



HAL
open science

Species turnover between age groups of horses and positive network of co-occurrences define the structure of horse strongylid communities: a meta-analysis

Michel Boisseau, Núria Mach, Marta Basiaga, Slawomir Kornaś, Tetiana Kuzmina, Claire Laugier, Guillaume Sallé

► To cite this version:

Michel Boisseau, Núria Mach, Marta Basiaga, Slawomir Kornaś, Tetiana Kuzmina, et al.. Species turnover between age groups of horses and positive network of co-occurrences define the structure of horse strongylid communities: a meta-analysis. 2022. hal-03794670

HAL Id: hal-03794670

<https://hal.inrae.fr/hal-03794670>

Preprint submitted on 3 Oct 2022

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial - NoDerivatives 4.0 International License

1 **Species turnover between age groups of horses and positive network of co-**
2 **occurrences define the structure of horse strongylid communities: a meta-**
3 **analysis**

4 Michel Boisseau^a, Núria Mach^b, Marta Basiaga^c, Sławomir Kornaś^c, Tetiana Kuzmina^d, Claire Laugier^e,
5 Guillaume Sallé^{a*}

6

7 a, INRAE, Université de Tours, ISP, Nouzilly, France

8 b, UMR 1313 GABI, INRAE, AgroParisTech, Université Paris-Saclay, Jouy-en-Josas, France

9 c, Department of Zoology and Animal Welfare, Faculty of Animal Science, University of Agriculture in
10 Kraków, 24/28 Mickiewicza Av., 30-059 Kraków, Poland

11 d, Department of Parasitology I. I. Schmalhausen Institute of Zoology NAS of Ukraine, Kyiv, Ukraine

12 e, Conseil Général de l'Alimentation, de l'Agriculture et des Espaces Ruraux, Ministère de l'Agriculture
13 et de l'Alimentation, Paris, France

14 *Corresponding author: guillaume.salle@inrae.fr

15

16

17

18

19 **Abstract**

20 Grazing horses are infected by a wide range of strongylid species mostly located in the large intestine.
21 Despite their impact on equine health and the emergence of drug resistant isolates, the phenology of
22 these nematodes has been poorly characterized and the rules structuring their assembly as a
23 community are not understood. Here, we compiled data on 46 equine strongylid species collected
24 worldwide at the regional or horse scales (upon deworming or after necropsy) to analyse their richness,
25 diversity and associated factors of variation. Worldwide, twelve species from the *Cylicocyclus* (n = 4),
26 *Cylicostephanus* (n = 3), *Coronocyclus* (n = 2) and *Cyathostomum* (n = 2) genera were found in at least
27 75% of sites. Geoclimatic conditions had a limited effect on strongylid communities, but reduced
28 species richness was found under the temperate European area. The recovery method did not affect
29 species richness and differences on the temporal and sampling effort scales between studies applying
30 either methods underpinned heterogeneous variances in community diversity.
31 At the horse level, rarefaction curves correlated poorly to parasite egg excretion, suggesting little
32 contribution of community diversity to this trait. Using a diversity partitioning approach, we found that
33 within-host diversity represented half of overall diversity underscoring the importance of host density
34 and environmental contamination to the diversity of strongylid communities. While this is expected to
35 erase diversity across communities, species turnover between age classes was the second most
36 important contributor to overall diversity (23.9%). This was associated with a network of positive co-
37 occurrences between the four most prevalent genera that we resolved at the anatomical niche level.
38 Altogether this pattern of β -diversity maintenance across age classes combined with positive co-
39 occurrences may be grounded by priority effects between major species. Our findings set the first
40 assembly rules of equine strongylid communities.

41 **Keywords:** community, co-occurrence, diversity, horse, parasite, strongylid

42 1. Introduction

43 Grazing horses are naturally infected by a diverse parasitic fauna including Strongylidae (Taylor et al.,
44 2012). They encompass subfamilies Cyathostominae and Strongylinae (Lichtenfels et al., 2008). The
45 use of anthelmintics has resulted in prevalence levels for *Strongylus* sp. that has the highest fatality
46 rate among Strongylinae (Herd, 1990), but has led to the emergence of resistant cyathostomin isolates
47 worldwide (Matthews, 2014). affects horse growth and the massive emergence of encysted larvae can
48 lead to fatal cases of cyathostomosis, the main cause of parasite mortality in young horses in
49 Normandy, France (Sallé et al., 2020). Despite their importance in veterinary medicine, the phenology
50 and the rules structuring the assembly of remain largely undefined due in part to the complexity of
51 collecting data from large mammalian hosts under experimental conditions.

52 Inter-individual variation exists in host susceptibility to infection, with significant variation across age
53 groups (Debeffe et al., 2016; Kornaś et al., 2015; Relf et al., 2013; Wood et al., 2013). In addition,
54 climatic conditions affect the development of the free-living stages on pastures (Nielsen et al., 2007)
55 and species reappearance following anthelmintic drug treatment varies across species (Kooyman et
56 al., 2016) suggesting that combinations of their phenology or genetic diversity increase their fitness.
57 Despite knowledge, variations in community structure in anatomical niches or environmental conditions
58 remain largely uncharacterised.

59 Better defining the patterns of variation in helminth communities could however greatly improve their
60 management in the field, i.e. by preventing the emergence of subdominant species following
61 elimination of more dominant ones (Lello et al., 2004) or anticipating consequences of infections by
62 other pathogens (Brosschot et al., 2021; Sweeny et al., 2020; Telfer et al., 2010). Data on helminth
63 community assembly have been gathered across a wide range of host species including humans
64 (Tchuem Tchuente et al., 2003), rabbit (Lello et al., 2004), rodent (Behnke, 2008; Dallas et al., 2019),
65 fishes (Poulin and Valtonen, 2002) with evidence of significant interactions between parasites. In
66 horses, community structure has been determined from necropsy reports or after deworming (Gawor,
67 1995; Kornas et al., 2011; Kuzmina et al., 2005; Kuzmina et al., 2011; Kuzmina et al., 2016; Lind et al.,

68 2003; Mfitilodze and Hutchinson, 1990; Ogbourne, 1976; Reinemeyer et al., 1984). However, these
69 reports remain often descriptive and limited exploration of the relationship between community
70 structure and host and environmental variables been conducted. For instance, variation in species
71 richness and diversity along the intestinal tract of horses has been reported (Collobert-Laugier et al.,
72 2002; Gawor, 1995; Ogbourne, 1976) and recent meta-analysis reported the impact of parasite
73 recovery method and geographical region on cyathostomin abundance or prevalence (Bellaw and
74 Nielsen, 2020). These studies however did not address how community structure, richness and
75 diversity varied with environmental factors of interest. In addition, co-occurrence patterns between
76 equine parasite species remain poorly characterized with contradictory findings reported (Sallé et al.,
77 2018; Stancampiano et al., 2010).

78 Here, we performed an exhaustive literature search to evaluate how equine strongylid communities
79 were shaped by environmental and host factors. In addition, we re-analyzed community data gathered
80 from individual horses to partition diversity across environmental factors and to build co-occurrence
81 networks at the anatomical niche level. Altogether, our results expand previous reports while opening
82 new perspectives on the parasite composition and structure.

83

84 2. **Materials and methods**

85 The python and R scripts used to collect and analyse the data and associated datasets will be
86 deposited at: https://github.com/MichelBoisseau37/script_meta_analyse upon manuscript
87 acceptance.

88

89 **2.1. Literature search and inclusion criteria**

90 Studies were gathered from the Medline database (<https://pubmed.ncbi.nlm.nih.gov/>) using "strongyl*
91 horse" as keywords, yielding a database of 1,032 articles. Article titles and abstracts were
92 subsequently screened for particular terms including all possible combinations of the following:
93 "intensity", "Intensity", "Abundance", "abundance", "Prevalence", "prevalence", "communit",

94 “Communit”, “Helminth” or “helminth”. To identify studies reporting strongylid burden, we also
95 included the main genera (“*Cylicocyclus*”, “*Cyathostomum*”, “*Cylicostephanus*”) as keywords. This
96 filtering step retained 300 papers. For these, titles and abstracts were read to determine whether
97 findings of strongylid species prevalence and abundance would be available.

98 To broaden our database, we also run manual search on google (<https://scholar.google.com/>) using
99 the keywords “prevalence abundance Cyathostominae horses” and “prevalence abundance
100 Strongylidae horse”. This query identified six additional studies. Additional studies were identified after
101 inspection of the references cited in three papers (Gawor, 1995; Ogbourne, 1976) that yielded
102 prevalence data from 13 additional studies.

103 To be included in the meta-analysis, reported articles had to meet several criteria. First, studies were
104 included in the dataset if they reported parasite prevalence (proportion of infected host), abundance
105 (the number of parasite individuals per host, including uninfected individuals) or relative abundance.
106 Second, published data were only included if the number of horses sampled was reported.

107 Finally, 39 papers met the aforementioned selection criteria (Supplementary Table 1). Among these, 16
108 papers presented both abundance and prevalence data, while others either only reported abundance
109 ($n = 5$) or prevalence ($n = 18$) data. Papers from Africa ($n = 2$) and Australia ($n = 3$) were included to
110 determine average relative abundance and prevalence across region-based studies but were not
111 considered in any further analyses as too few replicates were available to test for environmental effects.
112 For analytical purposes, papers were further divided into studies to account for the distinct
113 experimental groups considered. This process yielded 69 community sets for 46 species encountered
114 in at least one study.

115

116 **2.2. Parasite, environmental and host variables**

117 When available, metadata related to environmental factors were collected. Horse age was not
118 considered for the continent-wide study as only scarce and inaccurate data were available. The same
119 applied for horse breed or management type (wild or managed horses) that were reported in three

120 studies only. Following edition, information regarding continent, country, sampling effort (number of
121 horses sampled), the recovery method used (necropsy or deworming) were included.

122 Climate conditions were determined using the Köppen-Geiger climate classification
123 (<http://hanschen.org/koppen/#data>) matching the sampling area. The Köppen-Geiger climate
124 classification is encoded by a three- or two-letter system corresponding to the type of climate, rainfall
125 pattern and temperature range respectively. For studies with no precise location indicated, the closest
126 main city (state or country capital city) coordinates were used as a proxy (Supplementary Table 1). The
127 different climates were clustered into broad climatic categories, namely continental, tropical, temperate
128 and arid.

129

130 **2.3. Statistical analysis**

131 Data were analysed with R v. 3.6.3 and v. 4.0.2.

132 2.3.1. Analysis of worldwide strongylid community diversity

133 2.3.1.1. Considered regional community sets

134 Prevalence and relative abundance matrices were merged together into a single presence/absence
135 matrix (noted 0 and 1 for absence and presence respectively). Analyses were run on the species
136 presence/absence matrix for two reasons. First, this approach makes use of every available study (no
137 matter of the reported measure) thereby providing greater statistical power than an approach
138 considering abundance or prevalence data only. Second, most studies ($n = 29$ out of the 39 considered)
139 did not report any dispersion parameter of parasite abundance and prevalence preventing any
140 implementation of meta-analysis frameworks. In addition, because count data usually follow a negative
141 binomial distribution, reported arithmetic mean values can be poor descriptors of the parasite
142 population. This also speaks against sound estimation.

143 Because of the data structure, geoclimatic conditions (aggregated continent and climate zone) and the
144 recovery method (deworming or necropsy) were confounded. In addition, some conditions were under-
145 represented (less than five community sets) and not considered further. This included the community

146 sets from Asia ($n = 1$), Australia ($n = 3$), or sets from temperate Europe collected after deworming ($n =$
147 3).

148 As the deworming strategy was only applied in Europe, a continental European subset was considered
149 ($n = 33$ communities) to evaluate how the recovery method could affect the strongylid diversity.
150 Similarly, the geoclimatic zone effect was tested on communities recovered under the same necropsy
151 framework leaving 30 community sets that fell in temperate America ($n = 8$), tropical America ($n = 7$),
152 temperate Europe ($n = 11$) or continental Europe ($n = 5$).

153

154 2.3.1.2. Diversity analyses

155 Species richness was taken as an α -diversity estimator, and the β -diversity index was estimated with
156 the Jaccard index using the vegan R package. (version 2.5-7). Inter-sample distances were visualized
157 using the Non-Metric Scaling (NMDS) ordination method through the “metaMDS” function in the vegan
158 R package. This method generated fictitious axes maximizing the variance and allowing an optimal
159 visualization of distances. NMDS is the most appropriate representation because it minimizes
160 dissimilarities between similar communities and vice versa for more remote communities. NMDS was
161 run on the whole set of communities to establish how environmental factors (aggregated factor of the
162 geoclimatic zone and the recovery method), the sampling effort (number of horses) and the year of
163 publication were related to community structures.

164 A permutational multivariate analysis of variance (PERMANOVA) using distance matrix (adonis function
165 of the vegan R package) was then applied to identify the factors that would influence most the
166 community structure and to obtain the percentage of variance explained by the factors of interest,
167 i.e. sampling effort (number of horses), the geoclimatic zone and the recovery method using
168 appropriate subsets to control for environmental conditions.

169 The respective effects of the recovery method (tested on the European subset) or the geoclimatic area
170 (tested on the necropsy-based subset) on species richness were estimated using a linear model, fitting
171 sample size, year of publication and the factor of interest as fixed effects.

172

173 2.3.1.3. Effect of environmental factors on parasite prevalence across region-
174 based community sets

175 Last, we aimed to identify strongylid species whose presence or absence in a given set would be
176 associated with variation in the recovery method or the geoclimatic area. This was applied to the
177 aforementioned community subsets to account for the data structure. In each case, the presence or
178 absence of worm strongylid species was modelled using a logistic regression model, fitting an
179 interaction term between worm species and either the recovery method (European continental subset)
180 or the geoclimatic area (necropsy-based data). Because quasi-complete separation of species with
181 environmental factors of interest occurred (identified with the “detect_separation” function of the
182 *detectseparation* v0.1 package), we restricted this analysis to ten species. In addition, model
183 convergence could not be achieved in case the community set was fitted as a random effect. To deal
184 with this while accounting for inter-sets variation, we fitted the year of publication, sampling effort and
185 the species richness of the community as covariates.

186

187 2.3.2. Analysis of diversity at the horse level

188 These analyses were run on horse-based strongylid community data from three published studies,
189 either collected after the necropsy of 46 horses in France (Collobert-Laugier et al., 2002) or the
190 deworming of 48 horses in Poland (Sallé et al., 2018) and or 197 horses Ukraine (Kuzmina et al., 2016)
191 respectively.

192

193 2.3.2.1. Relationship between diversity and faecal egg count at the horse level

194 To evaluate how strongylid community diversity could relate to measured Faecal Egg Count (FEC) in
195 their host, we estimated Spearman’s correlations between FEC and species richness, or FEC and the
196 Gini-Simpson index which corresponds to the slope of the species accumulation curve at its basis and

197 relates to species dominance (Chase and Knight, 2013). This was not done on the necropsy data as
198 FEC measures were not available.

199

200 2.3.2.2. Impact of horse age on strongylid community diversity

201 To study the interactions between strongylid species within their hosts, we re-analyzed individual
202 necropsy report data collected in Normandy (Collobert-Laugier et al., 2002). Abundances of 20
203 strongylid species and other non- strongylid species (*Anoplocephala* sp., *Oxyuris* sp.) were quantified
204 within the caecum, ventral- and dorsal-colon of 46 horses. The data from 36 horses that had their age
205 registered and were infected by at least one parasite species were kept. To avoid spurious signals
206 associated with rare observations, we focused on the 15 species achieving at least 10% prevalence.
207 To capture the strongylid community assembly pattern across anatomical niches and age group, we
208 applied the hierarchical Bayesian joint species distribution model (Ovaskainen et al., 2017; Tikhonov et
209 al., 2017) implemented in the HMSC R package (v. 3.0-10) (Tikhonov et al., 2020). This model offers an
210 integrated framework to explicitly model species-species interactions while simultaneously accounting
211 for environmental covariates, species traits or their phylogenetic relationships (Ovaskainen et al., 2017;
212 Tikhonov et al., 2017). We used the same approach as described elsewhere (Abrego et al., 2020), fitting
213 two models that either accounted for the worm burden measured at the horse level or for additional
214 fixed effects including the anatomical niche and horse age. In both cases, horse and horse x anatomical
215 niche were fitted as random effects to account for inter-horse variation in species pairs co-occurrences
216 and for sample level variation not accounted for otherwise (Abrego et al., 2020). A hurdle-type model
217 was considered to account for the overdispersed nature of the count data, modeling log-transformed
218 positive species counts as normal and the binary presence or absence status using a probit-link
219 function (Abrego et al., 2020). Model predictive power was assessed by Area Under the Curve (AUC)
220 and R^2 parameters averaged across species for presence-absence and abundance data. Parameters
221 were estimated from 4,000 posterior samples collected every 100 iterations from four Markov Chain
222 Monte Carlo (MCMC) run for 150,000 iterations with a third discarded as burn-in.

223

224 2.3.3. Analysis at the anatomical niche level: species co-occurrences from necropsy
225 data

226 To study the interactions between strongylid species within their hosts, we re-analyzed individual
227 necropsy report data collected in Normandy (Collobert-Laugier et al., 2002). Abundances of 20
228 strongylid species and other non- strongylid species (*Anoplocephala* sp., *Oxyuris* sp.) were quantified
229 within the caecum, ventral- and dorsal-colon of 46 horses. The data from 36 horses that had their age
230 registered and were infected by at least one parasite species were kept. To avoid spurious signals
231 associated with rare observations, we focused on the 15 species achieving at least 10% prevalence.
232 To capture the strongylid community assembly pattern across anatomical niches and age group, we
233 applied the hierarchical Bayesian joint species distribution model (Ovaskainen et al., 2017; Tikhonov et
234 al., 2017) implemented in the HMSC R package (v. 3.0-10) (Tikhonov et al., 2020). This model offers an
235 integrated framework to explicitly model species-species interactions while simultaneously accounting
236 for environmental covariates, species traits or their phylogenetic relationships (Ovaskainen et al., 2017;
237 Tikhonov et al., 2017). We used the same approach as described elsewhere (Abrego et al., 2020), fitting
238 two models that either accounted for the worm burden measured at the horse level or for additional
239 fixed effects including the anatomical niche and horse age. In both cases, horse and horse x anatomical
240 niche were fitted as random effects to account for inter-horse variation in species pairs co-occurrences
241 and for sample level variation not accounted for otherwise (Abrego et al., 2020). A hurdle-type model
242 was considered to account for the overdispersed nature of the count data, modeling log-transformed
243 positive species counts as normal and the binary presence or absence status using a probit-link
244 function (Abrego et al., 2020). Model predictive power was assessed by Area Under the Curve (AUC)
245 and R^2 parameters averaged across species for presence-absence and abundance data. Parameters
246 were estimated from 4,000 posterior samples collected every 100 iterations from four Markov Chain
247 Monte Carlo (MCMC) run for 150,000 iterations with a third discarded as burn-in.

248

249 3. Results

250 3.1. Descriptive features of strongylid communities

251 The breakdown of publication counts for factors of interest is provided in Supplementary Table 2. As
252 can be seen in Figure 1A, studies using deworming were mostly carried out in Eastern Europe. Our
253 data showed an average number of horses per study of 28.65, ranging between 2 and 150 individuals.
254 Strongylid communities were dominated by a small number of species, *i.e.* *Cylicocyclus nassatus*,
255 *Cylicostephanus longibursatus*, *Cyathostomum catinatum* (Figure 1B, supplementary table 3) with an
256 average prevalence of $83.97\% \pm 3.38$. Along these three dominant species, a set of eight subdominant
257 species with an average prevalence of $54.96\% \pm 12.68$. On the other hand, 31 rare species had an
258 average prevalence of $8.93\% \pm 6.70$. These mean prevalence levels reported within horse populations
259 were slightly corroborated when considering the presence-absence pattern at the regional scale
260 (Figure 1C). Specifically, the dominant and subdominant species defined by an average prevalence of
261 25% or more at the horse population level were present in 75% of studies. Contrasting this broad
262 pattern, *Cylicostephanus goldi* was relatively frequent among horses but its presence varied widely
263 across studies (Figure 1C).

264 At the horse level, strongylid community diversity varied widely as illustrated by species accumulation
265 curves derived from the three available studies. The sampling of 95% of available diversity, required
266 22 horses (46% of total horse sample) for necropsy-based data from France (Collobert-Laugier et al.,
267 2002). On the contrary, strongylid communities gathered upon deworming reached the same threshold
268 for 25% of available horses (12 and 50 horses in the studies in Poland or Ukraine communities).

269 More diverse strongylid communities may reflect reduced drug usage and might be associated with
270 higher parasite egg excretion. However, limited agreement was found between Faecal Egg Count and
271 community diversity as measured by the rarefaction curves. Significant correlation was found between
272 FEC and species richness (Supplementary Figure 1) in strongylid communities gathered in Poland
273 (Spearman's $\rho = 0.31$, $P = 9 \times 10^{-3}$) but not in Ukraine (Spearman's $\rho = 0.03$, $P = 0.69$). Correlation
274 between FEC and the rarefaction curve at its basis (Gini index) was not significant in any case

275 (Spearman's $\rho = -0.14$, $P = 0.06$ and $\rho = 0.24$, $P = 0.1$ in Polish and Ukrainian data respectively),
276 suggesting that species dominance does not affect FEC.

277

278 **3.2. Diversity analysis of worldwide strongylid communities**

279 As a first step, an exploratory multivariate analysis was applied to our meta-community matrix (Figure
280 2). This analysis highlighted the outstanding contribution of the strongylid recovery method on
281 community diversity (Figure 2). This clustering also closely matched a gradient defined by differences
282 in the temporal and the sampling effort scales. Publication times ranged between 2005 and 2018 or
283 between 1935 and 2016 for deworming- and necropsy-based sets respectively. In addition,
284 significantly more horses were sampled in necropsy-based studies (36.84 and 18.9 horses sampled
285 respectively, $P = 2.2 \times 10^{-16}$).

286 The clear distinction between considered recovery methods was supported by the PERMANOVA ($P =$
287 3×10^{-4}) run on the strongylid communities gathered under the same geo-climatic conditions, i.e.
288 European continental climate. This was however associated with higher variance in deworming-based
289 methods ($F_{1,31} = 6.51$, $P = 0.02$, supplementary Figure 2).

290 Logistic regression models found significant variation for *Triodontophorus serratus* and *Strongylus*
291 *vulgaris* that were less often recovered in the five considered necropsy-based sets relative to the 28
292 deworming studies performed under the same geoclimatic conditions (interaction terms equal to -4.1
293 ± 1.96 , $P=0.03$ and -4.62 ± 2.00 , $P = 0.02$). Of note, quasi-complete separation with the recovery
294 method was found for 13 species, either showing complete absence (*Cylicostephanus. bidentatus*) or
295 complete presence (*Cylicocyclus ashworthi*, *Gyalocephalus capitatus*, *Cylicodontophorus bicoronatus*,
296 *Coronocyclus coronatus*, *Cylicocyclus. elongatus*, *Coronocyclus labratus*, *Poteriostomum*
297 *imparidentatum*, *Petrovinema poculatum*, *Parapoteriostomum mettami*, *Poteriostomum ratzi*,
298 *Cylicocyclus Radiatus*, *Triodontophorus tenuicollis*) in studies practicing necropsy (Supplementary
299 Figure 3). The species richness was however not significantly different between the recovery methods
300 ($P = 0.6$).

301 On the contrary, the ordination analysis found limited clustering associated with geoclimatic factors
302 (Figure 2). This was corroborated by PERMANOVA on the set of communities collected after necropsy
303 ($P = 1.9 \times 10^{-3}$). In that case, species richness was lower in the temperate European area (-6.57 , $P =$
304 5×10^{-3}). Apart from the 15 species with quasi-complete separation in one of the considered geoclimatic
305 zones (Supplementary Figure 4), we did not find significant evidence of variation in their prevalence
306 pattern across geoclimatic conditions.

307 To sum up, the recovery method was not associated with variation in species richness but it was
308 associated with significant turnover across communities, likely mirroring scale differences between the
309 considered sets. On the contrary, species turnover across geoclimatic areas was limited and reduced
310 richness was found under temperate Europe.

311

312 **3.3. Strongylid community dynamic across age groups**

313 Partitioning of overall γ -diversity across Ukraine operations revealed an outstanding contribution of
314 within-horse α -diversity (46.5% of overall diversity) relative to species turnover coefficients (Figure 3).
315 The observed value was however lower than under null model expectations (22.6% difference, $P=10^{-3}$)
316 ³). Species turnover between horse age-group was the second most important contributor to γ -diversity
317 (23.9% of γ -diversity), significantly higher than that expected from a null model ($P=10^{-3}$, Figure 3) and
318 higher than that explained by between-farm variation (20% of γ -diversity, Figure 3). Of note, observed
319 between-horse turnover did not depart from the simulated null distribution (0.06% difference, $P=0.4$,
320 Figure 3).

321 Because of the importance of age group in driving strongylid community structure, we also searched
322 for indicator species that would preferentially be associated with a given age group. This analysis found
323 that *C. labiatus* and *C. ashworthi* were preferentially encountered in two-years old horses, irrespective
324 of the considered coefficient (indicator values of 0.659, $P=7 \times 10^{-3}$, and 0.585, $P = 0.04$ or association
325 coefficient equal to 0.42, $P = 0.01$ and 0.41, $P = 0.01$ for *C. labiatus* and *C. ashworthi*, respectively).

326 Altogether these results suggest that horse-level community diversity and between age-class turnover
327 are the major contributors driving measured diversity under the European continental conditions
328 encountered across Ukraine operations.

329

330 **3.4. Co-occurrence of species within different anatomical niches**

331 Anatomical niche was the last scale examined to determine factors driving strongyloid community
332 structure. The predictive power of the modeling approach was high with a mean AUC of 0.88 and R^2
333 of 0.79 for models based on presence-absence or positive abundance data.

334 Raw or residual co-occurrences only supported positive interactions between the considered
335 strongyloid species (Figure 4). The correction for environmental covariates erased around half of
336 observed co-occurrences in both abundance (11 out of 26) and presence-absence (21 out of 42)
337 models. This was mostly due to correction for the anatomical niche that accounted for a third of total
338 variance while another third was equally shared across horse age and infection intensity (Table 1). Of
339 note, the presence of *C. goldi* was largely dependent on inter-horse variation but not its abundance
340 (supplementary Figures 5 and 6) suggesting limited expansion abilities once established in a host.

341 Residual models shared 12 positive co-occurrences involving *C. nassatus*, *C. labratus*, *C. coronatus*,
342 *Cylicostephanus minutus* and *C. catinatum* (Figure 4). However, more co-occurrences were found with
343 the presence-absence data. This suggests that strongyloid species tend to have similar co-infection
344 patterns but their simultaneous presence does not necessarily affect their respective relative
345 abundances.

346

347 **4. Discussion**

348 In this meta-analysis, we have explored the principles of strongyloid community structuring across
349 various scales, ranging from regional down to anatomical resolutions. Our findings define a community
350 with little rearrangements across contrasted geoclimatic conditions, whose diversity is mostly
351 grounded on within-host richness and turnover across-age groups. At the anatomical scale, strongyloid

352 communities are structured around positive co-occurrences fluctuating across anatomical niches. As
353 such, our results complement and expand the past work by Bellaw and Nielsen (Bellaw and Nielsen,
354 2020) with additional horse-based data and different approaches and interpretations of the findings.
355 While we tried to account for common sources of variation at the regional scale, this was often limited
356 by the paucity of available metadata or by the data structure of past published work. For instance, age
357 group was poorly annotated across studies and horse sex was seldom reported despite significant
358 species turnover described between both sexes (Sallé et al., 2018). In addition, the lack of appropriate
359 summary statistics in past published work constrained our analysis to a presence-absence framework
360 that certainly overlooks finer grain patterns found with abundance data (Brian and Aldridge, 2020). This
361 is mostly the result of the existing asymmetry between presence of the parasite - a single infected
362 individual is enough - and absence - that requires the whole species to be absent (Chase et al., 2019).
363 Reliance on this binary matrix defined at the regional level also prevented the use of other statistics
364 like the effective number of species (Chase and Knight, 2013) or individual based rarefaction (Chase et
365 al., 2018) to better account for differences in scales (geographical and temporal) across the considered
366 species sets. As a result, the significant species turnover across the recovery methods that we
367 observed as other authors (Bellaw and Nielsen, 2020) may simply reflect scale differences between
368 studies. We took care to correct for differences in sampling effort and restricted the analysis to the
369 same geoclimatic conditions. But the wider diversity variance found for strongylid communities
370 examined after horse deworming is also likely to reflect the higher abundance of this type of data
371 relative to the necropsy-based communities as the more communities are sampled, the more likely
372 that species recovery will vary (Chase et al., 2019).

373 Overall, our results highlighted a small group of abundant and widespread species *C. nassatus*, *C.*
374 *longibursatus* and *C. catinatum* that clearly separated from other species. Of note, these species were
375 consistently associated with drug resistance reports suggesting their phenology may have favoured
376 higher fitness towards the use of modern anthelmintics (van Doorn et al., 2014). It is unclear however
377 what phenotypic trait underpins their higher abundance and prevalence in horse populations. Their

378 respective fecundities estimated from the number of eggs found in female worms *utero* was among
379 the lowest suggesting their dispersal is unlikely to define a better ability to colonize their hosts (Sallé et
380 al., 2018).

381 In addition, analysis of horse-based data expanded our understanding of strongylid communities. First,
382 we found limited correlation between parasite egg excretion, a trait used to monitor infection in horses,
383 and the strongylid community diversity. This was against our hypothesis that more permissive hosts
384 would tolerate higher parasite burden, thereby increasing the community species richness. This lack
385 of linear relationship may mirror variations in respective strongylid species fecundities (Kuzmina et al.,
386 2012) and reflect possible density dependence effects that are known to occur in trichostrongylid
387 species infecting ruminants (Bishop and Stear, 2001).

388 Second, we found that overall diversity was mostly partitioned across within-host richness and species
389 turnover across horse age groups, and to a lesser-extent between farm variation. Species richness is
390 conditional upon availability of host resources, leading to higher richness in larger animal species or
391 for higher host density (Kamiya et al., 2014). Richness also depends on environment saturation in
392 infective larvae (Shmida and Wilson, 1985). The latter two aspects underscore the importance of
393 pasture hygiene and limited stocking density for the management of strongylid infection in grazing
394 livestock. Of note, environmental saturation is usually expected to erase species turnover among hosts
395 (Vannette and Fukami, 2017) but we identified a significant turnover between age groups. This turnover
396 remains to be fully dissected to disentangle between the respective effects of selection applied through
397 the mounting of an effective immune response as horses age (Debeffe et al., 2016; Kornaś et al., 2015;
398 Relf et al., 2013; Wood et al., 2013) or the drift arising from putative competitive processes between
399 species. To this respect, the increased species turnover across age groups may arise from priority
400 effects between strongylid species that would counterbalance the effect of environmental saturation in
401 eggs and infective larvae (Vannette and Fukami, 2017). Such priority effect would in turn yield a network
402 of positive co-occurrence between species as found for mixed infection of rodents by malaria species
403 (Ramiro et al., 2016).

404 Using data with anatomical niche resolution, we performed a first investigation of co-occurrence
405 between strongylid species. This analysis did not reveal any negative effects thereby corroborating
406 past results obtained after horse deworming in Poland conditions (Sallé et al., 2018). Of note, a few co-
407 occurrences were dependent on associated covariates including the anatomical niche or horse age
408 that may define optimal environmental conditions shared across parasite species. However, a core
409 network of positive interactions remained after accounting for environmental variations between
410 species from four distinct genera, namely *Cyathostomum* (*C. catinatum*), *Cylicocyclus* (*C. nassatus* and
411 *C. insigne*), *Cylicostephanus* (*C. minutus*) and *Coronocyclus* (*C. coronatus* and *C. labratus*). This would
412 be compatible with facilitative relationships between more divergent species that should be less prone
413 to resource competition (MacArthur and Pianka, 1966) although co-existence of the *Cylicocyclus* and
414 *Coronocyclus* members depart from this expectation. While the complexity of underpinning biological
415 processes may obscure the observed co-occurrences patterns (Blanchet et al., 2020), these
416 interactions were found across different parasite communities and with the lowest possible resolution
417 thereby supporting their biological meanings (Behnke, 2008). However, covariates like past infections
418 and treatment history were not accounted for and may contribute to this network. Besides, the
419 contribution of the resident gut microbiota may play a role in the definition of this particular structuring
420 as significant interactions between microbial and helminth communities are known to occur in horses
421 (Clark et al., 2018; Peachey et al., 2019; Walshe et al., 2020).

422 Altogether, priority effects between equine strongylid species in young horses may contribute to
423 enhance the species turnover found between age groups and to define the network of positive co-
424 occurrences between the most dominant genera. The simultaneous monitoring of both the microbial
425 and parasite communities using the latest barcoding approaches (Poissant et al., 2020) should provide
426 valuable contributions to challenge these hypotheses.

427

428 **Acknowledgements**

429 MBo and this work was supported by the Institut Français du Cheval et de l'Équitation (IFCE) and
430 Fonds Éperon grant.

431

432 **Figure and table legends**

433 **Figure 1. Descriptive features of worldwide equine strongylid communities**

434 A: World map of study distribution coloured by worm collection method, either following deworming
435 (green) or necropsy (orange).

436 B: Strongylid species (colored by their respective genera) relative abundances and prevalences
437 averaged across published mean estimates. Dotted lines distinguish between rare, subdominant and
438 dominant species in the left, middle and right quadrants respectively.

439 C: Comparison of observed species prevalence across studies (from the presence-absence matrix)
440 against the average published prevalence estimates. Colour and dotted lines match that used in B.

441

442 **Figure 2. Ordination applied to strongylid meta-community with fitted environmental variables**

443 Figure shows the first two axes from a NonMetric Dimensional Scaling (NMDS) ordination applied to
444 published metacommunity data. The colours correspond to the different types of environment
445 considered, coded as Continent – Climate – Recovery method. For visualization, species names were
446 abbreviated with correspondence given in Supplementary Table 3.

447

448 **Figure 3. Hierarchical diversity partitioning of overall diversity**

449 Within host alpha diversity and β -diversity (turnover) for the scales of interest (host, age class and farm),
450 observed (purple) and simulated (green) estimates are given in percentage of overall diversity
451 explained. Error bars materialize a 95% confidence interval of the simulated data.

452

453 **Figure 4. Raw and residual strongylid species co-occurrence matrices estimated from presence-**
454 **absence data or positive abundance data**

455 Co-occurrence matrices derived from presence-absence (left) or positive abundance (right) data are
456 represented for 15 strongylid species collected identified in 36 necropsied horses from Normandy,
457 France. In each case, off-diagonal elements describe the raw (above) and residual (below) co-
458 occurrences.

459

460 **Table 1. Proportion of variance explained by considered covariates in the residual co-occurrence**
461 **model**

462 For each covariate of interest (fixed or random effect), the proportion of variance explained is given for
463 the two model types considered using positive abundance data or presence-absence data.

464

Effect type	Covariate	Abundance data (%)	Occurrence data (%)
	Anatomical niche	31.2	32.7
Fixed effect	Horse age	14.6	12.9
	Worm burden	14.1	10.8
Random effect	Horse	12.8	33.4
	Horse x niche	27.3	19.1

465

466

467

468 **References**

- 469 Abrego, N., Roslin, T., Huotari, T., Tack, A.J., Lindahl, B.D., Tikhonov, G., Somervuo, P., Schmidt, N.M.,
470 Ovaskainen, O., 2020. Accounting for environmental variation in co-occurrence modelling reveals the
471 importance of positive interactions in root-associated fungal communities. *Molecular Ecology* 29,
472 2736-2746.
- 473 Behnke, J., 2008. Structure in parasite component communities in wild rodents: predictability, stability,
474 associations and interactions or pure randomness? *Parasitology* 135.
- 475 Bellaw, J.L., Nielsen, M.K., 2020. Meta-analysis of cyathostomin species-specific prevalence and
476 relative abundance in domestic horses from 1975–2020: emphasis on geographical region and
477 specimen collection method. *Parasites & Vectors* 13, 1-15.
- 478 Bishop, S., Stear, M., 2001. Inheritance of faecal egg counts during early lactation in Scottish Blackface
479 ewes facing mixed, natural nematode infections. *Animal Science* 73, 389-395.
- 480 Blanchet, F.G., Cazelles, K., Gravel, D., 2020. Co-occurrence is not evidence of ecological interactions.
481 *Ecology Letters* 23, 1050-1063.
- 482 Brian, J., Aldridge, D., 2020. Abundance data applied to a novel model invertebrate host sheds new
483 light on parasite community assembly in nature. *bioRxiv*.
- 484 Brosschot, T.P., Lawrence, K.M., Moeller, B.E., Kennedy, M.H., FitzPatrick, R.D., Gauthier, C.M., Shin,
485 D., Gatti, D.M., Conway, K.M., Reynolds, L.A., 2021. Impaired host resistance to *Salmonella* during
486 helminth co-infection is restored by anthelmintic treatment prior to bacterial challenge. *PLOS*
487 *Neglected Tropical Diseases* 15, e0009052.
- 488 Chase, J.M., Knight, T.M., 2013. Scale-dependent effect sizes of ecological drivers on biodiversity:
489 why standardised sampling is not enough. *Ecology letters* 16, 17-26.
- 490 Chase, J.M., McGill, B.J., McGlinn, D.J., May, F., Blowes, S.A., Xiao, X., Knight, T.M., Purschke, O.,
491 Gotelli, N.J., 2018. Embracing scale-dependence to achieve a deeper understanding of biodiversity
492 and its change across communities. *Ecology letters* 21, 1737-1751.

493 Chase, J.M., McGill, B.J., Thompson, P.L., Antão, L.H., Bates, A.E., Blowes, S.A., Dornelas, M.,
494 Gonzalez, A., Magurran, A.E., Supp, S.R., 2019. Species richness change across spatial scales. *Oikos*
495 128, 1079-1091.

496 Clark, A., Sallé, G., Ballan, V., Reigner, F., Meynadier, A., Cortet, J., Koch, C., Riou, M., Blanchard, A.,
497 Mach, N., 2018. Strongyle infection and gut microbiota: profiling of resistant and susceptible horses
498 over a grazing season. *Frontiers in Physiology* 9, 272.

499 Collobert-Laugier, C., Hoste, H., Sevin, C., Dorchies, P., 2002. Prevalence, abundance and site
500 distribution of equine small strongyles in Normandy, France. *Veterinary Parasitology* 110, 77-83.

501 Dallas, T.A., Laine, A.-L., Ovaskainen, O., 2019. Detecting parasite associations within multi-species
502 host and parasite communities. *Proceedings of the Royal Society B* 286, 20191109.

503 Debeffe, L., Mcloughlin, P.D., Medill, S.A., Stewart, K., Andres, D., Shury, T., Wagner, B., Jenkins, E.,
504 Gilleard, J.S., Poissant, J., 2016. Negative covariance between parasite load and body condition in a
505 population of feral horses. *Parasitology* 143, 983-997.

506 Gawor, J., 1995. The prevalence and abundance of internal parasites in working horses autopsied in
507 Poland. *Veterinary Parasitology* 58, 99-108.

508 Herd, R., 1990. The changing world of worms: The rise of the cyathostomes and the decline of
509 *Strongylus vulgaris*. *Compendium on Continuing Education for the Practicing Veterinarian* 12, 732-736.

510 Kamiya, T., O'Dwyer, K., Nakagawa, S., Poulin, R., 2014. What determines species richness of parasitic
511 organisms? A meta-analysis across animal, plant and fungal hosts. *Biological Reviews* 89, 123-134.

512 Kooyman, F., Van Doorn, D., Geurden, T., Mughini-Gras, L., Ploeger, H.W., Wagenaar, J., 2016.
513 Species composition of larvae cultured after anthelmintic treatment indicates reduced moxidectin
514 susceptibility of immature *Cylicocyclus* species in horses. *Veterinary parasitology* 227, 77-84.

515 Kornas, S., Basiaga, M., Kharchenko, V., 2011. Composition of the cyathostomin species in horses
516 with a special focus on *Cylicocyclus brevicapsulatus*. *Medycyna Weterynaryjna* 67.

- 517 Kornaś, S., Sallé, G., Skalska, M., David, I., Ricard, A., Cabaret, J., 2015. Estimation of genetic
518 parameters for resistance to gastro-intestinal nematodes in pure blood Arabian horses. *International*
519 *journal for parasitology* 45, 237-242.
- 520 Kuzmina, T., Kharchenko, V.A., Starovir, A., Dvojnjos, G., 2005. Analysis of the strongylid nematodes
521 (Nematoda: Strongylidae) community after deworming of brood horses in Ukraine. *Veterinary*
522 *Parasitology* 131, 283-290.
- 523 Kuzmina, T., Kornaś, S., Basiaga, M., Kharchenko, V., Vyniarska, A., 2011. Biodiversity of strongylids
524 (Nematoda: Strongylidae) communities in domestic horses from Poland and Ukraine. *Helminthologia*
525 48, 77-84.
- 526 Kuzmina, T., Lyons, E., Tolliver, S., Dzeverin, I., Kharchenko, V., 2012. Fecundity of various species of
527 strongylids (Nematoda: Strongylidae)—parasites of domestic horses. *Parasitology research* 111, 2265-
528 2271.
- 529 Kuzmina, T.A., Dzeverin, I., Kharchenko, V.A., 2016. Strongylids in domestic horses: Influence of horse
530 age, breed and deworming programs on the strongyle parasite community. *Veterinary parasitology*
531 227, 56-63.
- 532 Lello, J., Boag, B., Fenton, A., Stevenson, I.R., Hudson, P.J., 2004. Competition and mutualism among
533 the gut helminths of a mammalian host. *Nature* 428, 840-844.
- 534 Lichtenfels, J.R., Kharchenko, V.A., Dvojnjos, G.M., 2008. Illustrated identification keys to strongylid
535 parasites (Strongylidae: Nematoda) of horses, zebras and asses (Equidae). *Veterinary parasitology* 156,
536 4-161.
- 537 Lind, E.O., Eysker, M., Nilsson, O., Uggla, A., Höglund, J., 2003. Expulsion of small strongyle
538 nematodes (cyathostomin spp) following deworming of horses on a stud farm in Sweden. *Veterinary*
539 *parasitology* 115, 289-299.
- 540 MacArthur, R.H., Pianka, E.R., 1966. On optimal use of a patchy environment. *The American Naturalist*
541 100, 603-609.

- 542 Matthews, J.B., 2014. Anthelmintic resistance in equine nematodes. *International Journal for*
543 *Parasitology: Drugs and Drug Resistance* 4, 310-315.
- 544 Mfitilodze, M., Hutchinson, G., 1990. Prevalence and abundance of equine strongyles (Nematoda:
545 Strongyloidea) in tropical Australia. *The Journal of parasitology*, 487-494.
- 546 Nielsen, M.K., Kaplan, R.M., Thamsborg, S.M., Monrad, J., Olsen, S.N., 2007. Climatic influences on
547 development and survival of free-living stages of equine strongyles: implications for worm control
548 strategies and managing anthelmintic resistance. *The Veterinary Journal* 174, 23-32.
- 549 Ogbourne, C.P., 1976. The prevalence, relative abundance and site distribution of nematodes of the
550 subfamily Cyathostominae in horses killed in Britain. *Journal of Helminthology* 50, 203-214.
- 551 Ovaskainen, O., Tikhonov, G., Norberg, A., Guillaume Blanchet, F., Duan, L., Dunson, D., Roslin, T.,
552 Abrego, N., 2017. How to make more out of community data? A conceptual framework and its
553 implementation as models and software. *Ecology letters* 20, 561-576.
- 554 Peachey, L.E., Castro, C., Molena, R.A., Jenkins, T.P., Griffin, J.L., Cantacessi, C., 2019. Dysbiosis
555 associated with acute helminth infections in herbivorous youngstock—observations and implications.
556 *Scientific reports* 9, 1-16.
- 557 Poissant, J., Gavriiliuc, S., Bellaw, J., Redman, E.M., Avramenko, R.W., Robinson, D., Workentine, M.L.,
558 Shury, T.K., Jenkins, E.J., McLoughlin, P.D., 2020. A repeatable and quantitative DNA metabarcoding
559 assay to characterize mixed strongyle infections in horses. *International Journal for Parasitology*.
- 560 Poulin, R., Valtonen, E.T., 2002. The predictability of helminth community structure in space: a
561 comparison of fish populations from adjacent lakes. *International Journal for Parasitology* 32, 1235-
562 1243.
- 563 Ramiro, R.S., Pollitt, L.C., Mideo, N., Reece, S.E., 2016. Facilitation through altered resource availability
564 in a mixed-species rodent malaria infection. *Ecology letters* 19, 1041-1050.
- 565 Reinemeyer, C., Smith, S., Gabel, A., Herd, R., 1984. The prevalence and intensity of internal parasites
566 of horses in the USA. *Veterinary Parasitology* 15, 75-83.

- 567 Relf, V., Morgan, E., Hodgkinson, J., Matthews, J., 2013. Helminth egg excretion with regard to age,
568 gender and management practices on UK Thoroughbred studs. *Parasitology* 140, 641.
- 569 Sallé, G., Guillot, J., Tapprest, J., Foucher, N., Sevin, C., Laugier, C., 2020. Compilation of 29 years of
570 postmortem examinations identifies major shifts in equine parasite prevalence from 2000 onwards.
571 *International Journal for Parasitology*.
- 572 Sallé, G., Kornaś, S., Basiaga, M., 2018. Equine strongyle communities are constrained by horse sex
573 and species dispersal-fecundity trade-off. *Parasites & vectors* 11, 279.
- 574 Shmida, A., Wilson, M.V., 1985. Biological determinants of species diversity. *Journal of biogeography*,
575 1-20.
- 576 Stancampiano, L., Gras, L.M., Poglayen, G., 2010. Spatial niche competition among helminth parasites
577 in horse's large intestine. *Veterinary Parasitology* 170, 88-95.
- 578 Sweeny, A.R., Thomason, C.A., Carbajal, E.A., Hansen, C.B., Graham, A.L., Pedersen, A.B., 2020.
579 Experimental parasite community perturbation reveals associations between Sin Nombre virus and
580 gastrointestinal nematodes in a rodent reservoir host. *Biology Letters* 16, 20200604.
- 581 Taylor, M.A., Coop, R.L., Wall, R.L., 2012. *Veterinary parasitology*, Third edition
582 ed.
- 583 Tchuem Tchuente, L.A., Behnke, J., Gilbert, F., Southgate, V., Vercruyse, J., 2003. Polyparasitism
584 with *Schistosoma haematobium* and soil-transmitted helminth infections among school children in
585 Loum, Cameroon. *Tropical medicine & international health* 8, 975-986.
- 586 Telfer, S., Lambin, X., Birtles, R., Beldomenico, P., Burthe, S., Paterson, S., Begon, M., 2010. Species
587 interactions in a parasite community drive infection risk in a wildlife population. *Science* 330, 243-246.
- 588 Tikhonov, G., Abrego, N., Dunson, D., Ovaskainen, O., 2017. Using joint species distribution models
589 for evaluating how species-to-species associations depend on the environmental context. *Methods in*
590 *Ecology and Evolution* 8, 443-452.

591 Tikhonov, G., Opedal, Ø.H., Abrego, N., Lehtikoinen, A., de Jonge, M.M., Oksanen, J., Ovaskainen, O.,
592 2020. Joint species distribution modelling with the r-package Hmsc. *Methods in ecology and evolution*
593 11, 442-447.

594 van Doorn, D., Ploeger, H., Eysker, M., Geurden, T., Wagenaar, J., Kooyman, F., 2014. *Cylicocyclus*
595 species predominate during shortened egg reappearance period in horses after treatment with
596 ivermectin and moxidectin. *Veterinary parasitology* 206, 246-252.

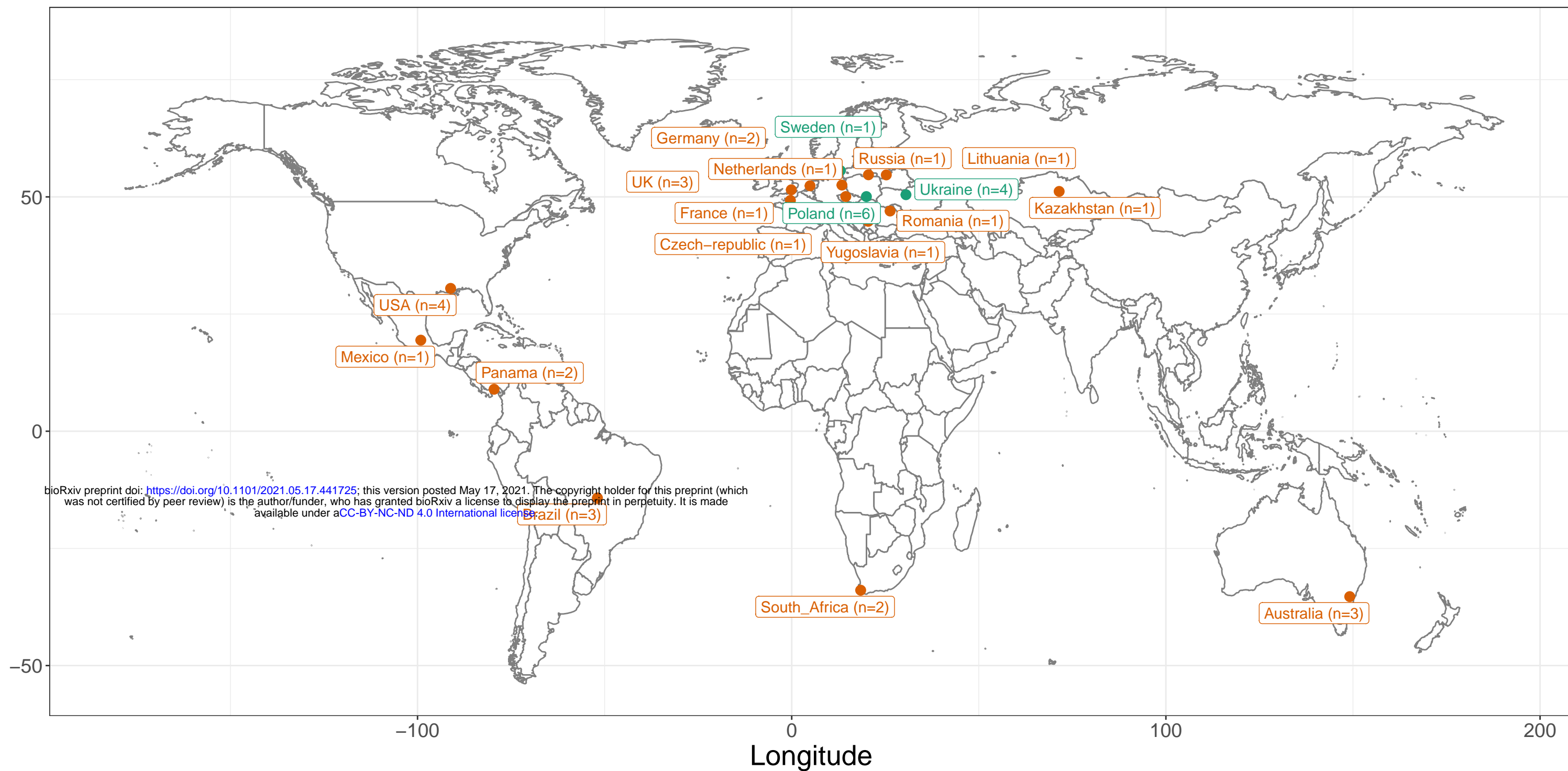
597 Vannette, R.L., Fukami, T., 2017. Dispersal enhances beta diversity in nectar microbes. *Ecology Letters*
598 20, 901-910.

599 Walshe, N., Mulcahy, G., Hodgkinson, J., Peachey, L., 2020. No Worm Is an Island; The Influence of
600 Commensal Gut Microbiota on Cyathostomin Infections. *Animals* 10, 2309.

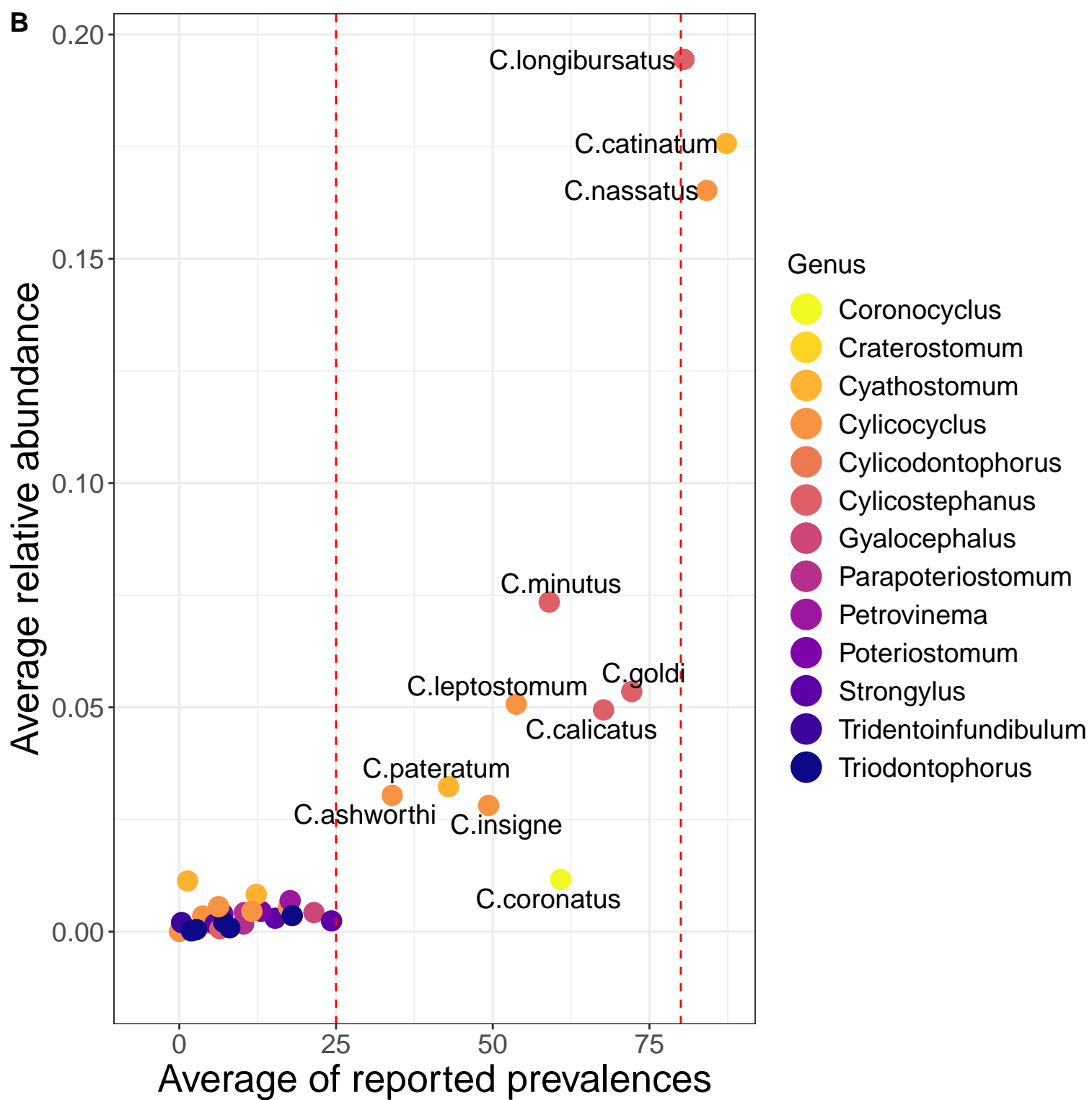
601 Wood, E., Matthews, J., Stephenson, S., Slote, M., Nussey, D., 2013. Variation in fecal egg counts in
602 horses managed for conservation purposes: individual egg shedding consistency, age effects and
603 seasonal variation. *Parasitology* 140, 115-128.

604

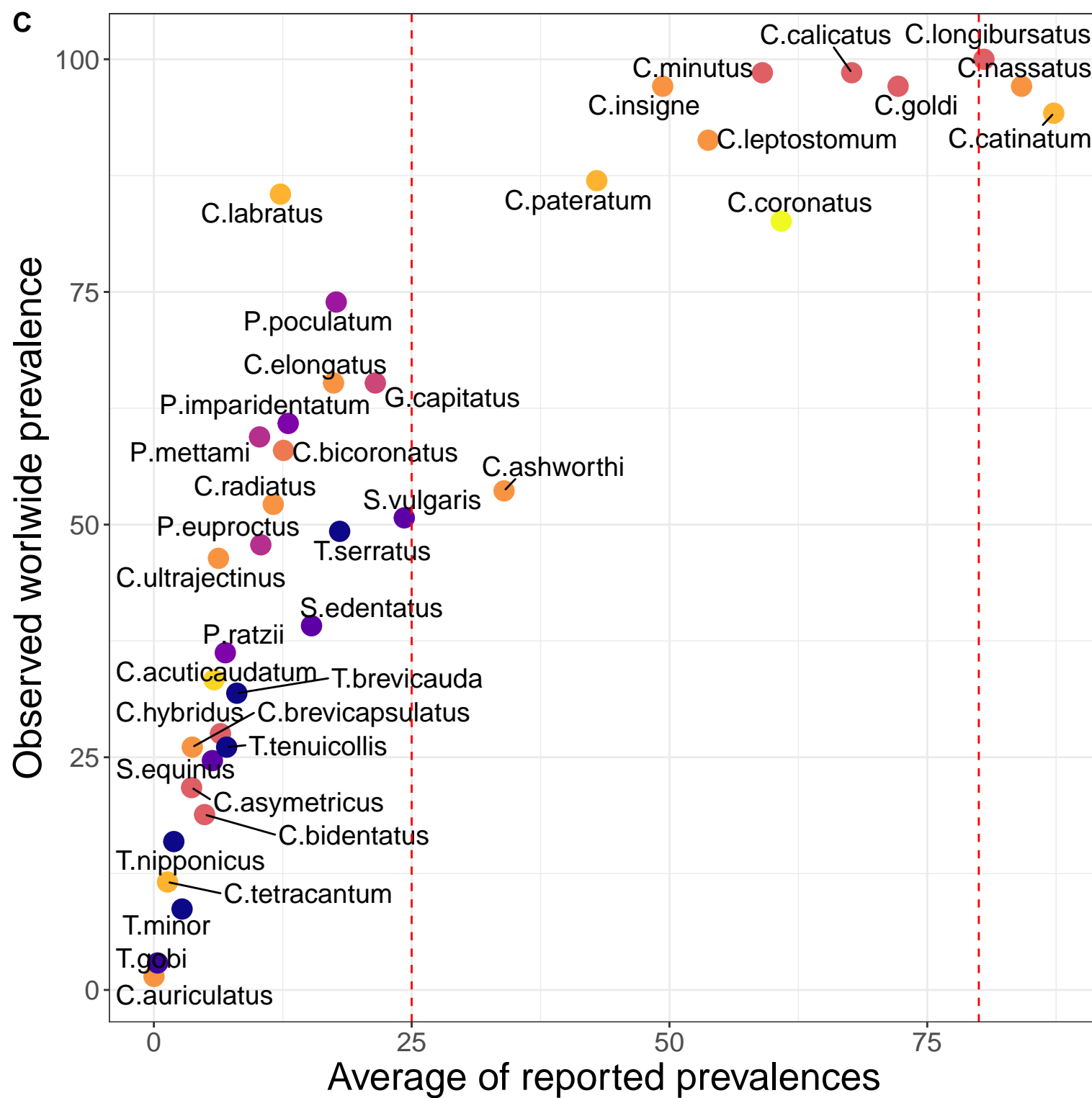
A

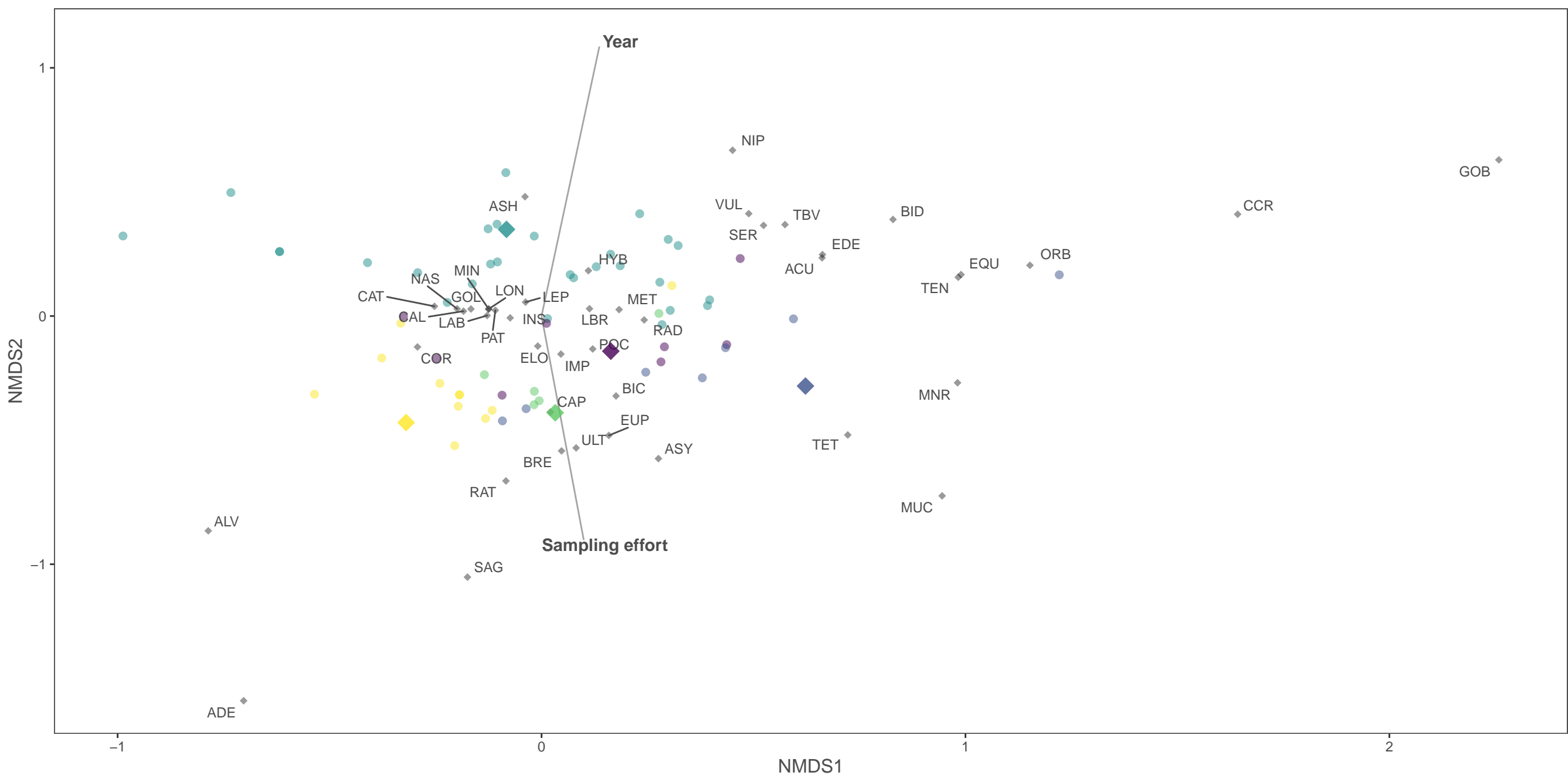


B



C





- America – Temperate – Necropsy
- America – Tropical – Necropsy
- Europe – Temperate – Necropsy
- Europe – Continental – Necropsy
- Europe – Continental – Deworming

