

158 The experimental design and sampling are summarized in **Fig. 1b**. Four Tm-ARCOL  
159 bioreactors were inoculated with fecal samples from the four infant donors and operated in  
160 parallel. Fermentations were carried out under continuous conditions for 22 days, after a 24-h  
161 batch microbial amplification. An eight-day microbiota stabilization phase was applied and the  
162 three last days of this period (days 6, 7 and 8) were chosen as control. Then, fourteen days of  
163 daily exposure to PE MPs (21 mg of MPs in 8 mL of a solution of deionized water containing  
164 0.01% (w/v) Tween 80) were carried out and the three last days (days 20, 21 and 22) were  
165 chosen as exposure days. Samples from the main bioreactors, termed luminal microbiota, were  
166 collected daily for microbiome characterization (storage at -80°C) and SCFA analysis (storage  
167 at -20°C). Additional samples from the mucin-alginate bead compartment, termed mucosal  
168 microbiota, were collected every other day (when mucin-alginate beads were replaced) for  
169 microbiota characterization. Beads were washed twice in sterile phosphate buffer saline -PBS-  
170 (pH 7.1) and stored at -80°C before downstream analyses. The medium surrounding mucin-  
171 alginate beads was also collected at the same times in case of insufficient amounts of DNA on  
172 the mucin-alginate beads. Additional luminal fermentation media were collected on day 8  
173 (control) and day 22 (exposure) for cell culture experiments, AhR activity measurement,  
174 scanning electron microscopy, Raman spectroscopy and volatolomic analysis. Samples were  
175 also collected daily from the atmospheric phase of the bioreactors to verify anaerobic conditions  
176 and determine gas composition. The daily extra volume of gas produced by microbial  
177 fermentation was also measured using a syringe connected to the gas bag.

## 178 **2.5 DNA extraction**

179 Genomic DNA was extracted from luminal samples and mucin-alginate beads (or surrounding  
180 medium) using the QIAamp Fast DNA Stool Mini Kit (12830-50, Qiagen, Germany) following  
181 the manufacturer's instructions with minor adjustments. Prior to DNA extraction, luminal  
182 samples were centrifuged (2000 g, 10 min, 4°C) and the pellet was mechanically disrupted  
183 using a bead beater (5 min, 20 beat/sec) with 300 mg of sterile glass beads (diameter ranging  
184 from 0.1 to 0.6 mm), incubated (70°C, 5 min) and centrifuged (12000 g, 1 min, 4°C). Mucin-  
185 alginate beads were subjected to the following modifications prior to DNA extraction: 10 min  
186 of incubation with citrate buffer (37°C), as previously described [48], before vortexing  
187 (maximal speed, 3 min) and centrifugation (8000 g, 1 min). DNA integrity was verified by  
188 agarose gel electrophoresis and Nanodrop 2000 analysis (Thermo Fisher Scientific, USA).  
189 DNA quantity was assessed using the Qubit dsDNA Broad Range Assay Kit (Q32851,  
190 Invitrogen, USA) with a Qubit 2.0 Fluorometer (Invitrogen, USA). Samples were stored at -  
191 20°C before gut microbiota analysis (qPCR and 16S metabarcoding).

## 192 **2.6 Total bacteria quantification by qPCR**

193 Total bacteria were quantified by qPCR using primers BAC338R and BAC516F with an  
194 hybridization temperature set at 58°C [49] (**Table 1**). Real-time PCR assays were performed  
195 on a Biorad CFX96™ Real-Time System (Bio-Rad Laboratories, USA) using the Takyon™  
196 Low Rox SYBR® 2X MasterMix blue dTTP kit (B0701, Eurogentec, Belgium). Each reaction  
197 was performed in duplicate in a final volume of 10 µL with 5 µL of Master Mix, 0.45 µL of  
198 each primer (10 µM), 1 µL of DNA sample (10 ng/µL) and 3.1 µL of ultra-pure water. The  
199 amplifications were performed as follows: 1 cycle at 95°C for 5 min, followed by 40 cycles at  
200 95°C for 30 s and 60°C for 1 min. A melting step was added to ensure the specificity of the  
201 primers. The standard curve was generated from 10-fold dilutions of bacterial DNA as  
202 described in [47].

## 203 **2.7 16S metabarcoding and data analysis**

204 The bacterial V3-V4 region of the 16S ribosomal DNA (rDNA) was amplified using primers  
205 V3\_F357\_N and V4\_R805 for bacterial fraction and Arch349F and Arch806R for  
206 methanogenic Archaea. Amplicons were generated using a Fluidigm Access Array followed by  
207 high-throughput sequencing on an Illumina MiSeq system performed at the Carver  
208 Biotechnology Center of the University of Illinois (Urbana, USA).

209 Bioinformatics analysis was performed by GeT-Biopuces platform (INSA/Toulouse  
210 Biotechnology Institute, Toulouse, France) with R software version 4.1.1 (2021-08-10) and  
211 rANOMALY package [50]. The demultiplexed raw sequence data were filtered by quality and  
212 the denoising process performed using DADA2 version 1.20.0 [51]. Reads with N bases or low  
213 phred quality score (under 2) were eliminated and reads under 100 pb length removed.  
214 Decontamination steps were carried out to filter out sequences corresponding to PhiX DNA  
215 used as a spike-in control for MiSeq runs and chimeric sequences were filtered out. Taxonomic  
216 affiliation of all amplicon sequence variants (ASVs) was performed with idtaxa function from  
217 DECIPHER package version 2.20.0 [52] using SILVA release 138 [53] and GTDB  
218 bac120\_arc122 [54] databases (60% bootstrap cut-off). To improve completeness of taxonomic  
219 affiliation, alignments were carried out using BLAST [55] (98% identity and coverage) on  
220 representative sequences of unassigned ASVs or assigned with incomplete taxonomy. A  
221 phylogenetic tree was constructed based on the representative ASVs sequences using functions  
222 from phangorn package version 2.7.1 [56].  $\alpha$ - and  $\beta$ -diversity indexes were calculated using the  
223 diversity\_alpha\_fun function from rANOMALY package [50]. The across-sample microbiota  
224 diversity was described using redundancy analysis (RDA) based on Bray-Curtis dissimilarities  
225 using vegan R-package version 2.5-7 [57]. The impact of the different parameters analyzed  
226 (total gas and SCFA production, gas and SCFA composition, luminal or mucosal sample types,

227 time and sex of donors) on dissimilarities among groups was evaluated using PERMANOVA  
228 permutation tests (999 permutations).

## 229 **2.8 Gas analysis**

230 The analysis of O<sub>2</sub>, N<sub>2</sub>, CO<sub>2</sub>, CH<sub>4</sub> and H<sub>2</sub> gases produced during the fermentation process in the  
231 atmospheric phase of the main bioreactor was performed using a HP 6890 gas chromatograph  
232 (Agilent Technologies, USA) coupled to a micro-TCD detector (Agilent Technologies, USA).  
233 Two series of columns, Molecular Sieve 5A and Porapack Q (Agilent Technologies, USA) were  
234 used. The composition of the gases was determined using calibration curves produced from  
235 ambient air (78.09 % N<sub>2</sub>, 20.95 % O<sub>2</sub>, 0.04 % CO<sub>2</sub>) and 3 gas mixtures A (5 % CO<sub>2</sub>, 5 % H<sub>2</sub>,  
236 90 % N<sub>2</sub>), B (19.98 % CO<sub>2</sub>, 80.02 % H<sub>2</sub>) and C (19.89 % CO<sub>2</sub>, 19.88 % CH<sub>4</sub>, 20 % H<sub>2</sub>, 40.23  
237 % N<sub>2</sub>).

## 238 **2.9 Analysis of short chain fatty acid composition**

239 2 mL of each luminal sample was centrifuged (5000 g, 15 min, 4°C) and 900 µL of the  
240 supernatant was diluted at 1/10 in 0.04 M H<sub>2</sub>SO<sub>4</sub> mobile phase, vortexed and filtered (pore size  
241 0.22 µm). The three major SCFAs (acetate, propionate and butyrate) were quantified by high-  
242 performance liquid chromatography (HPLC) (Elite LaChrom, Merck HITACHI, USA) coupled  
243 to a DAD diode. The HPLC column (150 x 7.8 mm) contained a negatively charged sulfonic-  
244 grafted polystyrene divinylbenzene stationary phase and carried an eluent containing acidified  
245 water. Data were obtained and analyzed by the EZChrom Elite software at 204 and 205 nm.  
246 SCFA concentrations were calculated from calibration curves established from solutions of  
247 known increasing concentration of acetate, propionate and butyrate (0, 10, 25 and 40 mM).

## 248 **2.10 Measurement of AhR activity**

249 The AhR activity of Tm-ARCOL luminal supernatants after exposure to PE MPs (day 22)  
250 relative to control conditions (day 8) was measured using a luciferase reporter assay method,  
251 as previously described [58]. H1L1.1c2 mouse hepatocellular carcinoma cells, containing a

252 stably integrated dioxin response element-driven firefly luciferase reporter plasmid  
253 pGudLuc1.1, were seeded in 96-well plates at  $10^5$  cells/well in Minimum Essential Medium  
254 Eagle Alpha Modification (MEM- $\alpha$ ) medium (with 10% (v/v) heat-inactivated fetal calf serum  
255 -FCS, 1% (v/v) penicillin/streptomycin, 1% (v/v) Geneticin G418) and cultured (37°C, 5%  
256 CO<sub>2</sub>) 24 h before stimulation with the Tm-ARCOL supernatants for 24 h. After incubation, the  
257 wells were washed with 100  $\mu$ L PBS (pH 7), and 50  $\mu$ L of Promega lysis buffer (pH 7.8) was  
258 added to each well. The plates were shaken for 1 h for cell lysis. After addition of 100  $\mu$ L of  
259 luciferase reagent (Promega, France), the luciferase activity was measured using a luminometer  
260 (Tecan, Switzerland). The experiments were performed in triplicate. All values were  
261 normalized for sample cytotoxicity using the Lactate Dehydrogenase Activity Assay (Promega,  
262 USA).

### 263 **2.11 Volatolomic analysis**

264 Volatile organic compounds (VOCs) from control (day 8) and exposed (day 22) luminal  
265 samples were analyzed by solid-phase microextraction (SPME) coupled with gas  
266 chromatography-mass spectrometry (GC-MS), as previously described [59,60]. Briefly, the day  
267 before analysis, 2.2 mL of saturated NaCl solution (360 g/L) was added to each sample (0.8  
268 mL). The mixing vials were closed under a N<sub>2</sub> flow, vortexed for homogenization and thawed  
269 for 24 h at 4°C. The following steps were then performed with an automated sampler (AOC-  
270 5000 Shimadzu, Japan): (i) preheating of the sample to 40°C for 10 min in the shaker (500 rpm),  
271 (ii) SPME trapping (75  $\mu$ m carboxen/polydimethylsiloxane, 23-gauge needle, Supelco, USA)  
272 of the VOCs for 30 min at 40°C, and (iii) thermal desorption at 250°C for 2 min in splitless  
273 mode in the GC inlet. Further analysis of VOCs was performed by GC/MS-full scan (GC2010,  
274 QP2010+, Shimadzu, Japan). VOCs were injected in a DB-5MS capillary column (60 m  $\times$  0.32  
275 mm  $\times$  1  $\mu$ m, Agilent, USA) according to the GC-MS parameters used by Defois and  
276 collaborators [59,60]. Provisional identification of VOCs was performed on the basis of mass

277 spectra, by comparison with mass spectra libraries (Wiley Registry 12th Edition / NIST 2020),  
278 and of retention indices (RI), by comparison with published RI values and with those of our in-  
279 house database. The peak area of the tentatively identified compounds was determined for each  
280 of the targeted molecules using a mass fragment selected for its specificity and lack from co-  
281 elution.

## 282 **2.12 Scanning electron microscopy analysis**

283 The luminal and mucosal samples of the Toddlers 2 and 3, one girl and one boy, were deposited  
284 on SEMPore filters (Jeol, Japan) and fixed for 12 h at 4°C in 0.2 M sodium cacodylate buffer  
285 at pH 7.4 containing 4% (w/v) paraformaldehyde and 2.5% (v/v) glutaraldehyde. The filters  
286 were washed for 10 min in sodium cacodylate buffer (0.2 M, pH 7.4) and post-fixed for 1 h  
287 with 1% (w/v) osmium tetroxide in the same buffer. The filters were then washed 20 min in  
288 distilled water. Gradual ethanol dehydration was performed from 25° to 100° (10 min each) and  
289 completed in hexamethyldisilazane (HMDS) for 10 min. The samples were mounted on stubs  
290 using adhesive carbon tabs and sputter-coated with gold-palladium (JFC-1300, Jeol, Japan).  
291 The analysis was carried out using a scanning electron microscope JSM-6060LV (Jeol, Japan)  
292 at 5kV in high vacuum mode.

## 293 **2.13 Raman spectroscopy mapping**

294 Mapping by Raman spectroscopy was carried out using a Raman spectrometer (Alpha300R  
295 Apyron, Witec, Germany) equipped with a confocal microscope. Exposed luminal samples (day  
296 22, Toddler 3) were vacuum filtered using an Anodisc filter (pore size 0.2 µm, diameter 25  
297 mm). The filters were excited with a Coherent sapphire laser (532 nm, laser power: 10 mW)  
298 and Raman spectra were recorded in backscattering configuration using a 50X objective (Zeiss,  
299 Germany) with a numerical aperture of 0.55 to achieve a theoretical spatial resolution of 590  
300 nm in the focal plane. Individual Raman spectra were recorded 10 times with an integration  
301 time of 1 s. For each filter, maps between 900 and 2500 spectra were recorded with a step of

302 1.33  $\mu\text{m}$  and 2  $\mu\text{m}$ , respectively. For each mapping, the recorded signal (Raman signal plus  
303 background) was integrated over a 20  $\text{cm}^{-1}$  spectral range of centered at 1294  $\text{cm}^{-1}$   
304 corresponding to the  $\text{CH}_2$  twisting vibration band of PE. The signal value was then  
305 automatically transformed by the software in a color scale (from black to bright yellow).

#### 306 **2.14 Intestinal cell lines and cell culture conditions**

307 The human colon adenocarcinoma Caco-2 cell line was obtained from the European Collection  
308 of Cell Cultures (ECACC, UK) and used at passage 58-62. The mucus-secreting colon  
309 adenocarcinoma HT29-MTX cell line was kindly provided by Dr. Thécla Lesuffleur (INSERM  
310 Lille, France) and used at passage 10-16. Both cell lines were maintained in cell culture medium  
311 (Dulbecco/Vogt modified Eagle's minimal essential medium -DMEM- without phenol red and  
312 4.5 g/L glucose), supplemented with 1% (v/v) penicillin/streptomycin, 1% (v/v) Non-Essential  
313 Amino Acids -NEAA-, 1% (v/v) Glutamax, and 10% (v/v) heat-inactivated fetal bovine serum  
314 -FBS- at 37°C in a humidified atmosphere of 5%  $\text{CO}_2$ /95% air. The medium was changed every  
315 two to three days and cells were split at 80% of confluence using Tryple express. The intestinal  
316 cells (90% Caco-2/10% HT29-MTX) were seeded in 24-well inserts with 1- $\mu\text{m}$  pore size  
317 (Millipore, Merck Millipore SAS, France) and maintained in the culture medium. From day 17,  
318 FBS of the culture medium was replaced by 1% (v/v) Insulin Transferrin Selenium (ITS) in  
319 order to limit any interference with the samples to be tested. After 18 days, Caco-2/HT29-MTX  
320 co-cultures were incubated apically for 24 h with Tm-ARCOL luminal samples taken on day 8  
321 (control) and day 22 (exposed), centrifuged (5000 g, 15 min, 4°C) and diluted 20-fold in cell  
322 culture medium (**Fig. 1c**). Experiments were performed with 2 biological replicates and 5  
323 technical replicates for each donor.

#### 324 **2.15 Cell viability**

325 In order to determine the potential cytotoxic effect of Tm-ARCOL luminal supernatants on  
326 Caco-2/HT29-MTX cells and to establish subtoxic conditions for testing on inserts, different

327 dilutions (1/2, 1/5, 1/10, 1/20, 1/30 and 1/50 in cell culture medium) were evaluated. Cell  
328 viability was measured in triplicate with CytoTox 96® non-radioactive cytotoxicity test  
329 (Promega, France) after 24 h of exposure. In 96-well plates, 25 µL of apical cell culture  
330 supernatants were diluted in 25 µL of 1% (v/v) ITS culture medium and mixed with 50 µL of  
331 substrate. The plates were then incubated for 15 min at room temperature in the dark. The  
332 reaction was stopped with 50 µL of stop solution (1M CH<sub>3</sub>COOH) and absorbance at 490 nm  
333 was measured with a Spark spectrophotometer (Tecan, Switzerland).

#### 334 **2.16 Intestinal permeability**

335 Starting on day 12, trans-epithelial electrical resistance (TEER) was continuously monitored in  
336 real time using the CellZScope2 device (NanoAnalytics, Germany). In addition, after a 24-h  
337 contact time with the 20-fold diluted luminal supernatants of Tm-ARCOL, the inserts were  
338 transferred to a new 24-well plate containing 1 mL of Hank's balanced salt solution -HBSS.  
339 Lucifer yellow -LY- (0.4 mg/mL, molecular weight 457 g/mol) and horseradish peroxydase -  
340 HRP- (0.4 mg/mL, molecular weight 44,000 g/mol) were dissolved in HBSS and added to the  
341 apical compartment. After 2 h at 37°C under an atmosphere of 5% CO<sub>2</sub>/95% air, samples were  
342 taken from the apical and basolateral compartments. For LY measurement, samples were  
343 analyzed in duplicate using a Spark multimode plate reader (Tecan, Switzerland). Excitation  
344 and emission wavelengths were 405 nm and 535 nm, respectively. Total HRP was determined  
345 by an ELISA assay. Briefly, 96-well black flat-bottom plates (Greiner, Dutcher, France) were  
346 coated overnight at 4°C with 50 µL of 10 µg/mL HRP mouse polyclonal (Abcam, France) in  
347 PBS. Plates were blocked with PBS-1% bovine serum albumin (BSA) before incubation with  
348 apical and basolateral samples collected from inserts. Rabbit polyclonal anti-HRP biotin  
349 (Abcam, France) was added at a concentration of 10 µg/mL before adding Fluorescein-5-  
350 isothiocyanate -FITC- conjugated streptavidin (Becton Dickinson, France) for 20 min and  
351 fluorescence intensity was measured at 485 nm/525 nm using a Spark microplate reader (Tecan,

352 Switzerland). The apparent permeability coefficient ( $P_{app}$  in cm/s) was determined by the  
353 following equation:  $P_{app} = C_b / (S \times C_0)$ , where  $C_b$  represents the level of accumulation of LY or  
354 HRP in the basolateral compartment after 2 h,  $S$  represents the surface of the membrane (0.33  
355 cm<sup>2</sup>) and  $C_0$  is the initial concentration of the marker (0.4 mg/mL -LY- or -HRP-) in the apical  
356 compartment.

### 357 **2.17 Interleukin-8 level measurement**

358 The apical and basolateral media of the co-cultures were collected after the contact time of 24  
359 h with the 20-fold diluted luminal supernatants of Tm-ARCOL and stored at -20°C until  
360 analysis. Interleukin-8 (IL-8) levels were measured using the Quantikine Human IL-8 ELISA  
361 kit (DY208-05, R&D Systems, USA) following the manufacturer's instructions. Briefly, in  
362 Maxisorp 96-well plates (Nunc, France), biotinylated Goat polyclonal anti IL-8 was added at a  
363 concentration of 10 ng/mL before adding HRP-conjugated streptavidin (Becton Dickinson,  
364 France) for 20 min. Absorbance intensity was measured at 450 nm and 540 nm using a Spark  
365 microplate reader (Tecan, Switzerland).

### 366 **2.18 Intestinal cell gene expression**

367 Total RNAs were extracted from cells for each insert using the AllPrep RNA Mini kit (Qiagen,  
368 France) according to the manufacturer's instructions. Quantity and quality of extracted RNA  
369 samples were evaluated using Nanodrop (Nanophotometer Implen-Dutscher, France) and  
370 Bioanalyzer (Agilent Technologies, USA), respectively. RNAs were reverse transcribed using  
371 enzyme IScript reverse transcription supermix (Biorad, France). Sample concentration was  
372 adjusted to 50 ng/μL. 384-well plates were filled by an Agilent Bravo Automated Liquid  
373 Handling Platform (Agilent Technologies, France). All wells contained 5 μL of the following  
374 mix: 2.5 μL of IQ SYBR green Supermix (Biorad, France), 1.5 μL of each primer set and 1 μL  
375 of cDNA sample. Amplification was performed using a ViiA7 Real-Time PCR System  
376 (Applied Biosystems-ThermoFisher Scientific, France). Thermal cycling conditions were as

377 follows: 3 min denaturation at 95°C followed by 40 cycles at 95°C for 15 s and 45 s at 60°C,  
378 and a melting curve step. The primer sequences of the targeted genes (genes encoding tight  
379 junction (TJ) proteins (*ZO-1*, *OCLN*) and genes involved in mucus synthesis (*MUC2*,  
380 *MUC5AC*) are listed in **Table 1**. Raw data that passed quality control were analyzed with  
381 LinRegPCR (version 2021.2) and then normalized against the expression of glyceraldehyde-3-  
382 phosphate dehydrogenase (*GAPDH*), a reference housekeeping gene.

### 383 **2.19 Statistical analysis**

384 Statistical analyses on gut microbiota activity (gas, SCFAs, AhR activity),  $\alpha$ -diversity indexes  
385 (number of observed ASVs and Shannon index) from metabarcoding data, as well as results  
386 from cell culture experiments, obtained in control and exposed groups, were performed using  
387 GraphPad Prism software version 9.3.1 for Windows (GraphPad Software, USA). Data normal  
388 distribution was verified by combining Anderson-Darling, D'Agostino & Pearson, Shapiro-  
389 Wilk and Kolmogorov-Smirnov tests and homoscedasticity was checked using the Fisher test.  
390 Then, appropriate statistical analysis was applied (either one-way ANOVA, Kruskal-Wallis or  
391 Welch's tests) and significance was considered for  $p < 0.05$ . For gut microbiota results,  
392 differential analyses (DESeq2, metagenomeSeq, metacoder) were performed on  
393 ExploreMetabar software using rANOMALY package [50,61]. Principal coordinate analysis  
394 (PCoA) of unweight UniFrac distances was performed and significance between groups was  
395 assessed with a permutational multivariate analysis of variance (PERMANOVA) using  
396 ADONIS (999 permutations) with non-parametric tests. Volatolomics data were processed  
397 using the Statistica Software (v.13) (StatSoft, France). Student's t-tests ( $p < 0.05$ ) were carried  
398 out for the VOC abundances determined in control (day 8) and exposed (day 22) groups and  
399 principal component analyses (PCA) were performed on the discriminant VOCs selected to  
400 visualize the structure of the data.

## 401 3. Results

### 402 3.1 Characteristics of PE MPs and interactions with the infant gut microbiota

403 DSC and TGA analysis of PE MPs showed the purity of the PE powder and characterized it as  
404 low-density PE. PE MP microscopy count indicated a size between 1 and 2  $\mu\text{m}$  for 50% of 156  
405 beads analysed [44]. Scanning electron microscopy of samples collected in the luminal phase  
406 and mucosal beads in Tm-ARCOL (**Fig. 2a**) showed colonization of the surface of PE MPs by  
407 various bacteria. In addition, Raman spectroscopy mapping of PE MPs in the luminal medium  
408 (**Fig. 2b**) after exposure provided the same spectrum as that obtained for the initial PE particles,  
409 indicating the absence of physicochemical changes to the MP surface during fermentation in  
410 Tm-ARCOL.

### 411 3.2 Impact of exposure to PE MPs on the infant gut microbiota composition

412 The stool microbiota of toddlers used for Tm-ARCOL inoculation was characterized by 16S  
413 metabarcoding (**Fig.S1**) and showed an absence of methanogenic Archaea and bacterial profiles  
414 highly donor-dependent, particularly at the family level. Notably, the stool of Toddler 1 was  
415 characterized by a high abundance of *Tannerellaceae*, while *Enterobacteriaceae* and  
416 *Akkermansiaceae* were found mainly in Toddler 2 and *Sutterellaceae* and *Oscillospiraceae* in  
417 Toddler 4.  $\alpha$ -diversity was also donor-dependent (data not shown), with the highest values for  
418 Toddlers 1 and 4 (107 and 145 ASVs, respectively) and the lowest for Toddlers 2 and 3 (54 and  
419 70, respectively). Some of these stool individual specificities was captured in Tm-ARCOL,  
420 such as *Tannerellaceae* in Toddler 1 and *Enterobacteriaceae* in Toddler 2 (**Fig. 3a**) and a  
421 clearly higher number of ASVs under control conditions of Toddler 4 relative to other toddlers  
422 (**Fig. 4a**). At the end of the stabilization phase in Tm-ARCOL (control), the bacterial abundance  
423 profiles at the family (**Fig. 3a**) and phylum (**Fig. S2a**) levels clearly indicated that the inter-  
424 individual variability observed in the initial donor stools was retained in the *in vitro* model, with  
425 the donor effect being the most important factor explaining luminal and mucosal  $\beta$ -diversity

426 (Fig. S2b). These data also highlighted clear differences between lumen and mucus-associated  
427 microbiota and, as expected, their stability at the end of the stabilization phase. After 14 days  
428 of daily exposure to PE MPs, clear shifts in bacterial populations were observed both in the  
429 luminal and mucosal compartments, and both at the phylum (Fig. S2a) and family (Fig. 3a)  
430 levels. Those changes were again donor-dependent. For instance, exposure to PE MPs resulted  
431 in disappearance of *Verrucomicrobia* (*Akkermansiaceae*) in Toddler 2 whereas opposite trends  
432 were seen in Toddler 4 and *Synergistetes* (*Dethiosulfovibrionaceae*) increased only in Toddler  
433 3. Exposure to PE MPs also induced in the luminal phase a loss in *Ruminococcaceae* in Toddlers  
434 1 and 2 while *Rikenellaceae* and *Sutterellaceae* decreased in Toddlers 2 and 4. In addition, the  
435 population of *Tannerellaceae* population, which was only present in Toddler 1, disappeared  
436 from the mucosal phase after exposure to MPs whereas *Enterobacteriaceae* increased in mucus-  
437 associated bacteria from Toddlers 2 and 4. Differential analysis of gut microbiota composition  
438 in all donors confirmed significant changes in bacterial populations. In the luminal phase,  
439 abundance of *Dethiosulfovibrionaceae*, *Planococcaceae*, *Enterobacteriaceae*, *Moraxellaceae*,  
440 *Acidaminococcaceae* and *Oscillospiraceae* increased significantly ( $p < 0.05$ ,  $n = 4$ ) after exposure  
441 to PE MPs, while the proportion of *Tannerellaceae*, *Rikenellaceae* and *Monoglobaceae*  
442 decreased (Fig. 3b).

443 Fewer populations were affected in the mucosal compartment, with an increased abundance of  
444 *Dethiosulfovibrionaceae*, *Planococcaceae*, *Acidaminococcaceae* and *Oscillospiraceae* but a  
445 loss in *Lachnospiraceae* and *Monoglobaceae* after exposure to PE MPs. The impact of such  
446 exposure on microbiota structure was confirmed by  $\beta$ -diversity analysis, which showed  
447 clustering between control and exposed samples (Fig. 4c), with a greater impact in Toddlers 1,  
448 2 and 3 compared to Toddler 4. Accordingly, MP treatment also resulted in a significant  
449 increase in observed ASVs ( $p < 0.01$ ) in the luminal microbiota only for Toddlers 1, 2 and 3.  
450 Similarly, in mucosal samples, the number of observed ASVs tended to increase in Toddlers 1

451 and 2, and significance was reached in Toddler 3 ( $p<0.05$ ). Finally, RDA analysis of bacterial  
452  $\beta$ -diversity, based on both 16S metabarcoding and metabolic activity data (see 3.3), indicated  
453 that exposure to PE MPs remained the most important parameter explaining variabilities.

### 454 **3.3 Impact of exposure to PE MPs on the metabolic activities of human infant** 455 **microbiota**

456 In parallel to microbiota composition, the impact of exposure to PE MPs was also monitored  
457 *via* gut microbial activities, first by monitoring fermentation gas and SCFAs daily (**Fig. S3**).  
458 Similarly to the microbiota composition described above, this analysis showed stable gas and  
459 SCFA profiles for each donor at the end of the stabilization period, with the exception of  
460 Toddler 3 due to technical problems. Exposure to PE MPs had very limited impact on total gas  
461 and SCFA production (**Fig. 5a** and **5b**). Regarding the production of major SCFAs (**Fig. 5c**),  
462 acetate tended to increase after exposure to PE MPs for Toddler 1, 2 and 3, while the opposite  
463 trend was observed for propionate. Butyrate production dropped after exposure to PE MPs for  
464 all donors with significance reached only for Toddlers 1 ( $p<0.0001$ ) and 4 ( $p<0.05$ ), and when  
465 data were averaged ( $p<0.0001$ ,  $n=4$ ). The ability of the gut microbiota to produce ligands  
466 capable to activate AhR was also determined in Tm-ARCOL supernatants after exposure to PE  
467 MPs (day 22) compared to control conditions (day 8) (**Fig. 6**). After exposure to PE MPs,  
468 luminal contents tended to have a lower ability to activate AhR in all donors except Toddler 2,  
469 suggesting a potential defect in the microbiota to produce metabolites known to activate this  
470 receptor involved in the regulation of intestinal homeostasis [40,62]. Finally, an additional  
471 analysis was performed on the microbial volatolome from the luminal phase (**Fig. 7**). 184 VOCs  
472 were detected in control and exposed samples (data not shown). The 26 discriminating VOCs  
473 between the two groups (**Table 2**) are hydrocarbons and oxygenated compounds (alcohols,  
474 aldehydes, ketones and esters). Volatolomic profiles based on these VOCs indicated a profound  
475 effect of exposure to PE MPs (**Fig. 7a**). The abundances of the 8 most discriminant VOCs ( $p$

476 0.001) are shown in **Table 2** and in **Fig. 7b**. In particular, the 6 hydrocarbon compounds  
477 (pentane, 3-methyl-; cyclopentane, 1,1,3-trimethyl-; 1-heptene, 2,4-dimethyl-; heptane, 2,3-  
478 dimethyl-; methyl C9-alkane - RI 814, methyl C9-alkane - RI 821 and methyl C9-alkane - RI  
479 861), two heptanone (2-heptanone and 2-heptanone, 4-methyl-), two alcohols (1-hexanol and  
480 1-heptanol) and two aldehydes (heptanal and octanal) were overproduced while 8 out of 12  
481 esters (among which propanoic acid, ethyl ester and butanoic acid) were underproduced after  
482 exposure to PE MPs.

### 483 **3.4 Impact of exposure to PE MPs on intestinal epithelium inflammation, permeability** 484 **and mucin synthesis as mediated by gut microbiota metabolites**

485 In order to explore whether alterations in the metabolic activities of the gut microbiota induced  
486 by exposure to PE MPs have consequences on the intestinal barrier, we determined the impact  
487 of Tm-ARCOL luminal supernatants on Caco2/HT29-MTX cells. The preliminary toxicity  
488 assay on intestinal cells with diluted Tm-ARCOL luminal supernatants showed no cytotoxicity  
489 for 1/20 and 1/50 dilutions (data not shown). Therefore, in the following cell experiments, the  
490 1/20 (v/v) dilution was used. Various parameters were measured to assess the impact of Tm-  
491 ARCOL luminal supernatants on trans- and para-cellular permeability. Gene expression of tight  
492 junction proteins, *OCLN* and *ZO-1*, was not significantly altered by Tm-ARCOL luminal  
493 supernatants of the exposed microbiota in Toddler 1, although a clear decreasing trend was  
494 observed for both genes in Toddler 1 (**Fig. S4a**). Trans-epithelial electric resistance ( $\Delta$ TEER)  
495 (**Fig. 8a**) as well as the translocation of Lucifer yellow (LY) (**Fig. 8b**) and horseradish  
496 peroxidase (HRP) (**Fig. 8b**), as indicators of para- and trans-cellular permeability, respectively,  
497 were unaffected by Tm-ARCOL luminal supernatants of the exposed microbiota. Interestingly,  
498 however, HRP and LY translocation tended to increase in Toddler 1, consistent with decreased  
499 *ZO-1* and *OCLN* gene expression. Finally, the expression of mucin-related genes, *MUC2* and  
500 *MUC5AC*, was measured without demonstrating any significant effect of Tm-ARCOL luminal

501 supernatants of the exposed microbiota (**Fig. S4b**). Nevertheless, *MUC2* gene expression  
502 tended to increase in 3 of the 4 donors (not detected in Toddler 4), while *MUC5AC* gene  
503 expression appeared to be slightly overexpressed in Toddlers 1 and 3 only. The inflammatory  
504 marker IL-8 (**Fig. 8c**) was also measured and the levels (70-100 pg/mL), detected in the apical  
505 and basolateral compartments, were not significantly impacted by the Tm-ARCOL luminal  
506 supernatants of the exposed microbiota.

#### 507 **4. Discussion**

508 Ingestion of MPs is increasingly recognized as a threat for human health. However, to date  
509 available data on the behavior of MPs throughout the human digestive tract are very limited,  
510 especially in the infant population. In this study, we showed for the first time, using an *in vitro*  
511 colonic model of toddler, that oral exposure to PE MPs alters the composition of gut microbiota  
512 by increasing potentially pathogenic populations of *Dethiosulfovibrionaceae*,  
513 *Enterobacteriaceae*, *Oscillospiraceae* and *Moraxellaceae* while decreasing *Monoglobaceae*.  
514 These changes in microbial populations have been associated with alterations in metabolic  
515 activity, as shown by altered VOC profiles and decreased butyrate production.

516 As an alternative to clinical trials, most of the few studies on the fate of MPs under human  
517 simulated digestive conditions have been conducted in *in vitro* models of the upper  
518 gastrointestinal tract, i.e. in gastric and small intestinal systems [63–66]. These studies  
519 demonstrated that artificial digestive fluids were not able to degrade MPs of five different  
520 polymers types (including PE) into smaller fragments or change their shape, size or texture, but  
521 described the formation of a biomolecular corona onto the surface of the particle surface  
522 [63,64]. Given this lack of physicochemical degradation of MPs in the upper gastrointestinal  
523 tract, in the present study we focused on interactions in the lower gut between PE MPs and  
524 microbiota, another major component of human gut homeostasis. Yet, only Tamargo and  
525 colleagues have studied interactions between human-derived gut microbiota and MPs, but using

526 polyethylene terephthalate -PET- and not PE, also of different size and shape ( $160 \pm 110 \mu\text{m}$ ,  
527 irregular morphology) [66]. We also recently performed a study in the Mucosal Artificial Colon  
528 (M-ARCOL) where we assessed the impact of repeated exposure to PE MPs on the composition  
529 and activity of the adult gut microbiota [44]. Here, as an extension of this work and in order to  
530 be as close as possible to the human situation, we have chosen to evaluate in Tm-ARCOL the  
531 effect of a physiological dose of MPs at 21 mg/day according to recent exposure data [67]. PE  
532 MPs were administrated daily for 14 days to reproduce chronic exposure, while Tamargo and  
533 colleagues evaluated the effect of a single dose of 166 mg PET MPs for 72 h. Finally, their  
534 study was carried out in adult digestive conditions whereas we have focused here on toddlers,  
535 due to the specificities of their digestive physiology and gut microbiota, which makes them a  
536 high-risk population regarding the ingestion of xenobiotics [13,26,68]. During the critical  
537 period of the first 1000 days, the gut microbiota of infants remains immature in terms of  
538 composition and functionalities [35]. This microbiota immaturity has been successfully  
539 reproduced in the Tm-ARCOL model [46], with lower microbial diversity, higher abundance  
540 of infant-related microbial populations (e.g. *Bifidobacteriaceae*, *Enterobacteriaceae*), higher  
541 acetate production and lower butyrate production in toddlers than in adults, consistent with *in*  
542 *vivo* data [69–71]. Repeated addition of PE MPs resulted in an increased  $\alpha$ -diversity in Tm-  
543 ARCOL, as previously reported in mouse fecal microbiota by Li and colleagues after 5-week  
544 exposure to MP PEs of 10-150  $\mu\text{m}$  in size [72]. This could be explained by the fact that PE MPs  
545 can provide a surface for the formation of biofilm-like microbial communities, as observed here  
546 by scanning electron microscopy and previously described with PET MPs in another human  
547 colon model [66]. The presence of MPs as providers of adhesion sites but also new carbon  
548 sources for microbes has been already described in an environmental context [73–75]. The  
549 second option does not seem feasible here since no degradation of MPs was observed using  
550 Raman spectroscopy analysis. In contrast to our results, Tamargo and colleagues reported a loss

551 of observed ASVs in the transverse and descending colonic compartments following a single  
552 addition of PET MPs [66], suggesting that the effect of MPs on human gut microbiota diversity  
553 would be, in addition to the exposure scenario, plastic-dependent (i.e. polymer type, size,  
554 shape). In our previous study of the adult microbiota, no change in bacterial diversity was  
555 observed after exposure to PE MPs, suggesting that the immature infant gut microbiota may be  
556 more sensitive to MPs than the mature microbiota [44]. When mice were exposed to PE MPs,  
557 the microbial shifts observed in fecal samples were study-dependent and not comparable to the  
558 changes observed in Tm-ARCOL [72,76], certainly due to the large discrepancies between  
559 human and rodent digestion and microbiome [45]. Interestingly, the human *in vitro* study by  
560 Tamargo and colleagues [66] on the exposure to PET MPs showed an increase in  
561 *Proteobacteria* and *Synergistetes* (phylum of *Dethiosulfovibrionaceae*) as observed in our  
562 study, suggesting that these populations could be considered as markers of exposure to MPs in  
563 the human gut microbiota. This hypothesis is reinforced by our recent results on the effect of  
564 PE MPs on the adult gut microbiota showing an increase, both in the lumen and mucus-  
565 associated microbiota, of *Enterobacteriaceae* and *Dethiosulfovibrionaceae* [44]. For the first  
566 time, we also showed in Tm-ARCOL that the impact of MPs can be different between the lumen  
567 and mucus-associated microbiota. This is particularly interesting as mucosal microbes appear  
568 to play an increasing role in human health and disease [77]. In addition, the size of the MPs  
569 seems to have a major role on gut microbiota shifts. The use of beads of the same size as in our  
570 study (1 to 10  $\mu\text{m}$ ) but coming from a different polymer (PS and polyvinyl chloride –PVC-) [78–80]  
571 [78–80] induced more similar effects on mouse fecal microbes (e.g. increase in  
572 *Enterobacteriaceae*) than larger PE particles [72,76]. Among the aforementioned cited  
573 populations associated to MP exposure, some families, such as *Dethiosulfovibrionaceae* and  
574 *Enterobacteriaceae*, are considered as pathobionts. The *Dethiosulfovibrionaceae* population  
575 has been linked to colorectal cancer [81], while *Enterobacteriaceae* containing well known

576 enteric pathogens, such as *Escherichia*, *Shigella*, *Campylobacter* or *Salmonella*, may be  
577 increased in irritable bowel syndrome -IBS- [82] and intestinal inflammatory diseases in  
578 humans [83]. Obviously, not all species in these families are pathogenic but most of them are  
579 commensal, highlighting the critical need to better characterize their potential pathogenic  
580 characteristics at a lower taxonomic level. It should be noted that the *Enterobacteriaceae*  
581 population appears to have a particular affinity for PE MPs. Indeed, *Escherichia coli* adheres  
582 to the surface of PE MPs within 30 s [84]. Similarly, in seawater, *Alpha* and  
583 *Gammaproteobacteria* have been identified as the core microbiome of biofilms associated with  
584 PE MPs [85].

585 Concerning the impact of exposure to PE MPs on the activity of the gut microbiota, apart from  
586 our recent study on the adult microbiota in M-ARCOL [44], no *in vitro* or *in vivo* study  
587 involving MPs has been conducted to date. Here, a decrease in butyrate production was notably  
588 observed following exposure to PE MPs. This has not been previously shown under adult  
589 colonic conditions [44], suggesting an infant-specific effect, which may be associated in this  
590 at-risk population with deleterious effects. In particular, the increase in butyrate production  
591 during infancy has been described as a key health milestone [69] as the effect of high  
592 concentrations of fecal butyrate in one-year old child has been shown to protect against allergies  
593 and asthma [86,87]. Of note, AhR activity also tended to be reduced following exposure to PE  
594 MPs, suggesting less production of AhR receptor ligands by the gut microbiota. This could be  
595 related to the decrease in abundance of some *Tannarellaceae*, like *Parabacteroides distasonis*,  
596 which are known to be producers of AhR ligands [88]. The ability of the gut microbiota to  
597 produce tryptophan-based AhR ligands influences the integrity of the gut epithelium as well as  
598 the development, function, production, and maintenance of several key mucosal immune cells  
599 and mediators [40,62,89,90]. A deficit in the production of AhR ligands by the microbiota has  
600 been reported in inflammatory bowel disease (IBD), celiac disease and obese patients compared

601 to healthy subjects [58,91,92]. As a novel approach, the activity of the gut microbiota was  
602 finally investigated in the present study by measuring VOCs. These compounds have important  
603 roles in biological interactions between organisms and are currently used as biomarkers of  
604 various human diseases, including cancer and gastrointestinal metabolic disorders [93,94].  
605 Marked differences between the abundances of VOCs before and after exposure to PE MPs  
606 were highlighted, characterized in particular by an underproduction of 8 ester compounds and  
607 an overproduction of 7 hydrocarbons. Again, the shifts in VOC profiles under infant colonic  
608 conditions were markedly different from those previously obtained under adult conditions [44].  
609 An increase in fecal esters has been observed in obese patients with non-alcoholic fatty liver  
610 disease [39], in IBS patients [95], but also after acute exposure *in vitro* of human faecal  
611 microbiota to benzo[*a*]pyrene [59]. This suggests that overproduction of esters could be an  
612 indicator of disturbances in the gut microbiota. Moreover, when detected in the breath, the  
613 formation of hydrocarbons/alkanes has advocated as a non-invasive marker of alcoholic  
614 hepatitis but also as a measure of lipid peroxidation observed repeatedly in the pathophysiology  
615 of IBD [96]. Our study also showed that among other modified VOCs, heptanone (2-heptanone  
616 and 2-heptanone, 4-methyl), two alcohols (1-hexanol and 1-heptanol) and two aldehydes  
617 (heptanal and octanal) were overproduced after exposure to PE MPs. These compounds were  
618 identified during low-density PE photodegradation in the environment [97], suggesting the  
619 possibility of PE degradation during our Tm-ARCOL fermentations, but again not consistent  
620 with Raman spectroscopy results. In light of these findings, the potential degradation of PE  
621 MPs by the human gut microbiota merits further investigation, although the restricted contact  
622 time in the human gut during gastrointestinal transit certainly does not promote plastic  
623 degradation, compared to longer exposure to soil or water microorganisms in the environment  
624 [98–100].

625 Considering the host response to exposure to PE MPs, we finally combined our *in vitro* colon  
626 model with co-cultured intestinal epithelial cells. Such a coupling has already been carried out  
627 to evaluate the cellular toxicity of virgin or digested PS or PE MPs [44,64] in adult but never  
628 in infant digestive conditions. Previous studies in mice orally exposed to PE MPs have reported  
629 deleterious effects on the intestinal barrier characterized by a low-grade inflammatory state  
630 [72,101] and conflicting results regarding the production of mucus [76,101]. This negative  
631 impact of MPs on the intestinal barrier was confirmed *in vitro*. Direct exposure of human  
632 intestinal Caco-2 cells to MPs led to oxidative stress for PE MPs [102] and increased epithelial  
633 permeability for PS MPs [64]. Here, for the first time under infant conditions, we investigated  
634 the potential indirect effect of exposure to PE MPs on human intestinal Caco-2/HT29-MTX  
635 cells, i.e. mediated by the metabolites of gut microbes. No significant effect was demonstrated,  
636 since we only observed for some toddlers a trend towards an increase in the expression of mucin  
637 genes as well as a non-significant decrease in the expression of genes encoding for tight junction  
638 proteins (occludin and ZO-1). Overall, this absence of significant deleterious effects suggests a  
639 mucus-protective effect of our co-culture model [103] and/or a loss of the putative response  
640 triggered by gut microbial metabolites on inflammation and disruption of the intestinal barrier  
641 due to the dilution of the Tm-ARCOL luminal supernatants.

642 Taken together, our results suggest that chronic exposure to PE MPs induced significant  
643 perturbations in the composition and activity of toddlergut microbiota that may be related to a  
644 “dysbiotic” state of the microbiota, even though increased bacterial diversity does not fit the  
645 accepted definition of dysbiosis [104]. This study also highlighted that the impact of PE MPs  
646 on the gut microbiota and the intestinal barrier *in vitro* was highly dependent on the donor. For  
647 example, exposure to MPs had no significant effect on the microbial diversity of Toddler 4 and  
648 induced greater deleterious effects on intestinal permeability in Toddler 1 than in the others.  
649 Such data could help to define which infants are high and low “responders” to ingesting MPs.

650 Our *in vitro* approach can definitively help for a better mechanistic understanding of the  
651 interactions between ingested micro- or even nano-sized plastics, the gut microbiota and mucus,  
652 considering inter-individual specificities. Further studies are needed to better investigate the  
653 effects of different plastic types, surface properties and sizes [105–108] on the structure and  
654 functions of the gut microbiota. The potential of the Tm-ARCOL model should be also  
655 extended to the simulation of other specific at-risk populations, such as neonates or pathological  
656 situations associated with gut microbial dysbiosis, such as obesity in infants [109]. Finally,  
657 since MPs may be the carrier of various chemical pollutants, heavy metals, antibiotics or  
658 pathogens [110–112], their bioavailability and/or their survival and biological effects could be  
659 studied using our *in vitro* approach, as well as their interactions with the human gut microbiota.

660

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675

676 **Financial interests**

677 The authors declare that the research was conducted in the absence of any commercial or  
678 financial relationships that could be considered as a potential conflict of interest.

679

680 **Data Availability Statement**

681 Raw data are available at NCBI under the Sequence Read Archive database in the BioProject  
682 n° PRJNA831093.

683

684 **Ethics approval**

685 This study is a non-interventional study with no additions to usual clinical cares. The protocol  
686 does not require approval from an ethics committee according to the French Public Health Law  
687 (CSP Art L 1121-1.1).

688

689

690

691 **Tables**

692

693 **Table 1. Primers used for qPCR for total bacteria load, Illumina sequencing for bacteria**  
 694 **and Archaeal populations and RT-qPCR to monitor intestinal gene expression**

695

Gene	Primer sequence 5'-3'	Target	Annealing temperature (°C)	References
<b>Primers qPCR</b>				
<i>16S</i>	F-GTATTACCGCGGCTGCTG R-ACTCCTACGGGAGGCAG	Total bacteria	58	Yu et al. 2005-
<b>Primers 16S for Illumina sequencing</b>				
<i>16S</i>	F-CCTACGGGNGGCWGCAG R-GACTACHVGGGTATCTAATCC	Bacteria		Klindworth et al. 2013
<i>16S archaea</i>	F-GYGCASCAGKCGMGAAW R-GGACTACVSGGTATCTAAT	Archaea	-	Takai et Horikoshi, 2000
<b>Primers RT qPCR</b>				
Gene	Primer sequence 5'-3'		Annealing temperature (°C)	References
<i>GAPDH</i>	F-GTGG AAGACTCATGACCACAG R-GCAGGGATGATGTTCTGGAGAG	Housekeeping gene	60	This study
<i>OCN</i>	F-CTGAAGCAAGTGAAGGGATCTG R-CCAACCATCTTCTTGATGTGTGA	Occludin	60	This study
<i>ZO-1</i>	F-AGCACAGTGCCTAAAGCTATTCC R-AAGATGGTCATACTGTGGTGGC	Zonula occludens 1	60	This study
<i>MUC2</i>	F-CAGTGCAGCATCCTCAAA R-GCAGAAGCACTCACAGTC	Mucin 2	60	This study
<i>MUC5AC</i>	F-CACACATGGGAAGCTGAG R-TCGGCAGTCAAAGAACAC	Mucin 5AC	60	This study

696

697 **Table 2. Volatile organic compounds significantly altered by chronic exposure to PE MPs**  
 698 **in Tm-ARCOL**  
 699 Values are the mean of abundances ( $\times 10^5$ ) of each candidate marker with its standard deviation  
 700 (in brackets).

Candidate markers	Tentatively identification <sup>a</sup>	m/z <sup>b</sup>	Exp. RI <sup>c</sup>	Ref. RI <sup>d</sup>	CAS number	p value	Control	Exposed	Under/over production compared to the control
<b>Hydrocarbons</b>									
Pentane, 3-methyl-	MS+RI	71	561	578	96-14-0	***	0.2 (74%)	3.9 (29%)	↗
Cyclopentane, 1,1,3-trimethyl-	MS	97	728	NA	4516-69-2	***	0.2 (51%)	3.3 (34%)	↗
Methyl C9-alkane - RI 814	MS	85	814	NA		***	0.5 (49%)	4.8 (22%)	↗
Methyl C9-alkane - RI 821	MS	71	821	NA		***	3.3 (42%)	17.6 (15%)	↗
1-heptene, 2,4-dimethyl-	MS+RI	70	841	855	19549-87-2	***	5.7 (34%)	23.9 (9%)	↗
Heptane, 2,3-dimethyl-	MS+RI	84	856	850	3074-71-3	***	0.7 (44%)	2.1 (17%)	↗
Methyl C9-alkane - RI 861	MS	71	861	NA		*	1.5 (59%)	5.1 (35%)	↗
<b>Esters</b>									
Propanoic acid, ethyl ester	MS+RI	57	710	708	105-37-3	***	1310.7 (26%)	278.9 (57%)	↘
Propanoic acid, 2-methyl-, ethyl ester	MS+RI	71	757	754	97-62-1	*	15.9 (45%)	4.9 (44%)	↘
Butanoic acid, ethyl ester	MS+RI	88	799	802	105-54-4	***	88.4 (15%)	20.1 (59%)	↘
Propanoic acid, propyl ester	MS+RI	57	807	804	106-36-5	**	203.7 (31%)	42.9 (55%)	↘
Butanoic acid, 3-methyl, ethyl ester	MS+RI	88	851	854	108-64-5	**	1.9 (36%)	0.6 (32%)	↘
Butanoic acid, propyl ester	MS+RI	89	892	899	105-66-8	**	31.4 (37%)	8.7 (30%)	↘
Propanoic acid, butyl ester	MS+RI	75	903	908	590-01-2	**	6.0 (3%)	3.7 (29%)	↘
Hexanoic acid, methyl ester	MS+RI	74	923	924	106-70-7	*	3.3 (63%)	26.1 (64%)	↗
Pentanoic acid, propyl ester	MS	71	1006	NA	141-06-0	**	6.2 (17%)	2.5 (41%)	↘
Heptanoic acid, methyl ester	MS+RI	74	1031	1021	106-73-0	*	0.3 (44%)	10.6 (65%)	↗
Heptanoic acid, ethyl ester	MS+RI	88	1093	1097	106-30-9	*	0.5 (67%)	15.1 (74%)	↗
Butanoic acid, hexyl ester	MS+RI	71	1193	1192	2639-63-6	**	0.8 (28%)	1.6 (23%)	↗
<b>Alcohols</b>									
1-hexanol	MS+RI	69	866	867	111-27-3	*	759.7 (27%)	2425.3 (50%)	↗
1-heptanol	MS+RI	70	979	970	111-70-6	**	127.8 (47%)	1644.9 (36%)	↗
<b>Aldehydes</b>									
Heptanal	MS+RI	81	900	899	111-71-7	**	6.0 (31%)	14.0 (26%)	↗
Octanal	MS+RI	84	1017	1003	124-13-0	**	4.0 (25%)	9.1 (25%)	↗
<b>Ketones</b>									
2-heptanone	MS+RI	58	887	891	110-43-0	*	17.5 (51%)	75.7 (59%)	↗
2-heptanone, 4-methyl-	MS+RI	58	942	936	6137-06-0	*	2.5 (35%)	4.4 (26%)	↗
2-decanone	MS+RI	58	1197	1192	693-54-9	**	5.6 (16%)	2.5 (28%)	↘

701  
702

703 <sup>a</sup> MS + RI, mass spectrum and RI agree with literature data; MS, mass spectrum agrees with literature  
 704 data; <sup>b</sup> Mass fragment used for area determination; (<sup>c,d</sup>) Retention indices on a DB5 capillary column  
 705 from experimental run (<sup>c</sup>) or bibliographic data (<sup>d</sup>); NA: Non-available; ↘: underproduction compared  
 706 to the control; ↗: overproduction compared to the control; \* $p < 0.05$ ; \*\* $p < 0.01$ ; \*\*\* $p < Bonferroni$   
 707 corrected values

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709 **Figure captions**

710 **Fig. 1. Outline of *in vitro* fermentations and cell culture experiments**

711 a) Schematic representation of the Tm-ARCOL gut model and the parameters used to reproduce  
712 the toddler colonic conditions. b) Tm-ARCOL experimental design and analysis. Four  
713 bioreactors inoculated with fecal samples from four different toddlers were run in parallel. After  
714 a stabilization phase, PE MPs were added daily into the bioreactor at a dose of 21 mg for 14  
715 days. The last three days of the stabilization period (days 6, 7, 8) were chosen as the control  
716 while the last three days of the treatment period (days 20, 21, 22) were selected as exposed  
717 days. Atmospheric and luminal fermentation media were sampled daily while mucin-alginate  
718 beads were collected every other day for further analysis of microbiota structure or activity. c)  
719 Cell culture experimental design and analysis. Caco-2 and HT29-MTX intestinal cells in co-  
720 culture were seeded using inserts. After an 18-day growth period, cells were exposed during 24  
721 h to day 8 or day 22 luminal supernatants of Tm-ARCOL and cytotoxicity, inflammation, mucin  
722 gene expression and intestinal permeability were evaluated.

723 *AhR*: Aryl Hydrocarbon Receptor; *ELISA*: Enzyme-Linked Immunosorbent Assay; *GC*: Gas  
724 Chromatography; *HRP*: Horse Radish Peroxidase; *HPLC*: High-Performance Liquid  
725 Chromatography; *LDH*: Lactate Dehydrogenase; *LY*: Lucifer Yellow; *MP*: Microplastic;  
726 *MUC2*: Mucin 2; *MUC5AC*: Mucin 5AC; *OCN*: Occludin; *PE*: Polyethylene; *SCFA*: Short  
727 Chain Fatty Acid; *TEER*: Trans-Epithelial Electrical Resistance; *ZO-1*: Zonula Occludens-1

728

729 **Fig 2. Scanning electron microscopy observations of PE MPs/microbiota interactions in**  
730 **Tm-ARCOL and Raman spectroscopy analysis of PE MPs at the end of exposed period**

731 a) Representative scanning electron microscopy images of luminal samples and mucin-alginate  
732 beads from Toddlers 2 and 3 before and after exposure to PE MPs. For the mucin-alginate  
733 beads, the analysis was performed after 48 h of colonization in the mucosal external  
734 compartment. The right panel shows a typical mucin-alginate bead and its colonization by gut  
735 microbiota (day 8). b) Example of Raman spectroscopy mapping of the surface of PE MP (the  
736 PE MP appears in yellow in the left image) in the day 22 luminal sample of Toddler 3; the  
737 Raman spectrum is compared with that obtained with the initial suspension of PE MPs.

738 *PE*: Polyethylene

739

740 **Fig 3. Effect of chronic exposure to PE MPs on luminal and mucosal bacteria in Tm-**  
741 **ARCOL**

742 a) Relative abundance of the 20 main bacterial populations at the family level during the control  
743 and exposed periods for each donor. M represents the medium surrounding the mucin-alginate  
744 beads. b) Differential Analysis (DESeq2, metagenomeSeq, metacoder,  $p < 0.05$ ) at the family  
745 level between control and exposed days (n=4).

746

747 **Fig 4. Effect of chronic exposure to PE MPs on  $\alpha$ - and  $\beta$ -diversity indexes of bacterial**  
748 **communities in Tm-ARCOL**

749 Observed ASVs (a) and Shannon (b) indexes for each donor and on average (n=4), expressed  
750 as means  $\pm$  SEM. Bacterial  $\beta$ -diversity (c) was also represented using a two-dimensional PCoA  
751 plot, based on the unweighted UniFrac distance matrix on luminal and mucosal samples (left)  
752 and a distance-based redundancy analysis (RDA), based on both 16S rDNA gene  
753 metabarcoding and metabolic activity data, applied to luminal samples and excluding the donor  
754 effect (right). Statistical differences were tested by one-way ANOVA with \*  $p < 0.05$ , \*\*  $p < 0.01$ ,  
755 \*\*\*  $p < 0.001$ , \*\*\*\*  $p < 0.0001$ .

756 *CONT: Control, EXP: Exposed*

757 **Fig 5. Impact of chronic exposure to PE MPs on gas and short chain fatty acid production**  
758 **in Tm-ARCOL**

759 a) Mean  $\pm$  SEM daily gas production in mL and (b) mean  $\pm$  SEM total SCFA production in  
760 mM, for each donor and in average (n=4). c) Mean  $\pm$  SEM daily acetate, propionate and butyrate  
761 production in mM for each donor and on average (n=4). Statistical differences were tested by  
762 one-way ANOVA with \*  $p < 0.05$ , \*\*  $p < 0.01$ , \*\*\*\*  $p < 0.0001$ .

763 *CONT: Control; EXP: Exposed; SCFA: Short Chain Fatty Acid*

764

765 **Fig 6. Impact of chronic exposure to PE MPs on the capacity of Tm-ARCOL luminal**  
766 **supernatants to activate Aryl Hydrocarbon receptor**

767 AhR activity is expressed as fold changes relative to the luciferase activity of the control period  
768 and after normalisation based on sample cytotoxicity (lactate dehydrogenase activity assay).  
769 All results are given for each donor and on average (n=4) and expressed as means  $\pm$  SEM.

770 *AhR: Aryl Hydrocarbon Receptor; CONT: Control; EXP: Exposed*

771

772 **Fig 7. Impact of chronic exposure to PE MPs on volatile organic compounds in the luminal**  
773 **compartment of Tm-ARCOL**

774 a) Principal component analysis (PCA) of the control and exposed days including the 26  
775 discriminant VOCs. b) Mean of abundance of the eight VOCs with the highest discriminating  
776 power. Abundances of selected VOCs are expressed in arbitrary units. \*\*\*  $p < 0.01$  Bonferroni  
777 corrected value.

778

779 **Fig 8. Impact of Tm-ARCOL luminal supernatants before and after chronic exposure to**  
780 **PE MPs on inflammation and intestinal barrier integrity using a Caco-2/HT29-MTX co-**  
781 **culture**

782 a) The difference between the average TEER before and after 24-h exposure to Tm-ARCOL  
783 luminal supernatants was calculated and termed  $\Delta$ TEER ( $\Omega \cdot \text{cm}^2$ ) as an indicator of paracellular  
784 permeability of the co-culture before and after exposure to MPs. b) LY transport (Papp in cm/s)  
785 and HRP Papp (in cm/s) as an indicator of paracellular and transcellular permeability,  
786 respectively. c) Mean IL-8 concentration (in pg/mL) in the apical and basolateral media of the  
787 cell co-culture. All results were given by donors, also on average (n=4) and expressed as means  
788  $\pm$  SEM.

789 *CONT: Control; EXP: Exposed; HRP: Horse Radish Peroxidase; IL-8: Interleukin-8; LY:*  
790 *Lucifer Yellow; Papp: Apparent Permeability coefficient; TEER: Trans-Epithelial Electrical*  
791 *Resistance*

792 **References**

- 793 [1] PlasticsEurope, The Facts 2021 - An analysis of European plastics production, demand  
 794 and waste data, (2021).  
 795 <[www.plasticseurope.org/application/files/9715/7129/9584/FINAL\\_web\\_version\\_Plastics\\_the\\_facts2019\\_14102019.pdf](http://www.plasticseurope.org/application/files/9715/7129/9584/FINAL_web_version_Plastics_the_facts2019_14102019.pdf)>.  
 796
- 797 [2] L. Wang, W.-M. Wu, N.S. Bolan, D.C.W. Tsang, Y. Li, M. Qin, D. Hou, Environmental  
 798 fate, toxicity and risk management strategies of nanoplastics in the environment: Current  
 799 status and future perspectives, *Journal of Hazardous Materials*. 401 (2021) 123415.  
 800 <https://doi.org/10.1016/j.jhazmat.2020.123415>.
- 801 [3] O.A. Vázquez, S. Rahman, An ecotoxicological approach to microplastics on terrestrial  
 802 and aquatic organisms: A systematic review in assessment, monitoring and biological  
 803 impact, *Environmental Toxicology and Pharmacology*. (2021) 103615.  
 804 <https://doi.org/10.1016/j.etap.2021.103615>.
- 805 [4] J.C. Prata, J.P. da Costa, I. Lopes, A.L. Andrady, A.C. Duarte, T. Rocha-Santos, A One  
 806 Health perspective of the impacts of microplastics on animal, human and environmental  
 807 health, *Science of The Total Environment*. 777 (2021) 146094.  
 808 <https://doi.org/10.1016/j.scitotenv.2021.146094>.
- 809 [5] J.P.G.L. Frias, R. Nash, Microplastics: Finding a consensus on the definition, *Marine  
 810 Pollution Bulletin*. 138 (2019) 145–147.  
 811 <https://doi.org/10.1016/j.marpolbul.2018.11.022>.
- 812 [6] J. Lehel, S. Murphy, Microplastics in the Food Chain: Food Safety and Environmental  
 813 Aspects, *Rev Environ Contam Toxicol*. 259 (2021) 1–49.  
 814 [https://doi.org/10.1007/398\\_2021\\_77](https://doi.org/10.1007/398_2021_77).
- 815 [7] T. Gouin, D. Cunliffe, J. De France, J. Fawell, P. Jarvis, A.A. Koelmans, P. Marsden,  
 816 E.E. Testai, M. Asami, R. Bevan, R. Carrier, J. Cotruvo, A. Eckhardt, C.N. Ong,  
 817 Clarifying the absence of evidence regarding human health risks to microplastic particles  
 818 in drinking-water: High quality robust data wanted, *Environment International*. (2020)  
 819 106141. <https://doi.org/10.1016/j.envint.2020.106141>.
- 820 [8] L.W.D. van Raamsdonk, M. van der Zande, A.A. Koelmans, R.L.A.P. Hoogenboom,  
 821 R.J.B. Peters, M.J. Groot, A.A.C.M. Peijnenburg, Y.J.A. Weesepeel, Current Insights  
 822 into Monitoring, Bioaccumulation, and Potential Health Effects of Microplastics Present  
 823 in the Food Chain, *Foods*. 9 (2020) 72. <https://doi.org/10.3390/foods9010072>.
- 824 [9] J.F. Provencher, G.A. Covernton, R.C. Moore, D.A. Horn, J.L. Conkle, A.L. Lusher,  
 825 Proceed with caution: The need to raise the publication bar for microplastics research,  
 826 *Science of The Total Environment*. 748 (2020) 141426.  
 827 <https://doi.org/10.1016/j.scitotenv.2020.141426>.
- 828 [10] E. Fournier, L. Etienne-Mesmin, S. Blanquet-Diot, M. Mercier-Bonin, Impact of  
 829 Microplastics in Human Health, in: T. Rocha-Santos, M. Costa, C. Mouneyrac (Eds.),  
 830 *Handbook of Microplastics in the Environment*, Springer International Publishing,  
 831 Cham, 2020: pp. 1–25. [https://doi.org/10.1007/978-3-030-10618-8\\_48-1](https://doi.org/10.1007/978-3-030-10618-8_48-1).
- 832 [11] H.A. Leslie, M.J.M. van Velzen, S.H. Brandsma, A.D. Vethaak, J.J. Garcia-Vallejo,  
 833 M.H. Lamoree, Discovery and quantification of plastic particle pollution in human blood,  
 834 *Environment International*. (2022) 107199.  
 835 <https://doi.org/10.1016/j.envint.2022.107199>.
- 836 [12] Y.S. Ibrahim, S. Tuan Anuar, A.A. Azmi, W.M.A. Wan Mohd Khalik, S. Lehata, S.R.  
 837 Hamzah, D. Ismail, Z.F. Ma, A. Dzulkarnaen, Z. Zakaria, N. Mustaffa, S.E. Tuan Sharif,  
 838 Y.Y. Lee, Detection of microplastics in human colectomy specimens, *JGH Open*. 5  
 839 (2021) 116–121. <https://doi.org/10.1002/jgh3.12457>.

- 840 [13] J. Zhang, L. Wang, L. Trasande, K. Kannan, Occurrence of Polyethylene Terephthalate  
841 and Polycarbonate Microplastics in Infant and Adult Feces, *Environ. Sci. Technol. Lett.*  
842 8 (2021) 989–994. <https://doi.org/10.1021/acs.estlett.1c00559>.
- 843 [14] N. Zhang, Y.B. Li, H.R. He, J.F. Zhang, G.S. Ma, You are what you eat: Microplastics  
844 in the feces of young men living in Beijing, *Sci Total Environ.* 767 (2021) 144345.  
845 <https://doi.org/10.1016/j.scitotenv.2020.144345>.
- 846 [15] P. Schwabl, S. Köppel, P. Königshofer, T. Bucsics, M. Trauner, T. Reiberger, B.  
847 Liebmann, Detection of Various Microplastics in Human Stool: A Prospective Case  
848 Series, *Ann Intern Med.* 171 (2019) 453–457. <https://doi.org/10.7326/M19-0618>.
- 849 [16] A. Ragusa, A. Svelato, C. Santacroce, P. Catalano, V. Notarstefano, O. Carnevali, F.  
850 Papa, M.C.A. Rongioletti, F. Baiocco, S. Draghi, E. D'Amore, D. Rinaldo, M. Matta, E.  
851 Giorgini, Plasticenta: First evidence of microplastics in human placenta, *Environment*  
852 *International.* 146 (2021) 106274. <https://doi.org/10.1016/j.envint.2020.106274>.
- 853 [17] T. Braun, L. Ehrlich, W. Henrich, S. Koepfel, I. Lomako, P. Schwabl, B. Liebmann,  
854 Detection of Microplastic in Human Placenta and Meconium in a Clinical Setting,  
855 *Pharmaceutics.* 13 (2021) 921. <https://doi.org/10.3390/pharmaceutics13070921>.
- 856 [18] J. Zhu, X. Zhang, K. Liao, P. Wu, H. Jin, Microplastics in dust from different indoor  
857 environments, *Sci Total Environ.* (2022) 155256.  
858 <https://doi.org/10.1016/j.scitotenv.2022.155256>.
- 859 [19] A. Ragusa, V. Notarstefano, A. Svelato, A. Belloni, G. Gioacchini, C. Blondeel, E.  
860 Zucchelli, C. De Luca, S. D'Avino, A. Gulotta, O. Carnevali, E. Giorgini, Raman  
861 Microspectroscopy Detection and Characterisation of Microplastics in Human  
862 Breastmilk, *Polymers (Basel).* 14 (2022) 2700. <https://doi.org/10.3390/polym14132700>.
- 863 [20] G. Kuttralam-Muniasamy, F. Pérez-Guevara, I. Elizalde-Martínez, V.C. Shruti, Branded  
864 milks – Are they immune from microplastics contamination?, *Science of The Total*  
865 *Environment.* 714 (2020) 136823. <https://doi.org/10.1016/j.scitotenv.2020.136823>.
- 866 [21] P.A. Da Costa Filho, D. Andrey, B. Eriksen, R.P. Peixoto, B.M. Carreres, M.E. Ambühl,  
867 J.B. Descarrega, S. Dubascoux, P. Zbinden, A. Panchaud, E. Poitevin, Detection and  
868 characterization of small-sized microplastics ( $\geq 5 \mu\text{m}$ ) in milk products, *Sci Rep.* 11  
869 (2021) 24046. <https://doi.org/10.1038/s41598-021-03458-7>.
- 870 [22] D. Li, Y. Shi, L. Yang, L. Xiao, D.K. Kehoe, Y.K. Gun'ko, J.J. Boland, J.J. Wang,  
871 Microplastic release from the degradation of polypropylene feeding bottles during infant  
872 formula preparation, *Nat Food.* (2020). <https://doi.org/10.1038/s43016-020-00171-y>.
- 873 [23] D. Li, L. Yang, R. Kavanagh, L. Xiao, Y. Shi, D.K. Kehoe, E.D. Sheerin, Y.K. Gun'ko,  
874 J.J. Boland, J.J. Wang, Sampling, Identification and Characterization of Microplastics  
875 Release from Polypropylene Baby Feeding Bottle during Daily Use, *JoVE.* (2021)  
876 62545. <https://doi.org/10.3791/62545>.
- 877 [24] Y. Su, X. Hu, H. Tang, K. Lu, H. Li, S. Liu, B. Xing, R. Ji, Steam disinfection releases  
878 micro(nano)plastics from silicone-rubber baby teats as examined by optical photothermal  
879 infrared microspectroscopy, *Nat Nanotechnol.* (2021). <https://doi.org/10.1038/s41565-021-00998-x>.
- 880 [25] K. Song, R. Ding, C. Sun, L. Yao, W. Zhang, Microparticles and microplastics released  
881 from daily use of plastic feeding and water bottles and plastic injectors: potential risks to  
882 infants and children in China, *Environ Sci Pollut Res.* (2021).  
883 <https://doi.org/10.1007/s11356-021-14939-7>.
- 884 [26] K. Sripada, A. Wierzbicka, K. Abass, J.O. Gimalt, A. Erbe, H.B. Röllin, P. Weihe, G.J.  
885 Díaz, R.R. Singh, T. Visnes, A. Rautio, J.Ø. Odland, M. Wagner, A Children's Health  
886 Perspective on Nano- and Microplastics, *Environ Health Perspect.* 130 (2022) 15001.  
887 <https://doi.org/10.1289/EHP9086>.
- 888

- 889 [27] S.J. Bosman, S.P. Nieto, W.C. Patton, J.D. Jacobson, J.U. Corselli, P.J. Chan,  
890 Development of mammalian embryos exposed to mixed-size nanoparticles, *Clin Exp*  
891 *Obstet Gynecol.* 32 (2005) 222–224.
- 892 [28] S. Bojic, M.M. Falco, P. Stojkovic, B. Ljubic, M. Gazdic Jankovic, L. Armstrong, N.  
893 Markovic, J. Dopazo, M. Lako, R. Bauer, M. Stojkovic, Platform to study intracellular  
894 polystyrene nanoplastic pollution and clinical outcomes: *Stem Cells and Plastic*  
895 *Pollution, Stem Cells.* (2020). <https://doi.org/10.1002/stem.3244>.
- 896 [29] J. Nie, Y. Shen, M. Roshdy, X. Cheng, G. Wang, X. Yang, Polystyrene nanoplastics  
897 exposure caused defective neural tube morphogenesis through caveolae-mediated  
898 endocytosis and faulty apoptosis, *Nanotoxicology.* (2021) 1–20.  
899 <https://doi.org/10.1080/17435390.2021.1930228>.
- 900 [30] T. Hua, S. Kiran, Y. Li, Q.-X.A. Sang, Microplastics exposure affects neural  
901 development of human pluripotent stem cell-derived cortical spheroids, *J Hazard Mater.*  
902 435 (2022) 128884. <https://doi.org/10.1016/j.jhazmat.2022.128884>.
- 903 [31] J. Hu, X. Qin, J. Zhang, Y. Zhu, W. Zeng, Y. Lin, X. Liu, Polystyrene microplastics  
904 disturb maternal-fetal immune balance and cause reproductive toxicity in pregnant mice,  
905 *Reproductive Toxicology.* (2021) S0890623821001581.  
906 <https://doi.org/10.1016/j.reprotox.2021.10.002>.
- 907 [32] T. Luo, C. Wang, Z. Pan, C. Jin, Z. Fu, Y. Jin, Maternal Polystyrene Microplastic  
908 Exposure during Gestation and Lactation Altered Metabolic Homeostasis in the Dams  
909 and Their F1 and F2 Offspring, *Environ. Sci. Technol.* 53 (2019) 10978–10992.  
910 <https://doi.org/10.1021/acs.est.9b03191>.
- 911 [33] T. Luo, Y. Zhang, C. Wang, X. Wang, J. Zhou, M. Shen, Y. Zhao, Z. Fu, Y. Jin, Maternal  
912 exposure to different sizes of polystyrene microplastics during gestation causes  
913 metabolic disorders in their offspring, *Environ Pollut.* 255 (2019) 113122.  
914 <https://doi.org/10.1016/j.envpol.2019.113122>.
- 915 [34] T. Yatsunencko, F.E. Rey, M.J. Manary, I. Trehan, M.G. Dominguez-Bello, M. Contreras,  
916 M. Magris, G. Hidalgo, R.N. Baldassano, A.P. Anokhin, A.C. Heath, B. Warner, J.  
917 Reeder, J. Kuczynski, J.G. Caporaso, C.A. Lozupone, C. Lauber, J.C. Clemente, D.  
918 Knights, R. Knight, J.I. Gordon, Human gut microbiome viewed across age and  
919 geography, *Nature.* 486 (2012) 222–227. <https://doi.org/10.1038/nature11053>.
- 920 [35] E. Fournier, C. Roussel, A. Dominicis, D. Ley, M.-A. Peyron, V. Collado, M. Mercier-  
921 Bonin, C. Lacroix, M. Alric, T. Van de Wiele, C. Chassard, L. Etienne-Mesmin, S.  
922 Blanquet-Diot, In vitro models of gut digestion across childhood: current developments,  
923 challenges and future trends, *Biotechnology Advances.* (2021) 107796.  
924 <https://doi.org/10.1016/j.biotechadv.2021.107796>.
- 925 [36] C. Bourlieu, O. Ménard, K. Bouzerzour, G. Mandalari, A. Macierzanka, A.R. Mackie,  
926 D. Dupont, Specificity of Infant Digestive Conditions: Some Clues for Developing  
927 Relevant In Vitro Models, *Critical Reviews in Food Science and Nutrition.* 54 (2014)  
928 1427–1457. <https://doi.org/10.1080/10408398.2011.640757>.
- 929 [37] K. Kalantar-Zadeh, K.J. Berean, R.E. Burgell, J.G. Muir, P.R. Gibson, Intestinal gases:  
930 influence on gut disorders and the role of dietary manipulations, *Nat Rev Gastroenterol*  
931 *Hepatol.* 16 (2019) 733–747. <https://doi.org/10.1038/s41575-019-0193-z>.
- 932 [38] D.J. Morrison, T. Preston, Formation of short chain fatty acids by the gut microbiota and  
933 their impact on human metabolism, *Gut Microbes.* 7 (2016) 189–200.  
934 <https://doi.org/10.1080/19490976.2015.1134082>.
- 935 [39] M. Raman, I. Ahmed, P.M. Gillevet, C.S. Probert, N.M. Ratcliffe, S. Smith, R.  
936 Greenwood, M. Sikaroodi, V. Lam, P. Crotty, J. Bailey, R.P. Myers, K.P. Rioux, Fecal  
937 Microbiome and Volatile Organic Compound Metabolome in Obese Humans With

- 938 Nonalcoholic Fatty Liver Disease, *Clinical Gastroenterology and Hepatology*. 11 (2013)  
939 868-875.e3. <https://doi.org/10.1016/j.cgh.2013.02.015>.
- 940 [40] B. Lamas, J.M. Natividad, H. Sokol, Aryl hydrocarbon receptor and intestinal immunity,  
941 *Mucosal Immunol*. 11 (2018) 1024–1038. <https://doi.org/10.1038/s41385-018-0019-2>.
- 942 [41] A. Andoh, Physiological Role of Gut Microbiota for Maintaining Human Health,  
943 *Digestion*. 93 (2016) 176–181. <https://doi.org/10.1159/000444066>.
- 944 [42] G.M.H. Birchenough, M.E.V. Johansson, Forming a mucus barrier along the colon,  
945 *Science*. 370 (2020) 402–403. <https://doi.org/10.1126/science.abe7194>.
- 946 [43] S.P. Claus, H. Guillou, S. Ellero-Simatos, The gut microbiota: a major player in the  
947 toxicity of environmental pollutants?, *Npj Biofilms Microbiomes*. 2 (2016) 16003.  
948 <https://doi.org/10.1038/npjbiofilms.2016.3>.
- 949 [44] E. Fournier, M. Leveque, P. Ruiz, J. Ratel, C. Durif, S. Chalancon, F. Amiard, M. Edely,  
950 V. Bezirard, E. Gaultier, B. Lamas, E. Houdeau, F. Lagarde, E. Engel, L. Etienne-  
951 Mesmin, S. Blanquet-Diot, M. Mercier-Bonin, Microplastics: what happens in the human  
952 digestive tract? First evidences in adults using in vitro gut models, *J. Hazard. Mater.*  
953 (2022) 130010. <https://doi.org/10.1016/j.jhazmat.2022.130010>.
- 954 [45] F. Hugenholtz, W.M. de Vos, Mouse models for human intestinal microbiota research: a  
955 critical evaluation, *Cell Mol Life Sci*. 75 (2018) 149–160.  
956 <https://doi.org/10.1007/s00018-017-2693-8>.
- 957 [46] E. Fournier, S. Denis, A. Dominicis, T. Van de Wiele, M. Alric, M. Mercier-Bonin, L.  
958 Etienne-Mesmin, S. Blanquet-Diot, A child is not an adult: development of a new  
959 in vitro model of the toddler colon, *Appl Microbiol Biotechnol*. (2022). *in press*
- 960 [47] C. Deschamps, E. Fournier, O. Uriot, F. Lajoie, C. Verdier, S. Comtet-Marre, M.  
961 Thomas, N. Kapel, C. Cherbuy, M. Alric, M. Almeida, L. Etienne-Mesmin, S. Blanquet-  
962 Diot, Comparative methods for fecal sample storage to preserve gut microbial structure  
963 and function in an in vitro model of the human colon, *Appl Microbiol Biotechnol*. 104  
964 (2020) 10233–10247. <https://doi.org/10.1007/s00253-020-10959-4>.
- 965 [48] S.H. Capone, M. Dufresne, M. Rechel, M.-J. Fleury, A.-V. Salsac, P. Paullier, M. Daujat-  
966 Chavanieu, C. Legallais, Impact of Alginate Composition: From Bead Mechanical  
967 Properties to Encapsulated HepG2/C3A Cell Activities for In Vivo Implantation, *PLOS*  
968 *ONE*. 8 (2013) e62032. <https://doi.org/10.1371/journal.pone.0062032>.
- 969 [49] Y. Yu, C. Lee, J. Kim, S. Hwang, Group-specific primer and probe sets to detect  
970 methanogenic communities using quantitative real-time polymerase chain reaction,  
971 *Biotechnol. Bioeng*. 89 (2005) 670–679. <https://doi.org/10.1002/bit.20347>.
- 972 [50] S. Theil, E. Rifa, rANOMALY: AmplicoN wOrkflow for Microbial community  
973 AnaLYsis, *F1000Res*. 10 (2021) 7. <https://doi.org/10.12688/f1000research.27268.1>.
- 974 [51] B.J. Callahan, P.J. McMurdie, M.J. Rosen, A.W. Han, A.J.A. Johnson, S.P. Holmes,  
975 DADA2: High-resolution sample inference from Illumina amplicon data, *Nat Methods*.  
976 13 (2016) 581–583. <https://doi.org/10.1038/nmeth.3869>.
- 977 [52] A. Murali, A. Bhargava, E.S. Wright, IDTAXA: a novel approach for accurate taxonomic  
978 classification of microbiome sequences, *Microbiome*. 6 (2018) 140.  
979 <https://doi.org/10.1186/s40168-018-0521-5>.
- 980 [53] C. Quast, E. Pruesse, P. Yilmaz, J. Gerken, T. Schweer, P. Yarza, J. Peplies, F.O.  
981 Glöckner, The SILVA ribosomal RNA gene database project: improved data processing  
982 and web-based tools, *Nucleic Acids Research*. 41 (2012) D590–D596.  
983 <https://doi.org/10.1093/nar/gks1219>.
- 984 [54] D.H. Parks, M. Chuvochina, C. Rinke, A.J. Mussig, P.-A. Chaumeil, P. Hugenholtz,  
985 GTDB: an ongoing census of bacterial and archaeal diversity through a phylogenetically  
986 consistent, rank normalized and complete genome-based taxonomy, *Nucleic Acids*  
987 *Research*. 50 (2022) D785–D794. <https://doi.org/10.1093/nar/gkab776>.

- 988 [55] S.F. Altschul, W. Gish, W. Miller, E.W. Myers, D.J. Lipman, Basic local alignment  
989 search tool, *J Mol Biol.* 215 (1990) 403–410. [https://doi.org/10.1016/S0022-2836\(05\)80360-2](https://doi.org/10.1016/S0022-2836(05)80360-2).  
990
- 991 [56] K. Schliep, A.J. Potts, D.A. Morrison, G.W. Grimm, Intertwining phylogenetic trees and  
992 networks, *Methods in Ecology and Evolution.* 8 (2017) 1212–1220.  
993 <https://doi.org/10.1111/2041-210X.12760>.  
994
- 995 [57] J. Oksanen, F.G. Blanchet, R. Kindt, P. Legendre, R. O’Hara, G. Simpson, P. Solymos,  
996 H. Stevens, H. Wagner, *Vegan: Community Ecology Package. R Package Version 1.17-8*, (2011).  
997
- 998 [58] B. Lamas, M.L. Richard, V. Leducq, H.-P. Pham, M.-L. Michel, G. Da Costa, C.  
999 Bridonneau, S. Jegou, T.W. Hoffmann, J.M. Natividad, L. Brot, S. Taleb, A. Couturier-  
1000 Maillard, I. Nion-Larmurier, F. Merabtene, P. Seksik, A. Bourrier, J. Cosnes, B. Ryffel,  
1001 L. Beaugerie, J.-M. Launay, P. Langella, R.J. Xavier, H. Sokol, CARD9 impacts colitis  
1002 by altering gut microbiota metabolism of tryptophan into aryl hydrocarbon receptor  
1003 ligands, *Nat Med.* 22 (2016) 598–605. <https://doi.org/10.1038/nm.4102>.  
1004
- 1005 [59] C. Defois, J. Ratel, S. Denis, B. Batut, R. Beugnot, E. Peyretailade, E. Engel, P. Peyret,  
1006 Environmental Pollutant Benzo[a]Pyrene Impacts the Volatile Metabolome and  
1007 Transcriptome of the Human Gut Microbiota, *Front Microbiol.* 8 (2017) 1562.  
1008 <https://doi.org/10.3389/fmicb.2017.01562>.  
1009
- 1010 [60] C. Defois, J. Ratel, G. Garrait, S. Denis, O. Le Goff, J. Talvas, P. Mosoni, E. Engel, P.  
1011 Peyret, Food Chemicals Disrupt Human Gut Microbiota Activity And Impact Intestinal  
1012 Homeostasis As Revealed By In Vitro Systems, *Scientific Reports.* 8 (2018) 11006.  
1013 <https://doi.org/10.1038/s41598-018-29376-9>.  
1014
- 1015 [61] E. Rifa, S. Theil, *ExploreMetabar: v1.0.0*, (2021).  
1016 <https://doi.org/10.5281/ZENODO.5245195>.  
1017
- 1018 [62] A. Agus, J. Planchais, H. Sokol, Gut Microbiota Regulation of Tryptophan Metabolism  
1019 in Health and Disease, *Cell Host Microbe.* 23 (2018) 716–724.  
1020 <https://doi.org/10.1016/j.chom.2018.05.003>.  
1021
- 1022 [63] V. Stock, C. Fahrenson, A. Thuenemann, M.H. Dönmez, L. Voss, L. Böhmert, A.  
1023 Braeuning, A. Lampen, H. Sieg, Impact of artificial digestion on the sizes and shapes of  
1024 microplastic particles, *Food and Chemical Toxicology.* (2019) 111010.  
1025 <https://doi.org/10.1016/j.fct.2019.111010>.  
1026
- 1027 [64] S. Liu, X. Wu, W. Gu, J. Yu, B. Wu, Influence of the digestive process on intestinal  
1028 toxicity of polystyrene microplastics as determined by in vitro Caco-2 models,  
1029 *Chemosphere.* 256 (2020) 127204. <https://doi.org/10.1016/j.chemosphere.2020.127204>.  
1030
- 1031 [65] P. Krasucka, A. Bogusz, E. Baranowska-Wójcik, B. Czech, D. Sz wajgier, M. Rek, Y.S.  
1032 Ok, P. Oleszczuk, Digestion of plastics using in vitro human gastrointestinal tract and  
1033 their potential to adsorb emerging organic pollutants, *Sci Total Environ.* (2022) 157108.  
1034 <https://doi.org/10.1016/j.scitotenv.2022.157108>.  
1035
- 1036 [66] A. Tamargo, N. Molinero, J.J. Reinoso, V. Alcolea-Rodriguez, R. Portela, M.A. Bañares,  
1037 J.F. Fernández, M.V. Moreno-Arribas, PET microplastics affect human gut microbiota  
communities during simulated gastrointestinal digestion, first evidence of plausible  
polymer biodegradation during human digestion, *Sci Rep.* 12 (2022) 528.  
<https://doi.org/10.1038/s41598-021-04489-w>.
- 1038 [67] K. Senathirajah, S. Attwood, G. Bhagwat, M. Carbery, S. Wilson, T. Palanisami,  
1039 Estimation of the mass of microplastics ingested - A pivotal first step towards human  
1040 health risk assessment, *J Hazard Mater.* 404 (2021) 124004.  
1041 <https://doi.org/10.1016/j.jhazmat.2020.124004>.  
1042
- 1043 [68] K.I. Ayeni, D. Berry, L. Wisgrill, B. Warth, C.N. Ezekiel, Early-life chemical exposome  
1044 and gut microbiome development: African research perspectives within a global

- environmental health context, *Trends in Microbiology*. 0 (2022).  
<https://doi.org/10.1016/j.tim.2022.05.008>.
- [69] M. Nilsen, C. Madelen Saunders, I. Leena Angell, M.Ø. Arntzen, K.C. Lødrup Carlsen, K.-H. Carlsen, G. Haugen, L. Heldal Hagen, M.H. Carlsen, G. Hedlin, C. Monceyron Jonassen, B. Nordlund, E. Maria Rehbinder, H.O. Skjerven, L. Snipen, A. Cathrine Staff, R. Vettukattil, K. Rudi, Butyrate Levels in the Transition from an Infant- to an Adult-Like Gut Microbiota Correlate with Bacterial Networks Associated with *Eubacterium Rectale* and *Ruminococcus Gnavus*, *Genes*. 11 (2020) 1245. <https://doi.org/10.3390/genes11111245>.
- [70] C.J. Hill, D.B. Lynch, K. Murphy, M. Ulaszewska, I.B. Jeffery, C.A. O’Shea, C. Watkins, E. Dempsey, F. Mattivi, K. Tuohy, R.P. Ross, C.A. Ryan, P.W. O’ Toole, C. Stanton, Evolution of gut microbiota composition from birth to 24 weeks in the INFANTMET Cohort, *Microbiome*. 5 (2017) 4. <https://doi.org/10.1186/s40168-016-0213-y>.
- [71] S. Wang, S. Zeng, M. Egan, P. Cherry, C. Strain, E. Morais, P. Boyaval, C.A. Ryan, E. M Dempsey, R.P. Ross, C. Stanton, Metagenomic analysis of mother-infant gut microbiome reveals global distinct and shared microbial signatures, *Gut Microbes*. 13 (2021) 1–24. <https://doi.org/10.1080/19490976.2021.1911571>.
- [72] B. Li, Y. Ding, X. Cheng, D. Sheng, Z. Xu, Q. Rong, Y. Wu, H. Zhao, X. Ji, Y. Zhang, Polyethylene microplastics affect the distribution of gut microbiota and inflammation development in mice, *Chemosphere*. 244 (2020) 125492. <https://doi.org/10.1016/j.chemosphere.2019.125492>.
- [73] E.R. Zettler, T.J. Mincer, L.A. Amaral-Zettler, Life in the “Plastisphere”: Microbial Communities on Plastic Marine Debris, *Environ. Sci. Technol*. 47 (2013) 7137–7146. <https://doi.org/10.1021/es401288x>.
- [74] K. Nanthini Devi, P. Raju, P. Santhanam, S. Dinesh Kumar, N. Krishnaveni, J. Roopavathy, P. Perumal, Biodegradation of low-density polyethylene and polypropylene by microbes isolated from Vaigai River, Madurai, India, *Arch Microbiol*. (2021). <https://doi.org/10.1007/s00203-021-02592-0>.
- [75] X. Chen, Y. Wang, S. Chen, Y. Sun, Q. Tan, Z. Ding, Y. Lu, Y. Yu, Microplastics as carbon-nutrient sources and shaper for microbial communities in stagnant water, *Journal of Hazardous Materials*. 420 (2021) 126662. <https://doi.org/10.1016/j.jhazmat.2021.126662>.
- [76] M. Djouina, C. Vignal, A. Dehaut, S. Caboche, N. Hirt, C. Waxin, C. Himber, D. Beury, D. Hot, L. Dubuquoy, D. Launay, G. Duflos, M. Body-Malapel, Oral exposure to polyethylene microplastics alters gut morphology, immune response, and microbiota composition in mice, *Environ Res*. (2022) 113230. <https://doi.org/10.1016/j.envres.2022.113230>.
- [77] L. Etienne-Mesmin, B. Chassaing, M. Desvaux, K. De Paepe, R. Gresse, T. Sauvaitre, E. Forano, T.V. de Wiele, S. Schüller, N. Juge, S. Blanquet-Diot, Experimental models to study intestinal microbes-mucus interactions in health and disease, *FEMS Microbiol Rev*. 43 (2019) 457–489. <https://doi.org/10.1093/femsre/fuz013>.
- [78] X. Chen, J. Zhuang, Q. Chen, L. Xu, X. Yue, D. Qiao, Chronic exposure to polyvinyl chloride microplastics induces liver injury and gut microbiota dysbiosis based on the integration of liver transcriptome profiles and full-length 16S rRNA sequencing data, *Sci Total Environ*. (2022) 155984. <https://doi.org/10.1016/j.scitotenv.2022.155984>.
- [79] D. Huang, Y. Zhang, J. Long, X. Yang, L. Bao, Z. Yang, B. Wu, R. Si, W. Zhao, C. Peng, A. Wang, D. Yan, Polystyrene microplastic exposure induces insulin resistance in mice via dysbacteriosis and pro-inflammation, *Sci Total Environ*. (2022) 155937. <https://doi.org/10.1016/j.scitotenv.2022.155937>.

- 1088 [80] S. Wen, Y. Zhao, S. Liu, Y. Chen, H. Yuan, H. Xu, Polystyrene microplastics  
1089 exacerbated liver injury from cyclophosphamide in mice: Insight into gut microbiota, *Sci*  
1090 *Total Environ.* (2022) 156668. <https://doi.org/10.1016/j.scitotenv.2022.156668>.
- 1091 [81] W. Liu, R. Zhang, R. Shu, J. Yu, H. Li, H. Long, S. Jin, S. Li, Q. Hu, F. Yao, C. Zhou,  
1092 Q. Huang, X. Hu, M. Chen, W. Hu, Q. Wang, S. Fang, Q. Wu, Study of the Relationship  
1093 between Microbiome and Colorectal Cancer Susceptibility Using 16SrRNA Sequencing,  
1094 *Biomed Res Int.* 2020 (2020) 7828392. <https://doi.org/10.1155/2020/7828392>.
- 1095 [82] R. Pittayanon, J.T. Lau, Y. Yuan, G.I. Leontiadis, F. Tse, M. Surette, P. Moayyedi, Gut  
1096 Microbiota in Patients With Irritable Bowel Syndrome-A Systematic Review,  
1097 *Gastroenterology.* (2019). <https://doi.org/10.1053/j.gastro.2019.03.049>.
- 1098 [83] B.-K. Zhu, Y.-M. Fang, D. Zhu, P. Christie, X. Ke, Y.-G. Zhu, Exposure to nanoplastics  
1099 disturbs the gut microbiome in the soil oligochaete *Enchytraeus crypticus*, *Environ.*  
1100 *Pollut.* 239 (2018) 408–415. <https://doi.org/10.1016/j.envpol.2018.04.017>.
- 1101 [84] A. Tang, X. Bi, J. Du, L. Rao, V. Vasanthakumar, Y.-B. Hu, M.-L. Fu, W. Sun, B. Yuan,  
1102 The effect of polyethylene microplastics on the disinfection of *Escherichia coli* by  
1103 sodium hypochlorite, *Sci Total Environ.* (2022) 155322.  
1104 <https://doi.org/10.1016/j.scitotenv.2022.155322>.
- 1105 [85] C. Tu, T. Chen, Q. Zhou, Y. Liu, J. Wei, J.J. Waniek, Y. Luo, Biofilm formation and its  
1106 influences on the properties of microplastics as affected by exposure time and depth in  
1107 the seawater, *Sci. Total Environ.* 734 (2020) 139237.  
1108 <https://doi.org/10.1016/j.scitotenv.2020.139237>.
- 1109 [86] C. Roduit, R. Frei, R. Ferstl, S. Loeliger, P. Westermann, C. Rhyner, E. Schiavi, W.  
1110 Barcik, N. Rodriguez-Perez, M. Wawrzyniak, C. Chassard, C. Lacroix, E.  
1111 Schmausser-Hechfellner, M. Depner, E. von Mutius, C. Braun-Fahrländer, A.M.  
1112 Karvonen, P.V. Kirjavainen, J. Pekkanen, J.-C. Dalphin, J. Riedler, C. Akdis, R. Lauener,  
1113 L. O'Mahony, High levels of butyrate and propionate in early life are associated with  
1114 protection against atopy, *Allergy.* 74 (2019) 799–809. <https://doi.org/10.1111/all.13660>.
- 1115 [87] M. Depner, D.H. Taft, P.V. Kirjavainen, K.M. Kalanetra, A.M. Karvonen, S. Peschel, E.  
1116 Schmausser-Hechfellner, C. Roduit, R. Frei, R. Lauener, A. Divaret-Chauveau, J.-C.  
1117 Dalphin, J. Riedler, M. Roponen, M. Kabesch, H. Renz, J. Pekkanen, F.M. Farquharson,  
1118 P. Louis, D.A. Mills, E. von Mutius, PASTURE study group, M.J. Ege, Maturation of  
1119 the gut microbiome during the first year of life contributes to the protective farm effect  
1120 on childhood asthma, *Nat Med.* 26 (2020) 1766–1775. <https://doi.org/10.1038/s41591-020-1095-x>.
- 1121 [88] X. Su, Y. Gao, R. Yang, Gut Microbiota-Derived Tryptophan Metabolites Maintain Gut  
1122 and Systemic Homeostasis, *Cells.* 11 (2022) 2296.  
1123 <https://doi.org/10.3390/cells11152296>.
- 1124 [89] H.M. Roager, T.R. Licht, Microbial tryptophan catabolites in health and disease, *Nat*  
1125 *Commun.* 9 (2018) 3294. <https://doi.org/10.1038/s41467-018-05470-4>.
- 1126 [90] V. Rothhammer, F.J. Quintana, The aryl hydrocarbon receptor: an environmental sensor  
1127 integrating immune responses in health and disease, *Nat Rev Immunol.* 19 (2019) 184–  
1128 197. <https://doi.org/10.1038/s41577-019-0125-8>.
- 1129 [91] B. Lamas, L. Hernandez-Galan, H.J. Galipeau, M. Constante, A. Clarizio, J. Jury, N.M.  
1130 Breyner, A. Caminero, G. Rueda, C.L. Hayes, J.L. McCarville, M. Bermudez Brito, J.  
1131 Planchais, N. Rolhion, J.A. Murray, P. Langella, L.M.P. Loonen, J.M. Wells, P. Bercik,  
1132 H. Sokol, E.F. Verdu, Aryl hydrocarbon receptor ligand production by the gut microbiota  
1133 is decreased in celiac disease leading to intestinal inflammation, *Sci Transl Med.* 12  
1134 (2020) eaba0624. <https://doi.org/10.1126/scitranslmed.aba0624>.
- 1135 [92] J.M. Natividad, A. Agus, J. Planchais, B. Lamas, A.C. Jarry, R. Martin, M.-L. Michel,  
1136 C. Chong-Nguyen, R. Roussel, M. Straube, S. Jegou, C. McQuitty, M. Le Gall, G. da  
1137

- 1138 Costa, E. Lecornet, C. Michaudel, M. Modoux, J. Glodt, C. Bridonneau, B. Sovran, L.  
1139 Dupraz, A. Bado, M.L. Richard, P. Langella, B. Hansel, J.-M. Launay, R.J. Xavier, H.  
1140 Duboc, H. Sokol, Impaired Aryl Hydrocarbon Receptor Ligand Production by the Gut  
1141 Microbiota Is a Key Factor in Metabolic Syndrome, *Cell Metab.* 28 (2018) 737-749.e4.  
1142 <https://doi.org/10.1016/j.cmet.2018.07.001>.
- 1143 [93] C.E. Garner, S. Smith, B. de Lacy Costello, P. White, R. Spencer, C.S.J. Probert, N.M.  
1144 Ratcliffem, Volatile organic compounds from feces and their potential for diagnosis of  
1145 gastrointestinal disease, *The FASEB Journal.* 21 (2007) 1675–1688.  
1146 <https://doi.org/10.1096/fj.06-6927com>.
- 1147 [94] R.P. Arasaradnam, J.A. Covington, C. Harmston, C.U. Nwokolo, Review article: next  
1148 generation diagnostic modalities in gastroenterology--gas phase volatile compound  
1149 biomarker detection, *Aliment Pharmacol Ther.* 39 (2014) 780–789.  
1150 <https://doi.org/10.1111/apt.12657>.
- 1151 [95] C.S. Probert, S. Reade, I. Ahmed, Fecal volatile organic compounds: a novel, cheaper  
1152 method of diagnosing inflammatory bowel disease?, *Expert Review of Clinical  
1153 Immunology.* 10 (2014) 1129–1131. <https://doi.org/10.1586/1744666X.2014.943664>.
- 1154 [96] C.S.J. Probert, I. Ahmed, T. Khalid, E. Johnson, S. Smith, N. Ratcliffe, Volatile organic  
1155 compounds as diagnostic biomarkers in gastrointestinal and liver diseases, *J  
1156 Gastrointestin Liver Dis.* 18 (2009) 337–343.
- 1157 [97] J. La Nasa, T. Lomonaco, E. Manco, A. Ceccarini, R. Fuoco, A. Corti, F. Modugno, V.  
1158 Castelvetro, I. Degano, Plastic breeze: Volatile organic compounds (VOCs) emitted by  
1159 degrading macro- and microplastics analyzed by selected ion flow-tube mass  
1160 spectrometry, *Chemosphere.* (2020) 128612.  
1161 <https://doi.org/10.1016/j.chemosphere.2020.128612>.
- 1162 [98] C. Akarsu, S. Özdemir, Y. Ozay, Ö. Acer, N. Dizge, Investigation of two different size  
1163 microplastic degradation ability of thermophilic bacteria using polyethylene polymers,  
1164 *Environ Technol.* (2022) 1–22. <https://doi.org/10.1080/09593330.2022.2071638>.
- 1165 [99] A. Tareen, S. Saeed, A. Iqbal, R. Batool, N. Jamil, Biodeterioration of Microplastics: A  
1166 Promising Step towards Plastics Waste Management, *Polymers (Basel).* 14 (2022) 2275.  
1167 <https://doi.org/10.3390/polym14112275>.
- 1168 [100] G. Giangeri, M.S. Morlino, N. De Bernardini, M. Ji, M. Bosaro, V. Pirillo, P. Antoniali,  
1169 G. Molla, R. Raga, L. Treu, S. Campanaro, Preliminary investigation of microorganisms  
1170 potentially involved in microplastics degradation using an integrated metagenomic and  
1171 biochemical approach, *Sci Total Environ.* (2022) 157017.  
1172 <https://doi.org/10.1016/j.scitotenv.2022.157017>.
- 1173 [101] H. Sun, N. Chen, X. Yang, Y. Xia, D. Wu, Effects induced by polyethylene microplastics  
1174 oral exposure on colon mucin release, inflammation, gut microflora composition and  
1175 metabolism in mice, *Ecotoxicology and Environmental Safety.* 220 (2021) 112340.  
1176 <https://doi.org/10.1016/j.ecoenv.2021.112340>.
- 1177 [102] R. Gautam, J. Jo, M. Acharya, A. Maharjan, D. Lee, K.C. Pramod Bahadur, C. Kim, K.  
1178 Kim, H. Kim, Y. Heo, Evaluation of potential toxicity of polyethylene microplastics on  
1179 human derived cell lines, *Sci Total Environ.* (2022) 156089.  
1180 <https://doi.org/10.1016/j.scitotenv.2022.156089>.
- 1181 [103] M. Hesler, L. Aengenheister, B. Ellinger, R. Drexel, S. Straskraba, C. Jost, S. Wagner,  
1182 F. Meier, H. von Briesen, C. Büchel, P. Wick, T. Buerki-Thurnherr, Y. Kohl, Multi-  
1183 endpoint toxicological assessment of polystyrene nano- and microparticles in different  
1184 biological models in vitro, *Toxicology in Vitro.* 61 (2019) 104610.  
1185 <https://doi.org/10.1016/j.tiv.2019.104610>.
- 1186 [104] S. Wei, M.I. Bahl, S.M.D. Baunwall, C.L. Hvas, T.R. Licht, Determining Gut Microbial  
1187 Dysbiosis - A review of applied indexes for assessment of intestinal microbiota

- 1188 imbalances, *Appl Environ Microbiol.* (2021) AEM.00395-21, aem;AEM.00395-21v1.  
1189 <https://doi.org/10.1128/AEM.00395-21>.
- 1190 [105] Y. Zhang, S. Wang, V. Olga, Y. Xue, S. Lv, X. Diao, Y. Zhang, Q. Han, H. Zhou, The  
1191 potential effects of microplastic pollution on human digestive tract cells, *Chemosphere.*  
1192 (2021) 132714. <https://doi.org/10.1016/j.chemosphere.2021.132714>.
- 1193 [106] V. Stock, L. Böhmert, G. Coban, G. Tyra, M.-L. Vollbrecht, L. Voss, M.B. Paul, A.  
1194 Braeuning, H. Sieg, Microplastics and nanoplastics: Size, surface and dispersant - What  
1195 causes the effect?, *Toxicol In Vitro.* (2022) 105314.  
1196 <https://doi.org/10.1016/j.tiv.2022.105314>.
- 1197 [107] W. Wang, J. Zhang, Z. Qiu, Z. Cui, N. Li, X. Li, Y. Wang, H. Zhang, C. Zhao, Effects  
1198 of polyethylene microplastics on cell membranes: A combined study of experiments and  
1199 molecular dynamics simulations, *J Hazard Mater.* 429 (2022) 128323.  
1200 <https://doi.org/10.1016/j.jhazmat.2022.128323>.
- 1201 [108] M. Vökl, V. Jérôme, A. Weig, J. Jasinski, N. Meides, P. Strohmriegl, T. Scheibel, R.  
1202 Freitag, Pristine and artificially-aged polystyrene microplastic particles differ in regard  
1203 to cellular response, *J Hazard Mater.* 435 (2022) 128955.  
1204 <https://doi.org/10.1016/j.jhazmat.2022.128955>.
- 1205 [109] Y.-J. Yang, Y.-H. Ni, Gut microbiota and pediatric obesity/non-alcoholic fatty liver  
1206 disease, *J Formos Med Assoc.* 118 Suppl 1 (2019) S55–S61.  
1207 <https://doi.org/10.1016/j.jfma.2018.11.006>.
- 1208 [110] N. Gao, L. Yang, X. Lu, Z. Duan, L. Zhu, J. Feng, A review of interactions of  
1209 microplastics and typical pollutants from toxicokinetics and toxicodynamics perspective,  
1210 *J Hazard Mater.* 432 (2022) 128736. <https://doi.org/10.1016/j.jhazmat.2022.128736>.
- 1211 [111] V. Budhiraja, A. Urh, P. Horvat, A. Krzan, Synergistic Adsorption of Organic Pollutants  
1212 on Weathered Polyethylene Microplastics, *Polymers (Basel).* 14 (2022) 2674.  
1213 <https://doi.org/10.3390/polym14132674>.
- 1214 [112] Z. Wang, J. Ding, R.M. Razanajatovo, J. Huang, L. Zheng, H. Zou, Z. Wang, J. Liu,  
1215 Sorption of selected pharmaceutical compounds on polyethylene microplastics: Roles of  
1216 pH, aging, and competitive sorption, *Chemosphere.* (2022) 135561.  
1217 <https://doi.org/10.1016/j.chemosphere.2022.135561>.
- 1218