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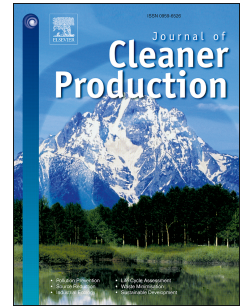
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CRedit author statement

Lennart Kokemohr: Writing - Original Draft; Conceptualization; Methodology; Software; Investigation; Data Curation; Visualization; **Neus Escobar:** Writing - Review & Editing; Methodological advice – Life Cycle Assessment; **Alexandre Mertens:** Writing - Review & Editing, Data and development of Belgium beef system; **Claire Mosnier:** Writing - Review & Editing, Data and development of French Beef system; **Giacomo Pirlo:** Writing - Review & Editing, Data and development of Italian Beef system; **Patrick Veysset:** Writing - Review & Editing, Data and development of French Beef system; **Till Kuhn:** Writing - Review & Editing; Methodological advice; Conceptualization; Supervision

Life Cycle Sustainability Assessment of European beef production systems based on a farm-level optimization model

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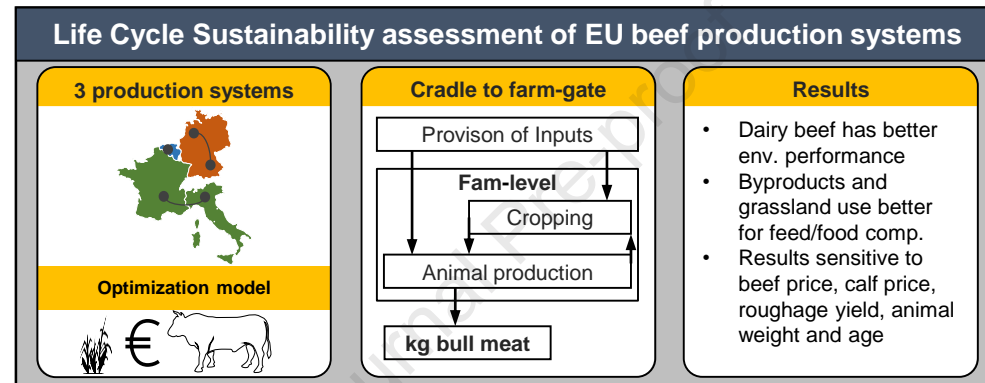
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1 **Life Cycle Sustainability Assessment of European beef production systems** 2 **based on a farm-level optimization model**

3 **Abstract**

4 The European Union (EU) is among the largest beef producers in the world. Besides the economic
5 turnover, beef production causes adverse environmental impacts such as climate change. The sector
6 is known for high heterogeneity in production systems, partly explained by different natural and
7 economic conditions. This study assesses the environmental, social, and economic performances
8 of three typical beef production systems in the EU at the farm level. The farm optimization model
9 FarmDyn is used in this study to carry out a Life Cycle Sustainability Assessment (LCSA) from
10 cradle to farm gate; combined with a sensitivity analysis on prices, yields and animal traits. The
11 assessed systems are a Belgian suckler cow farm that fattens its own offspring (BE); a system
12 where calves raised in a French suckler cow farm are fattened on a farm in Italy (FR-IT); and a
13 system where dairy bred calves from one farm are fattened on another farm, both located in
14 Germany (GE-GE). The functional unit is 1 kg of carcass weight from young bulls. In addition to
15 several environmental impact categories, the gross margin is estimated as an economic indicator.
16 The social performance is measured with on-farm workload differentiated by tasks, and human
17 calorie and protein conversion used for production. GE-GE performs better than the other systems
18 in the environmental indicators because emissions are partially allocated towards dairy production.
19 FR-IT shows the highest gross margin due to a higher beef price. BE and FR-IT use less human-
20 consumable feed, as both systems employ grasslands and by-products for animal feeding. The
21 sensitivity analysis identifies the price of beef and calves, the yield of roughage crops, and the
22 weight and age of animals as major factors influencing the results. FarmDyn proves useful to
23 perform LCSA of beef production on a farm-level as it integrates environmental, economic, and
24 social indicators in a consistent framework; while considering price effects and farmers' behaviour
25 in the context of farm heterogeneity and variability in management practices. Results thus provide
26 valuable information to inform not only farmers' decision but the debate of sustainable beef
27 production in the EU.

28 **Keywords:** *farm model; life cycle assessment; livestock; optimization model; sensitivity analysis;*
29 *sustainability*

30 **1. Introduction**

31 Livestock production causes 13% of the global greenhouse gas (GHG) emissions (Herrero et al.
32 2016), around 33% of nitrogen (N) pollution (Uwizeye et al. 2020) and uses more than 40% of
33 global arable land for feed production (Mottet et al. 2017). Concerns arise on the over-consumption
34 of meat as food, given the low calorie-conversion efficiency of livestock (Wilson et al. 2019).
35 According to Cassidy et al. (2013), an additional four billion people could be fed if all arable land
36 were used to directly grow food instead of fodder or biofuels. However, livestock production
37 contributes to the fight against hunger through the conversion of non-edible feedstuff into food for
38 human consumption (Smith et al. 2013). Furthermore, the livestock sector contributes to the
39 economy with a global production value of 1.2 trillion US\$ in 2018 (FAO 2020). Despite the
40 disadvantages of livestock production, the global consumption of livestock products has been rising
41 (FAO 2020) and plays a crucial role in reaching the United Nations' Sustainable Development
42 Goals (Mehrabi et al. 2020).

43 A large share of the global livestock production is concentrated in the European Union (EU), e.g.,
44 20% in 2018 (FAO 2020). In 2017, the EU-28 agricultural sector generated 10% of the region's
45 total GHG emissions with a production value of 170 billion €, with around 4 million people
46 employed in livestock farms (Peyraud and MacLeod 2020). Within the EU, cattle constitute the
47 largest share of the livestock population at around 50% of the total livestock units, with France,
48 Germany and Italy having the biggest herds (Cook 2020). Beef slaughtered in EU slaughterhouses
49 amounts up to 6.8 million tonnes carcass weight while the largest share is estimated for bulls¹ (34%),
50 followed by cows (30%) and heifers (16%) (EUROSTAT 2021). Bull meat production systems in
51 the EU are characterized by a high degree of heterogeneity. Systems differ by origin and breed of
52 the animals, age and weight at slaughtering as well as the kind and origin of feed used (Hocquette
53 et al. 2018). The highest stocking density of fattening farms can be found in the Benelux states and
54 Northern-Italy (Ihle et al. 2017).

55 A common methodology to examine the environmental sustainability of agri-food products is Life
56 Cycle Assessment (LCA) (Nguyen et al. 2010). The LCA framework can be extended to cover the
57 economic and social dimensions, i.e., through Life Cycle Costing (LCC) and social LCA (SLCA).
58 LCC is often applied to estimate costs and profits (Florindo et al. 2017), while SLCA aims to assess

¹ Non-castrated male bovine animals aged 1 year or more

59 impacts of production on the workforce, the local community, consumers, value chain actors, and
60 society (Achten et al. 2020). Life Cycle Sustainability Assessment (LCSA) provides an integrated
61 methodological framework based on the three-pillar concept of sustainability first mentioned in the
62 Brundtland report that combines LCA, LCC and SLCA (Zamagni 2012).

63 Several studies estimate environmental impacts of beef production in the EU, highlighting the role
64 of emissions from enteric fermentation, fodder production and manure management (e.g. Angerer
65 et al. 2021). Kamilaris et al. (2020) assessed the economic profitability of different beef production
66 scenarios alongside their environmental sustainability. Bragaglio et al. (2018) added the protein
67 conversion efficiency to account for the societal concern of feed vs. food competition in their LCA
68 of beef production in Italy. Yet, there are no examples of a LCSA application to European beef
69 production systems.

70 LCAs are generally conducted in a static setting, which does not consider the adaption of farmers
71 to changing conditions and their potential consequences (Lan and Yao 2019). In contrast,
72 mathematical modelling is a tool that captures decision-making, inter alia, in food production
73 systems (Djekic et al. 2018). For instance, farm models, like the FarmDyn model, focus on a farm-
74 scale analysis and are frequently used for assessing environmental impacts (Britz et al. 2021). Their
75 scope at the farm-level as the key decision-making unit allows capturing economic, environmental,
76 and social impacts of management scenarios and policies (Reidsma et al. 2018). In the LCA
77 context, optimization models can provide insights on changes of the environmental performance
78 of agricultural systems due to farmers' adaptation to changing conditions such as price or yield
79 changes (Veysset et al., 2010). By definition, bio-economic models capture not only biophysical
80 but also economic flows within and between farms and, therefore, are well suited to add the
81 economic dimension to LCA (Crosson et al., 2011). The advantages of optimization models can
82 also be utilized in large-scale sensitivity analysis (Pahmeyer et al. 2020). When carrying out LCA,
83 methodological choices and input data lead to uncertainty that affects the reliability of the results
84 and is commonly assessed by means of sensitivity and uncertainty analyses (Escobar et al. 2014).
85 However, the potential of bio-economic farm models to carry out both LCSA and LCA remains
86 underexplored.

87 The goal of this study is to assess the environmental, economic, and social performance of three
88 beef production systems in the EU within a LCSA framework. The FarmDyn model is applied to
89 assess sustainability trade-offs and benefits, while considering variability in prices, yields and

90 animal performance, as well as farmers' behaviour in the different geographical contexts. The
91 ultimate goal is to identify potential levers to increase the sustainability of typical EU beef
92 production systems on a farm-level, informing cleaner production strategies for farmers and policy
93 initiatives towards more sustainable beef production in the EU.

Journal Pre-proof

94 2. Materials and Methods

95 The LCSA is carried out according to the ISO standards 14040/44:2006 (ISO, 2006a, ISO, 2006b),
96 which include the following steps: goal and scope definition, life cycle inventory (LCI) analysis
97 and life cycle impact assessment (LCIA).

98 2.1 Goal and scope definition

99 The goal of this study is to compare the social, economic and environmental performance of three
100 typical beef production systems in the EU, as observed in major producing countries, namely
101 France, Germany, Italy and Belgium. The systems are defined from cradle to farm gate based on
102 data from one year (2017), covering several representative farms that were selected from the Agri
103 benchmark network (Chibanda et al. 2020), the International Farm Comparison Network (Hemme
104 et al. 2000) and the SustainBeef project (Mosnier et al. 2021). They were chosen for being
105 representative of dominant production systems in the EU. Impacts are calculated for each
106 production system and each farm within a system separately. The functional unit (FU) is one kg
107 carcass weight from slaughtered bulls. Carcasses from bulls constitute a different product
108 compared to other cattle (heifers, bullocks, cull cow), given the different product qualities and
109 prices. Co-products of bull production in the analysed systems are female calves (either sold, used
110 for replacement or sold as heifers, depending on the system) and cull cow beef. In dairy herds, milk
111 is also produced alongside the calves. Economic allocation is applied to allocate the impacts
112 between the co-products. It is the preferred method for allocation because the necessary
113 information on prices and economic flows is readily available in the used modelling framework.
114 Furthermore, the complexity of the systems makes it difficult to consistently define causal
115 relationships of physical flows throughout the different sub-steps (Mackenzie et al. 2017). The
116 allocation is thus based on revenues. The specific prices are taken from the farm data described
117 below. Where no exogenous market price exists, the optimization model is used to provide the
118 shadow prices for the economic allocation (Seidel & Britz 2020).

119 The three systems are described below. Key characteristics are summarized in table 1.

- 120 - The first system represents beef production in Wallonia, Belgium (BE). It consists of one
121 single farm that breeds and fattens (BE-BF) animals of the Belgian Blue breed on a mixed
122 diet of silage, beet pulp, and bought and self-produced concentrates. While suckler cows
123 are grazing during their lifetime, bulls are fattened indoors. Besides beef production, the

124 farm grows rapeseed, cereals and sugar beet as cash crops. 48% of the Belgium suckler
 125 cows are managed on farms with comparable herd size in Wallonia (Eurostat 2016).

126 - The second system (FR-IT) starts with a suckler cattle farm in the Massif Central, France
 127 (FR-IT-B). It keeps a herd of suckler cows of the Saler breed that are cross-bred with bulls
 128 of the Charolais breed. A portion of the herd is used to breed pure Salers-animals for
 129 replacement. The mountainous conditions only allow for permanent grasslands. Therefore,
 130 the feed consists of grazing, hay and bought concentrates. 16% of the French suckler cow
 131 herd is located in the Auvergne-Rhône-Alpes region (Eurostat 2016). The male offspring
 132 is transferred 800 km via lorry to Veneto (Italy) after weaning. The Italian farm (FR-IT-F)
 133 fattens the bulls with high daily weight gains (around 1.3 kg/day). The diet consists of maize
 134 silage as the main crop grown, beet pulp and concentrates. 31% of the bulls in Italy are
 135 managed on farms with comparable herd size in Northeast Italy (Eurostat 2016).

136 - The third system (GE-GE) starts with a dairy farm in Bavaria, Germany, which has a herd
 137 of Simmental Fleckvieh dairy cows (GE-GE-B). The farm produces milk, calves and grows
 138 fodder and cash crops, together with grasslands. Cows are fed a diet of maize and grass
 139 silage with complementation of concentrates. 16% of the German dairy cows are managed
 140 on farms with comparable herd size in Bavaria (Eurostat 2016). The 6-week-old male
 141 offspring is transported over 600 km via lorry to the North-West of Germany. The second
 142 farm (GE-GE-F) is involved in weaning, fattening and cash crop production. The weaning
 143 and fattening are based on a diet of maize silage and bought concentrates. 14% of the bulls
 144 in Germany are managed on farms with comparable herd size in North-Rhine-Westphalia
 145 (Eurostat 2016).

146 *Table 1 Overview on the systems and farms under analysis*

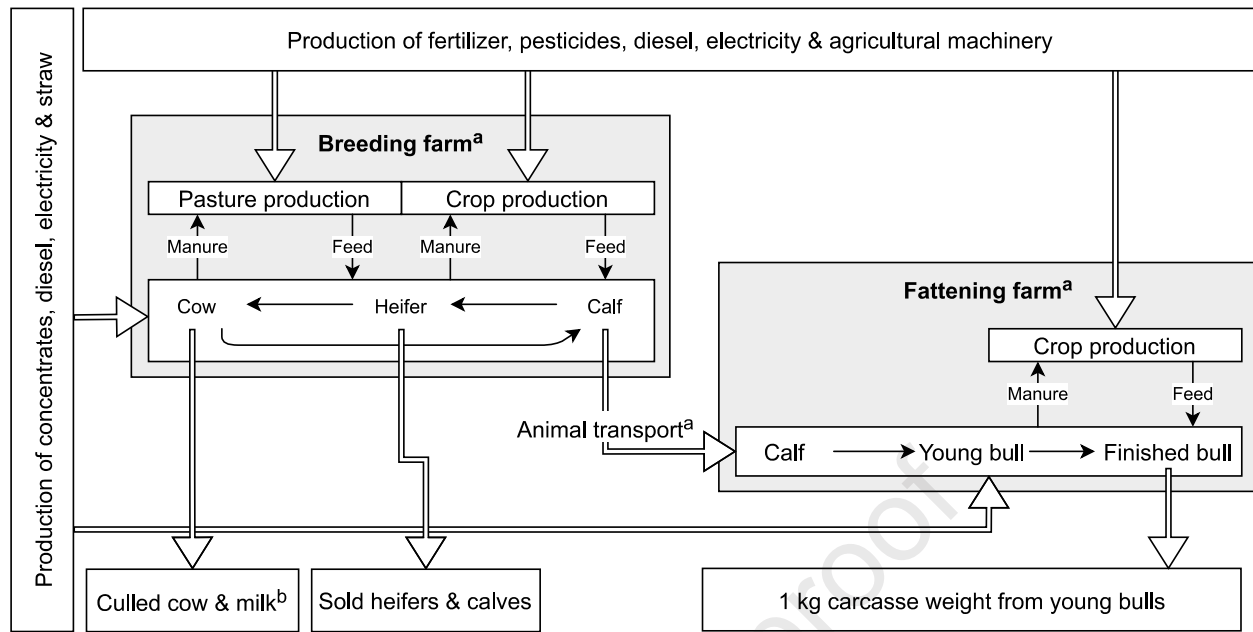
System	BE		FR-IT		GE-GE	
Farm ^a	BE-BF	FR-IT-B	FR-IT-F	GE-GE-B	GE-GE-F	
Country	Belgium	France	Italy	Germany	Germany	
Location	Wallonia	Massif Central	Veneto	Bavaria	North Rhine- Westphalia	
No. sold male animals per year ^b	56	38	324	48	280	
No. of cows	115	79	-	120	-	

Breed	Belgian Blue	Charolais Salers	& Charolais & Salers	Simmental	Simmental
Live weight at butchering ^c	640 kg	380-390 kg	700 kg	85 kg	720 kg
Age at selling ^d	20 months	9 months	17 months	1.5 months	18.7 months
Dress percentage	70 %	-	57 %	-	55 %
Arable land	49 ha	-	33 ha	39 ha	70 ha
Grassland	61 ha	96 ha	-	60 ha	-
Other activities generating co- products	cash crop	-	-	dairy, cash crop	cash crop

147 “a” Indices B and F stand for breeder and fatterer. “b” for breeding farms, this is the number of sold male
 148 calves, for fattening farms this is the number of butchered bulls. “c” for breeding farms, this is the weight
 149 at which the bull calves are transferred to the fattening farm. “d” for breeding farms, this is the age at transfer
 150 of bull calves, for fattening farms this is the age at butchering.

151 The system boundaries include all stages to deliver 1 kg of bull carcass weight from cradle to farm
 152 gate. As can be seen in figure 1, this refers to feed production (cultivation, seeding, fertilizing,
 153 pesticide application, liming and harvest), breeding (recreational activity in the herd, care taking
 154 of cows, heifers and calves), and fattening, as well as transport of animals between farms in FR-IT
 155 and GE-GE. Impacts associated with the production of agricultural inputs and services are included
 156 within the system boundaries, i.e., machinery production and operation, energy, concentrates,
 157 fertilizer and pesticide production.

158 In BE and the breeding farm in FR-IT, manure is handled as solid manure, whereas on the other
 159 farms, it is handled as liquid. In all systems, the amount of manure generated per FU is reused for
 160 fertilization and does not constitute a by-product from the system. Impacts from transport of the
 161 bulls to the slaughterhouse as well as from processing of the meat are excluded from the system
 162 boundaries.



163
 164 *Figure 1: System boundaries of the analysed beef production system. “a” in the Belgium system*
 165 *breeding and fattening are integrated in one farm which spares animal transport. “b” milk is only*
 166 *a co-product on the dairy farm of the German system*

167 2.2 Life cycle inventory

168 The LCI of the inputs and outputs entering and leaving the system boundaries is generated with the
 169 optimization model FarmDyn (Britz et al. 2014). FarmDyn captures economic as well as bio-
 170 physical processes. The model simulates farm management options, while the outcome represents
 171 the economically optimal distribution of agricultural activities and practices, maximizing the farms
 172 profit. FarmDyn was originally developed to enhance sustainability of agricultural systems and
 173 was recently expanded to depict cattle farming systems in the European context (Kuhn et al. 2020;
 174 Pahmeyer and Britz 2020). Each farm operates as an individual entity, which means that the farm
 175 program (including cash crop and dairy production) is optimized subject to boundary conditions
 176 such as prices or farm endowments. Farmers’ decisions include, inter alia, which animals to keep,
 177 how to feed them, which crops to grow and how to fertilize them. As for animal production,
 178 FarmDyn captures herd demographics (calving, raising periods, replacement, and selling) per
 179 month. The feed requirements are calculated using the methodology of the feed planning tool Zifo2
 180 (LfL 2016), by considering dry matter, fibre, protein, energy and nutrient intake as well as animal
 181 performance and lactation periods. The requirements can be met with a variety of bought and self-

182 produced feedstuff. The composition of nutrients in each feed is taken from LfL (2020). The
183 resulting feed use is shown in Table S1 of the Electronic Supplementary Material (ESM).

184 Crop production options are farm-specific by considering the respective yields, fertilizer needs and
185 land endowments. FarmDyna includes both cash and fodder crops, namely wheat, barley, rapeseed,
186 sugar beet, and maize silage. Grassland is differentiated by different means of harvest (silage, hay,
187 baling, grazing), seasonality, productivity and quality of the harvest.

188 On-farm emissions from the optimal activities after profit maximization are estimated according to
189 the methods specified in Table 2, including methane (CH₄), ammonia (NH₃), nitrogen oxides
190 (NO_x), nitrous oxide (N₂O), particulate matter emission (PM_{2.5}), nitrate (NO₃⁻) and phosphorus (P).
191 Emissions arising through the production of major farm inputs are based on the Ecoinvent database
192 version 3.6 (Wernet et al. 2016). These refer to the provision and transport of externally bought
193 feedstuff, bedding material, fertilizers, pesticides; as well as diesel used in agricultural machinery
194 for field and stable operations including cultivation, harvest, manure management and spreading.
195 The field and stable operations cover provision and operation of machines as well as energy
196 consumption. In FR-IT and GE-GE, impacts on the breeding farms are calculated per kg of live
197 weight of transferred animals, which are subsequently implemented as emission factors into the
198 optimization problem of the fattening farm.

199 Price data and work endowments are modelled based on the farm data from the Agri benchmark
200 network (Chibanda et al. 2020), the International Farm Comparison Network (Hemme et al. 2000)
201 and the SustainBeef project (Mosnier et al. 2021). Prices not covered in the above-mentioned
202 sources as well as work time requirements are taken from farm planning data (Achilles 2016). The
203 human-consumable share of protein and calorie content of the feedstuff and meat are based on
204 Laisse et al. (2016), Ertl et al. (2016) and Wilkinson (2011).

205 Table 2. On-farm emissions included in the environmental life cycle inventory and associated
206 estimation methods.

Source / Sub-source	Pollutant	Methodology	Tier ^a
Enteric fermentation	CH ₄	IPCC (2019)	2
Manure management	CH ₄	IPCC (2019)	2
	NH ₃ , N ₂ O, NO _x , N ₂	EEA (2016)	2
	Particulate matter	EEA (2013)	2

Pasture	CH ₄	IPCC (2019)	2
	NH ₃	EEA (2016)	2
	N ₂ O, NO _x , N ₂	IPCC (2019)	1
Field & Pasture / Manure application	NH ₃	EEA (2016)	2
	N ₂ O, NO _x , N ₂	IPCC (2019)	1
Field & Pasture / Fertilizer application	NH ₃	EEA (2016)	2
	N ₂ O, NO _x , N ₂	IPCC (2019)	1
Field / Lime application	CO ₂	IPCC (2019)	1
Field / Crop residues	N ₂ O, N ₂	IPCC (2019)	1
Field	Particulate matter	EEA (2016)	1
Field & Pasture	NO ₃ ⁻	Richner (2014)	
	P	Prasuhn (2006)	
Indirect N ₂ O	N ₂ O	IPCC (2019)	1

207 ^a In IPCC (2019) tiers represent three different levels of methodological complexity with tier 1 being the
208 basic method and tier 3 being the most complex method.

209 2.3 Life cycle impact assessment

210 The LCIA employs the ReCiPe methodology to quantify the following environmental impact
211 categories at the midpoint level (hierarchical perspective) (Huijbregts et al. 2017): global warming
212 potential (GWP), terrestrial acidification potential (TAP), freshwater eutrophication potential
213 (FEP), marine water eutrophication potential (MEP), particulate matter formation potential
214 (PMFP) and fossil fuel depletion potential (FDP). These have been identified as the most relevant
215 categories for the based on a comprehensive literature review of LCAs on beef production by de
216 Vries et al. (2015).

217 The economic performance is measured with the contribution margin (CM) per kg of carcass
218 weight. The CM is the revenues from a product deducted by variable costs to produce such product.
219 This includes revenues from sold beef, costs of buying concentrates, costs of producing roughages,
220 feed costs for rearing, operation and maintenance of machinery, costs of buying animals, variable
221 stable costs and other variable costs. Roughage production costs are measured based on the shadow
222 prices given by the model (Seidel & Britz 2020).

223 As for the social performance, working time (WT) on farm per FU is considered, differentiated by
224 type of work, i.e., feeding and taking care of the herd, work for calving, field work, stable
225 maintenance, fertilization and management and office work. Further social indicators considered
226 are the human-consumable calories (HCC) and protein (HCP) used to produce one kg carcass
227 weight. The indicators are included to represent the contribution of beef production to human
228 nutrition as this has been an ongoing societal debate (Mosnier et al. 2021).

229 2.4 Sensitivity analysis

230 FarmDyn allows performing a global all-at-once sensitivity analysis to examine the influence of
231 parametric uncertainty on the LCA results. The following parameters involved in the economic
232 optimization as well as allocation are varied: the beef price, the price of calves and weaned calves,
233 the milk price, and the price of concentrates. Additionally, the spatial and biological variability in
234 the systems is considered through variations in the yield of major roughage crops (grass and maize)
235 and animal parameters such as the weight and age at butchering, and the weight of weaned calves
236 (Table S2 in ESM). Using Latin Hypercube Sampling, a sample of 1,000 draws with
237 simultaneously changed levels of the aforementioned parameters is created, covering the full range
238 of possible factor level permutations. Because the distributions of the varied parameters are
239 unknown, uniform distributions without correlations are assumed. In FR-IT and GE-GE, the spatial
240 and temporal separation of the farms are considered by using separate sets of 1,000 draws on each
241 farm for crop yields and concentrate prices, respectively. The remaining parameters are similar on
242 the farms in the systems. For each draw, the management decisions on each farm are optimized
243 considering the changed parameters. The results of each optimized farm are combined in a single
244 data frame for each system and are then rescaled to have a mean of zero and a standard deviation
245 of one. This *standardization* allows the comparison of measurements that have different units. The
246 data frame is analysed through a regression analysis via ordinary least squares. The resulting
247 regression models are considered as meta-models and indicate the relative influence of the
248 parameters on the results.

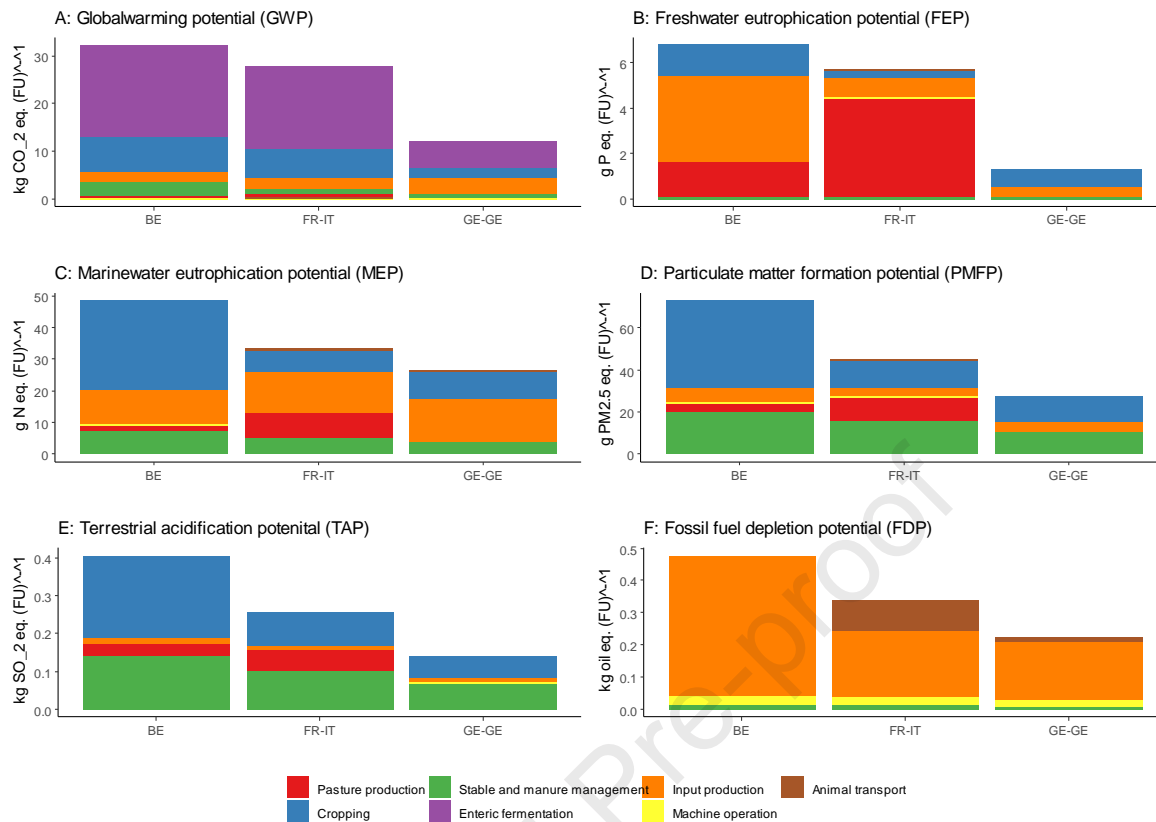
249 **3. Results**

250 3.1 Sustainability assessment

251 GE-GE shows the lowest values across all environmental impact categories, followed by FR-IT
252 and BE (Figure 2). BE has a GWP of 32.3 kg CO₂eq. per FU, compared to 27.7 kg in FR-IT and
253 12.0 kg in GE-GE. In the latter, impacts from the breeding stage are partially allocated to the co-
254 product milk. FR-IT performs better than BE due to the shorter lifespan of the animals. Enteric
255 fermentation constitutes the largest source of GWP across systems (46.5% - 62.4 %). Second
256 largest GHG emission sources are input production in GE-GE and FR-IT, and on-field emissions
257 in BE, all accounting for >20% of the GWP, respectively. This is due to the larger share of self-
258 produced feeds in BE. In FR-IT and GE-GE imported concentrates add emissions (included in
259 upstream input production).

260 The FEP sums up to 6.78 g P eq. per FU in BE, 5.67 g in FR-IT and 1.33 g in GE-GE. The greatest
261 contribution to FEP in BE is input production, specifically imported concentrates, with a share of
262 55.3%. In FR-IT, emissions from pastures (76.5%) dominate because of more grazing on the
263 breeding farm. In GE-GE, on-field emissions account for the largest share of FEP (62.4 %) as
264 maize silage is grown, which is prone to nutrient loss.

265 MEP is related to N leaching from fields and pasture, and NH₃ emissions from the concentrate
266 production and manure management. Total emissions of MEP sum up to 48.6 g N eq. per FU in
267 BE, 33.3 g in FR-IT and 26.3 g in GE-GE. In BE, crop production for self-produced feed accounts
268 for the largest share of the impact (58.7%). In FR-IT and GE-GE, the largest share is associated
269 with input production (>37%), specifically imported concentrates.



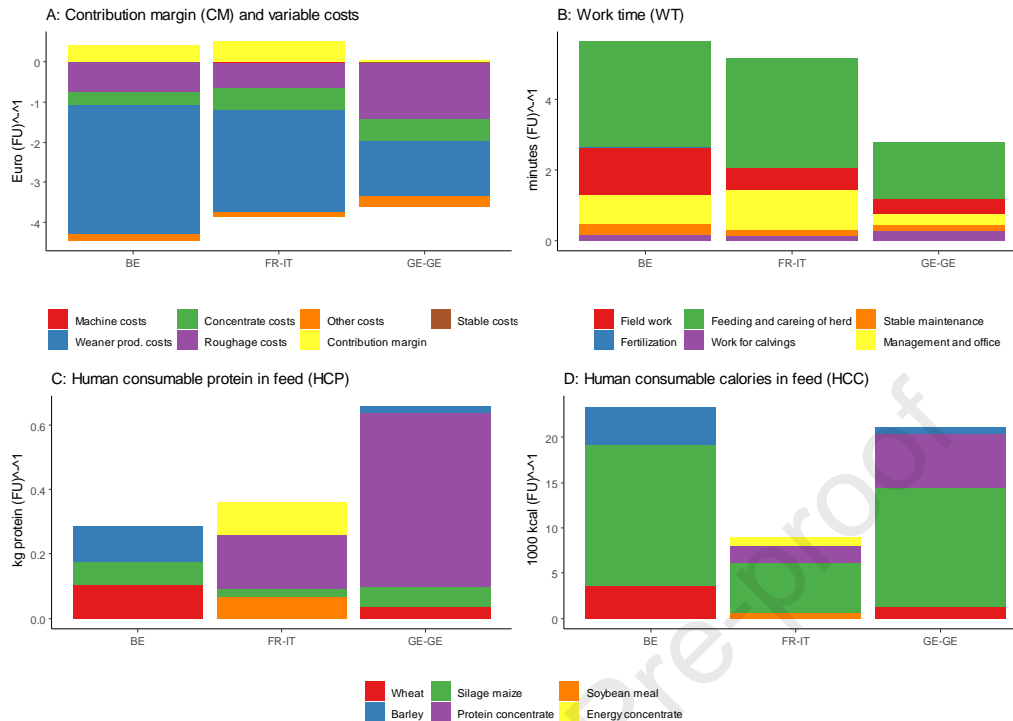
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271 *Figure 2 Environmental impacts of the beef production systems per kg of bull carcass. BE*
 272 *indicates the Belgium system, FR-IT the French-Italian system and GE-GE the German system.*
 273 *FU stands for 1 kg carcass weight from slaughtered young bulls*

274 The PMFP is estimated at 72.9 g in BE, 45.1 g in FR-IT and 27.3 g PM eq. per FU in GE-GE. The
 275 TAP sums up to 0.40 kg in BE, 0.26 kg in FR-IT and 0.14 kg SO₂ eq. per FU in GE-GE. Both
 276 PMFP and TAP are mainly caused by NH₃ emissions. Crop production and manure management
 277 are the prevailing emission sources in all systems. The allocation to the co-product milk leads to a
 278 better performance of GE-GE. FR-IT performs better than BE due to the shorter lifespan of the
 279 animals. The contribution of pastures to the PMFP and TAP in FR-IT is associated with the grazing
 280 in the breeding farm.

281 As for FDP, BE consumes 0.48 kg oil eq. per FU, followed by FR-IT (0.34) and GE-GE (0.23).
 282 Provision of inputs accounts for the largest share across systems. The transport of live animals in
 283 FR-IT contributes 28.1% to overall FDP compared to 7.11% in GE-GE because of a longer
 284 transport distance and higher weight of the transferred animals in FR-IT.

285 3.2 Economic and social indicators



286

287 *Figure 3 Economic and social indicators assessed with FarmDyn for the three systems. BE*
 288 *indicates the Belgium system, FR-IT represents the French-Italian system and GE-GE the*
 289 *German system. FU stands for 1 kg carcass weight from slaughtered young bulls*

290 The CM per FU is estimated at 0.39 € in BE, 0.50 € in FR-IT and 0.03 € in GE-GE. In BE and FR-
 291 IT, weanling production with suckler cows leads to the largest cost share with 71.6% and 66.0%,
 292 respectively. In GE-GE, calves are bought at a young age from dual-purpose dairy breeds resulting
 293 in lower costs (38.1%). In GE-GE, roughage production accounts for the largest share of costs with
 294 38.3%. Roughages are produced on arable land that bears opportunity costs because of the
 295 competition with cash crops. Feed concentrate costs are higher in systems with intensive fattening
 296 (FR-IT and GE-GE) because of the higher nutrient need for the higher weight gain.

297 As for the social performance, BE entails the highest workload with 5.63 minutes per FU, followed
 298 by FR-IT (5.17) and GE-GE (2.79). In GE-GE, less time is spent on calf production compared to
 299 BE and FR-IT because of the allocation towards milk production. The routine of sustaining the
 300 herd including feeding constitutes the largest share of workload, followed by field and management
 301 work. In BE, the WT is longer because cereals for feeding are produced on-farm. FR-IT entails

302 additional workload compared to BE and GE-GE because there are no shared efforts with other
303 farming branches, like management work.

304 All systems are net protein- and energy-consumers, meaning that more human-consumable protein
305 and energy are fed than produced. In BE, 0.29 kg human-consumable protein are fed per FU,
306 followed by FR-IT with 0.36 and GE-GE with 0.66. BE and FR-IT benefit from the high intake of
307 grass, which offers a source of protein non-edible by humans. GE-GE has the highest HCP. Here,
308 bulls receive maize as roughage. Since maize is rich in energy, diets must be balanced by adding
309 protein in the form of concentrates which have a high share of human consumable protein.

310 FR-IT has the lowest HCC at 8,900 human-consumable kcal in the feed per FU, followed by GE-
311 GE at 21,110 and BE at 23,300. The age of the animals determines the comparative result because
312 the energy required for maintaining their metabolism adds up over the lifetime of the animals. In
313 addition, the feeding of concentrates as energy supplement and the larger share of maize silage in
314 the ration further reduce the efficiency in BE and GE-GE. In FR-IT, beet-pulps (considered as non-
315 consumable by humans) are used to a larger extent, increasing the efficiency.

316 3.3 Sensitivity analysis

317 The regression output of all meta-models including R^2 , adjusted R^2 , Residual Std. Error and F-
318 Statistic is shown in the ESM table S3-S5. This sub-section focuses on GWP, CM, and WT, as
319 representation of the environmental, economic, and social dimension. The beta coefficients of the
320 regression models for GWP, CM and WT and the 95% confidence interval are shown in figure 4.

321 The beef price is among the factors with the greatest influence on the indicators. In all systems, a
322 higher beef price leads to a higher CM as this implies higher revenues. In BE, a higher beef price
323 leads to a higher GWP and WT because more emissions and work time are credited to beef
324 production in the allocation. In FR-IT, the beef price has little influence on the GWP and WT as
325 the fattening is limited by the endowment of stables and hence the herd size is constant with
326 increasing prices. Furthermore, it is a specialized fattening farm and no allocation is applied.

327 Variation in the animal weight impacts the performance of all systems. A share of the costs and
328 work tasks are constant per animal. When these are related to a higher weight per animal it results
329 in higher CM and lower WT per FU. A higher share of concentrates in the animals' ration is needed
330 to sustain the higher weight gain, causing additional emissions that increase GWP, e.g., in GE-GE.
331 However, the efficiency gain can outweigh these emissions, overall reducing GWP per FU, e.g., in
332 BE. With a higher weight, the revenues of animals increase. A higher revenue for bull calves leads
333 to higher emissions and time associated with the bull-calf production during the breeding stage due
334 to allocation. The higher price for the heavier calves bares higher costs on the fattening farm and
335 causes a lower CM. A higher price of calves and weaners can also lead to less bulls fattened due to
336 higher costs on the fattening farm, e.g. in GE-GE. Less bulls fattened implies that costs and labour
337 are distributed over less output, which decreases CM and increases WT per FU. Furthermore, the
338 self-produced roughages can be utilized better, which reduces GWP.

339 With a higher concentrate price, concentrates are used in smaller amounts, hence reducing GWP.
340 At the same time, the higher prices translate into higher feed costs, which slightly reduces the CM.
341 The smaller amount of concentrates increases the relative share of on-farm produced feed, which
342 increases the WT.

343 The impact of changes in yield of maize and grassland depends on how the yield is used: If
344 additional yield is used to replace low-emission concentrates, the GWP rises (e.g. in FR-IT), if it
345 is replaced with feedstuff with a high emission load the GWP decreases (e.g. maize yield in BE).

346 In all cases, increasing yields results in reduced feed costs and increased CM. WT increases with
347 higher amounts of self-produced feed. However, WT savings are also possible, when the land is
348 better utilized or the additional yield is utilized in grazing, which spares feeding time.

Journal Pre-proof



349

350 *Figure 4 Tornado diagram showing the influence of each parameter in the sensitivity analysis on the results in terms of global warming*
 351 *potential, contribution margin and working time. The standardized coefficients indicate the relative importance of each coefficient in the*

352 *related regressions. The unit of measurement is one standard deviation. The error bars indicate the 95% confidence intervals. Factors marked*
353 *with a '*' are specific to fattening farms.*

Journal Pre-proof

354 **4. Discussion**

355 The results suggest that the system fattening dairy breed bulls is favourable for the analysed
356 environmental indicators compared to the fattening of beef breed bulls. This is in line with previous
357 findings, for example Nguyen et al. (2010). Carbon sequestration through grassland production is
358 not considered, which could improve the performance of grass-based systems. However, recent
359 research by Hammar et al. (2022) found that a forage-grain beef system resulted in lower GWP
360 compared to an extensive grazing system even with consideration of carbon sequestration. Still,
361 cattle can be important to sustain current carbon pools under grassland (Conant et al. 2017). Huerta
362 et al. (2016) found extensive systems to outperform intensive systems in several environmental
363 impact categories indicating that the results depend on assumptions, used indicators, the location
364 and further characteristics of the analysed system.

365 A comparison of the results with the literature can be found in the ESM (S6) including information
366 on the FU and the scope of the respective studies. Here the FU is kg carcass weight from
367 slaughtered bulls without the consideration of slaughtering and retail. This inconsistency was
368 chosen as it allows the consideration of different dressing percentages of the different cattle breeds
369 while compromising on the comparability with other studies. However, the contribution of the
370 slaughtering and retail stage on the entire life cycle is reduced compared to the agricultural stage
371 (e.g., Huerta et al. 2016).

372 A major contribution of this study is that it includes indicators beyond the common environmental
373 impact categories in LCA to assess and compare the sustainability of beef farms under a LCSA
374 approach. The results show that the system with dairy breed bulls (GE-GE) has the lowest CM and
375 the highest HCP pointing at a trade-off between environmental and other sustainability indicators.
376 Kamilaris et al. (2020) found that intensive systems had a lower GWP, too, but their research shows
377 that intensive systems were more profitable. The contrasting results are caused by a higher beef
378 price in FR-IT and BE. A high HCP is also found in the literature (Bragaglio et al. 2018;
379 Wiedemann et al. 2015).

380 In this study, WT, HCC, and HCP are proposed as social indicators in the LCSA. Due to the novelty
381 of the approach, comparison to the existing literature is limited. The WT is calculated using German
382 farm planning data (Achilles 2016), which does not necessarily cover all particularities of the
383 analysed systems at the same level of detail as for environmental and economic indicators.

384 However, the data enables consideration of economies of scale of stables, different mechanization
385 levels and plot sizes. The WT indicator would benefit from a detailed representation of the work
386 types and a weighting of tasks by, for example, health hazards, employment potential or personal
387 fulfilment of the workers. In addition, WT spent in upstream processes like the production of inputs
388 should be included to gain insight on affected stakeholders outside the farming community and
389 align with the scope of the LCA. Other indicators of societal concern could be animal welfare or
390 human health (Paris et al. 2022). Implementing these kinds of indicators in FarmDyn is difficult as
391 quantifiable metrics and databases are not readily available.

392 The results indicate the potential of farm-level models in the application for LCSA as they offer
393 the technical detail to capture farm heterogeneity and present a framework to integrate economic
394 and social indicators. Another advantage is the utilization of the linear optimization to obtain
395 shadow prices where information on market prices is scarce, e.g., the costs of roughage production.

396 In the context of the sensitivity analysis, the farm-model captures the performance of the system
397 when conditions change. These conditions differ within systems and time, adding uncertainty to
398 the results. The model simulates farmers' decisions on production and management activities in
399 response to changing conditions. The sensitivity analysis points to the prices of beef and male
400 calves as influential parameters for the sustainability performance. Within the framework
401 proposed, higher prices tend to impact the systems through adjustments in the activities as well as
402 in allocation factors, which are estimated based on economic criteria. In view of the lack of
403 agreement on the allocation method (e.g. Wilfart et al. 2021), economic allocation is preferred here
404 over physical allocation, because the two major co-products obtained (meat and milk) have two
405 very distinct markets with stable demand for both, while prices are highly variable. FarmDyn
406 captures country-specific, detailed prices and economic flows, hence offering advantages to carry
407 out consistent economic allocation, relative to conventional LCA approaches. Furthermore,
408 physical allocation is not established for suckler cows because their milk is only used for weaning
409 and yields are unknown (Kyttä et al. 2022).

410 Finally, the study contributes to the debate on meat production and consumption in the EU,
411 considering multiple dimensions of sustainability. Despite declining consumption of beef meat in
412 the EU, production will likely not vanish (Hocquette et al. 2018). Levers to improve the
413 sustainability of existing production systems according to the results could be the efficient usage
414 of feedstuff non-edible by humans, e.g. industry by-products and grasslands and the integration of

415 dairy and beef production (van Selm et al. 2021). Decision-makers should be aware of farm
416 heterogeneity and the possibility of trade-offs between sustainability dimensions. Multi-criteria
417 decision-making (MCDM) tools offer the possibility to combine indicators in a single score and
418 choose options “close to the optimum” using subjective weights (Saeidi et al 2022). However, the
419 goal of this study is to compare the systems' performance and identify tradeoffs and hotspots in
420 each system among sustainability dimensions and not to rank systems. Performing MCDM analysis
421 would arguably come at the cost of losing detail and complexity and can result in misleading
422 conclusions.

423

424 **5. Conclusion**

425 The model FarmDyn is used to carry out a LCSA of three bull-beef production systems in major
426 producing EU countries including a comprehensive sensitivity analysis. Potential trade-offs
427 between different dimensions of sustainability are identified underlining the need to consider
428 economic and social indicators when comparing the sustainability of beef production. The dairy-
429 based bull fattening system shows better results in environmental indicators while economic
430 profitability, social indicators favoured the systems which utilized grasslands and industry by-
431 products in feeding. FarmDyn enabled the inclusion of price effects in the sensitivity analysis and
432 the economic allocation. Additional indicators would be needed to better represent the social
433 dimension of beef production, although this entails methodological challenges mainly related to
434 data availability. Future research should focus on the application to a larger farm sample to estimate
435 the extent of the observed findings and gain more representative results. The application of MCDM
436 could combine the indicators in a single score and help identifying favourable systems.

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444 **References**

- 445 Achilles, W. (2016): Betriebsplanung Landwirtschaft 2016/17. Daten für die Betriebsplanung in
446 der Landwirtschaft. 25th ed. Edited by Norbert Sauer. Darmstadt: Kuratorium für Technik und
447 Bauwesen in der Landwirtschaft (KTBL - Datensammlung, 2016/17).
- 448 Achten, W., Barbeau-Baril, J., Barros Telles Do Carmo, B., Bolt, P., Chandola, V., Corona
449 Bellostas, B., Dadhish, Y., Di Eusanio, M., Di Cesare, S., Di Noi, C., Eisfeldt, F. et al. (2020):
450 Guidelines for social life cycle assessment of products and organizations, pp. 138.
- 451 Angerer, V.; Sabia, E.; König von Borstel, U.; Gaulty, M. (2021): Environmental and biodiversity
452 effects of different beef production systems. *J. Environ. Manage.* 289, p. 112523. DOI:
453 10.1016/j.jenvman.2021.112523.
- 454 Bonnet, C.; Bouamra-Mechemache, Z.; Réquillart, V.; Treich, N. (2020): Viewpoint: Regulating
455 meat consumption to improve health, the environment and animal welfare. *Food Policy* 97, p.
456 101847. DOI: 10.1016/j.foodpol.2020.101847.
- 457 Bragaglio, A.; Napolitano, F.; Pacelli, C.; Pirlo, G.; Sabia, E.; Serrapica, F.; Serrapica, M.;
458 Braghieri, A. (2018): Environmental impacts of Italian beef production: A comparison between
459 different systems. *J. Clean. Prod.* 172, pp. 4033–4043. DOI: 10.1016/j.jclepro.2017.03.078.
- 460 Britz, W.; Lengers, B.; Kuhn, T.; Schäfer, D. (2014): A highly detailed template model for
461 dynamic optimization of farms. *Institute for Food and Resource Economics*, University of Bonn.
462 Model Documentation.
- 463 Britz, W.; Ciaian, P.; Gocht, A.; Kanellopoulos, A.; Kremmydas, D.; Müller, M.; Petsakos, A.;
464 Reidsma, P. (2021): A design for a generic and modular bio-economic farm model. *Agric. Syst.*
465 191, p. 103133. DOI: 10.1016/j.agsy.2021.103133.
- 466 Buendia, Eduardo; Tanabe, Kiyoto; Kranjc, Andrej; Jamsranjav, Baasansuren; Fukuda, Maya;
467 Ngarize, Sekai et al. (2019): 2019 Refinement to the 2006 IPCC Guidelines for National
468 Greenhouse Gas Inventories. Available online at [https://www.ipcc-](https://www.ipcc-nggip.iges.or.jp/public/2019rf/index.html)
469 [nggip.iges.or.jp/public/2019rf/index.html](https://www.ipcc-nggip.iges.or.jp/public/2019rf/index.html).
- 470 Cassidy, Emily S.; West, Paul C.; Gerber, James S.; Foley, Jonathan A. (2013): Redefining
471 agricultural yields: from tonnes to people nourished per hectare. *Environ. Res. Lett.* 8 (3),
472 p. 34015. DOI: 10.1088/1748-9326/8/3/034015.

- 473 Chibanda, Craig; Agethen, Katrin; Deblitz, Claus; Zimmer, Yelto; Almadani, Mohamad. I.;
474 Garming, Hildegard et al. (2020): The Typical Farm Approach and Its Application by the Agri
475 Benchmark Network. *Agriculture* 10 (12), p. 646. DOI: 10.3390/agriculture10120646.
- 476 Conant, Richard T.; Cerri, Carlos E. P.; Osborne, Brooke B.; Paustian, Keith (2017): Grassland
477 management impacts on soil carbon stocks: a new synthesis. *Ecol. Appl.* 27 (2), pp. 662–668.
478 DOI: 10.1002/eap.1473.
- 479 Cook, E. (Ed.). (2020). *Agriculture, Forestry and Fishery Statistics: 2020 Edition*. Eurostat
480 Publications Office of the European Union: Luxembourg. DOI: 10.2785/496803
- 481 Crosson, P., O’Kiely, P., O’Mara, F.P., Wallace, M., 2006. The development of a mathematical
482 model to investigate Irish beef production systems. *Agric. Syst.* 89, pp. 349-37. DOI:
483 10.1016/j.agsy.2005.09.008.
- 484 Djekic, Ilija; Sanjuán, Neus; Clemente, Gabriela; Jambrak, Anet Režek; Djukić-Vuković,
485 Aleksandra; Brodnjak, Urška Vrabič et al. (2018): Review on environmental models in the food
486 chain - Current status and future perspectives. *J. Clean. Prod.* 176 , pp. 1012–1025. DOI:
487 10.1016/j.jclepro.2017.11.241.
- 488 EEA (2013): *EMEP/EEA air pollutant emission inventory guidebook 2013*. Technical guidance
489 to prepare national emission inventories Eur. Environ. Agency, Tech. Rep. 12.
- 490 EEA (2016): *EMEP/EEA air pollutant emission inventory guidebook 2016*. Technical guidance
491 to prepare national emission inventories Eur. Environ. Agency, Tech. Rep. 21.
- 492 Ertl, P.; Knaus, W.; Zollitsch, W. (2016): An approach to including protein quality when
493 assessing the net contribution of livestock to human food supply. *Animal* 10 (11), pp. 1883–1889.
494 DOI: 10.1017/S1751731116000902.
- 495 Escobar Lanzuela, N., Ribal Sanchís, F.J., Rodrigo Señor, A.; Clemente Polo, G.; Pascual Vidal,
496 A.; Sanjuán Pellicer, N. (2015): Uncertainty analysis in the environmental assessment of an
497 integrated management system for restaurant and catering waste in Spain. *Int. J. Life Cycle. Ass.*
498 20 (2), pp. 244–262. DOI: 10.1007/s11367-014-0825-z.
- 499 Eurostat (2016): *Farm structure survey*. Eurostat. Available online at
500 [https://ec.europa.eu/eurostat/databrowser/view/EF_LSK_BOVINE__custom_3198514/default/ta](https://ec.europa.eu/eurostat/databrowser/view/EF_LSK_BOVINE__custom_3198514/default/table)
501 [ble](https://ec.europa.eu/eurostat/databrowser/view/EF_LSK_BOVINE__custom_3198514/default/table)

- 502 Eurostat (2021): Slaughtering in slaughterhouses - annual data. Eurostat. Available online at
503 [https://ec.europa.eu/eurostat/databrowser/view/APRO_MT_PANN__custom_668469/default/tab](https://ec.europa.eu/eurostat/databrowser/view/APRO_MT_PANN__custom_668469/default/table)
504 e.
- 505 FAO (2020) FAOSTAT Online Database. Available online at
506 <https://www.fao.org/faostat/en/#data/QV>.
- 507 Florindo, T. J.; Medeiros Florindo, G. I. B. de; Talamini, E.; Da Costa, J. S.; Ruviaro, C. F.
508 (2017): Carbon footprint and Life Cycle Costing of beef cattle in the Brazilian midwest. *J. Clean.*
509 *Prod.* 147, pp. 119–129. DOI: 10.1016/j.jclepro.2017.01.021.
- 510 Hammar, T.; Hansson, P.-A.; Rööös, E. (2022): Time-dependent climate impact of beef production
511 – can carbon sequestration in soil offset enteric methane emissions? *J. Clean. Prod.* 331, p.
512 129948. DOI: 10.1016/j.jclepro.2021.129948.
- 513 Hemme, T., Deblitz, C., Isermeyer, F., Knutson, R., & Anderson, D. (2000). The International
514 Farm Comparison Network (IFCN)-objectives, organisation and first results on international
515 competitiveness of dairy production. *Züchtungskunde* 72 (6), pp. 428-439.
- 516 Herrero, M.; Henderson, B.; Havlík, P.; Thornton, P. K.; Conant, R. T.; Smith, P. et al. (2016):
517 Greenhouse gas mitigation potentials in the livestock sector. *Nat. Clim. Chang.* 6 (5), pp. 452–
518 461. DOI: 10.1038/nclimate2925.
- 519 Hocquette, J.F.; Ellies-Oury, M.-P.; Lherm, M.; Pineau, C.; Deblitz, C.; Farmer, L. (2018):
520 Current situation and future prospects for beef production in Europe - A review. *Asian-Australas.*
521 *J. Anim. Sc.* 31 (7), pp. 1017–1035. DOI: 10.5713/ajas.18.0196.
- 522 Huerta, A. R.; Güereca, L. P.; Lozano, M. de la S. R. (2016): Environmental impact of beef
523 production in Mexico through life cycle assessment. *Resour. Conserv. Recy.* 109, pp. 44–53.
524 DOI: 10.1016/j.resconrec.2016.01.020.
- 525 Huijbregts, M. A. J.; Steinmann, Z. J. N.; Elshout, P. M. F.; Stam, G.; Verones, F.; Vieira, M. et
526 al. (2017): ReCiPe2016: a harmonised life cycle impact assessment method at midpoint and
527 endpoint level. *Int. J. Life Cycle. Ass.* 22 (2), pp. 138–147. DOI: 10.1007/s11367-016-1246-y.
- 528 Ihle, Rico; Dries, Liesbeth; Jongeneel, R. A.; Venus, Thomas; Wesseler, Justus (2017): Research
529 for AGRI Committee - The EU cattle sector: challenges and opportunities - milk and meat. Study.
530 Brussels: European Parliament.

- 531 ISO, 2006a. ISO 14040: Environmental Management. Life Cycle Assessment. Principles and,
532 Framework. International Organization for Standardization, Geneva, Switzerland.
- 533 ISO, 2006b. ISO 14044: Environmental Management. Life Cycle Assessment. Requirements and
534 Guidelines. International Organization for Standardization, Geneva, Switzerland.
- 535 Kamilaris, C.; Dewhurst, R. J.; Sykes, A. J.; Alexander, P. (2020): Modelling alternative
536 management scenarios of economic and environmental sustainability of beef finishing systems. *J.*
537 *Clean. Prod.* 253, p. 119888. DOI: 10.1016/j.jclepro.2019.119888.
- 538 Kuhn, T.; Enders, A.; Gaiser, T.; Schäfer, D.; Srivastava, A. K.; Britz, W. (2020): Coupling crop
539 and bio-economic farm modelling to evaluate the revised fertilization regulations in Germany.
540 *Agric. Syst.* 177, p. 102687. DOI: 10.1016/j.agsy.2019.102687.
- 541 Kytä, V.; Roitto, M.; Astaptsev, A.; Saarinen, M.; Tuomisto, H. L. (2022): Review and expert
542 survey of allocation methods used in life cycle assessment of milk and beef. *Int. J. Life Cycle*
543 *Ass.* 27, pp. 191-204. DOI: 10.1007/s11367-021-02019-4.
- 544 Laisse, S.; Rouillé, B.; Baumont, René; Peyraud, Jean-Louis (2016): Evaluation de la
545 contribution nette des systèmes bovins laitiers français à l’approvisionnement alimentaire
546 protéique pour l’être humain. *Rencontres Recherches Ruminants* 23, pp. 263–266.
- 547 Lan, K.; Yao, Y. (2019): Integrating Life Cycle Assessment and Agent-Based Modeling: A
548 Dynamic Modeling Framework for Sustainable Agricultural Systems. *J. Clean. Prod.* 238, p.
549 117853. DOI: 10.1016/j.jclepro.2019.117853.
- 550 LfL (2016): Zifo2–Zielwert Futteroptimierung. Poing, Bavaria: Bayerische Landesanstalt für
551 Landwirtschaft.
- 552 LfL (2020): Gruber Tabelle zur Fütterung in der Rindermast. 24th ed. Poing, Bavaria: Bayerische
553 Landesanstalt für Landwirtschaft.
- 554 Mackenzie, S. G.; Leinonen, I.; Kyriazakis, I. (2017): The need for co-product allocation in the
555 life cycle assessment of agricultural systems—is “biophysical” allocation progress? *Int. J. Life*
556 *Cycle. Ass.* 22 (2), pp. 128–137. DOI: 10.1007/s11367-016-1161-2.
- 557 Mehrabi, Z.; Gill, M.; van Wijk, M.; Herrero, M.; Ramankutty, N. (2020): Livestock policy for
558 sustainable development. *Nat. Food* 1 (3), pp. 160–165. DOI: 10.1038/s43016-020-0042-9.

- 559 Mosnier, C.; Jarousse, A.; Madrange, P.; Balouzat, J.; Guillier, M.; Pirlo, G. et al. (2021):
560 Evaluation of the contribution of 16 European beef production systems to food security. *Agric.*
561 *Syst.* 190, p. 103088. DOI: 10.1016/j.agsy.2021.103088.
- 562 Mottet, A.; Haan, C. de; Falcucci, A.; Tempio, G.; Opio, C.; Gerber, P. (2017): Livestock: On our
563 plates or eating at our table? A new analysis of the feed/food debate. *Glob. Food Sec.* 14, pp. 1–8.
564 DOI: 10.1016/j.gfs.2017.01.001.
- 565 Nguyen, T. L. T.; Hermansen, J. E.; Mogensen, L. (2010): Environmental consequences of
566 different beef production systems in the EU. *J. Clean. Prod.* 18 (8), pp. 756–766. DOI:
567 10.1016/j.jclepro.2009.12.023.
- 568 Pahmeyer, C.; Britz, W. (2020): Economic opportunities of using crossbreeding and sexing in
569 Holstein dairy herds. *J. Dairy Sci.* 103 (9), pp. 8218–8230. DOI: 10.3168/jds.2019-17354.
- 570 Paris, J. M. G.; Falkenberg, T.; Nöthlings, U.; Heinzl, C.; Borgemeister, C.; Escobar, N. (2022):
571 Changing dietary patterns is necessary to improve the sustainability of Western diets from a One
572 Health perspective. *Sci. Total Environ.* 811, p. 151437. DOI: 10.1016/j.scitotenv.2021.151437.
- 573 Peyraud, J.-L.; MacLeod, M. (2020): Study on future of EU livestock. How to contribute to a
574 sustainable agricultural sector? Luxembourg: Publications Office of the European Union.
- 575 Prasuhn, V. (2006): Erfassung der PO₄-Austräge für die Ökobilanzierung. Agroscope,
576 Switzerland.
- 577 Reidsma, P.; Janssen, S.; Jansen, J.; van Ittersum, M. K. (2018): On the development and use of
578 farm models for policy impact assessment in the European Union – A review. *Agric. Syst.* 159,
579 pp. 111–125. DOI: 10.1016/j.agsy.2017.10.012.
- 580 Richner, W.; Oberholzer, H. R.; Freiermuth Knuchel, R.; Huguenin, O.; Ott, S.; Nemecek, T.;
581 Walther, U. (2014): Modell zur Beurteilung der Nitratauswaschung in Ökobilanzen - SALCA-
582 NO₃. *Version 2*.
- 583 Saeidi, Parvaneh; Mardani, Abbas; Mishra, Arunodaya Raj; Cajas Cajas, Viviana Elizabeth;
584 Carvajal, Mercedes Galarraga (2022): Evaluate sustainable human resource management in the
585 manufacturing companies using an extended Pythagorean fuzzy SWARA-TOPSIS method. *J.*
586 *Clean. Prod.* 370, S. 133380. DOI: 10.1016/j.jclepro.2022.133380.

- 587 Seidel, C.; Britz, W. (2020): Estimating a Dual Value Function as a Meta-Model of a Detailed
588 Dynamic Mathematical Programming Model. *Bio-based and Applied Economics* 8 (1) pp. 75-99.
589 DOI: 10.13128/bae-8147.
- 590 van Selm, B.; Boer, I. J.M. de; Ledgard, S. F.; van Middelaar, C. E. (2021): Reducing greenhouse
591 gas emissions of New Zealand beef through better integration of dairy and beef production. *Agric.*
592 *Syst.* 186, p. 102936. DOI: 10.1016/j.agsy.2020.102936.
- 593 Smith, J.; Sones, K.; Grace, D.; MacMillan, S.; Tarawali, S.; Herrero, M. (2013): Beyond milk,
594 meat, and eggs: Role of livestock in food and nutrition security. *Animal Frontiers* 3 (1), pp. 6–13.
595 DOI: 10.2527/af.2013-0002.
- 596 Uwizeye, A.; Boer, I. J. M. de; Opio, C. I.; Schulte, R. P. O.; Falcucci, A.; Tempio, G. et al.
597 (2020): Nitrogen emissions along global livestock supply chains. *Nature Food* 1 (7), pp. 437–
598 446. DOI: 10.1038/s43016-020-0113-y.
- 599 Veysset, P.; Lherm, M.; Bébin, D. 2010. Energy consumption, greenhouse gas emissions and
600 economic performance assessments in French Charolais suckler cattle farms: Model-based
601 analysis and forecasts. *Agric. Syst.* 103, pp. 41–50. DOI: 10.1016/j.agsy.2009.08.005.
- 602 Vries, M. de; van Middelaar, C. E.; Boer, I.J.M. de (2015): Comparing environmental impacts of
603 beef production systems: A review of life cycle assessments. *Livest. Sci.* 178, pp. 279–288. DOI:
604 10.1016/j.livsci.2015.06.020.
- 605 Wernet, G.; Bauer, C.; Steubing, B.; Reinhard, J.; Moreno-Ruiz, E.; Weidema, B. (2016): The
606 ecoinvent database version 3 (part I): overview and methodology. *Int. J. Life Cycle. Ass.* 21 (9),
607 S. 1218-1230. DOI: 10.1007/s11367-016-1087-8.
- 608 Wiedemann, S.; McGahan, E.; Murphy, C.; Yan, M.-J.; Henry, B.; Thoma, G.; Ledgard, S.
609 (2015): Environmental impacts and resource use of Australian beef and lamb exported to the
610 USA determined using life cycle assessment. *J. Clean. Prod.* 94, pp. 67–75. DOI:
611 10.1016/j.jclepro.2015.01.073.
- 612 Wilfart, A., Gac, A., Salaün, Y., Aubin, J., Espagnol, S. (2021): Allocation in the LCA of meat
613 products: is agreement possible? *Clean. Env. Syst.* 2, p. 100028. DOI:
614 10.1016/j.cesys.2021.100028

- 615 Wilkinson, J. M. (2011): Re-defining efficiency of feed use by livestock. In *Animal* 5 (7),
616 pp. 1014–1022. DOI: 10.1017/S175173111100005X.
- 617 Wilson, N.; Cleghorn, C. L.; Cobiac, L. J.; Mizdrak, A.; Nghiem, N. (2019): Achieving Healthy
618 and Sustainable Diets: A Review of the Results of Recent Mathematical Optimization Studies.
619 *Adv. Nutr.* 10 Issue Supplement 4, pp. 389-S403. DOI: 10.1093/advances/nmz037.
- 620 Zamagni, A. (2012): Life cycle sustainability assessment. *Int. J. Life Cycle. Ass.* 17 (4), pp. 373–
621 376. DOI: 10.1007/s11367-012-0389-8.

Highlights (max. 85 characters per point total 5 points)

- 3 EU beef production systems are assessed with Life Cycle Sustainability Assessment
- The FarmDyn model allows consideration of price effects and farmers' behaviour
- Dairy-bull fattening shows better results in environmental indicators
- Socioeconomic indicators favoured the use of grasslands and by-products in feeding
- Results are sensitive to prices, yields and the animals' performance

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Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

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