Stream diatom biodiversity in islands and continents-A global perspective on effects of area, isolation and environment

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Short running title: Diatom in islands and continents

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67 **Abstract**

68 **Aim.** The species-area relationship (SAR) is one of the most distinctive biogeographic
69 patterns, but global comparisons of the SARs between island and mainland are lacking
70 for microbial taxa. Here, we explore whether the form of the SAR and the drivers of
71 species richness, including area, environmental heterogeneity, climate and
72 physicochemistry, differ between islands and similarly sized areas on mainland, referred
73 to as continental area equivalents (CAEs).

74 **Location.** Global.

75 **Major taxa studied.** Stream benthic diatoms.

76 **Methods.** We generated CAEs on six continental datasets and examined the SARs of
77 CAEs and islands (ISAR). Then, we compared CAEs and islands in terms of total
78 richness and richness of different ecological guilds. We tested the factors contributing to
79 richness in islands and CAEs with regressions. We used structural equation models to
80 determine the effects of area vs. environmental heterogeneity, climate and local
81 conditions on species richness.

82 **Results.** We found a non-significant ISAR, but a significant positive SAR in CAEs.
83 Richness in islands was related to productivity. Richness in CAEs was mainly dependent
84 on area and climate, but not directly on environmental heterogeneity. Species richness
85 within guilds exhibited inconsistent relationships with island isolation and area.

86 **Main conclusions.** Ecological and evolutionary processes shaping diatom island
87 biogeography do not depend on area at the worldwide scale probably due to the presence
88 of distinct species pool across islands. Conversely, area was an important driver of

diatom richness in continents, and this effect could be attributed to dispersal. Continents had greater richness than islands, but this was a consequence of differences in environmental conditions such as specific island climatic conditions. We stress the need for more island data on benthic diatoms, particularly from archipelagos, to better understand the biogeography of this most speciose group of algae.

Keywords

ecological guilds, freshwater diatoms, island biogeography, macroecology, species-area relationship, streams

Main Text

Introduction

A fundamental ecological law that describes how the number of species increases with area is the species-area relationship (SAR, Arrhenius, 1921). The SAR belongs to a few, truly robust generalizations in ecology detected in a wide range of ecosystems and taxa (Connor & McCoy, 1979; Rosenzweig, 1995; Drakare et al., 2006). Islands represent perhaps the most straightforward study setting to explore the SAR because of their well-defined area. Unlike most mainland habitat patches, islands are surrounded by an inhospitable matrix for continental taxa, which cannot be colonized and, consequently, cannot serve as a source of immigrants. This peculiar feature of islands inspired MacArthur & Wilson to develop the theory of island biogeography (MacArthur & Wilson, 1967), which has contributed enormously to modern biodiversity theory (Chase & Leibold, 2003), metapopulation biology (Hanski & Gaggiotti, 2004), community ecology (Mittelbach & McGill, 2019), landscape ecology (Farina, 2008) and biodiversity conservation (Prugh et al., 2008).

Island biogeography investigates how species richness on islands varies spatially and through time (Whittaker & Fernandez-Palacios, 2007). It postulates that larger and less isolated islands host more species than small and remote islands because larger area decreases extinction and proximity to mainland increases immigration. Larger islands may also encompass more species because they provide a larger target for immigration, higher habitat diversity (Lack, 1976) and have higher speciation rates (Whittaker & Fernandez-Palacios, 2007). Lastly, since island age affects diversification and erosion, it

120 may also determine species richness, which tends to be the highest in islands of
121 intermediate age according to the general dynamic model (Whittaker et al., 2008, 2017).

122 In the light of this knowledge, Chase et al. (2019) recently presented a framework for the
123 ecological mechanisms underlying the island SAR (ISAR). They suggested that passive
124 sampling (i.e. larger islands passively sample more individuals and species from the
125 regional pool than smaller islands), disproportionate effects (e.g. different colonization
126 and extinction rates in larger vs. smaller islands) and habitat heterogeneity (greater in
127 large islands) would be the main drivers of ISAR. Nevertheless, the major patterns and
128 drivers of island vs. mainland SAR are still poorly understood, particularly for the species
129 rich microorganisms.

130 Given the importance of environmental heterogeneity and dispersal on the SAR (Chase et
131 al., 2019), functional groups varying in resource utilization and dispersal can have
132 different SARs (Lomolino & Weiser, 2001; Báldi, 2008; Schrader et al., 2020). For
133 example, the SAR slope was steeper for specialist than for generalist bird species
134 (Matthews et al., 2014) and functional traits related to dispersal explained the SAR
135 variation in plant communities (Schrader et al., 2020). Thus, evaluating the SAR of
136 different ecological guilds may improve the knowledge of the niche- vs. dispersal-related
137 processes behind the SAR patterns. As functional diversity may have a distinct
138 (Jamoneau et al., 2018; Schrader et al., 2020) and even stronger response to
139 environmental variation than species diversity (Krause et al., 2014; Abonyi et al., 2018),
140 the SAR for different functional groups may elucidate how community assembly
141 processes operate through space and time (Tilman et al., 1997).

142 The ISAR has been tested with larger-bodied organisms, including terrestrial arthropods
143 (Simberloff & Wilson, 1969) and reptiles (Algar & Losos, 2011), birds (Kalmar &
144 Currie, 2006, 2007), vascular plants (Kreft et al., 2008), and fish (Sandin et al., 2008).
145 However, ISAR patterns are still poorly understood for microorganisms. Earlier
146 microbial field studies that used microcosms (Smith et al., 2005), lakes (Reche et al.,
147 2005), trees (Bell et al., 2005; Peay et al., 2007) or spring ecosystems (Teittinen &
148 Soininen, 2015) as surrogates of islands, reported significantly positive ISARs in almost
149 all systems (but see Teittinen & Soininen, 2015). However, investigations on
150 microorganismal diversity in real islands at a global scale are, to our knowledge, still
151 missing.

152 Rosenzweig (1995) hypothesized that islands should have lower local and regional
153 species richness than similarly sized continental regions due to isolation (lower mass- and
154 rescue effect), but steeper SAR slopes. This is because area tends to be a more critical
155 factor for biota on islands than on continents due to its stronger effects on extinction and
156 colonization (Kreft et al., 2008). However, in an extensive meta-analysis, Drakare et al.
157 (2006) did not find evidence for steeper SARs on islands (ISARs) than on mainland
158 across multiple species groups. The SAR patterns are typically explored within
159 archipelagos due to the presence of a common species pool, allowing assessment of the
160 pure area effect. However, there are also more general models for the SAR at the global
161 scale, searching for broader influences on the SAR (Kalmar & Currie, 2006; Kreft et al.,
162 2008; Triantis et al., 2015), including differences in evolutionary history (Rosenzweig,
163 1995).

164

165 Here, we adopted a similar perspective and investigated freshwater diatom SAR at a
166 worldwide scale, given that diatoms have large distributions (Finlay, 2002) and are
167 strongly controlled by environmental conditions (Soininen et al., 2016). We compared
168 SARs, total species richness, and species richness drivers between islands and
169 corresponding areas on five continents, referred to as continental area equivalents
170 (CAEs). For this comparison, we devised a novel method based on island-mainland pairs
171 (Fig. 1), assuming that terrestrial area is a good surrogate for area of freshwater habitat
172 (see Appendix S1 in Supporting Information). The CAEs corresponded to the sampling
173 area of 18 islands. We then examined (1) if SAR slopes differed between islands (ISAR)
174 and continents, (2) whether islands showed overall lower diatom species richness than
175 CAEs, (3) if species richness of island was related to environment, spatial isolation or
176 island age, and (4) whether habitat diversity, passive sampling or disproportionate effects
177 explained the SAR. We investigated these research questions separately for total diatom
178 species richness and species richness of ecological guilds, differing in dispersal capacity
179 and tolerance to nutrient limitation and disturbance (Passy, 2007, 2016), all expected to
180 influence the SAR (Matthews et al., 2014; Schrader et al., 2020).

181 **Materials and Methods**

182 *Biological and environmental datasets*

183 In total, we included 18 island datasets (Corsica, Cyprus, Guadeloupe, Iceland, Ireland,
184 Kauai, La Réunion, Martinique, Madeira, Majorca, Mayotte, New Caledonia, North New
185 Zealand, Oahu, Possession, São Miguel, Sardinia and South New Zealand) and six
186 continental datasets (China, Finland, France, French Guiana, Kenya and USA) in our
187 study (see Appendix S2). Diatoms were sampled from hard substrates (typically stones)

188 or macrophytes, generally during the low flow period (see Appendix S2 for details).
189 Although diatoms in some datasets were collected over several years, we did not expect a
190 substantial effect of interannual variation in our study, because we were interested in
191 regional diversity patterns and included environmental variables to account for this
192 potential variation.

193 Diatoms were cleaned with acid or hydrogen peroxide. A total of 400-700 diatom valves
194 were counted for each sampling site, which is sufficient for reliable estimates of total
195 diversity (Heino & Soininen, 2005). As the number of counted valves varied somewhat
196 among the samples, we studied if this would affect our richness estimates. We estimated
197 species richness with 300 valves and tested the correlation with the observed species
198 richness. We observed a very strong relationship between the estimated and the observed
199 species richness (R^2_{aj} = 0.98). Also, valve counts did not differ significantly between
200 islands and continents (Cliff test difference for large dataset, delta=-0.15). We thus
201 believe that the number of counted valves has only marginal impact on our richness
202 results.

203 Diatoms were generally identified up to species level, except in some rare case where
204 some of the valves were identified only to genus level (representing less than 5% of the
205 entire dataset). Homogenization of the taxonomy among regions was performed using the
206 OMNIDIA database (Lecointe et al., 1993, updated in November 2020). To ensure that
207 we have a proper estimate of the diversity, we i) evaluated the proportion of observed
208 species compared to the size of the species-pool in each region using basic Chao equation
209 (Chao, 1987) and calculated a 'corrected' species richness according to this ratio (i.e. the
210 observed species richness was increased relative to the proportion of missing species

211 estimated from the species pool) and ii) calculated a genus-based richness assuming that
212 genus level identification varies much less among diatomists than species identification.
213 We then ran analyses with observed species richness, corrected species richness and
214 genus richness (see Data analyses section). In total, our datasets comprised 1967 taxa,
215 further classified into four ecological guilds: low profile (species of short stature), high
216 profile (species of tall stature, typically filamentous, colonial or branched), motile
217 (species moving freely in the biofilm) and planktonic species (species not innate to the
218 benthos but originating from planktonic sedimentation) (Passy, 2007; Rimet & Bouchez,
219 2012; Soininen et al., 2016). Contrary to motile and high-profile species, low-profile
220 species are tolerant to nutrient limitation and disturbance and exhibit wider distributions
221 (Passy, 2016), suggesting potentially higher dispersal capabilities (Heino & Soininen,
222 2006). Planktonic species may indicate important features of the sites such as low current
223 velocity and large rivers.

224 Physico-chemical data of each sampling site included pH, conductivity ($\mu S.cm^{-1}$), total
225 phosphorus ($mg.l^{-1}$) and water temperature (°C), with the exception of Finland and
226 Possession island (with no water temperature data) and Ireland, Kenya and New Zealand
227 (with no total phosphorus data). Physico-chemical data were collected up to two months
228 before the diatom sampling. Climate data were obtained from WorldClim database at 0.5
229 minutes resolution (Hijmans et al., 2005), including annual precipitation (mm),
230 seasonality in precipitation (%), annual temperature (°C), and temperature seasonality
231 (standard deviation of monthly mean temperatures). For each sampling site, we also
232 extracted elevation from the Global Multi-resolution Terrain Elevation Data 2010
233 (Danielson & Gesch, 2011) and computed terrain slope as a proxy for current velocity.

For islands, we determined age of formation from the literature (see Appendix S2) and isolation using the isolation index of Dahl (Dahl, 1991, Gillespie et al., 2008). This index (equation 1) is based on the sum of square root distances to the nearest equivalent or larger island (d_i), the nearest island group or archipelago (d_a), and the nearest continent (d_c).

$$\textit{Isolation index} = \sqrt{(d_i)} + \sqrt{(d_a)} + \sqrt{(d_c)} \quad (1)$$

Creation of continental area equivalents (CAEs)

For a reliable comparison of species-area relationships between islands and continents, which are vastly different in size, we generated CAEs, comparable in size to the islands by taking subsets of the continental data (see Algar & Losos, 2011) for a related approach). The method used to create these CAEs (Fig. 1) was as follows.

We first computed the geographical centroid of each island and calculated $D_{c\text{-}i}$, a vector representing the Euclidean distance between the centroid and each island sample site *i*. Second, for each continent, we calculated D_{jj} the Euclidean distance matrix between sample sites *j*. All Euclidean distances were calculated from geographical coordinates expressed in a projected geographical system adapted for each region (see Appendix S2). Third, we treated all continental sites as candidate CAE centroids and calculated $D_{v\text{-}j}$ representing the Euclidean distance between the candidate CAE centroid *v* and all other *j* continental sites. We then computed a matrix $DD_{v\text{-}j,c\text{-}j}$ (equation 2), which represented the absolute difference between i) the distance between a candidate CAE centroid and all other sites in the focal continent ($D_{v\text{-}j}$, i.e. raw of the matrix D_{jj}) and ii) the distance between the island centroid and all other sites in the focal island ($D_{c\text{-}i}$).

$$DD_{v\text{-}j,c\text{-}i} = |D_{v\text{-}j} - D_{c\text{-}i}| \quad (2)$$

Note that the minimum value of $DD_{v\text{-}j,c\text{-}i}$ is theoretically 0, indicating that the distance between a centroid and an island site *i* is identical to the distance between a CAE centroid and a continent site *j*. Thus, smaller $DD_{v\text{-}j,c\text{-}i}$ equates to similar distances between an island centroid and island sites and the distances between a candidate CAE centroid and continental sites. We then assigned for each centroid-island site distance a unique corresponding CAE centroid-continent site distance ($\Delta_{c\text{-}i,v\text{-}j}$, i.e. the minimum value of the column of $DD_{v\text{-}j,c\text{-}i}$, equation 3).

$$\Delta_{c\text{-}i,v\text{-}j} = min(DD_{.,c\text{-}i}) \quad (3)$$

Then, we considered that the CAE centroid could be considered as the centroid of a CAE only if at least N=15 of the selected CAE centroid-continent site distances $\Delta_{c\text{-}i,v\text{-}j}$ were below a threshold value (θ) set to 5 km. Thus, theoretically, the number of sites in each CAE could vary between 15 and the total number of sites in each island. Note that because Kauai and Possession islands have less than 15 sites, N was set to 10 for the creation of their CAEs. Finally, to avoid pseudoreplication within the sites of CAEs, we selected for each continent-island pair only CAE separated by a distance of at least twice the mean distance between centroids and their corresponding sites. Due to this procedure, the size of CAEs could be, in some rare case, much smaller than the corresponding island size.

The CAEs, corresponding to the sampling area of an island, were successfully created in all continents (see appendix S3). Kenya was an exception because we were unable to create CAEs corresponding to Corsica, Iceland, Ireland, New Caledonia, Sardinia, North

and South New Zealand, which were larger in size. Also, following our methodology, it was not possible to create CAEs corresponding to the island of Mayotte in USA and Finland, the island of São Miguel in Finland and USA and Possession Island in China and USA because their continental sites were more spread out than the island sites. Consequently, the total number of continent-island pairs for creating CAEs was 96.

Randomization procedure for calculation of species richness and other environmental variables

For each continent-island pair, we randomly selected 15 sites within the CAEs and 15 sites within each island (10 sites for Kauai and Possession and their respective CAEs; 20 iterations) to achieve comparable sampling effort for islands and CAEs. For each random subset, we calculated species richness as the total number of species observed among the 15 sites, and area from the convex hull around these 15 sites. We found that the areas in islands estimated using convex hulls were good surrogates for whole island areas (see appendix S4). We also calculated median values for each environmental variable for the 15 sites in the CAEs and islands and computed their environmental heterogeneity as the multivariate dispersion of all environmental variables using the average distance of all samples to the sample centroid in the multivariate space with the *betadisper* function in the *vegan* package. Environmental variables used in the analyses and computation of environmental heterogeneity were selected because they are known to be important for stream diatom distributions (Soininen, 2007; Soininen et al., 2016). For the computation of environmental heterogeneity in the Kenya and Ireland dataset, we respectively used total nitrogen and orthophosphate concentrations instead due to the lack of total

phosphorus data (none of the nutrient concentration was used in the computation of heterogeneity for New Zealand and Possession islands due to missing data).

Data analyses

We conducted separate analyses for CAEs and islands to examine the relationship between species richness and area (SAR). We used linear mixed models (LMMs) for CAEs to account for continental influences that may underlie differences in species pools and the potential lack of independence among CAEs, given that multiple CAEs were created within a continent (i.e. continents were included as random factors). We performed traditional linear models for islands. We tested SAR with three commonly used models (DeMalach et al., 2019), including power (Arrhenius model), logarithmic (Gleason model) and Michaelis-Menten, and selected the best model based on the lowest Akaike Information Criterion. We also tested relationships between area and the 'corrected' species richness (according to the size of the species pool) and genus richness to ensure that the sampling effort or the taxonomic resolution did not influence our results.

To test for passive sampling, we estimated species richness from rarefaction curves based on species occurrence. For each CAE and island, we pulled at random 15 sites and randomly selected 130 species occurrences without replacement, thus ensuring that the maximum occurrence of each species did not exceed 15. Species richness was then estimated from the 130 occurrences and used to generate the SAR, which was fit with mixed models for CAEs and traditional linear models for islands. According to Chase et al. (2019), failure to detect SAR using this estimation of species richness would suggest that SAR is caused by passive sampling only. However, the reverse is not true, and

significant SAR observed with this estimation of species richness does not necessarily prove the absence of passive sampling (Chase et al., 2019).

Then, to test for the effect of area on species richness after controlling for environmental variation, we first computed global LMMs for total and guild species richness and eight environmental variables (pH, conductivity, elevation, annual temperature, annual precipitation, temperature seasonality, precipitation seasonality and environmental heterogeneity). Models were constructed using the median values of species richness as the response variable and median environmental variables obtained from the subsampling procedure as explanatory variables (N = 851, i.e. one value for each 833 CAEs and each 18 island). Prior to analyses, explanatory variables were log-transformed to improve normality when necessary and standardized, but we did not treat for multicollinearity here, as this does not affect the fit of the model. Second, residuals from these regressions were regressed against log-transformed area with LMMs for CAEs and simple linear models for islands.

The number of islands in our study is comparable to the number of islands in many other SAR studies (see data used in Matthews et al., 2019) but admittedly not very high (N=18) for a study at the worldwide scale (Kalmar & Currie, 2006). Therefore, we performed a sensitivity test with our continental datasets to determine the number of CAEs required for observing a significant SAR, acknowledging that the number of islands and CAEs necessary to detect a SAR may be different. We used the median values of the species richness and area obtained from the randomization procedure for each continent-island pair (N=96), and randomly sampled (1000 times) K continent-island pairs. Each time we fit the SAR with the best SAR model (logarithmic) and extracted the probability (*P*) of

observing a significant SAR, as well as the median values of model coefficients. We varied K from 11 to 96, i.e. the total number of continent-island pairs available in our dataset. We then identified the minimum number of 'islands' needed to observe a significant SAR with our data (P>95%). We performed these analyses with both traditional linear models and LMMs (e.g. assuming a common species pool).

We compared species richness of islands and species richness of their corresponding CAEs with Cliff's non-parametric effect size statistic (Romano et al., 2006; Tecchio et al., 2016), due to the large number of data points resulting from the randomization procedure (i.e., decreasing variance around the mean). We also used Cliff's tests to compare the species richness of each ecological guild between CAEs and islands.

To compare species richness of CAEs and islands after removing the effect of environment, we computed LMMs as above but also included all the values of random subsampling (x20, N=17020). We therefore used a nested design in the random factors of the models, so that subsampling values are nested within each continent/island. Residual richness values were then extracted from the models and compared between CAEs and islands with Cliff's tests.

We used linear mixed models for CAEs and traditional linear models for islands to examine the relationship between species richness, environmental heterogeneity, the median of all environmental variables and the median values of latitude and longitude. Environmental explanatory variables were log-transformed to improve normality when necessary and we run separate regression models with each environmental factor and species richness to avoid multicollinearity. We also tested for non-linear relationships separately with all environmental variables with the same procedure.

Finally, to disentangle the possible drivers of the SAR for continents, we implemented piecewise structural equation modeling (SEM, Lefcheck, 2016) using linear mixed models with continental dataset as a random factor. We could not implement such models for islands due to an insufficient number of data points. We assumed an *a priori* model (Fig. 2) predicting species richness as directly influenced by area, environmental heterogeneity (as defined above), local environmental conditions and climatic conditions. We used conductivity and elevation as predictors of local conditions, temperature seasonality and annual precipitations as predictors of climate, as they were significant predictors of species richness in global LMMs and exhibited low collinearity in pairwise correlations tests (see Appendix S5). We assumed that the effect of area on species richness could also be indirect through environmental heterogeneity, according to the habitat diversity hypothesis (Lack, 1976). Finally, we also assumed that temperature seasonality and precipitation are directly influenced by elevation. We included a correlation between temperature seasonality and precipitation as well as between conductivity and precipitation (see Appendix S5). We used the Fisher's *C* statistic to test the consistency of the theoretical model with the data. All analyses were run for total richness and separately for richness of each ecological guild.

All analyses were conducted with R (R Core Team, 2019) using packages 'vegan' (Oksanen et al., 2019), 'spatstat' (Baddeley et al., 2015), 'raster' (Hijmans, 2019), 'sf' (Pebesma, 2018), 'lmerTest' (Kuznetsova et al., 2017), 'lme4' (Bates et al., 2015), 'effsize' (Torchiano, 2020) and 'piecewiseSEM' (Lefcheck, 2016).

Results

391 *SAR patterns*

392 We found a significant positive SAR for total species richness in CAEs, but not in islands
393 (Fig. 3a). The best model describing the SAR in CAEs was the logarithmic model (see
394 Appendix S6). The observed R^2 values were relatively low compared to values usually
395 observed for islands but comparable to those found in continental areas (Kreft et al.,
396 2008). Similar results emerged with rarefied richness (see Appendix S7), 'corrected'
397 species richness given the size of the species pool (see Appendix S8), genus richness (see
398 Appendix S9) and also after removing the effect of environmental variation (see
399 Appendix S10).

400 The sensitivity analysis revealed that a minimum of 52 continent-islands pairs is needed
401 to observe a significant SAR with our data. This number dropped to 16 when using mixed
402 models with continent (a surrogate for the species pool) as a random effect (see Appendix
403 S11).

404 About half of the 1967 identified species belonged to the motile guild (see Appendix
405 S12), followed by the high profile and low-profile guilds. Planktonic species and species
406 with variable guilds represented a minor part of the communities. Species richness within
407 all guilds was significantly and positively related to area in CAEs (Fig. 3b-e) and this
408 relationship persisted for all but the high-profile guild after controlling for the
409 environment (see Appendix S10).

410 *Comparison of species richness of islands and CAEs*

411 Overall, species richness was significantly lower in the islands than in the respective
412 CAEs for more than 50% of all continent-island pairs ($N = 96$) (Fig. 5a). Similar results

413 emerged for the guilds, especially for the planktonic guild, where over 70% of the
414 comparisons had significantly higher species richness in CAEs. The only exception was
415 the low-profile guild whose species richness tended to be higher in islands (ca. 60%).
416 Importantly, however, when environmental variation was accounted for, the species
417 richness differences between CAEs and islands disappeared in more than 80% of cases
418 (Fig. 5b).

419 *Ecological variables driving species richness*

420 In islands, we found significant relationships between species richness and isolation for
421 total, low profile species richness (U-shaped pattern) and planktonic guild (negative
422 linear pattern and a weak non-linear pattern) (Fig. 4, see Appendix S13). There was no
423 relationship between richness and age of island for the total community or any of the
424 ecological guilds (see Appendix S13). Apart from isolation, total species richness in
425 islands was significantly related only to phosphorus concentration. Species richness of
426 guilds was also significantly related to some other environmental variables depending on
427 the guild considered (see Appendix S13).

428 Total species richness in CAEs was significantly related to environmental heterogeneity,
429 pH, conductivity, phosphorus concentration, all climatic variables and longitude (see
430 Appendix S13). The piecewise SEM models (Lefcheck, 2016) disentangled the effects of
431 the influencing factors and demonstrated that diatom species richness in CAEs was
432 related to area, habitat heterogeneity, physicochemistry, elevation, and climate. The data
433 fitted well the *a priori* model (Fig. 6) for total species richness and species richness of all
434 ecological guilds. The marginal R^2 (variance explained by the fixed effects only) for total
435 species richness was 0.71 and varied between 0.14 (for low-profile species) to 0.59 (for

436 motile species). In the SEMs, area explained species richness independently, without any
437 indirect effect through environmental heterogeneity, except for the motile species
438 richness where the effect of environmental heterogeneity was negative. Indeed, total
439 species richness and richness of the motile guild were mainly driven by area and
440 temperature seasonality (Fig. 6a, d). Low-profile species richness was only explained by
441 area and precipitation (Fig. 6b). High-profile species richness was mainly explained by
442 area, elevation and conductivity, while climate had no direct effect (Fig. 6c). Finally,
443 planktonic species richness was solely determined by elevation and was thus the only
444 group without a significant relationship with area.

445 **Discussion**

446 Here, we conducted the first comparative analysis of island vs. mainland species-area
447 relationship for microbes, providing insight into the roles of area, environmental
448 heterogeneity, isolation and island age on species richness patterns. We showed for
449 freshwater diatoms that: (i) there was a significant SAR in continents but not in islands
450 (except for high profile), (ii) regional species richness was higher in continents than in
451 islands, but this difference was explained entirely by environmental conditions (iii) the
452 effect of isolation varied among diatom guilds and (iv) area and median environmental
453 conditions but not environmental heterogeneity were significant predictors of diatom
454 richness. Next, we will discuss the main findings in more detail and highlight our major
455 conclusions about total community and guild richness.

456 *Drivers of species richness in islands*

457 The lack of a significant SAR in islands may be due to low sample size (N = 18) or may
458 represent a real biogeographical pattern. Sensitivity analyses performed for CAEs

459 revealed that 16 islands are needed to detect a SAR given a common species pool. This
460 result is consistent with numerous studies on other organisms, reporting ISAR for a
461 relatively small number of islands within archipelagos (Matthews et al., 2019). However,
462 at a global scale, a much higher number of islands (N=52, Appendix S11) may be
463 required for detection of diatom ISAR.

464 The absence of diatom ISAR may have evolutionary and ecological causes. First, diatoms
465 may have distinct species pools across the globe (Soininen et al., 2016) and differences in
466 island area may not be sufficient to predict richness on islands that differ greatly in
467 species pool. As the size of the species pool influences the shape of the SAR (Catano et
468 al., 2021), future analyses on archipelagos will be essential for determining whether
469 ISAR exists for diatoms (but see Jüttner et al., 2018). Second, environmental
470 heterogeneity, which increased with island size (Fig. 3, and see Appendix S5) and is
471 recognized as an important driver of SAR (Lack, 1976; Chase et al., 2019), had no direct
472 impact on island species richness. Third, island richness was related only to total
473 phosphorus, suggesting that productivity is a key factor explaining island diatom species
474 richness at this scale. Note however, that due to data availability, only phosphorus
475 concentration was considered as a resource factor for explaining species richness. The
476 consideration of other nutrient resources, known to influence diatom diversity (e.g.
477 nitrogen, iron, Passy, 2007, Soininen, 2007), may improve the understanding of diatoms
478 species richness in islands.

479 We found that isolation might have some effect on species richness in islands. Two of the
480 most isolated islands (Oahu and New Caledonia) actually showed high species richness,
481 resulting in a U-shaped relationship between species richness and isolation for total and

low-profile species richness. Oahu and New Caledonia still had the highest species richness when the latter is corrected by species pool but the U-shaped relationship is only marginally significant ($p = 0.09$, see Appendix S8). Greater speciation in the most isolated islands, which have many endemic species, e.g. New Caledonia has been dubbed "Galapagos of diatoms" (Moser et al., 1998), may explain their higher richness considering that endemic and total species richness are typically correlated (Kallimanis et al., 2010). We could, however, not exclude the fact that some other unmeasured environmental factors, particularly related to islands conditions, may also be responsible for this pattern. Finally, our finding further suggests that the biogeographical drivers of diatom richness on real islands are trait dependent.

Following Rosenzweig (1995), we hypothesized that islands would harbour lower species richness than continents due to diminished dispersal and rescue effects. While species richness was indeed lower in islands compared to continents, this difference disappeared when we accounted for environmental differences. Thus, annual precipitation, higher in islands than continents, was associated with lower species richness (see Appendix S10), likely because of its positive effect on current velocity, and subsequently, shear stress (Heino & Soininen, 2007).

Drivers of species richness in continents

We tested whether SARs in continents could result from passive sampling and environmental heterogeneity, which are major drivers of the SAR (Lack, 1976; Rosenzweig, 1995; Stein et al., 2014; Chase et al., 2019). Surprisingly, species richness in continents was not directly explained by environmental heterogeneity in the SEM.

Although area was strongly related to environmental heterogeneity (but poorly related to other environmental variables, see Appendix S5), none of the SEM models showed a direct effect of habitat heterogeneity on either total or guild species richness, except for motile species. For the latter, the direct effect of habitat heterogeneity was negative, contrary to the results observed in univariate regressions (see Appendix 13) due to the strong collinearity between area and heterogeneity. Given that we still observed a significant SAR with species richness estimated from the rarefaction curves, passive sampling cannot be completely ruled out (Chase et al., 2019). However, the impact of area on richness in continents might also be due to disproportionate effects, including dispersal, extinction and speciation. While extinction and speciation have been less studied in diatoms, dispersal and mass effects (whereby species maintain their presence in unfavorable conditions via immigration, Shmida & Wilson, 1985) were shown to have a notable influence on regional to subcontinental diatom communities (Soininen, 2007; Jamoneau et al., 2018; Leboucher et al., 2020). For continental diatoms, larger areas may thus increase the probability of immigration from the surrounding landscape, particularly for species with high dispersal capabilities (mass-effect species), thereby increasing CAE's diversity.

Environmental factors, such as nutrients, climate and elevation, were also important predictors of total and guild species richness. Total species richness decreased with temperature seasonality, as did the species richness of motile species, which represented ca. 50% of the whole community (see Appendix S12). As motile species are generally warm-water species (Pound et al., 2021) and high seasonality occurs in colder areas, it is possible that motile guild richness was limited by unfavorable temperatures. Species

richness of the high profile and planktonic guilds was the lowest at high elevation. For high-profile species, high elevation is stressful due to increased current velocity and probability for dislodgement. For planktonic species, high elevations do not provide sufficient habitat, given that these species require large water bodies. Species richness of the low-profile guild is positively influenced by annual precipitation probably because this guild is tolerant to physical disturbance (Passy, 2007), which should increase its richness in the community.

Conclusions

We examined diatom ISARs and compared them with the SARs of similarly sized continental area equivalent across five continents. Contrary to most previous studies, we did not find significant ISAR for total species richness but detected significant relationships of richness with total phosphorus. These results imply that diatom richness in islands is not related to area but is controlled by productivity. However, the lack of ISAR may be due to distinct species pool across islands in our study. Species richness was typically higher in continental areas than in similarly sized islands, most probably due to differences in climate and related environmental conditions, such as current velocity. The significant SAR for continents may originate from disproportionate effects, such as mass effect, but not from environmental heterogeneity. Isolation influenced the richness of the whole community and some diatom guilds in islands. These finding indicate that there are important differences in richness responses to island properties among ecological guilds and between the community level and the functional level. Finally, the proposed new method for species-area comparisons between islands and continental area equivalents will advance research on biogeography of islands vs.

550 mainland. We advocate obtaining global diatom data, particularly from archipelagos to
551 better understand the drivers of island species diversity.

Figures legend

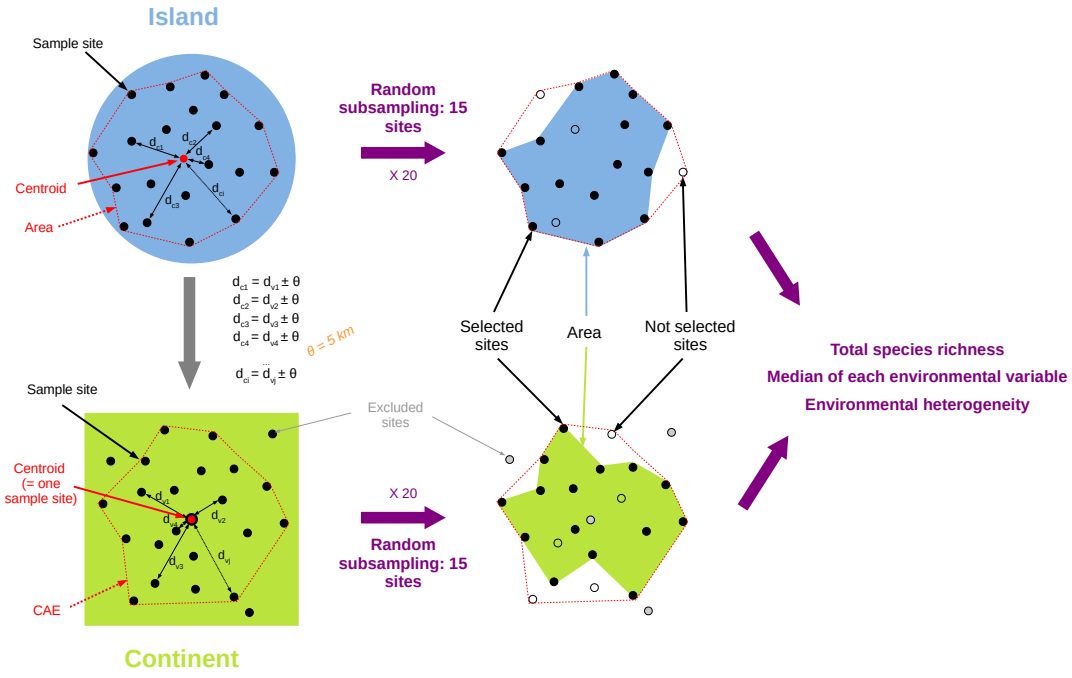
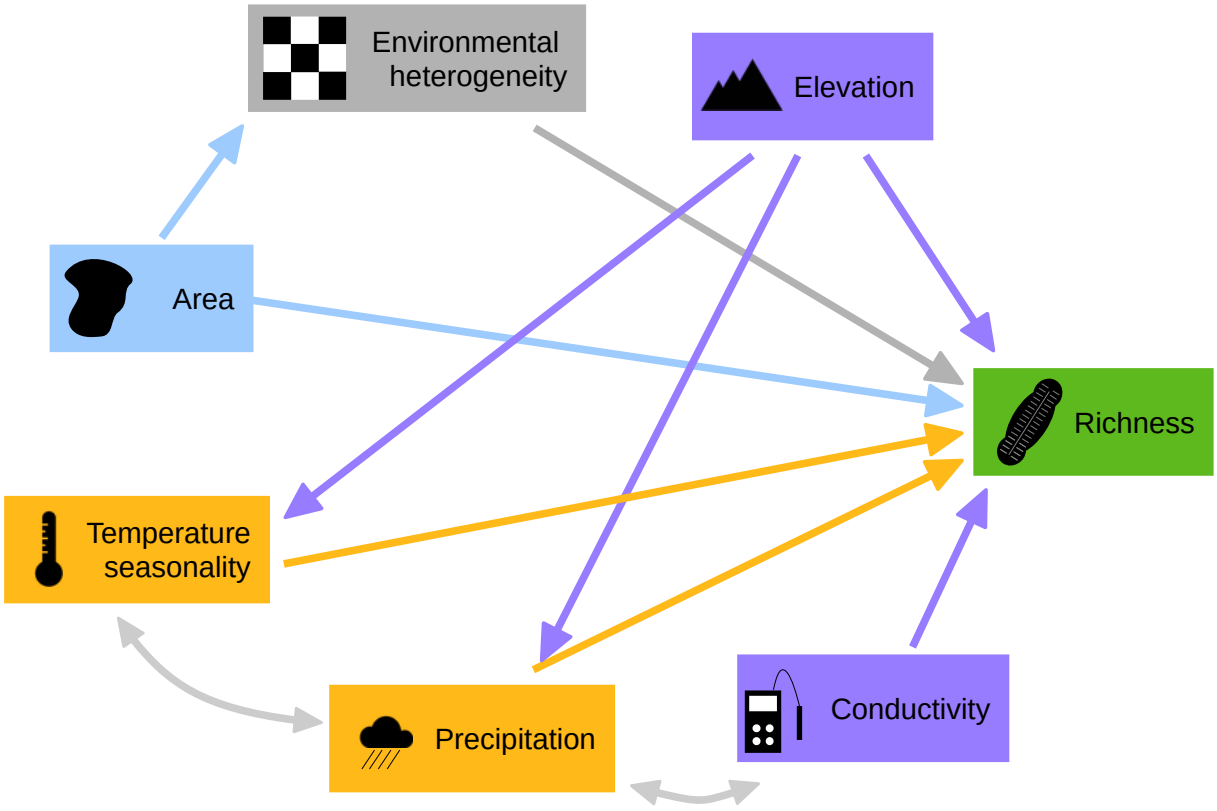
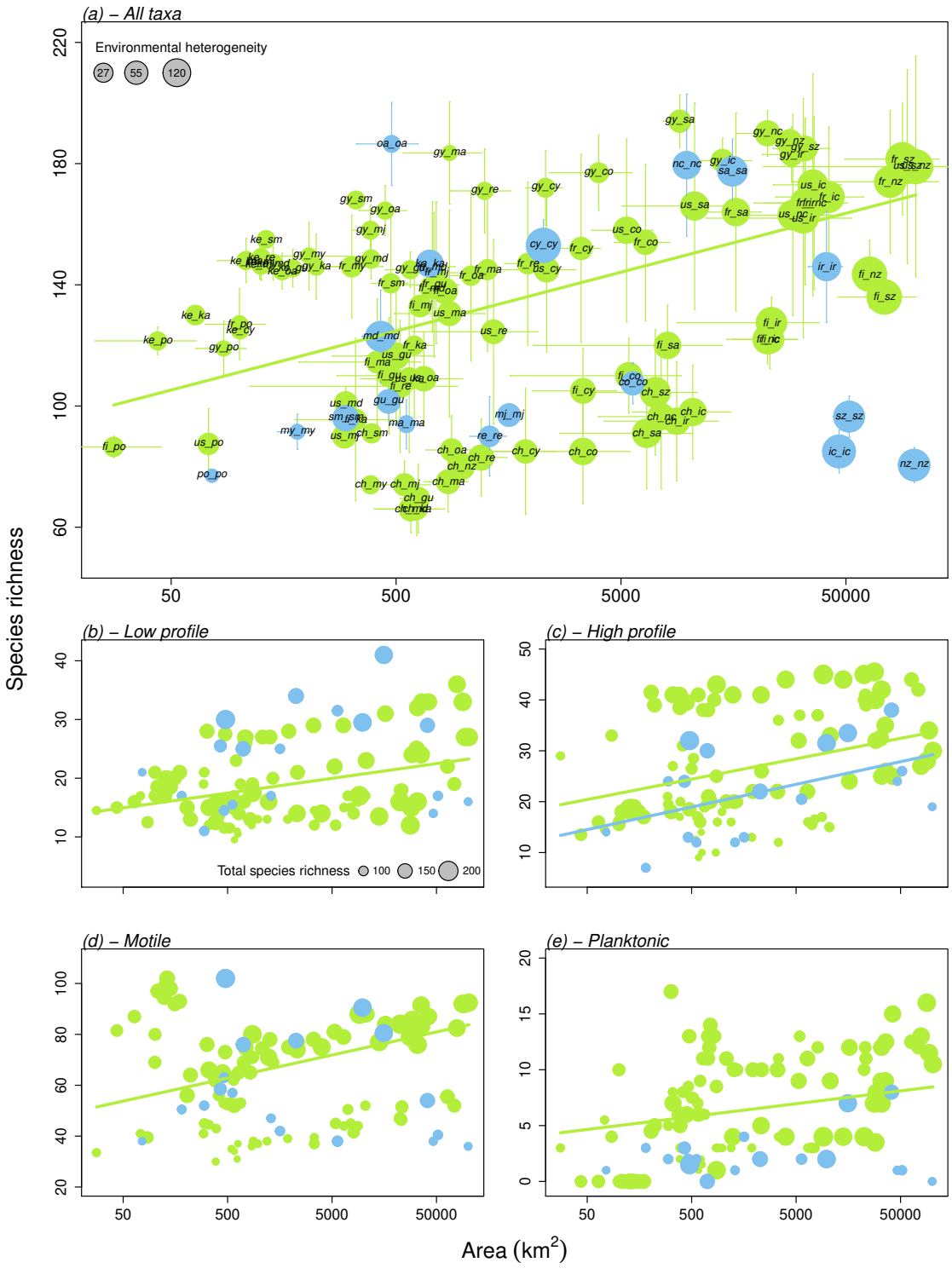


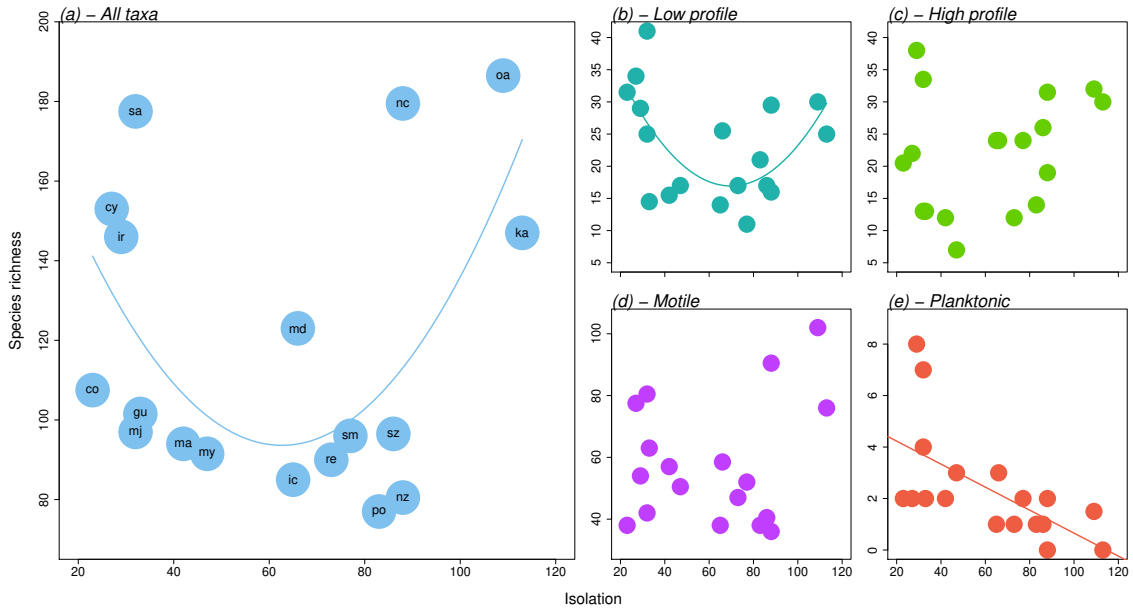
Figure 1: Descriptive diagram of the methods. Diagram describing the methodological process used for creation of continental area equivalents (CAE) and subsampling of both islands and CAEs.



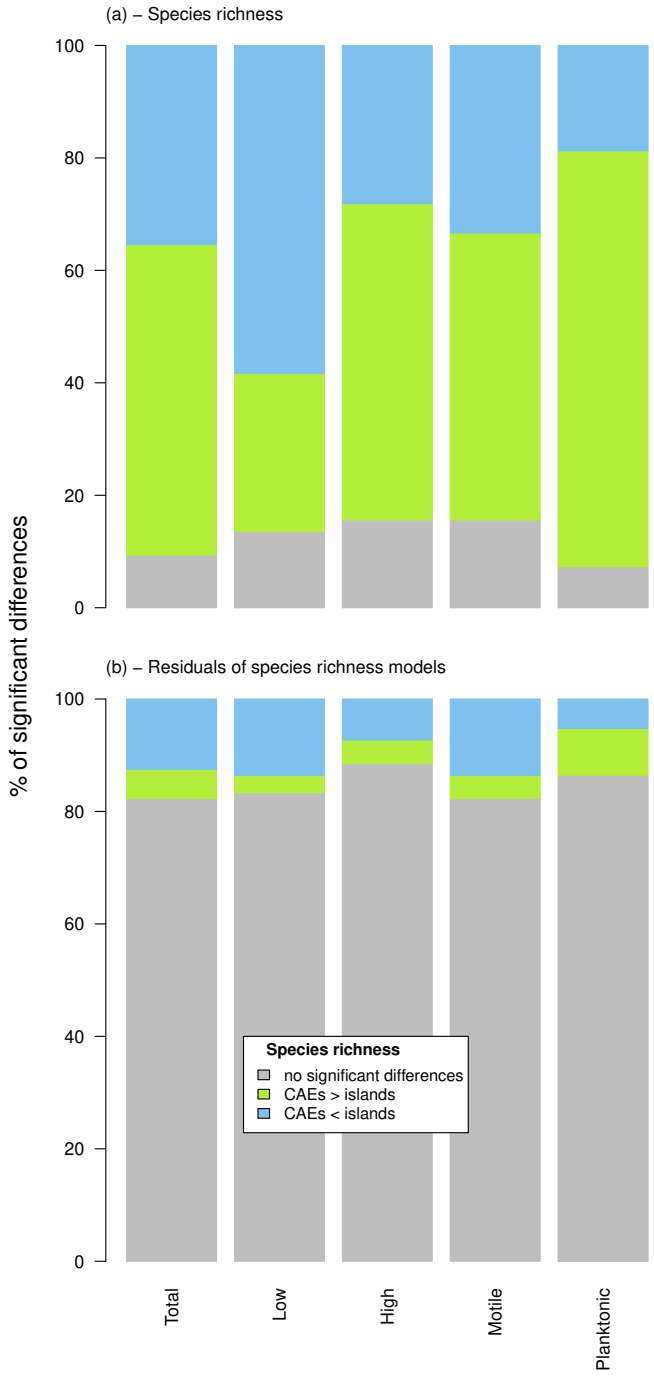
558 **Figure 2. A priori model explaining diatom species richness.** Species richness is
559 modeled as a function of area, environmental heterogeneity, local environmental
560 conditions (purple) and climate (orange).


(a) – All taxa
Environmental heterogeneity
27
55
120
(b) – Low profile
Total species richness
100
150
200
(c) – High profile
(d) – Motile
(e) – Planktonic
Species richness
Area (km2)
50
500
5000
50000

562 **Figure 3. Species-area relationships for continents and islands.** Species-area
563 relationships for continental area equivalents (CAE) (in green, N=96) and islands (in
564 blue, N=18) for total species richness (a), and richness of low profile (b), high profile (c),
565 motile (d) and (e) planktonic species. Green regression lines represent significant linear
566 fits in mixed models for CAEs: richness = 8.44x + 72.24, R^2_m = 0.22 for total species
567 richness, 1.09x + 10.69, R^2_m = 0.13 for low-profile species, 1.74x + 13.61, R^2_m = 0.11 for
568 high-profile species, 3.94x + 38.36, R^2_m = 0.11 for motile species and 0.50x + 2.69, R^2_m =
569 0.05 for planktonic species, where x = log(area). The blue regression line represents
570 significant linear fit for high-profile species of islands: 1.93x + 6.95, R^2_{aj} = 0.23. Dot
571 sizes are proportional to environmental heterogeneity (in log) for all taxa (a) and
572 proportional to total species richness for functional groups (b-e). Error bars represent
573 standard deviation estimated from the subsampling procedure. Text in dots indicate the
574 dataset used for computing species richness and area. For example, 'fr_my' indicates the
575 position of Mayotte CAE in France. Continental datasets are indicated by 'fr' for France,
576 'us' for US, 'fi' for Finland, 'ch' for China, 'ke' for Kenya and 'gy' for French Guiana
577 and islands indicated by 'ic' for Iceland, 'co' for Corsica, 'gu' for Guadeloupe, 'ma' for
578 Martinique, 're' for La Réunion, 'my' for Mayotte, 'nz' for North New Zealand, 'sz' for
579 South New Zealand, 'nc' for New Caledonia, 'ka' for Kauai, 'oa' for Oahu, 'po' for
580 Possession, 'cy' for Cyprus, 'ir' for Ireland, 'md' for Madeira, 'mj' for Majorca, 'sm' for
581 São Miguel and 'sa' for Sardinia.



583 **Figure 4. Relationships between island species richness and isolation.** Relationship
584 between total species richness (a) and species richness of each ecological guild (b-e) with
585 island isolation for islands ($N = 18$). Significant linear and quadratic relationships
586 ($p<0.05$) are shown by regression fits (only the fit with the lower AIC is shown if both
587 are significant, see Appendix S13): $0.03x^2 – 3.79x + 212.32$, $R^2_{aj} = 0.26$ for total richness,
588 $0.01x^2 – 0.96x + 50.44$, $R^2_{aj} = 0.32$ for low-profile and $-0.04x + 5.12$, $R^2_{aj} = 0.34$ for
589 planktonic species. For island names, see Fig. 1. Isolation is based on index defined by
590 Dahl (Dahl, 1991).

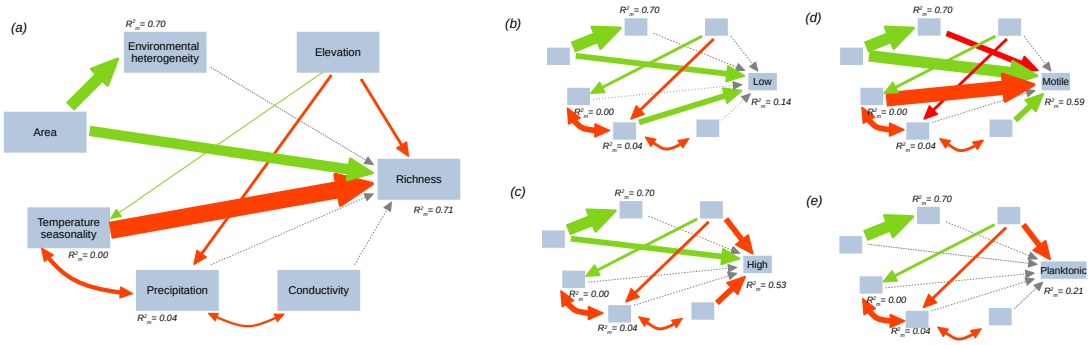


592 **Figure 5. Comparison of species richness between continental area equivalents**

593 **(CAEs) and islands.** Percentage of significant and non-significant tests ($N = 96$

594 continent-island pairs) between CAEs and islands for species richness (a) and species

614 richness residuals (b). Tests were performed for total and guild species richness. Species
615 richness residuals were estimated from linear mixed models with species richness as the
616 dependent variable, and pH, conductivity, elevation, mean annual temperature and
617 precipitation, temperature and precipitation seasonality and environmental heterogeneity
618 as explanatory variables and continent as a random factor. Comparisons of values (i.e.,
619 species richness or residuals of species richness) were performed with Cliff's test,
620 whereby tests with delta >0.33 indicated significant differences (Romano et al., 2006).
621



623 **Figure 6. Structural equation models explaining species richness in continental area**
624 **equivalents (CAEs).** Structural equation models for total species richness (a), low profile
625 (b), high profile (c), motile (d) and planktonic (e) species richness in continents (N = 96
626 CAEs). Green and red arrows represent significant positive and negative relationships,
627 respectively, whereas gray-dashed arrows represent non-significant relationships. Arrow
628 widths are proportional to the standardized regression coefficients and R^2_m values
629 represent marginal R^2 from a linear mixed model. All models fitted well the *a priori*
630 model, i.e. the model including all shown causal relationships (Fisher's C = 14.99, df =
631 14, p = 0.38 for all models).

Data Availability Statement

Data are available under the following link: https://doi.org/10.57745/ZPBSLT

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637 **Biosketch**

638 Aurélien Jamoneau is a researcher in community ecology at INRAE institute of Cestas-
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640 and temporal scales. Janne Soininen in a professor in spatial environmental research at
641 the University of Helsinki. He is interested in large-scale community ecology and
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