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RESEARCH PAPER

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Pinus pinea: a natural barrier for the insect vector of the pine wood nematode?

Inge van Halder^{1*} , Alberto Sacristan^{2,3} , Jorge Martín-García^{2,3} , Juan Alberto Pajares^{2,3^} and Hervé Jactel¹

Abstract

Key message: In mixed stands of *Pinus pinaster* and *Pinus pinea*, fewer insect vectors of the pinewood nematode (PWN) were captured than in pure *P. pinaster* stands. This finding has practical implications for PWN disease management, including the recommendation to improve the diversity of maritime pine plantations and to conserve stone pines in infected areas.

Context: The PWN is an invasive species in European pine forests, being vectored by the longhorn beetle *Monochamus galloprovincialis*. The presence of less preferred host trees may disrupt the insect vector dispersal and slow the spread of the disease.

Aims: The aim of the study was to compare the abundance of *M. galloprovincialis* in pure stands of *Pinus pinaster*, a preferred host tree, pure *P. pinea* stands, a less preferred host, and mixtures of these two species.

Methods: We selected 20 mature pine stands varying in % *P. pinaster* and % *P. pinea* in Spain. In each stand, we installed 3 pheromone traps to catch *M. galloprovincialis*. We related trap catches to stand and landscape composition.

Results: The level of capture of *M. galloprovincialis* was highest in pure *P. pinaster* stands and decreased with increasing proportion of *P. pinea*.

Conclusions: The presence of stone pine mixed with maritime pine significantly reduces the local abundance of the PWN insect vector. The most plausible mechanism is that *P. pinea* emits odors that have a repulsive effect on dispersing beetles.

Keywords: *Monochamus galloprovincialis*, *Bursaphelenchus xylophilus*, *Pinus pinaster*, Stone pine, Umbrella pine, Repellent, Mixed forests

1 Introduction

The pinewood nematode (PWN) *Bursaphelenchus xylophilus* (Steiner and Buhner) is one of the worst threats to coniferous forests in Europe. It is known to cause

massive tree mortality wherever it has been established as in Japan, Korea, China, and more recently in Portugal (Rodrigues et al. 2015). It has been introduced several times in Spain but has been successfully eradicated so far (<https://gd.eppo.int/reporting/article-6227>). Like most invasive pest species, it is very difficult to control. All stages of the invasion process must therefore be taken into account in a holistic manner to try to limit its impacts, from its transport and introduction in a new area, its establishment and proliferation, to its spread across the country (Chapple et al. 2012). In this context, the status of host or non-host plant for the invasive pest

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[^]Juan Alberto Pajares is deceased.

In memory of the late colleague Juan Pajares.

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is crucial because it is relevant to each of these steps. The list of host plants is used to identify the pathways of the transport of goods or commodities that support the stages of dissemination of the non-native species (Meurisse et al. 2019). Surveillance and detection are more effective when applied to habitats of the host plants. The most common method of eradication of non-native pests is host plant removal (Liebhold and Kean 2019). A promising method of non-native pest control relies on the mixtures of host and non-host plant species to enhance associational resistance effects (Guyot et al. 2015). Moreover, the presence of patches of non-host species can slow the spread of invasive pests (Rigot et al. 2014; Nunes et al. 2021).

In all countries where it is present or has been introduced, the pinewood nematode is carried and transmitted by longhorn beetles of the genus *Monochamus*. In Europe, it has so far only been detected in *Monochamus galloprovincialis* (Olivier) (Naves et al. 2015). The question of which host plant to monitor and possibly to remove in order to eradicate the pinewood nematode is therefore primarily concerned with the host plants of its insect vector. Common European pine species like maritime pine (*Pinus pinaster* Ait.), black pine (*Pinus nigra* Arnold), Scots pine (*Pinus sylvestris* L.), and Aleppo pine (*Pinus halepensis* Mill.) are well-established host trees of the pine sawyer beetle *M. galloprovincialis* (Appen-dix). However, the status of the stone (or umbrella) pine *Pinus pinea* (L.) as host of *M. galloprovincialis* is more doubtful (EFSA, 2012). In a laboratory choice test, adult *M. galloprovincialis* fed on fresh shoots of *P. pinea*, but it was the least preferred European pine species (Naves et al. 2006). In the same experiment, oviposition tests in the lab showed that very few eggs were laid on *P. pinea* bolts, and no offspring emerged from the bolts (Naves et al. 2006). In another laboratory experiment, Sanchez-Husillos et al. (2013) observed regular feeding on fresh *P. pinea* twigs (although always less than on twigs of other main European pine species). They also obtained some emergences of young adults from *P. pinea* logs, but less than 15% of egg laying resulted in the emergence of offspring on *P. pinea*. In a non-choice test in the laboratory, *M. galloprovincialis* was also found to be able to feed on 2-year-old potted pine trees with no significant difference between *P. pinea* and *P. pinaster* (Gonçalves et al. 2020). Yet, although Portugal has large areas of stone pine forests (www.euforgen.org) and the nematode has been present there for more than 20 years (Mota et al. 1999), no pinewood nematode dieback and no major attack of *M. galloprovincialis* have ever been reported on *P. pinea* in this country. Discrepancies between the results of laboratory experiments and *in natura* observations may occur and be due to differences in pine sawyer behavior

in the lab and the field or the sensitivity/attractiveness of mature living trees compared to laboratory plants. Surprisingly, no field experiments have been conducted to date to address these questions.

To fill this knowledge gap, we developed a field trial in the Valladolid area (Castilla y León, Spain) where *P. pinea* and *P. pinaster* co-occur in large mature forests. In particular, we were interested in testing the hypothesis that the presence of *P. pinea* in mixed *P. pinaster* stands would reduce trap catch levels compared to catches in nearby pure *P. pinaster* forests, and that this reduction was proportional to the percentage of *P. pinea* in the mixture.

2 Material and methods

2.1 Study area and stand selection

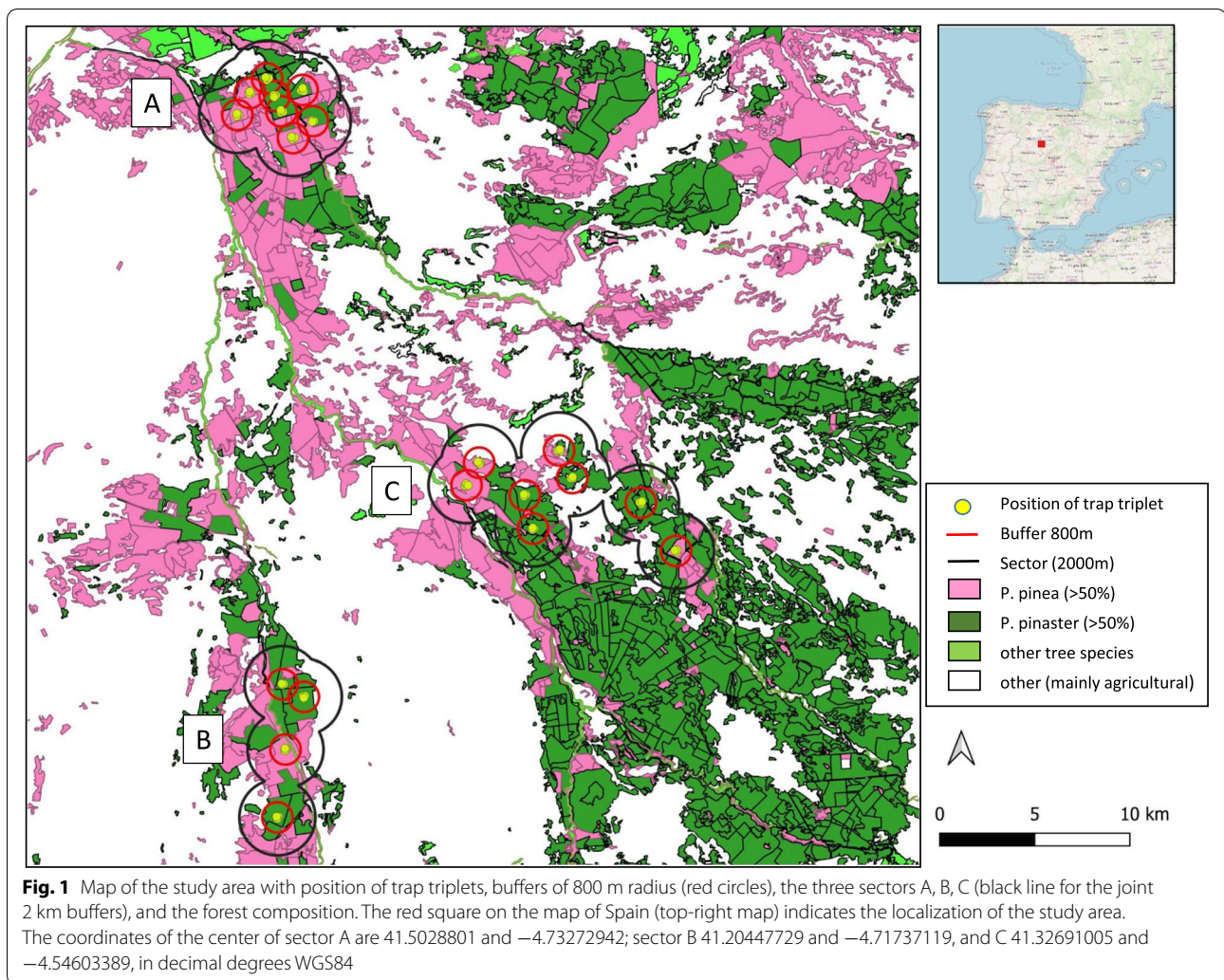
The study was set up in a forest region south of Valladolid, in the center of Spain. The forests in this region are dominated by pine forests composed of pure *P. pinea* stands, pure *P. pinaster* stands, and mixtures of these two species. The management of these stands is focused on production of edible pine seeds and resin in *P. pinea* or *P. pinaster* stands, respectively. The region has a continental climate and sandy soils. The PWN has never been detected in this region.

Within this region, we selected three sectors (see Fig. 1). Within each sector, we selected four or eight pine stands varying in their relative proportion of *P. pinaster* and *P. pinea*, from 0 to 100 % (i.e., pure *P. pinaster* stands, pure *P. pinea* stands, and mixtures of both in various proportions). In total, 20 stands were selected. *P. pinaster* stands have in general a higher tree density than *P. pinea* stands, but since the latter tree species has a denser and bigger canopy, we considered that light conditions in both stand types were comparable. Stand management includes removal of dead trees within 1 or 2 years maximum. The amount of dead trees was not measured but was in any case very low and similar between stand types. There was no resin collection in the selected *P. pinaster* stands. The age of the different stands was similar, ranging from 60 to 80 years.

2.2 Insect sampling

In each stand, three pheromone traps (Cross Vane® type; Alvarez et al. 2015a) were positioned with a mean distance of 80–100 m between traps, which is comparable to the attraction range of these traps (90–125 m, Jactel et al. 2019). Only four traps, in four different stands, were located at 100 m from the forest edge; all other traps were at more than 200 m from the edge. Traps were attached 2 m high on metal poles, at about 10 m from the nearest tree.

The 60 traps were installed on 22–24 May 2019 and equipped with attractive Galloprotect 2D lures for *M.*



galloprovincialis (Alvarez et al. 2016). These lures contain two bark beetle kairomones (ipsenol and 2-methyl-3-buten-2-ol) and the pheromone monochamol (2-undecyloxy-1-ethanol). The traps were emptied every 2 to 3 weeks until 25 October 2019, and the dispensers were changed twice (26 July and 11 September). The trap collection cup contained an insecticide to kill the beetles, but no liquid. For the analyses we used, the number of *M. galloprovincialis* beetles per sex and per trap cumulated over the whole season. Trap catches can be considered a proxy for the abundance of the insect vector during the flight season.

2.3 Tree composition at different scales around the traps

We characterized the pine tree composition at different scales around the traps using a GIS (QGIS 3.16.7, QGIS Development Team). We created buffers of 50 m to 800 m around traps, to analyze at which spatial scale

a possible effect of tree composition on beetle catches might play a role.

We created around each trap a 50 m radius circular buffer to estimate the tree species composition in the direct surroundings of the traps. We counted within this buffer the number of *P. pinaster* trees and *P. pinea* trees using aerial photographs from 2018. The two pine species clearly differ in their crown shape and can therefore be recognized on aerial photos, which was confirmed by several field visits. For each 50 m buffer, we calculated then the % *P. pinea* among the total number of trees (the sum of *P. pinea* and *P. pinaster* equaled 100%, no other tree species being present in the sampled plots). We also counted the number of trees of the two pine species in a larger buffer of 200 m radius, created around the center of the three traps per stand. To characterize the tree composition at different landscape scales, we created buffers of 400 m, 600 m, and 800 m radii around the center of each

trap triplet. We used land cover maps of the third Spanish forest inventory, which provided for each forest stand the percentage of the three most abundant tree species. This inventory was carried out between 1997 and 2007, so we verified on aerial photos of 2018 the composition of the stands. We calculated the % *P. pinea* and % *P. pinaster* in each buffer. For that, we estimated the surface covered by a given pine species by multiplying, for each forest stand intercepted by the buffer, the % of the pine species in that stand by the surface of that stand included in the buffer. Then, we summed up all the surfaces covered by each pine species and divided it by the buffer area to obtain the % cover of each pine species. For each buffer size separately, the % *P. pinea* and % *P. pinaster* were negatively correlated. The % *P. pinea* in the 50 m buffer was positively correlated with the % *P. pinea* in the 200, 400, 600, and 800 m buffer (see Table 2 and 3 in Appendix). The composition of the landscape in the buffers was largely dominated by pine stands (*P. pinaster* or *P. pinea*), which represented 100%, 97%, 94%, and 91% of the buffer area for the 200 m, 400 m, 600 m, and 800 m buffers, respectively.

Finally, we created around each trap a 2 km radius buffer and joined these buffers per sector, creating one map per sector (Fig. 1). We calculated the % *P. pinea* and % *P. pinaster* for each of the three sectors using the same method as for the landscape buffers. The % *P. pinea* and % *P. pinaster* varied between the three sectors. The % *P. pinea* was 52.4%, 36.3%, and 24.1% for sectors A, B, and C respectively, and the % *P. pinaster* was 26.2%, 31.5%, and 35.6% for sectors A, B, and C, respectively. This resulted in a *pinea/pinaster* ratio of 2.0, 1.1, and 0.7, respectively. The remaining area of each sector was occupied by other land uses, mostly agricultural land (Fig. 1).

2.4 Analyses

We started the analyses at the 50-m scale and analyzed the effect of the following variables on the number of *M. galloprovincialis* beetles caught per trap: *sex* (as factor: male, female), % *P. pinea*, and the interaction *sex**% *P. pinea*. To select the most appropriate model, we first compared three different types of models based on count data with these explanatory variables: a generalized linear mixed model (glmm) with a Poisson error structure and two glmm models with a negative binomial error structure (i.e., linear or quadratic parameterization (Hardin & Hilbe, 2007)). Comparison between models was based on the value of the Akaike's information criterion corrected for small sample sizes (AICc). Next, we compared the best of these three models based on count data with a linear mixed model (Gaussian error structure) on log-transformed count data of beetles (natural logarithm plus one). Since AIC values cannot be compared between models with transformed and non-transformed response data, we visually checked

various assumptions (normality of residuals, normality of random effects, linear relationship, homogeneity of variance, multicollinearity) between these two models, using the R package "performance".

We used mixed models to take into account the experimental set up, where we had count data of males and females for the same trap, where three traps were positioned in the same stand, and where stands were in three different sectors. We therefore added a random effect of trap within stand within sector to all models. The model structure thus takes into account the overlapping buffers for the three traps per stand.

Among the three models based on count data, the two models with the negative binomial error distribution had the lowest AICc values (AICc 776.4 and 778.3) compared to the Poisson model (AICc 782.5). Comparison of the residuals of the glmm with the negative binomial error distribution and the lmm model on log-transformed data showed that the latter had a better distribution of residuals, and this Gaussian model was therefore used in the analyses.

Next, we used for the selected model type a model simplification procedure removing nonsignificant variables while applying marginality principle, where the principal effects were not removed if involved in a significant interaction. We visually checked model residuals of the selected model. For the best model, R^2 values were calculated to estimate the variance explained by fixed effects (marginal R^2 , R^2_m) and by fixed plus random effects (conditional R^2 , R^2_c) (Nakagawa and Schielzeth, 2013). We ran these models for different buffer sizes (up to 800 m) to analyze if the effect observed differed with buffer size.

Since we had very different beetle catches between sectors, we ran the same model as selected above but including sector and the interaction term between sector and % *P. pinea* in the fixed part of the model. This allowed us to analyze if the same effect of % *P. pinea* on beetle captures was observed in each sector (i.e., no significant interaction between sector and % *P. pinea*).

All analyses were carried out in R 4.1.2 (R Core Team, 2021). The following functions and libraries were used: package "glmmTMB" for mixed models, package "performance," and "see" for model residuals, function `r.squaredGLMM` from package "MuMin" for calculation of R^2 and ANOVA from package "car" or significance of selected variables. The dataset used for the analyses is available in a repository (Van Halder et al. 2022).

3 Results

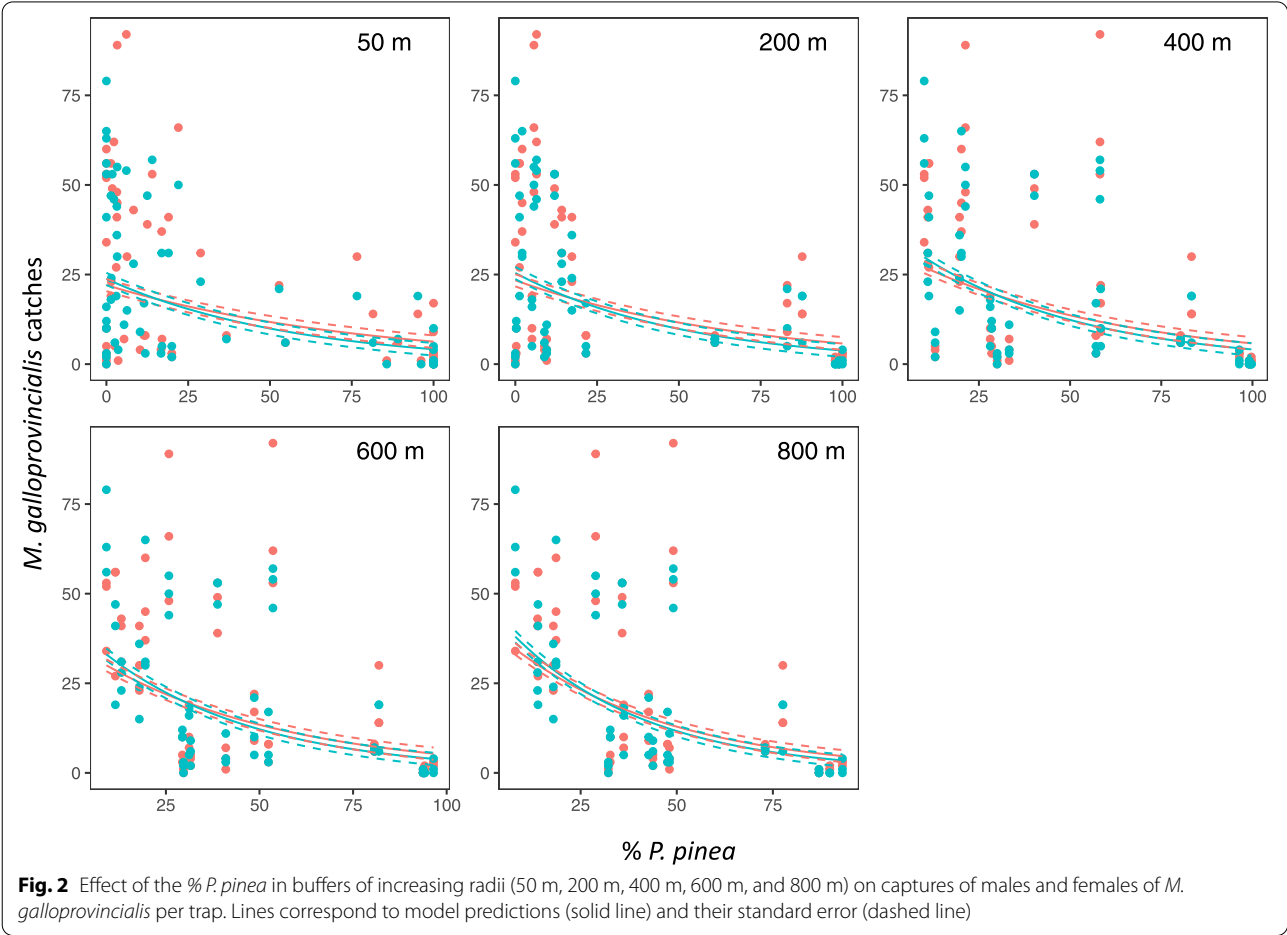
We captured a total of 2707 *M. galloprovincialis* beetles in the 60 traps, 1289 males and 1418 females. The number of beetles varied between 0 and 146 individuals per trap, with a mean of 45.1 individuals.

For the Gaussian model with the % *P. pinea* measured in a buffer of 50 m, we retained all variables, i.e., % *P. pinea*, *sex*, and their interaction, since the interaction was significant. The marginal and conditional R^2 of this model were respectively 0.221 and 0.931. There was a significant negative effect of % *P. pinea* at 50 m on the number of

M. galloprovincialis trapped (Table 1), with very low captures in pure stands of *P. pinea* (Fig. 2). The variable *sex* was included in the model but was not significant. The significant interaction between % *P. pinea* and *sex* indicated a more important negative effect of % *P. pinea* for males than for females (Fig. 2).

Table 1 Results of the linear mixed model for each buffer size around the traps. Explanatory variables were % *P. pinea*, *sex*, and their interaction. Significance of each explanatory variable is given (*P*-value), and for the % *P. pinea*, the estimate is also provided. R^2m corresponds to the variance explained by the fixed effects, R^2c to the variance explained by fixed and random effects (trap within stand within sector). The AICc value of each model is also provided

Buffer size (m)	% <i>P. pinea</i>		Sex	Sex*% <i>P. pinea</i>	R^2m	R^2c	AICc
	Estimate	<i>p</i> -value	<i>p</i> -value	<i>p</i> -value			
50	−0.0126	5.35e-07***	0.2186	0.0017**	0.221	0.931	190.4
200	−0.0140	3.53e-06***	0.2184	0.0017**	0.248	0.934	195.1
400	−0.0172	4.01e-05***	0.2321	0.0152*	0.238	0.927	200.7
600	−0.0197	2.20e-05***	0.2365	0.0310*	0.264	0.923	201.1
800	−0.0234	8.34e-06***	0.2404	0.0577	0.308	0.919	200.9



For the different buffer sizes up to 600 m, the results were comparable to the model with the buffer of 50 m. Only at 800 m the interaction between *sex* and % *P. pinea* was no longer significant. The R^2 values increased slightly with buffer size (Table 1).

In the model including sector and the interaction sector*% *P. pinea*, variables retained in the selected model were % *P. pinea*, *sex*, sector, and the interaction % *P. pinea***sex*. The marginal and conditional R^2 of this model were respectively 0.724 and 0.934. There was a significant effect of sector (χ^2 44.9, $P < 0.0001$) with fewer catches in sector A (with the greatest *pinea/pinaster* ratio) than in sectors B or C (no significant difference between them, Figure 3 in Appendix), but no significant interaction between sector and % *P. pinea*. For this model, variable selection was comparable for the different buffer sizes (see Table 4 in Appendix).

4 Discussion

Our field experiment demonstrated a clear effect of native pine species composition on the trap captures of *M. galloprovincialis*, the insect vector of the pine wood nematode in southern Europe. The level of *M. galloprovincialis* captures in pheromone traps was high in pure mature *P. pinaster* stands, very low in pure mature *P. pinea* stands, and decreased in mixed forests of the two pine species with increasing proportion of *P. pinea*.

Two hypotheses, not mutually exclusive, can be proposed to explain the pattern of different capture levels of *M. galloprovincialis* as a function of the relative amount of the two pine species in the forests. They relate to the use of pine species as breeding substrate and/or emitter of chemical cues for the selection of suitable habitat.

First, one can assume that *P. pinea* emit odors that have a repellent effect on flying adults of *M. galloprovincialis*, reducing their probability of being attracted and caught by pheromone traps in *P. pinea* stands. Numerous studies have shown that herbivorous insects use the odors of their host plants as cues and more specifically that conifer beetles are strongly attracted by volatile terpenic compounds (Raffka 2014; Seybold et al. 2006; Tunset et al. 1993). For example, it has been shown that *M. galloprovincialis* is attracted by alpha pinene (Pajares et al. 2004; Ibeas et al. 2007), a monoterpenic compound. This compound is even used in attractive lures in combination with the pheromone (Alvarez et al. 2016). Conversely, non-host volatiles (NHVs) are used by insect herbivores as olfactory signals indicating the unsuitability of non-host plants or habitats (Zhang and Schlyter 2004; Campbell and Borden 2009). The main difference in the terpene profile of *P. pinea* compared to other European native pines is the high proportion of limonene (and the consequently low proportion of alpha and beta pinene)

(Roussis et al. 1995; Rodrigues et al. 2017; Gaspar et al. 2020). Electrophysiological tests showed that limonene is well detected by the olfactory sensillae of the antennae of *M. galloprovincialis* (Alvarez et al. 2015b, 2016). Limonene is known as a common repellent or feeding deterrent for many conifer insects (Nordlander 1990; Ibrahim et al. 2001; Romón et al. 2017). In a laboratory bioassay, Sanchez-Husillos et al. (2013) showed that limonene applied to cut twigs of *P. pinaster* resulted in lower feeding activity by *M. galloprovincialis* than on control shoots. Limonene is also reported as a feeding deterrent for *Monochamus alternatus*, the insect vector of PWN in Asia (Fan and Sun 2006). In two field trials testing different combinations of pheromones and kairomones, the addition of limonene reduced the capture of *M. galloprovincialis* in pheromone traps by 25% and 46%, respectively (Alvarez et al. 2016). It is therefore likely that *P. pinea* trees do emit a high concentration of limonene into the atmosphere, and that this volatile monoterpene acts as a repellent for the pine sawyer beetle and/or as an odor masking the attractive alpha and beta pinene emitted by *P. pinaster*. Likewise, the higher catches in the pure *P. pinaster* stands may be linked with the attractive odors emitted by this tree species.

The second explanatory hypothesis for *M. galloprovincialis* capture patterns is that they reflect the amount of breeding resource, i.e., the quantity of *P. pinaster*, which is negatively correlated with the quantity of *P. pinea*. Several laboratory studies have shown that although *M. galloprovincialis* can feed on fresh shoots of *P. pinea*, the number of oviposition events on *P. pinea* trunks is lower than on *P. pinaster*, and the number of emerging offspring greatly reduced (Sanchez-Husillos et al. 2013) or zero (Naves et al. 2006). In our study, the average level of capture of *M. galloprovincialis* should then be proportional to the amount of suitable breeding substrate, i.e., weakened *P. pinaster* trees, assumed to be correlated with the area of *P. pinaster* stands in the landscape. This is consistent with the observations in the different buffers and with the much lower captures in sector A (9.8 individuals/trap) where the proportion of *P. pinaster* was lowest. However, the average number of catches per trap was higher in sector B (90.4) than in sector C (57.7), while they have a similar cover of *P. pinaster* (31.5% and 35.6%). Moreover, *M. galloprovincialis* has great flight capabilities, being able to travel several kilometers (David et al. 2014; Robinet et al. 2019). It is therefore virtually capable of covering the entire sector studied. It is also known that it does not respond to the attraction of the pheromone until it is sexually mature, which takes more than 2 weeks (Etxebeste et al. 2016; Robinet et al. 2019). If we suppose that *M. galloprovincialis* disperses from its emerging site and *P. pinea* and *P. pinaster* do not exert

a repellent or attractive effect, then the beetles should have been found at a similar level of abundance in both *P. pinea* and *P. pinaster* stands of the same sector. In other words, if the main reason for the lower number of *M. galloprovincialis* captures in *P. pinea* stands was the lower amount of breeding substrate, this would imply that emerging insects remain for more than 15 days in the stand where they were produced. This seems less likely than the repellent effect of *P. pinea* (mediated by the release of limonene), but further experimentation will be necessary to decide between the relative contribution of the proposed hypotheses.

Interestingly, several recent studies have suggested that *P. pinea* is also more resistant or tolerant to the pine wood nematode *B. xylophilus* than other European pines like maritime or Scots pines (Nunes da Silva et al. 2015; Rodrigues et al. 2017; Pimentel et al. 2020). Recent laboratory studies have even shown that PWN-infected *P. pinea* saplings were able to reduce nematode infection to near zero (Estorninho et al. 2022). Additionally, Gaspar et al. (2020) found that the least infested *P. pinaster* trees had the highest concentration of limonene, while Liu et al. (2017) observed a higher limonene-synthase expression in resistant than in susceptible Masson pines (*Pinus massoniana* Lamb.).

We captured more females than males in our pheromone traps, which is consistent with the observation of the same biased sex ratio in traps baited with the same type of lure (Alvarez et al. 2016). This difference is likely due to the additional attraction of females by the aggregation pheromone emitted by *M. galloprovincialis* males. However, we lack information on the effect of limonene on the respective behavior of males and females. Thus, it is difficult to explain the significant interaction between the sex of trapped insects and the rate of *P. pinea* in pine plantations.

The two possible explanations we put forward to explain the pattern of lower *M. galloprovincialis* numbers in maritime pine and stone pine mixtures fit well with the theory of associational resistance (Jactel et al. 2021), which predicts lower abundance and damage of insect herbivores in mixed forests than in pure forests. According to the underlying host concentration hypothesis (Hambäck & Englund, 2005), specialized herbivorous insects, such as the pine sawyer beetle, would be more likely to immigrate, less likely to emigrate, and therefore spend more time feeding and breeding in habitat patches with a higher concentration of host resources, such as pure maritime pine stands. A second hypothesis, that of “host apparency,” states that host plants surrounded by heterospecific neighbors would be less apparent, i.e., more difficult to spot, locate, and colonize due to a disruption of

visual (plants less tall than their non-host neighbors; Castagneyrol et al. 2013) or olfactory (neighbors emitting non-host volatiles; Jactel et al. 2011) cues used by insect herbivores to find a favorable host.

Our findings have two practical implications. First, the EU Commission implementing decision of 26 September 2012 on “emergency measures to prevent the spread within the Union of *Bursaphelenchus xylophilus*” requires precautionary felling of all susceptible plants in the infested zone and in a zone with a radius of 500 m around the plants infested with PWN, with all *Pinus* species considered “susceptible plants.” Here, we argue that it would be unnecessary to cut and remove *P. pinea* trees in a PWN-infected landscape because field observations in Portugal indicate that these trees do not develop pine wilt disease and seem not to be used for breeding by the insect vector while producing valuable ecosystem services. Secondly, as a preventive measure, it would be interesting to maintain or foster *P. pinea* regeneration or plantations to slow the spread of the PWN by repelling the insect vectors. *P. pinaster* stands could also be better protected by mixing them with *P. pinea* or by planting hedgerows of *P. pinea* around them (Dulaurent et al. 2012).

5 Conclusions

The number of *M. galloprovincialis* captures in pheromone traps was highest in *P. pinaster* stands and decreased with increasing proportion of *P. pinea* in mixed stands. This pattern can be due to a lower breeding resource, i.e., *P. pinaster* trees, by a repellent effect of *P. pinea* odors on the dispersal behavior of *M. galloprovincialis*, or by a combination of both mechanisms. Based on these results, we argue to conserve stone pines in PWN-infested areas, even in case of an eradication program, and to promote this tree species in forest landscapes within risk areas to slow the spread of the disease.

Appendix

List of publications supporting the status of host plant for *Monochamus galloprovincialis* *Pinus nigra*

Akbulut, S. (2009). Comparison of the reproductive potential of *Monochamus galloprovincialis* on two pine species under laboratory conditions. *Phytoparasitica*, 37(2), 125–135.

Akbulut, S., Keten, A., & Stamps, W. T. (2008). Population dynamics of *Monochamus galloprovincialis* Olivier (Coleoptera: Cerambycidae) in two pine species under laboratory conditions. *Journal of pest science*, 81(2), 115–121.

Akbulut, S., Keten, A., Baysal, I., & Yüksel, B. (2007). The effect of log seasonality on the reproductive potential

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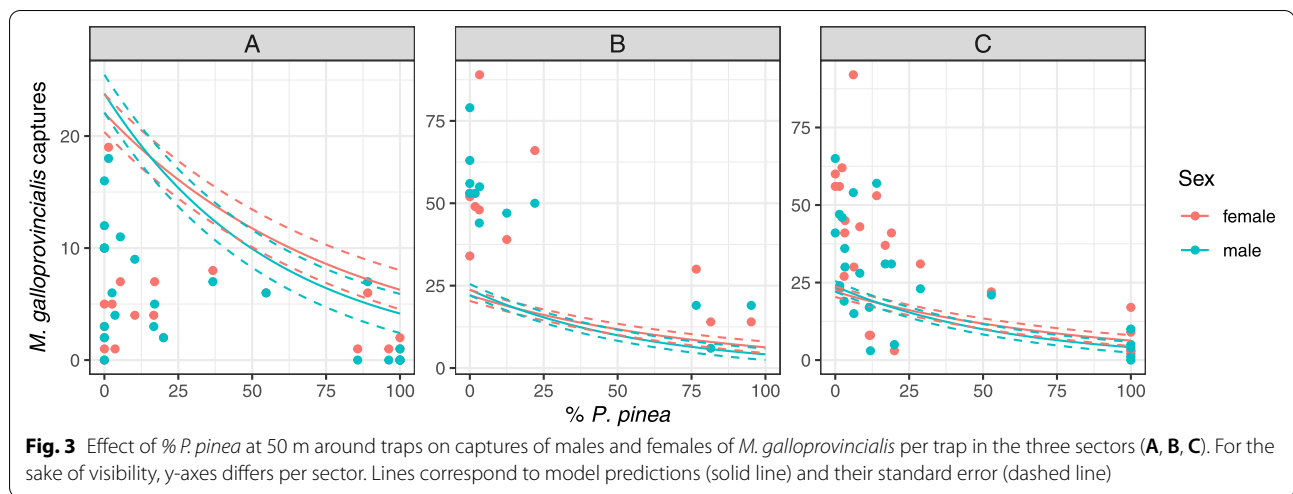


Table 2 Pearson's correlation coefficients and the associated *P*-value of the *t*-test for the relationship between the % *P. pinea* and the % *P. pinaster* at different buffer sizes around the traps

Variables	Pearson's <i>r</i>	<i>p</i> -value
% <i>P. pinea</i> — % <i>P. pinaster</i> , buffer 50 m	−1.000	2.20E-16
% <i>P. pinea</i> — % <i>P. pinaster</i> , buffer 200 m	−1.000	2.20E-16
% <i>P. pinea</i> — % <i>P. pinaster</i> , buffer 400 m	−0.989	2.20E-16
% <i>P. pinea</i> — % <i>P. pinaster</i> , buffer 600 m	−0.948	2.20E-16
% <i>P. pinea</i> — % <i>P. pinaster</i> , buffer 800 m	−0.897	2.20E-16

Table 3 Pearson's correlation coefficients and the associated *P*-value of the *t*-test for the relationship between the % *P. pinea* in the 50 m buffer and the % *P. pinea* at larger buffer scales

Variables	Pearson's <i>r</i>	<i>p</i> -value
% <i>P. pinea</i> 50 m — % <i>P. pinea</i> 200 m	0.969	2.20E-16
% <i>P. pinea</i> 50 m — % <i>P. pinea</i> 400 m	0.863	2.20E-16
% <i>P. pinea</i> 50 m — % <i>P. pinea</i> 600 m	0.854	2.20E-16
% <i>P. pinea</i> 50 m — % <i>P. pinea</i> 800 m	0.828	3.21E-16

Table 4 Results of the best linear mixed model for each buffer size around the traps (with sector in the fixed part of the model). Significance of each explanatory variable is given (*P*-value), and for the % *P. pinea*, the estimate is also provided. R^2_m corresponds to the variance explained by the fixed effects and R^2_c to the variance explained by fixed and random effects (trap within stand)

Buffer <i>P. pinea</i> (m)	% <i>P. pinea</i>		Sector <i>p</i> -value	Sex <i>p</i> -value	Sex*% <i>P. pinea</i> <i>p</i> -value	R^2_m	R^2_c
	Estimate	<i>p</i> -value					
50	−0.0126	1.220e-07***	1.786e-10***	0.219	0.001695**	0.724	0.934
200	−0.0139	6.122e-07***	2.827e-11***	0.218	0.001661**	0.736	0.936
400	−0.0167	1.274e-05***	1.149e-08***	0.232	0.01515*	0.707	0.932
600	−0.0190	7.353e-06***	8.564e-08***	0.236	0.03095*	0.712	0.931
800	−0.0224	2.911e-06***	9.006e-07***	0.240	0.0577	0.720	0.930

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Code availability

The R-code used for the analysis is available from the corresponding author on reasonable request.

Authors' contributions

HJ, JAP, and AS designed the study. AS did the site selection and fieldwork. IVH and HJ did the analyses and the first draft. HJ, IVH, AS, and JMG reviewed and edited the text. The authors read and approved the final manuscript.

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Availability of data and materials

The dataset used for the current study is available at <https://doi.org/10.15454/JXFGPI> (van Halder et al. 2022).

Declarations

Ethics approval and consent to participate

The trapped *Monochamus galloprovincialis* beetles are not protected at a national or international level.

Consent for publication

The authors read and approved the final manuscript.

Competing interests

The authors declare that they have no competing interests.

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