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A Marginal Abatement Cost Curve for Greenhouse gases attenuation by additional carbon storage in French agricultural land

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40 **Highlights :**

41 Additional carbon storage, net GHG budget, and cost of 8 agricultural practices

42 MACC shows abatement potential of 39-59 MtCO₂e.yr⁻¹ for carbon price 55-250 €.tCO₂e⁻¹

43 Key practices: agroforestry, hedges, cover crops, grasslands in crop sequences

44 No “one size fits all strategy” due to heterogeneity across regions and practices

45 French agricultural carbon sink is 5 times higher than carbon neutrality target

46

47 **Abstract**

48 Following the Paris agreement at COP21, the European Union (EU) set a carbon neutrality

49 objective by 2050, and so did France. The French agricultural sector can contribute as a carbon

50 sink through carbon storage in biomass and soil, in addition to reducing GHG emissions. The

51 objective of this study is to quantitatively assess the additional storage potential and cost of a

52 set of eight carbon-storing practices. The impacts of these practices on soil organic carbon

53 storage and crop production are assessed at a very fine spatial scale, using crop and grassland

54 models. The associated area base, GHG budget, and implementation costs are assessed and

55 aggregated at the region level. The economic model BANCO uses this information to derive

56 the marginal abatement cost curve for France and identify the combination of carbon storing

57 practices that minimizes the total cost of achieving a given national net GHG mitigation target.

58 We find that a substantial amount of carbon, 36 to 53,5 MtCO₂e yr⁻¹, can be stored in soil and

59 biomass for reasonable carbon prices of 55 and 250 € tCO₂e⁻¹, respectively (corresponding to

60 current and 2030 French carbon value for climate action), mainly by developing agroforestry

61 and hedges, generalising cover crops, and introducing or extending temporary grasslands in

62 crop sequences. This finding questions the 3-5 times lower target retained for the agricultural

63 carbon sink by the French climate neutrality strategy. Overall, this would decrease French GHG
64 emissions by 8 to 11,7% respectively.

65

66 Key words: soil organic carbon sequestration, climate change mitigation, greenhouse gas,
67 carbon neutrality, agriculture, abatement cost

68 1. Introduction

69 The Intergovernmental Panel on Climate Change (IPCC) demonstrated that in order to keep
70 global warming below +1.5°C compared to the pre-industrial period, it would be necessary to
71 achieve global carbon neutrality by 2050 (Allen et al., 2019). Achieving carbon neutrality
72 implies both drastically reducing greenhouse gas emissions and increasing the terrestrial CO₂
73 sink, through changes in land use and agricultural and forestry practices that promote carbon
74 sequestration in soils and woody biomass. The EU targets to balance 500-600 MtCO_{2e}.yr⁻¹ of
75 residual emissions with carbon sequestration, 50-85% of which is expected to come from soils
76 and biomass (European Commission, 2018). France is planning to rely even more on biological
77 carbon sequestration, which is expected to offset 80% of its residual 80 MtCO_{2e}.yr⁻¹ in 2050
78 (MTES, 2020). However, in this planning exercise, the role of agricultural soils is modest –
79 storing 10 MtCO_{2e}.yr⁻¹ – putting a heavy and likely unrealistic pressure on the forestry sector
80 (I4CE, in prep). A thorough assessment of the carbon storage potential of agricultural soils, via
81 the implementation of climate smart practices, and of the related costs is therefore needed in
82 order to design realistic pathways towards carbon neutrality in 2050, both at French and
83 European levels.

84 Several studies have estimated the cost efficiency of carbon storage practices. Most of these
85 studies focused on no-till or reduced tillage (De Cara et al., 2006, Pautsch et al. 2001, Feng et

86 al. 2000, 2002, 2006 and Kurkalova et al. 2006; Moran et al., 2011, Frank et al., 2015), which
87 have recently been demonstrated to have little to no effect on soil organic carbon (SOC) in
88 temperate regions when the entire soil profile is considered (Haddaway et al., 2017). Other
89 studies have estimated the cost of storing more SOC by introducing temporary grasslands or
90 alfalfa in typical Australian farms (Kragt et al., 2012), converting arable land to permanent
91 grasslands in some US states (Antle et al., 2001), or reducing summer fallow and increasing
92 adoption of conservation tillage in US wheat and corn systems (Antle et al., 2007).

93 The existing literature presents two shortfalls. First, in most cases, previous studies focus on a
94 single practice, thus falling short of providing an estimate of the total SOC storage potential at
95 regional or national scale. Second, they often neglect to estimate the total GHG budget of
96 practices. While carbon storage is a major component of carbon neutrality strategies, promoting
97 practices for which carbon storage is offset by increased GHG emissions would be obviously
98 counter-productive. Moreover, carbon pricing mechanisms such as cap-and-trade systems or
99 carbon offsetting schemes incentivize the total GHG budget, rather than its sub-components.
100 Likewise, some studies have analysed the mitigation potential of some practices that are
101 favourable to carbon storage, such as introduction of cover crops, agroforestry and better
102 grassland management (Pellerin et al. 2017, Fellmann et al. 2021) or land-use change (Mosnier
103 et al. 2019). Nonetheless, these studies fall short of estimating the large-scale potential for
104 carbon storage and rely on rough assumptions for the effects of these practices on carbon
105 storage.

106 The objectives of this study are to assess the carbon storage potential, the net GHG balance and
107 the cost efficiency of a large range of agricultural practices in order to refine the French and
108 European carbon neutrality strategies.

109 Marginal Abatement Cost Curves (MACCs) which inform on the costs of an additional unit of
110 emission reduction at any given total abatement level (Huang et al. 2016) provide useful

111 information to develop climate change policies. They help identifying the mitigation measures
112 that should be prioritized to meet climate mitigation targets and assessing the necessary carbon
113 price to reach them. Povellato et al. (2007), Kuik et al. (2009), Vermont and De Cara (2010),
114 Eory et al. (2018) reviewed different approaches used to derive MACCs. Vermont and De Cara
115 (2010) divide them in three broad categories: (i) supply-side, micro-economic models among
116 which are micro-econometric and mathematical programming models, (ii) equilibrium models
117 that simultaneously model supply and demand in one or more sectors of the economy, and (iii)
118 engineering cost approaches. The first two approaches represent the behaviour of economic
119 agents (producers and/or consumers) and typically estimate the mix and the share of adoption
120 of each mitigation practice for a given carbon price. Engineering models generally estimate the
121 cost of mitigation practices one by one and sort them by increasing cost-effectiveness (Pellerin
122 et al., 2017, Eory et al., 2018). Except from Biggar et al. (2013) where results are disaggregated
123 by the main US regions and by crop type, MACCs from engineering models generally aggregate
124 mitigation potential and cost per measure at the national level.

125 Here we estimate the MACC of all SOC storing practices which are relevant in mainland
126 France. To do so, we use the BANCO model (Bamière et al., 2017) which is in-between supply-
127 side models and engineer-type approaches. The mitigation potential per hectare, costs and
128 maximum area of application of each practice are estimated for each region, based on detailed
129 crop and grassland biophysical simulations or on literature and expert knowledge. An
130 optimization procedure then allows to determine the cost-minimizing allocation of the effort
131 across practices and regions to achieve various climate mitigation targets at national level. We
132 demonstrate that 36 to 53,5 MtCO₂e.yr⁻¹ - the equivalent of 8 to 11,7% of current French GHG
133 emissions - can be stored in the soil and biomass of agricultural land for reasonable carbon
134 prices of 55 and 250 €.tCO₂e⁻¹.

135 2. Methodology

136 2.1. Selection of SOC storing practices

137 The SOC storing practices were identified based on a literature review for the three major land-
138 use types: croplands including temporary grasslands, permanent grasslands and forests. Land-
139 use changes between these three types were not considered. Eight carbon storing practices were
140 found relevant for France ([Table 1](#)): 1) spatial or temporal expansion of cover crops; 2)
141 mobilization of new exogenous organic carbon resources currently not applied on agricultural
142 soils; 3) replacement of silage maize with temporary grasslands; 4) agroforestry; 5) hedges; 6)
143 moderate intensification of extensive grasslands; 7) grassland grazing instead of mowing; and
144 8) grass cover of vineyards. No/low-tillage practices are excluded in the present study according
145 to the most recent meta-analysis showing that, in the temperate context, no-tillage results in a
146 redistribution of SOC over the soil profile with little to no increase in total SOC when the entire
147 soil profile is considered (Haddaway et al., 2017; Ogle et al., 2019).

148 These eight practices also happen to be the most relevant for total carbon storage (including
149 biomass) on agricultural land (cropland and grassland). Regarding forest soils, existing
150 scientific evidence only provides one clear recommendation: avoiding whole-tree harvesting –
151 including remnants and stumps (Achat et al., 2015; Mayer et al., 2020). Fortunately, this type
152 of harvesting remains exceptional in France and accordingly, no alternative SOC storing
153 practice was assessed for forest soils. For these two reasons, our estimate covers the total carbon
154 storage, soil and biomass, of agricultural land. For more details on the literature review and
155 selection process, see Pellerin et al. (2020).

156

	General description	Major additional working operations and investments	Carbon storage potential from literature review	References	Method used to evaluate the carbon storage potential in this study
Expansion of cover crops in croplands	Cover crops temporal extension (0.5-4 months) in areas where cover crops are already planted and cover crops spatial expansion in areas where cover crops are not present	Purchase of seeds, sowing, and mechanical destruction. Irrigation at seeding where needed and possible	313 ± 313 kg C ha ⁻¹ yr ⁻¹	Justes et al. (2013), McDaniel et al. (2014), Poeplau & Don (2015), Lal (2015), Constantin , et al. (2010)	Crop simulations at high spatial resolution with a process-based crop model (STICS)
New organic C inputs in croplands	Application in croplands of compost or digestate from organic waste and sludge from waste water treatment plants, both from sources that are not currently incinerated or buried in landfills.	Purchase, transport and spread of new organic C inputs Reduction of synthetic fertilization	100 kg C ha ⁻¹ yr ⁻¹ (sewage sludge), 100 kg C ha ⁻¹ yr ⁻¹ (liquid manure), 300 kg C ha ⁻¹ yr ⁻¹ (manure), 500 kg C ha ⁻¹ yr ⁻¹ (compost)	Zavattaro et al (2017), Morvan et al (2013), Powlson et al (2012)	Crop simulations at high spatial resolution with a process-based crop model (STICS)
Expansion of temporary grasslands in croplands	Extension (1-2 extra years) of existing temporary grasslands in crop rotations and substitution of fodder maize by temporary grasslands in crop rotations.	Reduction of synthetic fertilization Change in animal feed rations Change in crop-specific technical operations	~130 to 500 kg C ha ⁻¹ yr ⁻¹ (first 10 years after the plantation of a grassland)	Conant et al (2001), Franzluebbers et al (2014), Johnston et al (2017), Creme et al (2020)	Crop simulations at high spatial resolution with a process-based crop model (STICS)
Agroforestry in croplands	Plantation of trees <i>Juglans regia</i> x <i>nigra</i> , <i>Prunus avium</i>) in croplands (75 trees/ha).	Tree plantation and maintenance Timber harvest	250 kg C ha ⁻¹ of UAA.yr ⁻¹ (-230; +730) in the soil (cropland only) 900 kg C ha ⁻¹ yr ⁻¹ (-430; +1350) in the biomass	Cardinael et al (2017), Lorenz & Lal(2014), Kim et al, (2016), De Stefano & Jacobson (2017), Feliciano et al (2018), Cardinael et al (2018), Chatterjee et al (2018), Shi et al (2018), Pardon et al (2017), Drexler et al. (2021)	Assessment based on literature, national resolution
Hedges in croplands	Plantation of trees at the border of croplands (49 to 68 linear m/ha).	Tree plantation and maintenance Timber harvest	750 kg C.ha ⁻¹ of hedges.yr ⁻¹ (490; 1020) in the soil (cropland only) 240 kg C .ha ⁻¹ of UAA.yr ⁻¹ (-120; +370) in the biomass		Assessment based on literature, national resolution

			(authors' calculations)		
Moderate intensification of extensive grasslands	Increase of synthetic fertilization (here +50 kg N.ha ⁻¹) in permanent grasslands having a low initial level of synthetic fertilization	Increase use of synthetic fertilizers	0-210 ±70 kg C ha ⁻¹ yr ⁻¹	Conant et al.(2017), Sandermann et al. (2015), Abdalla et al. (2018), Franzluebbers &Stuedemann (2009), Poepleau et al (2018).	Permanent grassland simulations at high spatial resolution with a process-based grassland model (PaSim)
Grazing instead of mowing in permanent grasslands	Substitution of mowing by direct grazing in permanent grasslands whose management is mixed (based on grazing and mowing) Substitution of 1 our 2 cuts per year.	Change in synthetic and organic fertilization Change in harvest operations of hay and silage	From 111 ± 11 to 380 ± 20 kg C.ha ⁻¹ yr ⁻¹ depending on biomass removal	Pineiro et al. (2010) ; McSherry & Ritchie, (2013) ; Soussana & Lemaire (2014) ; Lu et al. (2017) ; Zhou et al. (2017) ; Abdalla et al. (2018) ; Eze et al. (2018) ; Byrnes et al. (2018)	Permanent grassland simulations at high spatial resolution with a process-based grassland model (PaSim)
Grass cover of vineyards (winter or permanent cover)	Permanent or winter grass cover of vineyards except for Cognac and Mediterranean vineyards and in stony ground areas	Plantation of grass cover and purchase of seeds Mechanical destruction (winter grass cover) Removal of chemical destruction (permanent grass cover) Increased use of synthetic fertilizers	From 160 (winter) 490 (permanent) kg C ha ⁻¹ yr ⁻¹	Constantin et al. (2012), Arrouays et al. (2002)(2002)	Assessment based on literature, national resolution

Table 1: Key characteristics of the eight carbon storing practices considered in this study

157

158

159 **2.2. General assumptions and baseline**

160 The additional carbon storage and cost incurred by the adoption of a storing practice are
161 measured relative to a reference situation, over a 30-years simulation period, assuming constant
162 market context and cropping systems corresponding to the average 2009-2013 period. Climate
163 conditions are also assumed to be the same for all scenarios and correspond to the 1983-2013
164 period. The land-use shares of croplands, permanent grasslands and forests are assumed to be
165 constant (no land-use changes over time nor between scenarios).

166 **2.3. Additional carbon storage assessment and GHG budget**

167 **2.3.1. Carbon storage**

168 One originality of the study lies in its high spatial resolution. The simulations were performed
169 at the scale of 30 966 homogeneous agricultural pedoclimatic units (PCU, size < 8x8km), each
170 one being characterized by its local climate, dominant soil type(s), 1 to 3 dominant cropping
171 and grassland systems (crop/grassland sequences identified in French Land Parcel
172 Identification System and current crop/grassland management practices identified in national
173 surveys), and initial SOC stock from soil inventory data (Mulder et al., 2016) (see Table A 1 for
174 a summary of data sources).

175 The simulations were carried out using two process-based ecosystem models, STICS (Brisson
176 et al, 2003, pre-version 10) for arable crops and PaSim (Ma et al, 2015, version 5.3) for
177 permanent grasslands. Both models include an explicit representation at a daily time step of the
178 water, nitrogen, and carbon cycles in the soil, plant growth, and account for the effect of the
179 multiple pedoclimatic factors (e.g. radiation, temperature, precipitation, detailed soil properties)
180 and management practices that drive these processes. They both provide multiple outputs
181 including crop and grass production, change in SOC stock, nitrogen leaching, and GHG
182 emissions (NH₃ and N₂O emissions, enteric CH₄). PaSim simulates the SOC dynamics over the

183 whole soil depth while STICS consider it only for the first 30 cm. To estimate carbon storage
184 over the entire soil profile (0-100 cm), STICS results for the first 30 cm were extrapolated using
185 the function proposed by Balesdent et al (2018) (Table A 3). Details regarding the modelling
186 approach are provided in Graux et al., (2020), Launay et al., (2021a) and the simulation outputs
187 are accessible online (Launay et al., 2021; Martin, 2021).

188 In order to minimize modelling bias, we focus on the additional carbon storage, calculated as
189 the difference between the simulated C stock under C storing practices and the simulated C
190 stock under current management practices (i.e. baseline), after a 30 years period:

191 **additional C storage** (tC ha⁻¹ yr⁻¹) = $\frac{(final\ C\ stock\ storing\ practice - final\ C\ stock\ baseline)}{simulation\ length\ in\ years}$

192 Current soil-crop models are not able to simulate the carbon stock changes resulting from
193 agroforestry, hedges and grass cover of vineyards. The average value for additional storage
194 from the literature review (cf. [Table 1](#)) was retained for these three practices, without
195 considering spatial heterogeneity.

196 **2.3.2. Greenhouse gases budget**

197 For each storing practice, a complete GHG budget was also calculated. Carbon sequestration in
198 soil and biomass, N₂O emissions, nitrate leaching, and NH₃ volatilization were simulated by
199 STICS and PaSim for croplands and grasslands, respectively. The emissions associated to
200 changes in fertilizer manufacturing, fuel consumption and substitution of carbon-intense
201 materials and energy by wood use were also estimated.

202 We assumed that farmers buy or sell the differences of fodder induced by the new practice, thus
203 maintaining both animal feed and animal production levels. Consequently, there is no variation
204 in enteric methane and manure management emissions.

205 The detailed values and sources of emission factors are available in Bamière et al. (2021).

206 **2.4. Implementation cost assessment**

207 Implementation costs are calculated as the difference between the storing practice and the
208 current practice. A negative cost represents a gain for the farmer. We account for overhead
209 variations (purchase of inputs, crop management operations impacting labour, machinery or
210 fuel, etc.), dedicated investments, and revenue changes associated with production changes
211 (yield variation, change in land allocation - e.g. crop area substituted with trees or hedges, wood
212 sales, etc.), excluding any “optional subsidy” (e.g. Common Agricultural Policy payments, agri-
213 environmental measures, local subsidies). For the storing practices implying cash flows varying
214 over time (agroforestry and hedges), we compute a constant annuity with a 4.5% discount rate
215 (Quinet et al. (2013)).

216 The technical changes resulting from the implementation of storing practices (e.g. changes in
217 input and labour use, land allocation, ...) are derived from the literature or, when not available,
218 from expert knowledge. All technical changes and the corresponding sources of information
219 are documented in Bamière et al. (2021).

220 In order to minimize modelling bias, simulated yields were used to estimate variation
221 coefficients that were applied in percentage to the reference yield of each crop obtained from
222 national statistics. Absolute simulated yields were only used for grass due to the lack of national
223 statistics on grass yield. In addition, crop species which are not explicitly simulated by STICS
224 are assumed to undergo the same yield and inputs variations as the most similar simulated crop
225 (for example, durum wheat, is associated with winter wheat).

226 Gross margin losses resulting from changes in land allocation (e.g. agroforestry tree line
227 footprint) are estimated based on data from the Farm Accountancy Data Network (FADN).

228 When the changes in land allocation are crop-specific (e.g. substitution of fodder maize with
229 grass, which can in turn change the share of other crops such as soft wheat or rapeseed in a crop

230 sequence), we use crop-specific gross margins. The gross margins per crop are not directly
231 available in the French FADN. Prices and yields are available for each crop, but input costs
232 (e.g. seeds, fertilizer, ...) are only reported at the farm level. We therefore estimate crop-specific
233 costs using a linear regression on the farms' land allocation to distribute the expenses among
234 crops. This work is carried out in each of the regions. When, for a given crop, the estimates are
235 not significantly different from zero at a 5% level (or if the sample size is lower than 30 farms),
236 we use the gross margin of the same crop in a neighbouring region or, if not available, the one
237 estimated at the national level. Because a large part of fodder crops is self-consumed, yield
238 and/or prices for alfalfa, grass and fodder maize are not available in FADN data and were
239 obtained from annual agricultural statistics (Table A 1).

240 Any yield variation specific to fodder is assumed to be compensated by a substitute feed ration
241 with the same energy and protein level and with a similar fill value, resulting in unchanged milk
242 and meat production. Feed purchases or sales are adjusted accordingly.

243 ***2.5. Area base of carbon storing practices and aggregation of results at the*** 244 ***region level***

245 Technical criteria on the applicability of a given storing practice are derived from the literature
246 and, when not available, from expert knowledge (Table A 2). They are used to quantify the
247 maximum potential area base of each selected practice. These criteria can lead to the exclusion
248 of certain crops, crop sequences, or soil types for a given storing practice (e.g. no cover-crop
249 for intercropping period <2 months, no agroforestry if soil depth <1m or plot size <1ha).

250 STICS and PaSim only simulate the dominant cropping/grassland systems in each pedo-
251 climatic unit (PCU). In order to ensure the representativeness of the results at regional and
252 national levels, a three-stage spatial aggregation procedure is implemented. First, of the area of
253 each "dominant cropping system" is upscaled so that the area simulated by the models equals

254 the agricultural area of the whole PCU. Then, knowing the weight of each PCU in the region,
255 an aggregation is carried out at the regional level. Finally, a crop-specific correction factor is
256 applied to match the regional areas of each crop from the Annual Agricultural Statistics.

257 Area bases, regional costs, and regional additional storage potentials of each practice are
258 provided in Table A4 to A7.

259 ***2.6. Cost-effective allocation of the additional carbon storage effort***

260 In the context of climate change mitigation, rewarding carbon storage regardless of the GHG
261 budget of practices makes little sense. The cost-effective allocation of the net GHG abatement
262 (net GHG mitigation plus additional carbon storage) effort across practices and regions is
263 therefore determined, based on the per hectare cost for farmer and net GHG budget, as well as
264 the potential applicability of each practice in each region.

265 For that purpose, we use the BANCO model (Bamière et al, 2017), which optimizes the uptake
266 level ($\sum_c X_{r,p,c}$, in ha) for each practice p in each region r , to minimize the total mitigation cost
267 to achieve a national mitigation target, considering compatibility constraints between practices
268 (e.g. for agroforestry, tree rows are no more available for other SOC storing practices). The
269 total cost TC (Eq.1) and the total mitigation are determined by the sum for all regions (r),
270 practices (p), and crops (c) of the actual uptake level of the measure ($X_{r,p,c}$) times the associated
271 unitary costs ($uc_{r,p,c}$) and unitary abatement ($ua_{r,p,c}$), respectively. The constraints are: (i)
272 compliance with a total mitigation target TM (Eq.2), (ii) compliance with the maximum area
273 base ($\overline{X_{r,p,c}}$) for each tuple (region, practice, crop) (Eq.3), and (iii) competition between
274 measures for the use of land at the region level (Eq.4). Equation 4 reads as, for instance, the
275 total number of hectares of a given crop c used by mutually exclusive practices ($I_{p,p',c} = 1$) in
276 a given region must not exceed the total initial area of crop c in this region ($\overline{X_{r,c}}$), minus the area
277 of crop c converted to another land use ($X_{r,p',c} * luc_{r,p',c}$, $luc_{r,p',c}$ being the land use conversion

278 coefficient). For example, the wheat area concerned by cover cropping plus the wheat area
 279 converted to tree rows due to agroforestry or hedges implementation, must not exceed the total
 280 wheat area of the region (based on national statistics).

281 The crop index is a modelling proxy for crop sequences, in order to accurately account for the
 282 actual uptake potential of practices, because some practices are not applicable to all crop
 283 sequence (e.g. cover crops can sometimes be implanted after wheat and sometimes not,
 284 depending on the sowing date of the following crop) or to the whole area of a given crop
 285 sequence (e.g. minimum soil depth of 1m for agroforestry). However, results are only presented
 286 at the practice x region grain because storing practices cannot in practice be implemented on a
 287 single crop.

288 Net emitting practices in a given region are also excluded from the cost-effective allocation of
 289 the mitigation effort.

290

$$291 \quad \min_{X_{r,p,c}} TC = \sum_{r,p,c} uc_{r,p,c} \times X_{r,p,c} \quad \text{Eq. 1}$$

$$292 \quad \text{s.t.} \quad \sum_{r,p,c} ua_{r,p,c} \times X_{r,p,c} = TM, (\lambda) \quad \text{Eq.2}$$

$$293 \quad 0 < X_{r,p,c} \leq \overline{X_{r,p,c}}, \forall(r, p, c) \quad \text{Eq.3}$$

$$294 \quad \sum_{p' \geq p} (X_{r,p',c} * I_{p,p',c} * luc_{r,p',c}) \leq \overline{X_{r,c}}, \forall(r, p, c) \quad \text{Eq.4}$$

295 By varying the national mitigation target and reporting the associated marginal cost (λ in
 296 $\text{€} \cdot \text{tCO}_2\text{e}^{-1}$, dual price of Eq.2), we are able to depict a marginal abatement cost curve.

297 The model is written in GAMS, non-linear programming and solved with the CONOPT solver.

298 3. Results

299 3.1. Cost and efficiency per storing practice

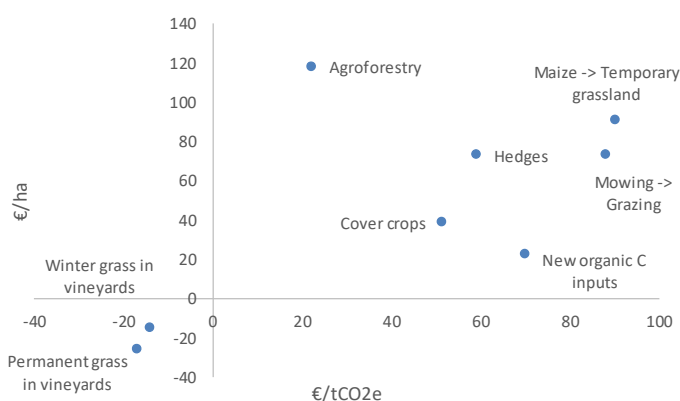
300 SOC storage represents the bulk of the net GHG budget for most practices ([Table 2](#)[Table-2](#), [Table](#)
301 [3](#)[Table-3](#)). The three exceptions are agroforestry and hedges, for which biomass storage is the
302 most important component, and the moderate intensification of permanent grasslands, for which
303 the increase in fertilisation and related N₂O emissions cancels out the benefits of SOC storage.
304 The practices with the highest potential for additional SOC storage per hectare (tCO₂e.ha⁻¹.yr⁻¹)
305 ¹) are agroforestry, grass cover of vineyards, and the replacement of mowing by grazing, with
306 more than 1 tCO₂e.ha⁻¹.yr⁻¹. They are followed by the expansion of temporary grasslands and
307 cover crops and the moderate intensification of extensive grasslands, with circa 0.75 to 0.80
308 tCO₂e.ha⁻¹.yr⁻¹, and finally new organic C inputs in cropping systems and hedges.

309 On the whole, additional carbon storage is costly to farmers. Four practices, grass cover of
310 vineyards, new organic resources, moderate intensification of permanent grasslands, and cover
311 crops expansion, have moderate average implementation costs (from -26 to 39 €.ha⁻¹ on average
312 at the national level, see [Table 3](#)[Table-3](#)). They are dominated by additional input or machinery
313 costs (e.g. mineral N fertilizer in grasslands, seeds and sowing of the cover crops, purchase and
314 delivery of green waste compost), which are only partly compensated for by an increase in
315 production (e.g. increased grass yield, increase in yield for some crops in the case of cover
316 crops, slight increase or stabilisation of yields over time for the new organic resources).

317 Practices such as grazing instead of mowing, the expansion of temporary grasslands in crop
318 rotations, agroforestry, and hedges have higher average implementation costs (73 to 118 €.ha⁻¹
319 ¹, see [Table 2](#)[Table-2](#)). The last three practices all imply some crop substitution, thus decreasing
320 the share of cash crops. The resulting loss of revenue is neither compensated for by wood sales
321 (agroforestry and hedges), nor by inputs or machinery savings (temporary grasslands). As for

322 the substitution of mowing by grazing, it implies a decrease in hay and grass silage stocks, the
 323 cost of which generally not being offset by the increase in grazed fodder nor by the savings in
 324 harvesting costs.

325



326 *Figure 1 National average implementation costs of practices carbon storing, in relation to their abatement cost (based on net*
 327 *GHG budget).*

328

329 What matters more for decision makers is the cost-efficiency of the various storing practices.

330 Overall, abatement costs per tCO₂e are correlated with implementation costs per hectare (Figure
 331 1Figure 1). However, agroforestry and hedges are among the most expensive practices per
 332 hectare, but become the second and fourth cheapest respectively on a per tCO₂e basis, thanks
 333 to their storage potential (mostly woody biomass). To the contrary, the merit of new organic
 334 resources is greatly diminished. In the end, the most cost-efficient practices are (on average at
 335 the national level) grass cover of vineyards, agroforestry, cover crops, and hedges. Their
 336 abatement costs per tCO₂e are lower than or close to the current French carbon target value¹

¹ Value that serves as a benchmark for evaluating public investment projects in France. It gives the threshold below which we consider that the implementation of a storing practice is beneficial for society as a whole, without prejudging the public policy instrument to be implemented. [The Value for Climate Action \(strategie.gouv.fr\)](https://strategie.gouv.fr)

337 (i.e. 55€·tCO₂e⁻¹). Nevertheless, after excluding the practices that were net emitters in some
338 regions, nearly all C storing practices in the remaining regions have an abatement cost that fits
339 below the carbon target price set by France to achieve carbon neutrality by 2050 (i.e. 250
340 €·tCO₂e⁻¹, Quinet et al. (2019)).

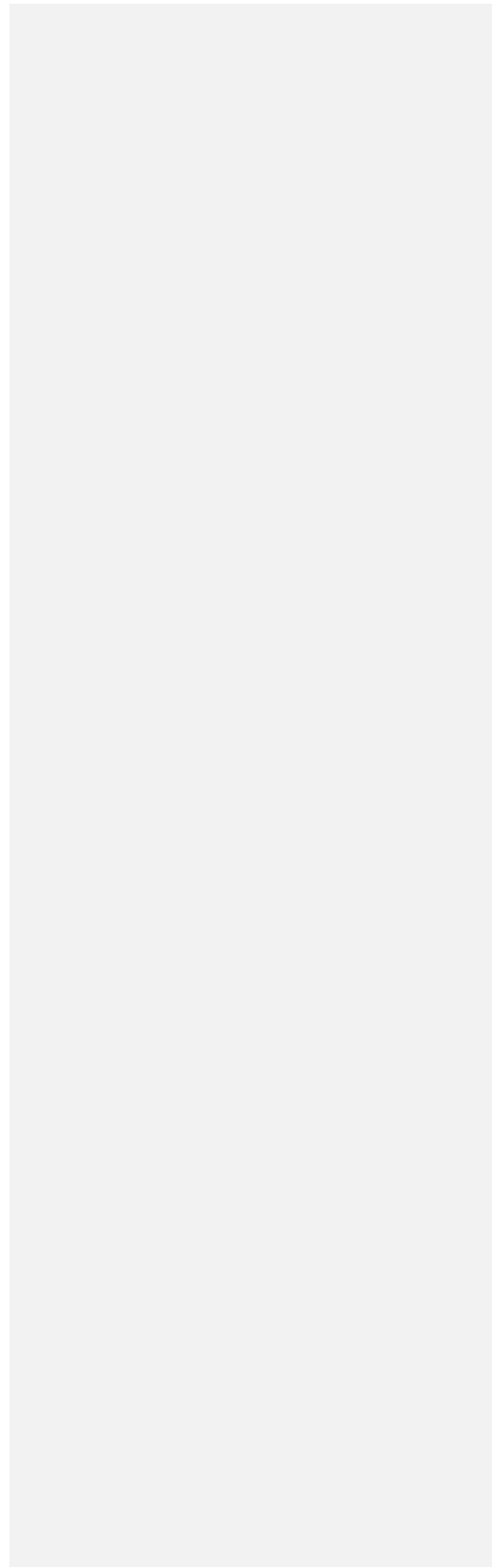
341 As we can see in [Table 2](#), there is great inter-regional variability in the cost-efficiency of
342 SOC storing practices, arising from the regional implementation cost and/or net abatement
343 potential per hectare. This heterogeneity is mainly due to the heterogeneity of land allocation,
344 yield potential, and cropping and grassland systems between the different agricultural
345 production basins, as well as the yield variation level following the uptake of a practice. Some
346 regions in northern France are important producers of high gross margin crops such as sugar
347 beets and potatoes. Practices that lead to a decrease in cultivated land, such as agroforestry,
348 hedges, and introduction of temporary grasslands in crop sequences are then particularly costly
349 in these regions.

350 Extending the duration of existing cover crops is cheaper than introducing them in crop
351 sequences, making the cover crops lever cheaper in regions where extending duration is more
352 often feasible. The cost of introducing cover crops also increases with the share of grain maize
353 which requires the use of more expensive cover crops seeds such as faba beans and vetches,
354 and also often leads to an increase in irrigation (see Launay et al., 2021). These additional
355 expenses are not always offset by an increase in production.

356 The regional variability of the new organic C inputs implementation cost depends mostly on
357 the type of organic resources available in the region (purchase price, delivery and spreading
358 costs): digestates and sludge composts from wastewater treatment plants are currently nearly
359 costless for farmers, unlike bio- and green-waste composts.

360 Finally, the heterogeneity in crop and grassland systems and in pedo-climatic conditions across
361 regions also influences the potential for additional SOC storage and the net GHG budget of the
362 various practices (see Launay et al, 2021, for a comprehensive discussion).

363



364
365
366

Table 2 National average results per storing practice (in bold). Regional extrema are provided in brackets. Average additional SOC storage is reported for the whole soil profile (0-100 cm). The net GHG budget includes additional C storage in soil and biomass and other GHGs mitigation. A negative/positive GHG budget corresponds to a net carbon sequestration/ a net emission, respectively. (*Excluding net emitting regions)

Storing practices	Potential applicability	Additional SOC storage	Net GHG budget	Cost for farmer	SOC storage cost	Abatement cost
	(Mha)	(tCO ₂ e ha ⁻¹ yr ⁻¹)	(tCO ₂ e ha ⁻¹ yr ⁻¹)	(€ ha ⁻¹ yr ⁻¹)	(€ tCO ₂ e ⁻¹)	(€ tCO ₂ e ⁻¹)
Expansion of cover crops	16,03	-0,775 (-1,392; -0,148)	-0,736 (-1,340; -0,141)	39 (12; 147)	49 (19; 301)	51 (20,3; 304,9)
New organic C inputs	1,46	-0,359 (-0,722; -0,066)	-0,324 (-0,668; -0,111)	22,6 (-92; 269)	63 (-127; 933)	70 (-137,7; 1 195,9)
Expansion of temporary grasslands	6,63	-0,785 (-2,747; 0,010)	-0,903 (-3,014; -0,087)	91 (-41; 314)	116 (-66; 455)	90 (-47,8; 269,7)
Agroforestry	5,33	-1,432 (-1,718; -0,706)	-5,306 (-5,629; -4,493)	118 (63; 179)	82 (53; 105)	22 (12,7; 32,0)
Hedges	8,83	-0,115 (-0,144; -0,056)	-1,236 (-1,385; -0,974)	73 (54; 87)	633 (549; 987)	59 (52,8; 65,9)
Moderate intensification of extensive grasslands	3,94	-0,747 (-1,118; -0,116)	0,010 (-0,326; 1,131)	28 (12; 38)	35 (16; 324)	101 (49,3; 136,1) *
Grazing instead of mowing (perm. grass.)	0,09	-1,349 (-1,962; -0,111)	-0,986 (-1,149; -0,173)	73 (-85; 146)	55 (-761; 141)	88 (-491,0; 293,3)
Grass cover of vineyard s permanent	0,15	-1,704 (-2,212; -1,301)	-1,534 (-1,892; -1,256)	-26 (-27; -22)	-15 (-21; -11)	-17 (-21,8; -13,4)
Grass cover of vineyard s in winter	0,41	-1,100 (-1,100; -1,100)	-1,087 (-1,087; -1,087)	-15 (-15; -15)	-14 (-14)	-14 (-14,0; -14,0)

367

Table 3: Detail of the net GHG mitigation potential of each storing practice. The national average results per practice are in bold, the regional extrema are provided in brackets. Average additional SOC storage is reported for the whole soil profile (up to 1 m). A negative/positive GHG budget corresponds to a net carbon sequestration/ a net emission, respectively.

Storing practices	Potential applicability	Additional SOC storage	Biomass C storage	Other GHGs	Net GHG budget	Total additional SOC storage potential	Total abatement potential
	(Mha)	(tCO ₂ e ha ⁻¹ yr ⁻¹)	(tCO ₂ e ha ⁻¹ yr ⁻¹)	(tCO ₂ e ha ⁻¹ yr ⁻¹)	(tCO ₂ e ha ⁻¹ yr ⁻¹)	(M tCO ₂ e yr ⁻¹)	MtCO ₂ e yr ⁻¹
Expansion of cover crops	16,03	-0,775 (-1,392; -0,148)	0,000	0,023 (-0,005;0,089)	-0,736	12,43	-12,06
New organic C inputs	1,46	-0,359 (-0,722; -0,066)	0,000	0,035 (-0,045;0,104)	-0,324	0,53	-0,47
Expansion of temporary grasslands	6,63	-0,785 (-2,747;0,010)	0,000	-0,219 (-0,484; -0,097)	-0,903	5,21	-6,66
Agroforestry	5,33	-1,432 (-1,718; -0,706)	-3,300	-0,574 (-0,616; -0,485)	-5,306	7,63	-28,26
Hedges	8,83	-0,115 (-0,144; -0,056)	-0,893 (-0,995; -0,716)	-0,228 (-0,247; -0,203)	-1,236	1,02	-10,91
Moderate intensification of extensive grasslands	3,94	-0,747 (-1,118; -0,116)	0,000	0,784 (-1,118; -0,116)	0,010	2,95	0,15
Grazing instead of mowing (perm. grass.)	0,09	-1,349 (-1,962; -0,111)	0,000	0,517 (-0,061;0,961)	-0,986	0,12	-0,7
Grass cover of permanent vineyards	0,15	-1,704 (-2,212; -1,301)	0,000	-1,100 (0,046;0,320)	-1,534	0,26	-0,23
Grass cover of vineyards in winter	0,41	-1,100 (-1,100; -1,100)	0,000	0,013 (0,013;0,013)	-1,087	0,45	-0,45

371 **3.2. Maximum storage potential**

372 Assuming the additivity of our eight practices, the maximum cumulated technical potential for
373 carbon storage in soil and biomass amounts to 56 MtCO₂e.yr⁻¹ at the national level, of which
374 30.6 MtCO₂e.yr⁻¹ are in the soil. 98% of this storage potential comes from five practices: the
375 expansion of cover crops and temporary grasslands and the moderate intensification of
376 permanent grasslands, thanks to their large area base, and agroforestry and hedges, which also
377 benefit from a high storage potential per hectare. The maximum cumulated technical abatement
378 potential is 59 MtCO₂e.yr⁻¹.

379 Now, accounting for interactions between practices and the cost-effective allocation of the net
380 GHG abatement effort, the maximum economic abatement potential is 58,9 MtCO₂e.yr⁻¹,
381 corresponding to 77% and 13% of the agricultural and the national GHG emissions,
382 respectively. This net GHG abatement arises at 91% from additional carbon storage in soil (28.2
383 MtCO₂e.yr⁻¹) and in biomass (25,5 MtCO₂e.yr⁻¹) (see [Figure 3](#) and [Table 4](#)). Four
384 practices account for 97% of the net abatement: agroforestry (48%), hedges (18,5%), the
385 expansion of cover crops (19.6%), and, to a lesser extent, the expansion of temporary grasslands
386 (11%).

387 **3.3. Marginal abatement cost curve**

388 The marginal abatement cost curve depicted by means of the cost-effective allocation model
389 BANCO is analysed at four interesting points: (A) with no incentive to store carbon in soils;
390 for the current (B) and 2030 (C) carbon target price set by the French government; and (D) the
391 maximum abatement ([Figure 2](#)).

392

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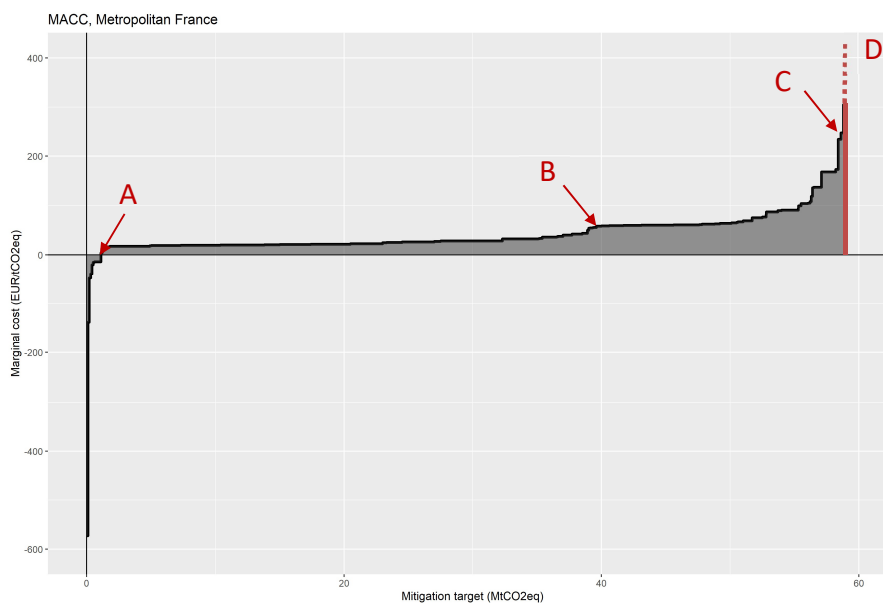
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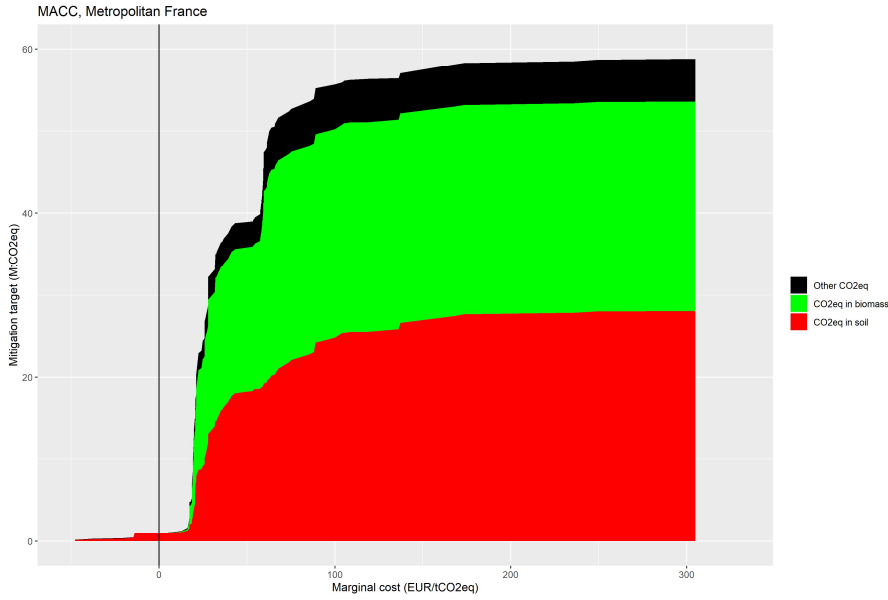
Scenario	(A) No incentive	(B) 55 € tCO ₂ e-1	(C) 250 € tCO ₂ e-1	(D) Maximum abatement
Total net GHG abatement (M tCO₂e yr⁻¹)	1,08	39,52	58,72	58,90
of which (in %)				
additional soil carbon storage	98%	47%	48%	48%
additional carbon sequestration in biomass	0%	45%	43%	43%
other GHG abatement	2%	8%	9%	9%
Total cost to farmers (M€ yr⁻¹)	-45,9	900,5	2452,9	2538
Marginal abatement cost (€ tCO₂e⁻¹)	-4,66	54,4	249	1 328
Storing practices contribution (M tCO₂e yr⁻¹) :				
Expansion of cover crops	0,000	8,137	11,432	11,526
New organic C inputs	0,176	0,280	0,441	0,474
Expansion of temporary grasslands	0,162	1,836	6,455	6,484
Agroforestry	0	28,259	28,259	28,259
Hedges	0	0,213	11	10,913
Moderate intensification of extensive grasslands	0	0,053	0,349	0,349
Grazing instead of mowing permanent grasslands	0,020	0,020	0,071	0,073
Grass cover of vineyards (permanent)	0,233	0,233	0,233	0,233
Grass cover of vineyards (winter)	0,448	0,448	0,448	0,448

394 Table 4: Total net GHG abatement, total cost to farmers, marginal abatement cost, and contribution of the storing practices
395 at (six) points of the marginal abatement cost curve.



396 Figure 2: Marginal abatement cost curve for mainland France: net GHG abatement (MtCO₂e yr⁻¹) on X axis; marginal
397 abatement cost (€ tCO₂e-1) on Y axis.

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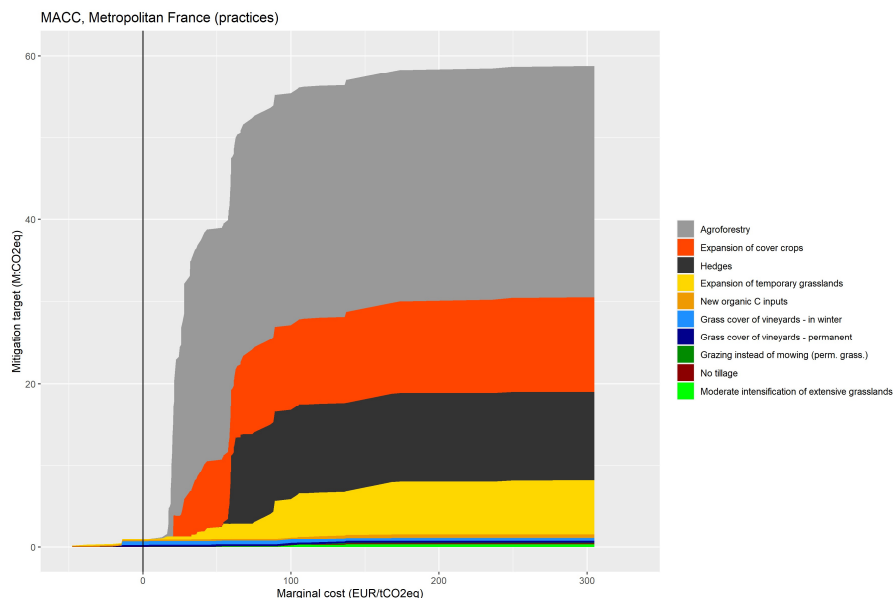


399 *Figure 3 : Marginal abatement cost curve: contribution of SOC carbon storage, biomass C storage, and other GHGs to the*
400 *total abatement target (Y axis), depending on the marginal abatement cost (€/tCO2e-1, on X axis).*

401

402

403



404 Figure 4: Marginal abatement cost curve: contribution of SOC storing practices to the total abatement target (Y axis),
405 depending on the marginal abatement cost (€ tCO₂e⁻¹, on X axis).

406

407 The maximum abatement potential is obtained for a total cost for farmers of 2 538 M€.year⁻¹
408 and a marginal cost (i.e. the cost of the last ton of CO₂e abated) of 1 328 € tCO₂e⁻¹, which is far
409 more expensive than the current (55 € tCO₂e⁻¹) and the 2030 (250 € tCO₂e⁻¹) target carbon
410 prices.

411 For a cost of carbon varying from 55 to 250 € tCO₂e⁻¹, it is possible to avoid the emission of
412 39,5 to 58,7 MtCO₂e.yr⁻¹. The associated additional carbon storage in soil and biomass amounts
413 to 36,3 to 53,2 MtCO₂e.yr⁻¹, including 18,6 to 28,1 MtCO₂e.yr⁻¹ for the sole additional SOC
414 storage (Figure 3). And the total cost for farmers would range from 900 to 2 453 M€.yr⁻¹.
415

416 As for the practices to be deployed at the national level, agroforestry and cover crops are
417 essential whatever the net mitigation target and the carbon price (Figure 4). The

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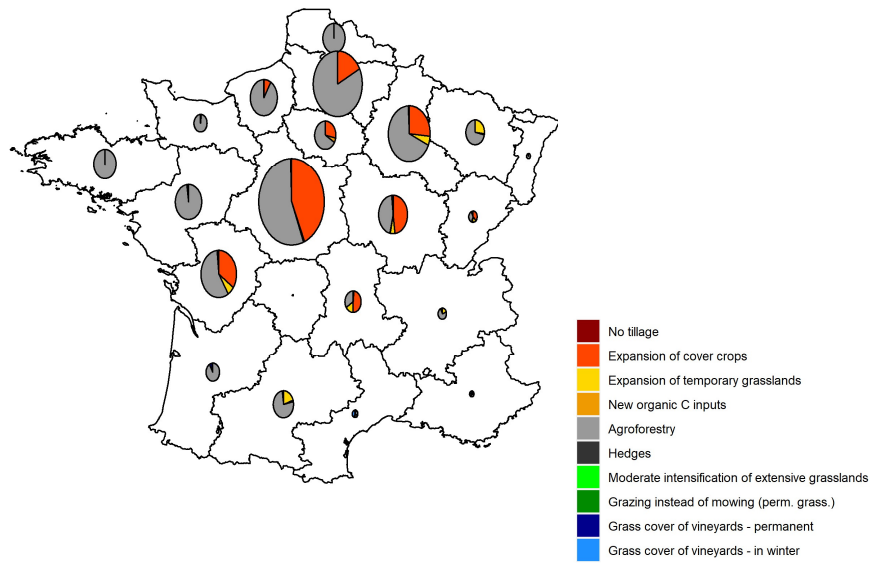
418 expansion of temporary grasslands has a wide abatement cost range depending on the regional
419 context and the practice is deployed progressively. Hedges become essential only for high
420 national mitigation targets.

421 The regional breakdown of the national mitigation target shows that there is no “one-size-fits-
422 all” solution, but rather a combination of good practices at the right place (Figure 5 and
423 6). For a given mitigation target (or carbon price), the contribution of each region varies both
424 in absolute value and in composition (i.e. type of practice implemented). Our results thus
425 provide useful information for the design of cost-effective policies at the region level, among
426 which the territorial climate-air-energy plans that the regions have to set up.

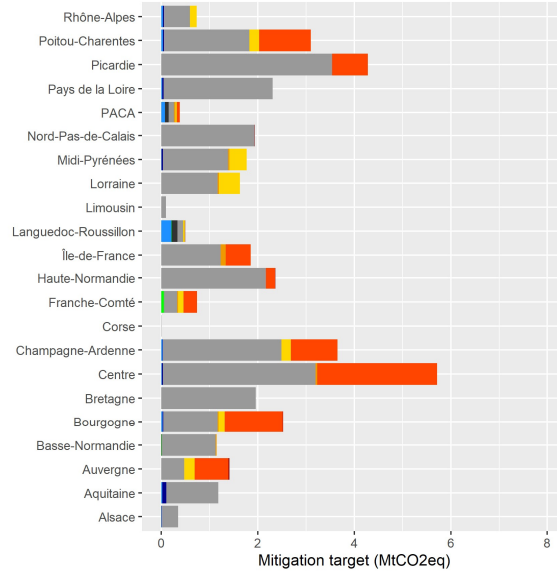
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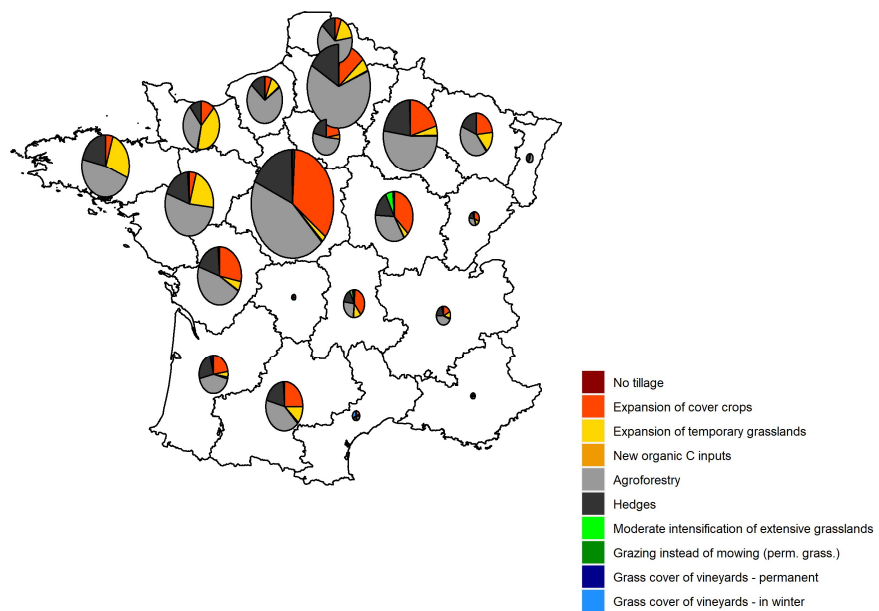


Regional breakdown of national mitigation target

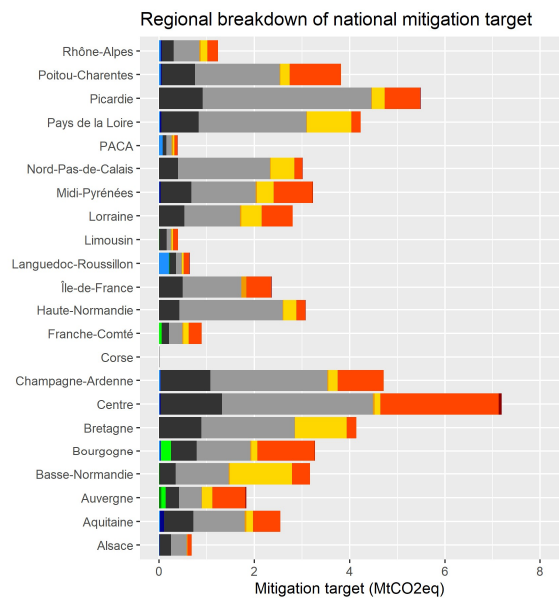


(a)

429 *Figure 5 : Regional carbon storage allocation, in MtC yr⁻¹, detailed per storing practice for the current (201,7 € tC⁻¹ or 55€ tCO₂e⁻¹) target carbon price (i.e. value for climate action) in France.*



(b)



431 Figure 6 : Regional carbon storage allocation, in MtC yr⁻¹, detailed per storing practice for the 2030 (b: 917 € tC⁻¹ or 250€ tCO₂e⁻¹) target carbon price (i.e. value for climate action) in France.

432

433 **4. Discussion**

434 ***4.1. Comparison with previous studies***

435 Our results show an overall net GHG abatement potential of 58,7 MtCO₂e.yr⁻¹ at a cost lower
436 than 250 €.tCO₂e⁻¹, when implementing agricultural practices selected on their a priori ability
437 to increase SOC storage in France. In the UK (agricultural area 33% smaller than France), Eory
438 et al (2015) estimated that 3,8 MtCO₂e.yr⁻¹ could be avoided for a cost lower than 225 £.tCO₂e⁻¹
439 ¹. In Ireland (agricultural area 5 times smaller than France), Teagasc (2012) estimated that less
440 than 3,1 MtCO₂e.yr⁻¹ could be avoided at a cost lower than 150 €.tCO₂e⁻¹ as opposed to 57
441 MtCO₂e for the same marginal cost in our study. Although these estimates consider all
442 mitigation practices while ours is restricted to carbon storage, they are comparatively much
443 lower than our estimates for France. Indeed, out of our four major practices, they only consider
444 cover crops but Teagasc (2012) limits its area base to spring barley and Eory et al (2015) does
445 not quantify the SOC storage benefit of the practice. Eory et al (2015) also notes the high
446 potential of agroforestry but does not quantify it in its MACC. For France, Pellerin et al. (2017)
447 and Fellmann et al. (2021, as part of a more comprehensive EU-level assessment) also
448 considered several mitigation practices and reported an abatement potential of about 32
449 MtCO₂e yr⁻¹ and 17,1 MtCO₂e yr⁻¹, respectively, at a cost lower than 250 € tCO₂e⁻¹. There
450 again, our abatement potential estimate is comparatively much higher: Pellerin et al. (2017) did
451 not consider grass/maize substitution and limited the cover crops potential to extension in space
452 (while we are also considering extension in time). Regarding agroforestry and hedges, the
453 estimates are similar except that Pellerin et al. (2017) arbitrarily limits the adoption to 7% of
454 the area base. Fellmann et al. (2021) addressed different practices (except cover crops) and did
455 not account for their carbon storage potential nor for CO₂ emissions.

456 It is difficult to compare our results with the other existing studies which tend to focus on a
457 single practice. Despite this, two results seem to stand out from the literature (Table A8 **Erreur !**
458 **Source du renvoi introuvable.Erreur ! Source du renvoi introuvable.**) regardless of the
459 method used or the territory studied: i) conservation agriculture, based among others on reduced
460 tillage and better crop rotations management, stands out for low carbon storage targets many in
461 studies prior to Haddaway et al (2017); ii) drastic land use changes such as afforestation
462 dominate for higher carbon storage targets (Newell et Stavins 2000, Plantinga et al. 1999,
463 Stavins 1999, Lubowsky et al. 2006). For instance, in the USA, Lubowsky et al (2006)
464 estimated that the afforestation of about 25 % of each state's agricultural land could increase
465 carbon sequestration by 750 MtC.yr⁻¹ for a cost of 105 €.tC⁻¹.yr⁻¹. However, such large land use
466 changes would compromise food security in the absence of dietary changes or waste reduction
467 (Muller et al., 2017; Springmann et al., 2018) which is why they were not considered in our
468 initial screening for carbon storage practices.

469 ***4.2.Negative implementation costs***

470 As shown in **Table 2Table-2**, only “grass cover of vineyards” presents negative implementation
471 costs in all regions, due to a decrease in crop management operations. This finding is consistent
472 with the observation that a large share of French vineyards are already covered with grass (50%
473 of the potential for permanent cover and 80 % of the potential for winter cover in 2013, Pellerin
474 et al. 2020). However, these negative implementation costs probably also point out the existence
475 of non-monetary barriers to adoption (e.g. labour availability constraints) and the possible
476 importance of some unaccounted costs (e.g., transaction costs, potential impacts on wine yield
477 and quality). In addition, they are sensitive to the hypotheses on the type of machinery used on
478 vineyards.

479 **4.3. Practices not considered**

480 Because this study focused on farming practices able to increase carbon storage, a few
481 mitigation levers typical of the agriculture and food sector such as shifting to plant-based diets
482 or optimizing nitrogen fertilization (Arneth et al., 2019) were not considered. No-till farming
483 and reduced tillage have been excluded due to their little potential for SOC storage in a
484 temperate context (see introduction). The effect of conservation agriculture on SOC stocks is
485 considered to arise essentially from the associated cover crops in temperate regions (Autret et
486 al., 2016). Similarly, biochar was not retained because of its questioned potential in the
487 temperate context (Arneth et al., 2019) and because to our knowledge it is currently not
488 practiced at all in France. Moderate intensification of pastures turns out to be excluded in
489 several regions as increased N₂O emissions from fertilizers more than offset the additional SOC
490 storage. Most importantly with regards to the French and European carbon neutrality targets,
491 the sylvo-pastoral potential for storage in tree biomass has been neglected. Using the same
492 criteria as for agroforestry on arable soil, it would add a maximal area base of 2,25 Mha and a
493 maximal biophysical potential of 7,4 MtCO₂e.yr⁻¹.

494 Last but not least of major carbon storage levers, wetland and peatland restoration was not
495 included in this study. In the European context, peatland restoration has been estimated to avoid
496 emissions between 2 and 34 tCO₂e.ha⁻¹.yr⁻¹ depending on climate, land-use and the extent of
497 degradation (Pellerin et al., 2020; Barthelmes, 2018). Moreover, the EU potential for reducing
498 emissions through peatland restoration has been estimated at around 109 MtCO₂e yr⁻¹
499 (Barthelmes, 2018). Due to the dearth of data on the abatement costs related to this practice in
500 France, let alone their spatial heterogeneity, it was not possible to include it in this study. With
501 only 140 kha of organic soils and 3,2 MtCO₂e.yr⁻¹ of reported emissions from wetland or
502 peatland drainage (CITEPA, 2020), this shortfall does not undermine our main conclusions: for

503 carbon prices higher than 28 €.tCO₂e⁻¹, the estimated abatement potential is more than 10 times
504 higher than the 3,2 MtCO₂e.yr⁻¹ maximum potential for wetland and peatland restoration.

505 ***4.4. Main limitations of the study***

506 In this study, we analysed which combinations of practices in each region would achieve GHG
507 national mitigation targets at the lowest cost, considering the area base and relative abatement
508 cost values of eight carbon storing practices at the regional level. Two types of limits were
509 identified: a first set related to uncertainty of biophysical models and a second set related to the
510 simplifying economic assumptions.

511 First, although the predictive value of STICS was shown to compare well with long term trials
512 (Clivot et al., 2020), model outputs remain uncertain, in particular for the simulation of
513 grassland management for which the simulated average is consistent with the literature but not
514 the simulated range (Pellerin et al., 2020). For instance, the ability of STICS to simulate SOC
515 dynamics in field crop rotations that include temporary grasslands is probably lower than for
516 pure arable cropping systems because STICS has rarely been evaluated for such mixed cropping
517 systems. The potential for additional SOC storage could therefore have been overestimated. In
518 PaSim, permanent grasslands management in each pedoclimatic unit was not fully adapted to
519 the local pedoclimatic condition. For instance, mowing and grazing dates were adapted with
520 temperatures (degree days) but do not account for the bearing capacity of soils in case of
521 grazing, thus potentially leading to an over-valuation of grazed grass. In addition, the models
522 do not account for the reduced need for fertilization nor for the higher water retention capacity
523 of soil, when the soil organic matter increases. By omitting potential savings of inputs such as
524 N fertilizer, this study could have overestimated the cost of the storing practices or
525 underestimated the associated GHG mitigation. The most important limit, which is also the
526 most challenging to address, is that these models have only been scarcely validated on field

527 trials for the specific practices considered (Levvasseur et al., 2021). For details on models and
528 simulation plan limitations, see Graux et al., (2020), Launay et al., (2021) and Pellerin et al
529 (2020).

530 For agroforestry and hedges, for which no model was available, the limit of our study lies in
531 the lack of spatial heterogeneity in the regional carbon storage potential per hectare. While very
532 diverse agroforestry systems and types of hedgerows exist in France (e.g. 22 types of hedgerows
533 in France, <https://afac-agroforesteries.fr/typologie-nationale-des-haies/>), we defined only one
534 scenario for each of these two practices, due to a limited number of observations about carbon
535 sequestration in the literature (Cardinael et al 2017; 2018; Mayer et al 2022). For instance, we
536 considered agroforestry systems with hybrid walnut and wild cherry trees for timber production.
537 However, more diverse agroforestry systems with production of other goods such as fuelwood,
538 fruits, nuts or even honey could potentially improve the economic performance of these systems
539 and reduce the cost of carbon per hectare. This heterogeneity is nevertheless accounted for in
540 the uncertainty range. Agroforestry could also be practiced on more marginal land with
541 shallower soils (<1m depth) and with a higher decrease in crop yield, but the performance in
542 terms of carbon sequestration and economic return could not be tested due to an absence of
543 data.

544 Economic estimates also suffer from several simplifying assumptions. First, our cost
545 estimations assume that overheads related to changes in cultivation operations per hectare are
546 constant across farms. In reality, economies of scale are likely to occur, especially for
547 mechanisation costs which depend on farm size, level of machine utilisation, machine
548 characteristics and type of ownership. Further analysis would be needed to explore the impacts
549 of these practices on different farm types. Second, costs were estimated for stable reference
550 prices, without the simulation of market feedbacks. Likewise, crop areas and livestock
551 production were maintained at levels close to those observed during the reference period, but

552 crop yields varied from -26% to +18% depending on practices (Launay et al., 2021), and forage
553 production was not kept constant. As crop supply is modified, it is very likely that market prices
554 would also be modified but the economic model used in this study does not account for this
555 effect. This in turn raises the issue of potential GHG emissions leakage outside of France (Frank
556 et al., 2015), although we tried to limit this issue by excluding major land-use changes from the
557 set of storing practices.

558 Eventually, the adoption of storing practices by farmers will depend on the type of public
559 policies implemented. These policies have both direct effects (e.g., the impact of introducing a
560 tax on net GHG budgets of farms) and indirect effects (induced by the adjustment of prices to
561 market equilibrium). A tax is for instance likely to lead to a decrease in ruminant numbers, an
562 increase in legumes and would induce profound changes in markets, farmers' incomes and food
563 prices (Mosnier et al., 2019; Tang et al., 2018). Further studies considering the full range of
564 market effects would be needed to test which public policies would be most effective in
565 achieving the storage goals tested in this study.

566 Last but not least, one must bear in mind that the additional carbon storage allowed by a storing
567 practice is i) «finite», i.e. it stops when a new carbon stock equilibrium is reached after a certain
568 number of years; and ii) «non-permanent», i.e. the storing practice should continue, even once
569 the equilibrium is reached, to prevent soil carbon release. For example, using “long-term
570 average” formula mandated by carbon offset standards to deal with the cyclical nature of
571 biomass changes in trees (e.g. Label Bas Carbone, 2020) would almost half the “physical”
572 biomass storage potential we estimated over the first 30 years.

573 ***4.5. Possible improvements and perspectives***

574 Our study is a first step towards the assessment of the synergy and trade-offs between carbon
575 storage and GHG emissions reduction. It also raises the issue of the permanence of carbon

576 storing practices, and how to ensure it, as achieving the annual additional storage potential
577 estimated in our study involves their continuous implementation for 30 years. Moreover,
578 additional storage must be considered in the context of climate change, with significant impacts
579 not only on carbon dynamics (Crowther et al. 2016), but also on land use, production systems
580 and practices.

581 Future work should include: i) simulations carried out under several climatic scenarios; ii)
582 robustness or sensitivity analysis of the cost-effective strategy accounting for inter-annual price
583 variations as well as the uncertainty surrounding additional carbon storage estimates. The next
584 step would be to include our detailed results on practices (yields, technical costs variations, etc.)
585 in microeconomic supply models, to assess more accurately the opportunity cost of practices
586 and various abatement strategies and to help designing cost-effective climate policies. Research
587 on how to overcome the barriers to the adoption of the cost-effective practices by farmers is
588 also needed.

589 **5. Conclusion**

590 Eight potential carbon storing practices relevant to the French metropolitan territory were
591 identified based on a literature review: cover crops; new carbon inputs (e.g. sludge);
592 replacement of silage maize with temporary grasslands; moderate intensification of extensive
593 grasslands (+50kgN.ha⁻¹); animal grazing instead of mowing; agroforestry; hedges; and grass
594 cover of vineyards. The carbon storage potential of the first 5 practices was simulated at a very
595 fine spatial scale, together with their total GHG budget, considering the spatial heterogeneity
596 in pedoclimatic conditions and cropping techniques. No model being available for the last 3
597 practices, their carbon storage and net mitigation potential was assessed based on a
598 comprehensive literature review and expert knowledge. After assessing the implementation cost

599 at the regional scale, the potential applicability, and the net mitigation potential of each carbon
600 storing practice, we integrated them into an economic model (BANCO).

601 We find a potential for net GHG abatement of 58,9 MtCO₂e.yr⁻¹ for a total cost for farmers of
602 circa 2.5 G€.yr⁻¹, which would offset 13% of national GHG emissions and 77% of the
603 agricultural sector emissions. 99,7% of this abatement potential can be achieved at a lower cost
604 than 250 €.tCO₂e⁻¹ (i.e. the 2030 target carbon price in France). The abatement potential mostly
605 arises from additional carbon storage in soil (28,2 MtCO₂e.yr⁻¹, i.e. 48%) and biomass (43%),
606 and 98% of the total additional SOC storage potential is found in arable soils, where initial SOC
607 stocks are low. Reaching high mitigation targets mostly relies on the full deployment of four
608 key practices which add up to 97% of the net abatement potential: agroforestry, cover crops,
609 hedges, and the expansion of temporary grasslands – at the expense of silage maize – in crop
610 rotations. This finding supports the strategy of the European Commission which includes these
611 four practices in its list of six major storing practices, in addition to afforestation and peatland
612 restoration (European Commission, 2021).

613 The associated additional carbon sink in soil and biomass amounts to 53,6 MtCO₂e per year
614 and is 435% higher than the 10 MtCO₂e yr⁻¹ storage objective assigned to agricultural land in
615 the climate neutrality strategy (MTES, 2020). Adding the sylvo-pastoral potential would further
616 raise this additional carbon sink to 60 MtCO₂e per year. This figure offers a whiff of optimism
617 into the otherwise bleak assessment of the chances that France meets its 2050 climate neutrality
618 target (e.g. Haut Conseil pour le Climat, 2021). This good news should be tempered by the fact
619 that our estimate of “additional carbon storage” likely comes on top of currently unaccounted
620 emissions from cropland soils: the French national inventory currently reports a net
621 sequestration of 1 MtCO₂e.yr⁻¹ whereas French cropland soils are most likely net emitters, with
622 an order of magnitude estimated at 0.19 tC.ha⁻¹.yr⁻¹ (Pellerin et al, 2020).

623 The fine resolution of both agronomic and economic estimates also shows the importance of
624 taking into account the biophysical and agricultural specificities of each region for the design
625 of a cost-effective policy, for there is not “one good carbon storing practice” to increase carbon
626 storage in soils, rather a combination of good practices at the right place. Our results therefore
627 provide useful information for policy makers, on the potential and cost of carbon storage at a
628 fine spatial scale. They can be used in the frame of the CAP reform for the design of the national
629 voluntary eco-scheme, for instance. A policy aiming at supporting additional SOC storage in
630 arable land must not come at the expense of the preservation of high existing carbon stocks in
631 permanent grasslands and forests.

632 Finally, agriculture is at the center of several challenges (e.g. water quality, biodiversity
633 conservation, food security, bioeconomy, and indeed climate change). Given the high cost of
634 some practices such as agroforestry and hedges, and the fact that the practices studied provide
635 services other than just storing carbon in soil, there is a need to ensure coherence between the
636 existing policies and, ideally, bundles of ecosystemic services should be accounted for in an
637 integrated policy.

638

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974 Appendices

975 *A.1 Methodology, data, and data sources*

Type of calculation	Data requirements	Data sources
Carbon storage and GHG budget	Soil data (characteristics and use)	Geographic database for land use in France on a scale of 1/1 000 000 (BDGSF, INRA Infosol) Mulder et al. (2016) for initial SOC stocks, 90*90m grid
	Climate data	SAFRAN data base, 8x8 km grid (Meteo France, processed by INRA Agroclim)
	Crop sequences	Derived from the French Land Parcel Identification System (INRA ODR)
	Cropping practices Permanent grasslands management	Crop practices survey, 2006 and 2011 (SSP) Nitrate Directive 2012 for intercropping management Permanent grasslands survey (SSP) ISOP system (Information and objective monitoring of grasslands)
	GHG emission factors and equations, Emissions induced upstream/downstream	IPCC 2006, French Inventory (CITEPA), Carbone® database (ADEME)
Implementation costs	Crop reference prices and yields Crop gross margins	FADN (2009-2013 ; SSP), Annual agricultural statistics (2009-2013 ; SSP), coefficient from General Association of Corn Producers for fodder maize
	Input prices	Eurostat, national statistics
	Crop management operations costs	CUMA (machinery cooperative) third-party service delivery scale (FNCUMA, APCA)
Potential applicability	Crop areas and livestock numbers	Annual agricultural statistics (2009-2013 ; SSP)
	Limiting soil characteristics	BDGSF (INRA Infosol)

976 *Table A 1 Data sources per calculation type*

977

SOC storing practices	Implementation criteria
Expansion of cover crops	- Extension of the covers in place - Insertion of new cover crops in all fallows lasting more than two months

Expansion of temporary grasslands	<ul style="list-style-type: none"> - Extension (1-2 extra years) of existing temporary grasslands in crop rotations and substitution of fodder maize by temporary grasslands in crop rotations. - Replacement of silage maize by three years of temporary grasslands in crop rotations
New organic C inputs	<ul style="list-style-type: none"> - Random application of new products in rotations that were not receiving organic fertilization
Moderate intensification of extensive grasslands	<ul style="list-style-type: none"> - Supply of 50 kgN/ha/year for unfertilized or low-fertilized (< 50kgN/ha/year) permanent grassland
Grazing instead of mowing permanent grasslands	<ul style="list-style-type: none"> - For intensively mown permanent grasslands (4 cuts), 2 cuts are substituted by grazing events - For highly used permanent grasslands (2 cuts + 2 grazing events), 1 cut is substituted by 1 grazing event
Grass cover of vineyards (permanent)	<ul style="list-style-type: none"> - Vineyards with bare soils, winter grass cover, or grass cover every other inter-row - Excluding Mediterranean and Cognac vineyards (potential yield loss too important) - Soil stoniness > 35% and/or soil with stones > 7,5cm diameter
Grass cover of vineyards (winter)	<ul style="list-style-type: none"> - Bare inter-rows
Agroforestry	<ul style="list-style-type: none"> - Croplands \geq 1 hectare - Soil depth \geq 1m - No hydromorphic soils
Hedges	<ul style="list-style-type: none"> - Group of fields \geq 8 hectares - Soil depth \geq 50 cm

978 *Table A 2 Implementation criteria of the SOC storing practices*

979

SOC storing practices	Soil carbon stock conversion factor from 0-30cm to 0-100cm soil horizon
Expansion of cover crops	1,70603359
Expansion of temporary grasslands	1,68125666
New organic C inputs	1,71418257
Moderate intensification of extensive permanent grasslands	1,21033939
Grazing instead of mowing permanent grasslands	1,36695288
Grass cover of vineyards (permanent and winter)	1,886
Agroforestry	1,886
Hedges	1,886

980 *Table A 3. Factor to convert carbon stock of the 0-30cm soil horizon into carbon stock of the whole soil profile (0-100cm horizon)*

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A.2 Area base, storage potential, net GHG budget and cost per storing practice and region

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Table A 4 Area base of each practice in each region, in hectare. Adapted from [dataset] Bamière et al. (2021).

Region code	Region name	Cover crops	New organic resources	Expansion of temporary grasslands	Agroforestry	Hedges	Moderate intensification of permanent grasslands	Grazing instead of mowing	Grass cover of vineyards : permanent	Grass cover of vineyards : winter
11	Ile-de-France	508 260	149 706	993	219 634	417 204	20 028	0	0	0
21	Champagne-Ardenne	1 162 875	58 142	83 092	438 119	851 270	190 031	0	2 742	23 758
22	Picardie	1 062 450	29 014	106 841	628 755	745 768	10 486	0	0	0
23	Haute-Normandie	375 442	37 845	109 100	387 684	346 455	33 888	0	0	0
24	Centre	1 854 283	131 111	259 262	583 058	1 078 759	180 776	0	10 265	10 427
25	Basse-Normandie	658 484	37 405	453 383	208 961	272 307	219 105	11 419	0	0
26	Bourgogne	947 045	52 447	153 418	207 839	435 212	642 999	0	9 065	19 971
31	Nord-Pas-De-Calais	543 215	18 543	171 550	345 223	303 814	26 323	0	0	0
41	Lorraine	681 569	68 410	184 786	210 220	428 605	205 823	0	0	0
42	Alsace	207 262	31 320	21 089	60 254	168 356	44 665	0	9 667	800
43	Franche-Comte	268 794	44 407	92 110	52 882	124 325	213 932	0	0	0
52	Pays de la Loire	1 389 317	48 522	1 211 196	472 101	655 169	280 570	0	16 757	15 529
53	Bretagne	1 274 339	34 142	1 311 486	411 626	700 419	68 802	351	0	0
54	Poitou-Charentes	1 296 033	81 914	476 073	345 809	558 279	130 814	0	11 127	30 692
72	Aquitaine	750 086	141 558	319 528	208 245	480 029	172 507	0	61 242	18 798
73	Midi-Pyrenees	1 402 295	138 804	675 886	272 197	522 445	328 574	635	18 796	4 353
74	Limousin	280 049	78 884	277 854	20 581	126 645	337 215	4 335	0	0
82	Rhone-Alpes	518 855	127 926	316 425	109 564	198 489	399 390	2 879	11 827	27 152
83	Auvergne	544 957	22 610	295 257	97 266	228 192	340 210	53 861	0	0
91	Languedoc-Roussillon	186 462	48 395	59 343	23 207	113 108	38 788	14 420	0	191 483
93	Provence-Alpes-Cote-Azur	114 555	83 851	56 236	22 361	71 056	56 273	0	0	69 569
94	Corse	na	na	na	494	3 484	na	na	na	na

985

Table A 5 Implementation cost, per hectare of area base, of each practice in each region (€ ha⁻¹). Note: a negative cost represents a gain for the farmer, a positive cost a shortfall, compared to the baseline scenario. Adapted from [dataset] Bamière et al. (2021).

Region code	Region name	Cover crops	New organic resources	Expansion of temporary grasslands	Agroforestry	Hedges	Moderate intensification of permanent grasslands	Grazing instead of mowing	Grass cover of vineyards : permanent	Grass cover of vineyards : winter
11	Ile-de-France	28,6	-92,0	19,8	139,0	71,9	26,3			
21	Champagne-Ardenne	27,5	41,4	131,3	125,1	71,6	24,0		-25,5	-15,2
22	Picardie	30,9	29,2	233,5	156,9	76,2	29,2			
23	Haute-Normandie	18,9	50,8	197,0	144,6	75,5	24,0			
24	Centre	27,6	8,6	49,3	114,4	69,9	26,0		-25,7	-15,2
25	Basse-Normandie	38,6	28,1	263,4	107,6	73,0	24,2	-38,1		
26	Bourgogne	36,1	16,3	-40,6	92,7	69,9	32,5		-25,3	-15,2
31	Nord-Pas-De-Calais	21,7	229,0	313,7	178,8	86,5	37,7			
41	Lorraine	68,2	21,5	102,4	105,6	71,3	14,4			
42	Alsace	146,9	52,2	31,2	133,1	84,7	17,5		-27,4	-15,2
43	Franche-Comte	27,3	-21,6	4,8	90,4	72,0	12,1			
52	Pays de la Loire	24,4	214,5	67,9	95,2	71,3	29,2		-25,8	-15,2
53	Bretagne	38,4	268,9	142,8	91,7	75,1	31,0	-84,7		
54	Poitou-Charentes	29,5	28,0	13,9	96,0	73,9	31,9		-26,9	-15,2
72	Aquitaine	109,4	32,2	112,2	101,4	75,3	31,3		-27,4	-15,2
73	Midi-Pyrenees	43,7	-6,5	19,2	85,7	71,1	25,9	-7,2	-22,2	-15,2
74	Limousin	40,0	30,8	28,7	72,7	71,6	38,2	145,8		
82	Rhone-Alpes	26,2	25,2	24,5	91,7	74,3	20,3	94,6	-25,4	-15,2
83	Auvergne	51,5	48,1	11,4	84,5	71,8	34,8	112,6		
91	Languedoc-Roussillon	38,4	22,8	-5,3	68,7	64,6	35,9	-3,4		-15,2
	Provence-Alpes-Cote-									
93	Azur	11,9	-12,6	16,5	63,3	58,4	37,5			-15,2
94	Corse	na	na	na	67,4	54,1	na	na		na

Table A 6 Net GHG budget per hectare of area base for each practice in each region ($tCO_2e\ ha^{-1}$). Note : a negative value means a net removal of carbon from the atmosphere compared to the baseline scenario. Adapted from [dataset] Bamière et al. (2021).

Region code	Region name	Cover crops	New organic resources	Expansion of temporary grasslands	Agroforestry	Hedges	Moderate intensification of permanent grasslands	Grazing instead of mowing	Grass cover of vineyards : permanent	Grass cover of vineyards : winter
11	Ile-de-France	-1,079	-0,668	-0,087	-5,606	-1,190	0,156			
21	Champagne-Ardenne	-0,869	-0,396	-2,448	-5,607	-1,223	0,101		-1,885	-1,087
22	Picardie	-0,749	-0,494	-2,654	-5,629	-1,225	0,192			
23	Haute-Normandie	-0,528	-0,488	-2,616	-5,585	-1,231	0,033			
24	Centre	-1,340	-0,265	-0,470	-5,436	-1,199	0,030		-1,755	-1,087
25	Basse-Normandie	-0,606	-0,554	-2,960	-5,332	-1,244	0,022	-0,811		
26	Bourgogne	-1,299	-0,368	-0,848	-5,420	-1,240	-0,326		-1,892	-1,087
31	Nord-Pas-De-Calais	-0,335	-0,495	-3,014	-5,593	-1,313	0,119			
41	Lorraine	-1,008	-0,422	-2,372	-5,536	-1,241	0,135			
42	Alsace	-0,482	-0,111	-0,410	-5,511	-1,385	0,264		-1,256	-1,087
43	Franche-Comte	-1,055	-0,414	-1,239	-5,230	-1,266	-0,246			
52	Pays de la Loire	-0,141	-0,259	-0,791	-4,789	-1,198	0,246		-1,725	-1,087
53	Bretagne	-0,163	-0,225	-0,850	-4,750	-1,269	0,579	-0,173		
54	Poitou-Charentes	-0,841	-0,171	-0,423	-5,129	-1,264	0,398		-1,514	-1,087
72	Aquitaine	-0,797	-0,228	-0,451	-5,173	-1,286	0,199		-1,359	-1,087
73	Midi-Pyrenees	-0,593	-0,215	-0,523	-4,953	-1,228	0,100	-1,149	-1,586	-1,087
74	Limousin	-0,368	-0,196	-0,106	-4,493	-1,217	0,169	-0,497		
82	Rhone-Alpes	-0,424	-0,212	-0,453	-4,912	-1,277	0,122	-0,681	-1,826	-1,087
83	Auvergne	-1,307	-0,289	-0,738	-4,887	-1,228	-0,256	-0,901		
91	Languedoc-Roussillon	-0,635	-0,215	-0,531	-5,092	-1,187	0,497	-0,724		-1,087
	Provence-Alpes-Cote-									
93	Azur	-0,584	-0,330	-0,449	-5,000	-1,107	1,131	0		-1,087
94	Corse				-4,507	-0,974				

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Table A 7 Detail of the net GHG budget per practice and region : additional SOC storage (SOC, tCO₂e ha⁻¹), additional carbon storage in biomass (B, tCO₂e ha⁻¹), variation of GHG emissions (GHG, tCO₂e ha⁻¹), per hectare of area base. Note : a negative value means an increase in carbon storage or a decrease in GHG emission compared to the baseline scenario. (ECC : expansion of cover crops ; NOR : new organic resources ; ETP : expansion of temporary grasslands ; AF : agroforestry ; H : hedge ; MIPG : moderate intensification of permanent grasslands ; GIM : grazing instead of mowing ; GCVP/W: grass cover of vineyards permanent/winter) Adapted from [dataset] Bamière et al. (2021).

Region code	ECC		NOR		ETG		AF			H			MIPG		GIM		GCVP		GCVW	
	SOC	GHG	SOC	GHG	SOC	GHG	SOC	B	GHG	SOC	B	GHG	SOC	GHG	SOC	GHG	SOC	GHG	SOC	GHG
11	-1,099	0,020	-0,722	0,054	0,010	-0,097	-1,718	-3,300	-0,588	-0,128	-0,833	-0,228	-0,644	0,800						
21	-0,883	0,014	-0,444	0,048	-1,964	-0,484	-1,702	-3,300	-0,606	-0,130	-0,858	-0,234	-0,583	0,685			-2,201	0,317	-1,100	0,013
22	-0,746	-0,003	-0,570	0,077	-2,356	-0,298	-1,713	-3,300	-0,616	-0,131	-0,857	-0,237	-0,477	0,668						
23	-0,539	0,011	-0,543	0,056	-2,314	-0,302	-1,670	-3,300	-0,615	-0,129	-0,865	-0,237	-0,648	0,680						
24	-1,392	0,052	-0,294	0,028	-0,307	-0,163	-1,558	-3,300	-0,578	-0,120	-0,851	-0,227	-0,731	0,761			-2,015	0,261	-1,100	0,013
25	-0,625	0,020	-0,594	0,040	-2,564	-0,397	-1,460	-3,300	-0,572	-0,120	-0,896	-0,229	-0,705	0,727	-1,237	0,427				
26	-1,372	0,073	-0,396	0,028	-0,615	-0,233	-1,555	-3,300	-0,565	-0,125	-0,888	-0,227	-1,108	0,783			-2,212	0,320	-1,100	0,013
31	-0,337	0,002	-0,600	0,104	-2,747	-0,268	-1,699	-3,300	-0,594	-0,142	-0,934	-0,237	-0,516	0,635						
41	-1,033	0,025	-0,461	0,039	-1,929	-0,442	-1,635	-3,300	-0,601	-0,129	-0,877	-0,235	-0,498	0,633						
42	-0,488	0,006	-0,066	-0,045	-0,269	-0,142	-1,597	-3,300	-0,613	-0,144	-0,995	-0,247	-0,451	0,715			-1,301	0,046	-1,100	0,013
43	-1,144	0,089	-0,469	0,055	-1,009	-0,230	-1,367	-3,300	-0,563	-0,117	-0,921	-0,228	-0,741	0,495						
52	-0,148	0,007	-0,312	0,053	-0,618	-0,173	-0,971	-3,300	-0,519	-0,087	-0,895	-0,216	-0,619	0,865			-1,972	0,248	-1,100	0,013
53	-0,163	0,000	-0,288	0,063	-0,632	-0,218	-0,914	-3,300	-0,536	-0,089	-0,957	-0,224	-0,479	1,058	-0,111	-0,061				
54	-0,878	0,037	-0,202	0,032	-0,229	-0,194	-1,279	-3,300	-0,549	-0,112	-0,927	-0,225	-0,615	1,013			-1,671	0,157	-1,100	0,013
72	-0,837	0,040	-0,272	0,044	-0,247	-0,205	-1,292	-3,300	-0,582	-0,114	-0,938	-0,234	-0,699	0,898			-1,449	0,090	-1,100	0,013
73	-0,587	-0,005	-0,232	0,017	-0,358	-0,165	-1,139	-3,300	-0,514	-0,100	-0,912	-0,216	-0,644	0,744	-1,962	0,813	-1,774	0,188	-1,100	0,013
74	-0,372	0,003	-0,209	0,013	0,007	-0,113	-0,707	-3,300	-0,485	-0,073	-0,935	-0,209	-0,765	0,934	-1,032	0,535				
82	-0,466	0,041	-0,210	-0,002	-0,266	-0,187	-1,048	-3,300	-0,563	-0,098	-0,949	-0,230	-0,550	0,672	-1,151	0,469	-2,117	0,292	-1,100	0,013
83	-1,311	0,004	-0,356	0,067	-0,523	-0,215	-1,071	-3,300	-0,516	-0,096	-0,915	-0,216	-1,118	0,863	-1,319	0,418				
91	-0,673	0,038	-0,267	0,051	-0,328	-0,203	-1,292	-3,300	-0,500	-0,106	-0,871	-0,210	-0,428	0,925	-1,684	0,961			-1,100	0,013
93	-0,629	0,045	-0,367	0,037	-0,261	-0,188	-1,153	-3,300	-0,547	-0,089	-0,801	-0,217	-0,116	1,247					-1,100	0,013
94							-0,706	-3,300	-0,501	-0,056	-0,716	-0,203								

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A3 Literature review of SOC storage opportunity costs

Study	Scope of the GHG assessment ²	Approach ³	SOC storing practices	Study area	Cost (€ tC ⁻¹ ⁴)	Amount of carbon stored in MtC yr ⁻¹ ⁵
Antle et al., 2001	Soil	EM	Conversion of arable land to grassland	Montana zone 1	80 – 415	0,15 – 1,35
			Continuous cropping system (no fallow)	Montana zone 1	20 – 105	1,86 – 4,32
			Conversion of arable land to grassland	Montana zone 2	85 – 440	0,46 – 0,75
			Continuous cropping system (no fallow)	Montana zone 2	25 – 115	0,88 – 2,15
Antle et al., 2007	Soil	EM	Conservation agriculture (wheat system)	Centre US	0 – 170	0 – 0,5
			Conservation agriculture (corn-soybean system)	Center US	0 – 170	0 – 0,7
			Reduction of fallow land	Center US	0 – 170	0 – 0,9

² Soil carbon sequestration only or complete GHG balance.

³ Methodology based on a mathematical programming model (MP), an econometric model (EM), a partial/general equilibrium model (PEM, GEM), an "engineering" type model (ING). These approaches can be coupled or not with other types of models (e.g. biophysical).

⁴ In euros per ton of carbon. The ranges correspond to the costs associated with the different levels of carbon sequestration presented in the next column. These costs are calculated relative to a reference situation in which agricultural practices remain unchanged. Dollar-Euro exchange rate of sept. 2018.

⁵ In million tons of carbon per year (unless otherwise noted in parentheses). This is the additional amount of carbon sequestered in the studied area compared to the baseline scenario.

De Cara et al., 2006	GHG	MP	No-till or reduced tillage	EU 17	25 – 125	2 - 7
Feng et al., 2006	Soil	EM	Conservation agriculture	Iowa	0 – 170	0 – 5
			Land retirement	Iowa	0 – 170	0 – 20
Frank et al., 2015	Soil	PEM	No-till or reduced tillage	Europe 27	10 – 100	2 - 10
Kragt et al., 2012	Soil	MP	Introduction of grasslands in crop rotations	Western Australia	5 – 132	0,01 – 0,04
Lubowski et al., 2006	Soil	EM	Afforestation	US	0 – 105	0 – 750
McCarl et al., 2001	Soil	PEM	Conservation agriculture	US	12	60
			Afforestation	US	46	200
Paustch et al., 2001	Soil	EM	Conservation agriculture	Iowa	0 – 515	0 – 2
Pellerin et al., 2013	GHG	ING	No tillage	France	2	1
			Introduction of cover crops	France	43	0,3
			Agroforestry and hedgerows	France	4	0,4
			Extension of the duration of temporary grasslands	France	60	0,4
Plantinga et al., 1999	Soil	EM	Afforestation	Maine	0 – 95	0 – 5
			Afforestation	South Carolina	0 – 40	0 – 16
			Afforestation	Wisconsin	0 – 65	0 – 60