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Comparing ecological and evolutionary variability within datasets

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9 Running Head: Comparing variation within datasets

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14

15 **ABSTRACT**

16 Many key questions in evolutionary ecology require the use of variance ratios such as
17 heritability, repeatability, and individual resource specialization. These ratios allow
18 researchers to understand how phenotypic variation is structured into genetic and non-
19 genetic components, to identify how much organisms vary in the resources they use or how
20 functional traits structure species communities. Understanding how evolutionary and
21 ecological processes differ among populations and environments therefore often requires
22 the comparison of these ratios across groups (i.e. populations, sexes, species). Inference
23 based on comparisons of ratios can be limited, however. Variance ratios can remain the
24 same across group despite very different values in the numerator and denominator
25 variances. Moreover, evolutionary ecologists are most often interested in differences in
26 specific variance components among groups rather than in differences in variance ratios
27 *per se*. Recommendations for how to infer whether groups differ in variance are not clear in
28 the literature. Using simulations, we show how questions regarding the estimation of
29 variance components and their differences among groups can be answered with Linear
30 Mixed Models (LMMs). Frequentist and Bayesian frameworks have similar abilities to
31 identify differences in variance components. However, variance differences at higher levels
32 of organization can be difficult to detect with low sample sizes. We provide tools to conduct
33 power analyses to determine the appropriate sample sizes necessary to detect differences
34 in variance of a given magnitude. We conclude by supplying guidelines for how to report
35 and draw inferences based on the comparisons of variance components and variance ratios

36 **SIGNIFICANCE STATEMENT**

37 Many critical questions in ecology and evolution use variance ratios, such as repeatability,
38 heritability, or individual resource specialization, to make inferences about ecological and
39 evolutionary processes. In many cases these inferences rely on the comparison of variance
40 ratios among datasets (populations, sexes, or environments). In this article, we show that
41 current approaches of drawing inferences about group differences from comparisons of
42 ratios are inappropriate because ratios can differ due to differences in the numerator,
43 denominator, or both. We investigated how questions regarding differences in variance
44 ratios and constituent variance components can be evaluated using Linear Mixed Model
45 approaches (LMMs) and provide guidance for appropriate sampling schemes under
46 different scenarios and discuss common pitfalls associated with estimation of differences in
47 variance component among datasets.

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49 Running Head: Comparing variation within datasets

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51 Keywords: Repeatability, animal personality, individual variation, mixed models, individual
52 niche specialization, functional traits

53

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57 Sciences at North Dakota State University.

58 **Conflicts of interest/Competing interests**

59 The authors declare no conflict of interest

60 **Availability of data and material**

61 All code and data for simulations is available on the Open Science Framework's project for
62 this article: <https://osf.io/5aw42/>

63 **Code availability**

64 All code and data for simulations is available on the Open Science Framework's project for
65 this article: <https://osf.io/5aw42/>

66 **Author contribution**

67 Each author contributed equally to the design, analysis and writing of the manuscript.

68 **Ethics approval**

69 Not applicable

70 **Consent to participate**

71 Not applicable

72 **Consent for publication**

73 Not applicable

74

75 INTRODUCTION

76 Our understanding of many evolutionary and ecological processes is underpinned
77 by an estimation of variance ratios (Table 1). For example, the reporting of repeatability
78 has become pervasive in behavioral studies as it summarizes the amount of variation in
79 behavior attributable to differences among individuals. Informally these differences among
80 individuals can be thought of as differences in their average behaviors. Repeatability then
81 can be interpreted as how much of the overall variation is attributable to individual
82 differences

83 Use of variance ratios like repeatability spans a broad swath of evolutionary ecology
84 (Table 1). This includes the most well-known variance standardized ratio: heritability, and
85 extends to interest in community ecology regarding the distribution of functional trait
86 variation expressed within versus among populations or species (Violle et al. 2012).

87 While useful for understanding the relative magnitude of variation, variance ratios
88 can be highly misleading when compared between groups (Houle 1992; Wilson 2018).
89 Comparisons of variance ratios are only narrowly interpretable because these ratios can
90 differ when numerators differ, when denominators differ, or when both differ. Indeed,
91 variance ratios can be equal despite having different numerators and denominators values.
92 Put another way, differences between groups in ratios like repeatability are not
93 informative as to absolute differences in the magnitudes of variation observed.

94 **Table 1** Examples variance ratios found in the the ecological and evolutionary literature

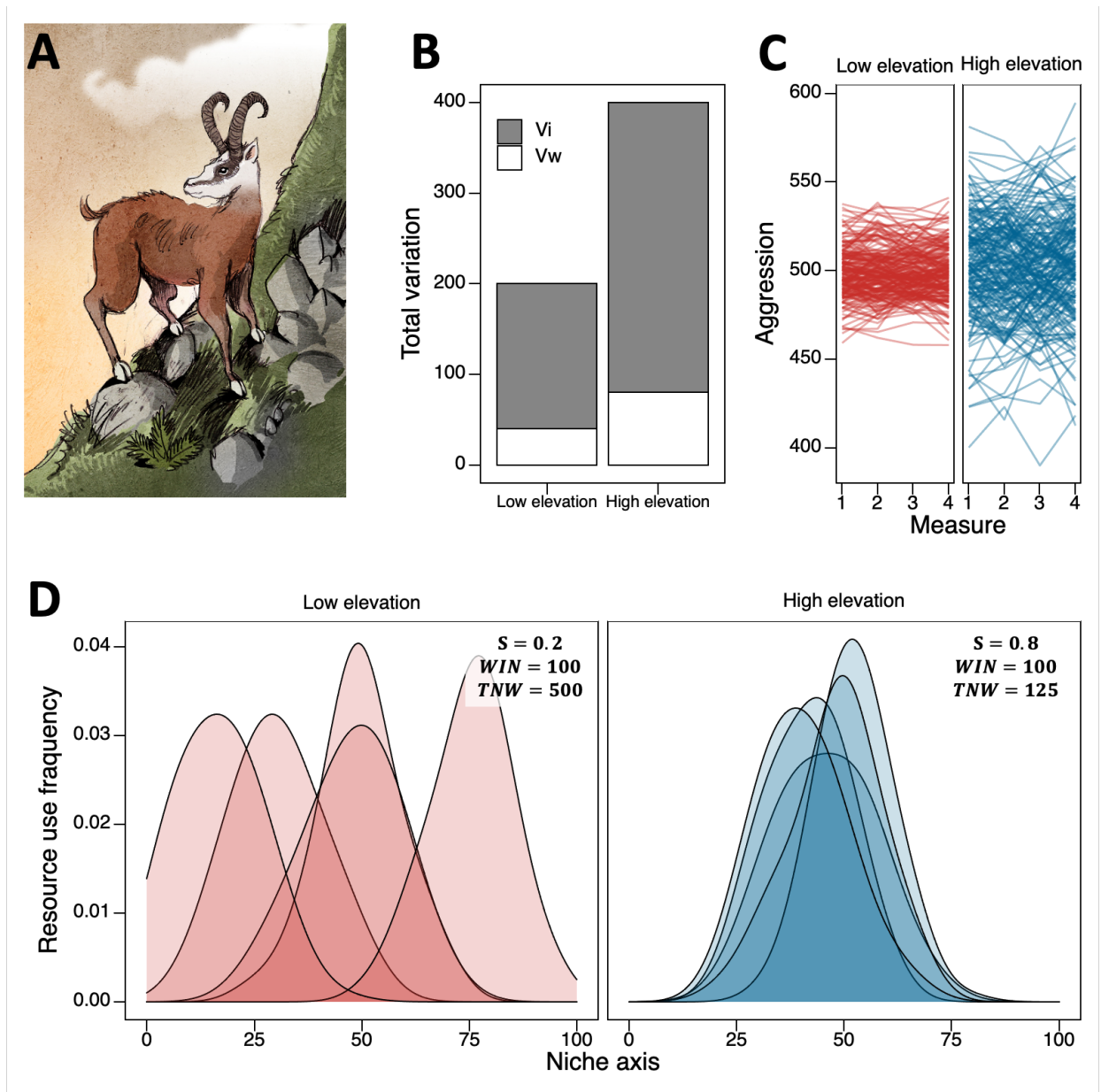
Discipline	Variance ratio	Definition	Description	References
Quantitative Genetics	<i>Heritability</i>	$h^2 = Va / Vp$	The proportion of variation attributable to additive genetic variance (Va)	(Mousseau and Roff 1987)
Behavioral Ecology	<i>Adjusted Repeatability</i>	$R_A = Vi / (Vi + Vw)$	The proportion of variation attributable to among-individual differences (Vi) relative to either the total variation ($Vi + Vf + Vw$) or after adjusting for fixed-effects ($Vi + Vw$)	(Lessells and Boag 1987)
	<i>Unadjusted Repeatability</i>	$R_U = Vi / (Vi + Vf + Vw)$		(Nakagawa and Schielzeth 2010)
Ecology	<i>Individual Niche Specialization</i>	$S = WIC / TNW$	The proportion of variation attributable to within-individual preference in niche (WIC) (usually expressed as standard deviations)	(Bolnick et al. 2002)
Community Ecology	<i>T-ratios</i>	$T_{IP/IC} = V_{IP} / V_{IC}$	The proportion of variation attributable to within-population variance (V_{IP}) relative to the community variance (V_{IC})	(Violle et al. 2012)
		$T_{IC/IR} = V_{IC} / V_{IR}$	The proportion of variation attributable to community variance (V_{IC}) relative to the regional pool variance (V_{IR})	(Violle et al. 2012)

95

96 Legend: Va : additive genetic variance in a trait, Vi : among-individual variance in trait, Vw : within-individual (i.e. residual)
 97 variance in a trait, WIC : within-individual variance in niche preference, TNW : Total niche width, T_{IP} : total amount of trait
 98 variation in a community, V_{IP} : within-population variance in trait, V_{IC} : community variance in trait, V_{IR} : regional pool variance

99 To further illustrate the inferential limits of variance ratios, consider the following
100 scenario: researchers are studying the behaviors and dietary habits of two populations of
101 the mythical Dahu (*Dahu desterus*; Fig. 1A) at different elevations. These elusive creatures
102 have shorter hind-legs on their left side, thus only allowing for clockwise movement
103 (Chartois and Claudel 1945; Jacquat 1995). While measuring aggressive interactions,
104 researchers find no differences in means between populations and similar behavioral
105 repeatabilities ($\tau = 0.8$; Fig. 1B). Put another way, the same relative amount of variation is
106 attributable to individuals in each population. The researchers notice, however, that there
107 are large differences in the among- and within-individual variances of each population. Had
108 researchers only examined repeatabilities and mean differences they would
109 inappropriately conclude that the populations are behaviorally equivalent. Instead, the
110 actual variance estimates reveal that individuals from the high-altitude population are very
111 distinct from one another in their aggressive tendencies while, at low-altitude, individuals
112 show little departure from the population average (Fig. 1B, C).

113 These researchers are also curious as to whether the harsher climate at the top of
114 the mountain range leads to a narrower dietary breadth. Researchers predict that
115 individual resource specialization will be higher in the low elevation population, as *D.*
116 *desterus* have more food options to choose from. To the researcher's surprise, they find
117 much higher individual resource specialization in the high-altitude population: $S_1 = 0.2$, $S_2 =$
118 0.8 . Upon examining the specific values of among- and within-individual variation in niche,
119 they find that these differences are a result of the high elevation population having a much
120 narrower total niche width (Fig. 1D) while the within-individual variation in niche
121 preference is equal between populations.



123

124 **Fig. 1** Reliance on variance ratios can lead to misleading inferences. (A) The elusive Dahu
 125 (*Dahu dexterus*) in its natural environment. (B) Two populations of Dahus living at different
 126 elevations do not differ in their repeatability of aggressive interactions (τ). (C) By plotting
 127 the individual aggression scores over the course of multiple measurements, it is clear that
 128 individuals are more distinct in their aggressive behavioral strategies at high elevation.
 129 This inference cannot be made by investigating repeatability alone. (D) The two population
 130 have very different resource specialization indices (S). A more accurate inference is that
 131 individuals do not differ in niche width (WIN), it is instead the total niche width (TNW) that
 132 is narrower in the high-altitude population. Code available here: <https://osf.io/5aw42/>.
 133 Illustration: [Philippe Semeria](#) (CC BY 3.0 license)

134 This means that it is the difference in diet preference among individuals that drives
135 the difference between the two populations. With more varied resources available at low
136 elevation, each individual can specialize along the total niche axis, yet the breadth of diet
137 preference within-individuals is the same between populations.

138 For both traits, exclusive reliance on ratios would have led to either inappropriate
139 or incomplete inferences (i.e. inappropriately concluding behavioral equivalence and
140 incompletely recognizing the basis of differences in apparent specialization). Due to these
141 problems with interpretations of variance ratios (Houle 1992; Dochtermann and Royauté
142 2019), what would be of greater use to researchers is to instead evaluate differences in
143 specific variance components.

144 *A statistical framework for comparing variance components*

145 The statistical procedures necessary for the estimation of variance components and ratios
146 within a single population have been the subject of much attention (e.g. mixed models for
147 repeatability: Dingemanse and Dochtermann 2013; animal models for heritability: Wilson
148 et al. 2010; individual niche specialization: Bolnick et al. 2002; Coblenz et al. 2017;
149 functional trait variation: Nakagawa and Schielzeth 2012; Violle et al. 2012; Carmona et al.
150 2016). There is also a long history in quantitative genetics regarding the comparison of
151 variances and covariance structures among groups (Shaw 1991; Arnold and Phillips 1999;
152 Roff 2002; Roff et al. 2012; Aguirre et al. 2014). Unfortunately, these quantitative genetic
153 approaches have been poorly disseminated across fields (but see (Dochtermann and Roff
154 2010; White et al. 2020). Here we describe and investigate methods for detecting
155 differences in variance components amongst groups. Specifically, we compare the strength

156 and weaknesses of three statistical approaches: comparison of confidence intervals, model
157 comparison with AIC, and Bayesian estimation of the difference in variance components.
158 While this selection of methods encompasses very different philosophical approaches to
159 data analysis, all three are routinely used in the estimation of repeatability and other ratios.

160 We consider a scenario where a phenotypic attribute, y , is measured repeatedly for
161 individual organisms occupying one of two different environments (E1 and E2) and in
162 which variation occurs among and within-individuals (V_I and V_W respectively). In the
163 following sections we focus on differences in individual variation and repeatability. Note,
164 however, that this scenario can also be expanded to the comparison of diet specialization
165 for individuals occupying different environments or how functional traits vary among and
166 within species in two different environments.

167 An easy way to compare these variance components and their ratios ($\tau = V_I/(V_I +$
168 $V_W)$) is to estimate the variance components for each environment in separate statistical
169 models. We can then test for differences in variances and ratios by environment based on
170 whether estimate confidence intervals overlap or not. While straightforward, this method
171 suffers from two key limitations. First, basing inference on the overlap of 95 % confidence
172 intervals is overly conservative (Barr 1969), especially when sample size is low. It is
173 instead whether the confidence interval for the *difference* in variances excludes 0 that is
174 relevant for drawing inferences. This difference cannot be directly estimated from the
175 approach we have described. However, statistical significance can still be assessed by
176 comparing the overlap of the 83% confidence intervals for variance components, a
177 threshold that provides a better approximation for an $\alpha = 0.05$ for the null hypothesis of no
178 difference (Schenker and Gentleman 2001; Austin and Hux 2002; MacGregor-Fors and

179 Payton 2013; Hector 2021). Second, by estimating variance components in separate
 180 statistical models, the hierarchical structure of the data, i.e. the variance components
 181 nested within the environments, has been broken. As a result, potential average differences
 182 in the traits of interest are not appropriately tested.

183 Instead, we suggest that a more appropriate procedure would be the use of a Linear
 184 Mixed Model (LMM) where the among- and within-individual variance is estimated for
 185 each environment within the same statistical model. This statistical model can be described
 186 by the following equation:

$$187 \quad y_{ij} = \beta_0 + \beta_1 \text{Environment} + ID_{0i} + e_{0ij} \quad (\text{equation 1})$$

$$188 \quad ID_{0i} \sim MVN(0, \Omega_{ID}); \quad \Omega_{ID} = \begin{bmatrix} V_{ID} E_1 & 0 \\ 0 & V_{ID} E_2 \end{bmatrix}$$

$$189 \quad e_{0ij} \sim MVN(0, \Omega_e); \quad \Omega_e = \begin{bmatrix} V_e E_1 & 0 \\ 0 & V_e E_2 \end{bmatrix}$$

190 where y_{ij} describes the phenotypic traits for the i th individual and j th observation. ID_{0i} , is
 191 the deviation from an overall intercept, β_0 , for the i th individual. β_1 represents the
 192 regression coefficient for the fixed effect of environment (here a contrast coefficient). The
 193 random intercepts and residual variance (e_{0ij}) both follow a multivariate normal
 194 distribution, and Ω_{ID} and Ω_e , are the variance-covariance matrices at the among- and
 195 within-individual levels respectively.

196 The diagonal elements of these matrices represent the among- and within-
 197 individual variances in each environment: E_1 and E_2 . The off-diagonal elements represent
 198 the cross-environment correlation (set to 0 if individuals are only ever evaluated in one of
 199 the two environments). This formulation has the advantage of allowing considerable
 200 flexibility in the specification of the statistical models considered (Dingemanse and

201 Dochtermann 2013). LMMs are now available for most statistical software and their
202 generalized extensions can accommodate non-normal error distributions (Table 2).

203 Upon fitting LMMs, several methods are then available to determine whether a
204 variance ratio or components of the ratio differ by environment. Specific hypotheses of
205 which variance component differs across environment can be easily tested via model
206 comparison. For example, a model where only the among-individual variance differs by
207 environment can be compared to a null model where the among and within- individual
208 variance are kept constant across developmental environments (Royauté et al. 2019).
209 These models can be estimated within a frequentist framework via restricted maximum
210 likelihood or a Bayesian framework and suitable decision criteria can be used to determine
211 which model best fits the data. In the case of restricted maximum likelihood estimation, it is
212 also possible to use likelihood ratio tests to compare these models. Note however that the
213 proper degrees of freedom to apply to each model is unclear and additional care should be
214 taken when using this method (Pinheiro and Bates 2006; Santostefano et al. 2016). We
215 recommend calculating these degrees of freedom by considering each variance component
216 as a full parameter for more conservative testing (see also the tutorial in ESM3).

217 **Table 2** Packages and softwares allowing to test for differences in variance components using Linear Mixed Models (LMM) along with
 218 parameter estimation method (maximum likelihood (ML), restricted maximum likelihood (REML), hierachical likelihood (H-ML) or
 219 Bayesian framework) and inference method (Likelihood Ratio tests (LRT), AIC, bootstrapping or credible interval for ΔV). This list is not
 220 comprehensive and is instead based on widely-used commercial softwares and R packages
 221

Package or software	Free or commercial	Estimation	Testing method	Among-unit variance by group	Residual variance by group	Distributions handled	Comments	Reference
ASREmL	Commercial	ML/REML	LRT, AIC, bootstrapping	Yes	Yes	Gaussian		(Gilmour et al. 2015)
SAS	Commercial	ML/REML	LRT, AIC, bootstrapping	Yes	Yes	Gaussian, Poisson, Binomial		SAS Institute Inc.
nlme	Free	ML/REML	LRT, AIC, bootstrapping	Yes	Yes	...		(Pinheiro and Bates 2006)
lme4	Free	ML/REML	LRT, AIC, bootstrapping	Yes	No	Gaussian, Poisson, Binomial		(Bates et al. 2015)
glmmTMB	Free	ML/REML	LRT, AIC, bootstrapping	Yes	Yes	...		(Brooks et al. 2017)
hglm	Free	H-ML	LRT, AIC, bootstrapping	Yes	Yes	...	Residual variance modelled as Gamma distribution	(Rönnegård et al. 2010)
R-INLA	Free	Approximate Bayesian	credible intervals for ΔV	Yes	Yes	...		(Lindgren and Rue 2015)
MCMCglmm	Free	Bayesian	DIC, credible intervals for ΔV	Yes	Yes	Gaussian, Poisson, Binomial		(Hadfield 2010)
brms	Free	Bayesian	WAIC, LOO, credible intervals for ΔV	Yes	Yes	...	Residual variance modelled as log-normal distribution	(Bürkner 2017)

222 In many cases, researchers are also interested in whether the difference in variance
223 components have a biologically meaningful effect. In other words, when asking questions
224 about whether variance components vary between environments, we are mostly interested
225 in the *magnitude of the difference* in these components across environments. While model
226 comparison of LMMs can help us understand whether a statistically detectable difference is
227 observable across environments, the magnitude of the difference can only be determined
228 by examining the difference in variance components among environment: ΔV estimated as
229 $V_{E2} - V_{E1}$ in our case. When the trait of interest is expressed as standard deviation units (i.e.
230 mean centered and scaled to the standard deviation of the dataset across all populations
231 and environments), this difference can be considered an effect size for the magnitude of the
232 difference among variance components, thus making comparisons across studies possible
233 (Royauté et al. 2015; Hamilton et al. 2017; Royauté and Dochtermann 2017). Note that ΔV
234 could also be expressed on a ratio scale (V_{E2}/V_{E1}) or on a log-additive scale ($\log(V_{E2}) - \log$
235 (V_{E1})). We will return to the topic of statistical significance vs. appropriate effect sizes later
236 in the paper. For now, we simply consider ΔV on an additive scale with data expressed in
237 standard unit deviations because it allows the most straightforward interpretation and
238 functions in cases where a variance component is zero or approaching zero. ΔV can be
239 calculated from the maximum likelihood estimates in a frequentist framework but
240 calculation of the uncertainty around this estimate is not straightforward and requires
241 additional steps such as bootstrapping. In a Bayesian framework, the calculations are much
242 simpler given that the distribution of ΔV can be directly estimated by taking the difference
243 in the posterior distribution of $V_{E2} - V_{E1}$. The posterior mode of ΔV can then be interpreted

244 as the estimated strength of ΔV , with credible intervals representing the precision around
245 this estimate.

246 In summary, approaches based on LMM and their generalized extensions allow
247 great flexibility and are well suited to study questions related to how variation in
248 phenotypic traits varies at multiple levels of organization. In the next section, we describe
249 the performance of LMMs to detect differences in variance components.

250

251 **METHODS**

252 The simulations described below focus on interpretation in the context of behavioral
253 repeatability. However, it is worth noting again that inferences about the ability to estimate
254 and detect differences in variances generalizes to the components of the ratios described in
255 Table 1.

256 *Data simulations*

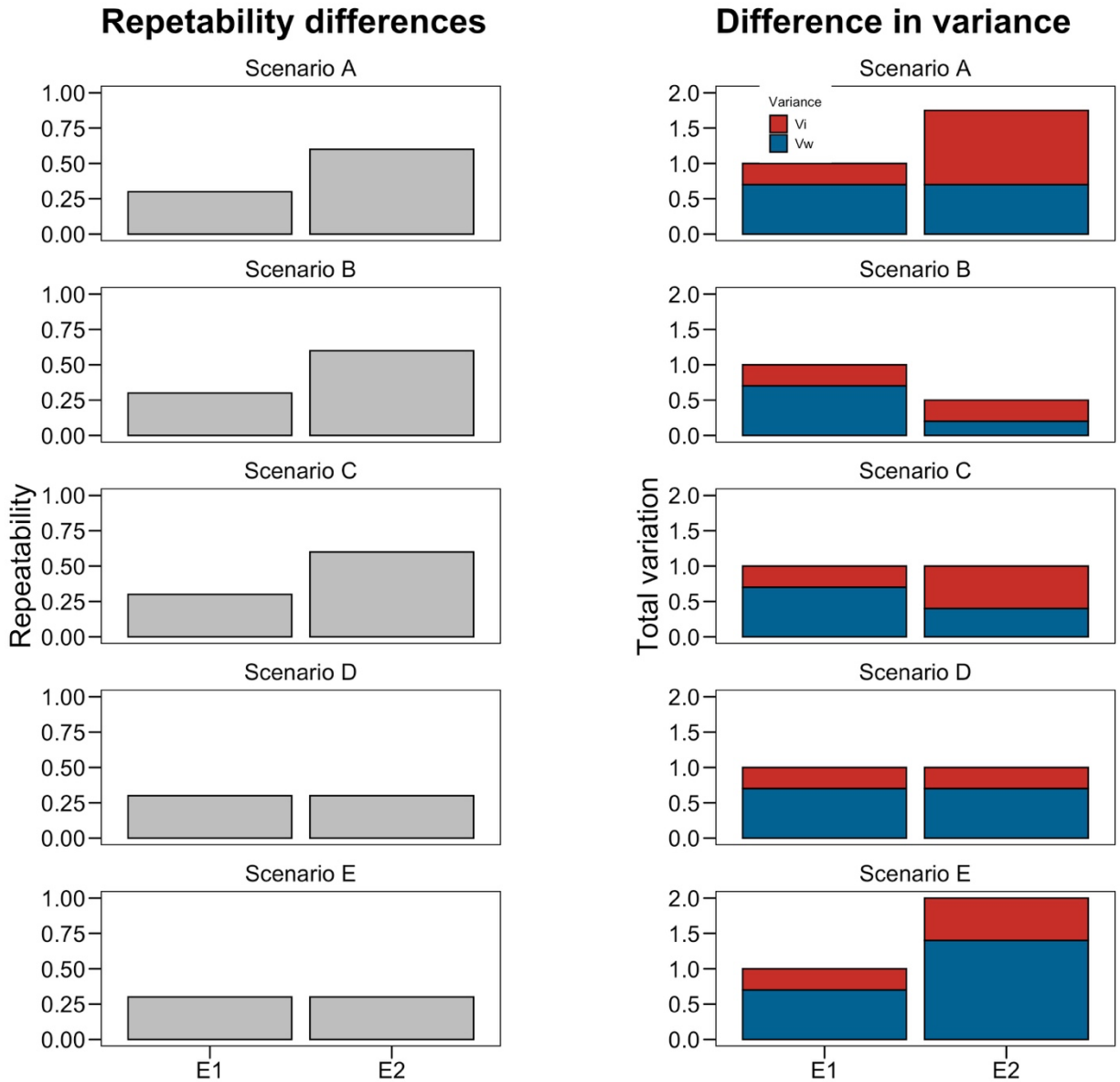
257 To compare the performance of statistical procedures for detecting differences in variance
258 components and variance ratios, we performed a series of simulations based on the
259 scenarios illustrated in Fig. 2. In these scenarios a phenotypic attribute y is measured in
260 two different environments (E1 and E2) and variation occurs among and within individuals
261 (V_I and V_W respectively). In scenarios A through C the repeatability (τ) differs by an equal
262 amount between the two environments ($\Delta\tau = 0.3$), but the underlying driver of this
263 difference is either due to a difference in the among-individual variance (A), in the within-
264 individual variance (B) or in both the among and within-individual variance (C). Note that

265 for scenario C, the total variance remains the same between environments. In scenarios D
266 and E, we explore cases where the variance ratios are equal among environment, either
267 because all variance components are equal as well (D) or in spite of differences in all
268 variance components (E) (see Table S1 for exact values for all parameters).

269 Using the R statistical environment (R Core Team 2020), we generated 500 datasets for
270 each of the following combinations:

- 271 • Sample size varying from 20 to 200 individuals by increments of 20 for each
272 environment (sample size was equal between the two environments)
- 273 • Number of repeated measures taken on each individual varying from 2 to 6
274 repeated measures by increments of 1
- 275 • Five different scenarios of known difference in variance ratios as described in Fig. 1
276 and Table S1.

277 Each dataset was simulated by sampling from a Gaussian distribution for the random
278 (among-individual values) and the error (within-individual) terms. This resulted in a total
279 of 125,000 datasets on which we tested three different statistical procedures to detect
280 differences in variance components and variance ratios. We provide all R code for data
281 generation and analysis in the Electronic Supplementary Materials (ESM1).



282

283 **Fig. 2** Scenarios used in simulations detailing how differences or lack of difference in
 284 repeatability (right-side column) can arise from different patterns in the underlying
 285 variance components (left-side column; exact values can be found in Table S1). Scenarios
 286 A-C correspond to cases where the total variation differs between two environments (E1
 287 and E2) due to differences in the among-individual variance (V_i , A), the within-individual
 288 variance (V_w , B) or both (C). Scenarios D-E indicate cases where the ratios remain constant
 289 across environments, because all variance components are identical (D) or in spite of
 290 variance component being different among environments (E)

291

292 *Comparison of confidence interval overlap from separate mixed models*

293 We first compared the overlap of 83 % confidence intervals for variance component when
294 estimated from separate linear mixed models. We specified one mixed model for
295 environment 1 and one for environment 2. These models are a simplified version of the one
296 presented in equation (3):

297 $y_{ij} = \beta_0 + ID_{0i} + e_{0ij}$ (equation 2)

298 $ind_{0i} \sim \mathcal{N}(0, V_{ID});$

299 $e_{0ij} \sim \mathcal{N}(0, V_e)$

300 The experimental units in the environment of interest are included as random effects and
301 no additional fixed effect are needed. Upon fitting these models, we computed 83 %
302 confidence intervals for the among and within-individual variance. Datasets where these
303 intervals did not overlap were considered as statistically different.

304 *Frequentist LMM with AIC model comparison*

305 Our second approach was to fit the LMM approach described above and test for the for the
306 significance of the difference in among- and within-individual variance using likelihood
307 ratio tests. Specifically, we specified four different mixed models corresponding to the four
308 different possibilities by which variance components may differ (Royauté et al. 2019;
309 Bucklaew and Dochtermann 2021):

- 310 • Model 1: a null model where the among (V_I) and within-individual variance (V_W)
311 was kept constant among environments.

- 312 • Model 2: a model where only the among-individual variance differs among
313 environments, while the within-individual variance is kept constant ($V_I \neq$ & $V_W =$)
- 314 • Model 3: a model where only the within-individual variance differs among
315 environments while the among-individual variance is kept constant ($V_I =$ & $V_W \neq$)
- 316 • Model 4: a model where both the among and within-individual variance were
317 allowed to vary among environments ($V_I \neq$ & $V_W \neq$)

318 For each dataset combination, we then compared each model's Aikake's Information
319 Criterion value (AIC). AIC allows the comparison of relative fit of statistical models and
320 models with lower AIC values indicate better support relative to competing models. These
321 simulations and this analytical framework are similar to previously used approaches (Shaw
322 1991; Jenkins 2011; Tüzün et al. 2017). These models were specified using the *nlme*
323 package for mixed models (Pinheiro and Bates 2006) using Restricted Maximum
324 Likelihood (REML).

325 *Bayesian LMM and difference in variance components*

326 We next fit a mixed model where variances among and within units were allowed to vary
327 between environments (as in model 4 described above) to each randomly generated
328 dataset. We calculated the posterior mode for the difference in variance components
329 (calculated as $\Delta V = V_{E2} - V_{E1}$) and estimated the 95 % credible intervals based on the
330 Highest Posterior Density of this distribution. 95 % credible intervals excluding 0 were
331 taken to indicate statistically detectable differences in variance components among
332 environments. All models were run with the *MCMCglmm* package (Hadfield 2010) using
333 default iteration settings to shorten computing time (13000 iterations, 3000 burn-in

334 iterations and thinning interval of 10 iterations). We used priors that were minimally
335 informative for the variance components (See ESM1 and ESM3 for prior specification and a
336 discussion on priors).

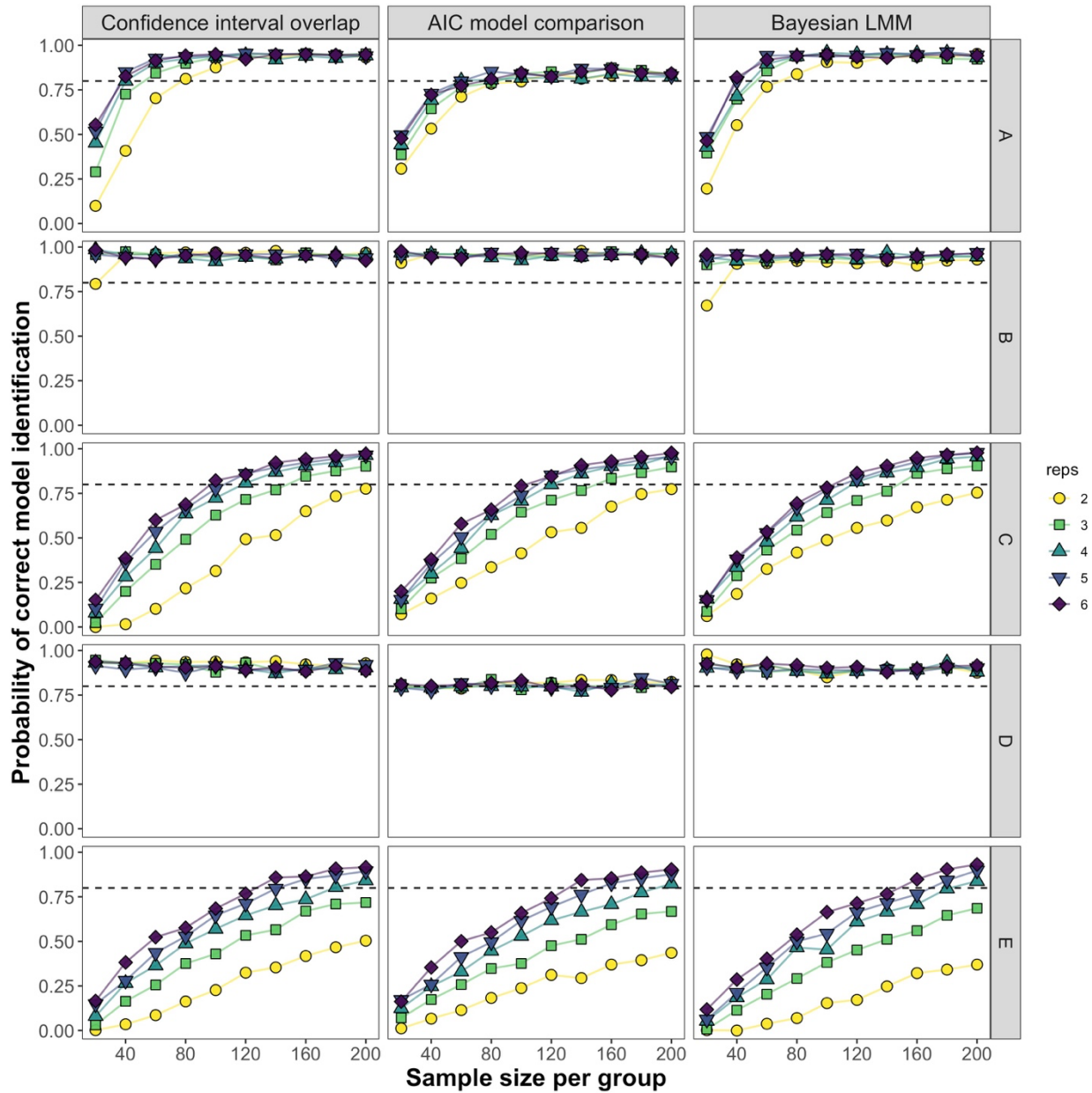
337 *Probability of correct model identification, precision, bias and accuracy estimations*

338 We calculated the probability of detecting the model with the correct difference in variance
339 components (hereafter “abridged” to probability of correct model identification), precision,
340 relative bias and accuracy under each scenario and sampling design to compare the
341 performance of maximum likelihood and Bayesian mixed models. For Method 1 (overlap of
342 83 % intervals), we assigned values of 1 when significant differences in variance
343 components were detected in directions predicted by the data generating process, and 0
344 otherwise. For Method 2, we calculated the probability of correct model identification as
345 the proportion of times the model with the lowest AIC matched the generating model. For
346 Method 3, we calculated whether a given model detected a difference in variance
347 components based on the overlap of the 95 % credible intervals of the ΔV posterior
348 distribution with 0. As in Method 1, we then assigned values of 0 or 1 based on whether the
349 detected difference matched with the data generation process of the corresponding
350 scenario. We calculated the probability of correct model identification as the proportion of
351 analyzed datasets in which we detected differences in the direction predicted by each
352 scenario and statistical method. Precision, indicating the similarity of the results produced
353 by simulations with a given scenario, was calculated as the difference between 25 % and 75
354 % quantiles of estimates (van de Pol 2012). To calculate the relative bias (in %) for each
355 statistical approach by scenario, we calculated the mean difference between the expected

356 value and the value observed in each of the 500 simulations. Finally, we report the root
357 mean square of error (RMSE) for each scenario and sample sizes. This metric calculates
358 how close estimates are to the expected values and serves as an estimate of the accuracy of
359 each statistical approach by scenario.

360 **RESULTS**

361 The probability of correctly detecting differences in variance components did not differ
362 substantially between frequentist and Bayesian methods of estimation (Fig. 3). The highest
363 probability of correct model identification was observed for cases where the variance ratio
364 differs as a result of changes to the within-individual variance (scenario B) or when
365 variation remained equal between environments (scenario D). The statistical power to
366 differentiate between alternative scenarios (i.e. scenarios A, C and E) was lower, especially
367 with small sample sizes and low number of repeated measures (Fig. 3). Importantly, no
368 statistical method seemed to outperform all others across scenarios. Our results are
369 consistent with previous simulations showing that the among-individual variance
370 component is particularly difficult to estimate at small sample sizes (Dingemanse and
371 Dochtermann 2013).



372

373 **Fig. 3** Effect of sampling design on the probability of correct model identification by
 374 scenario type and statistical modeling approach. Each point represents the probability of
 375 detecting the correct differences in variance averaged over 500 simulated datasets for a
 376 given sample size (n : number of individuals measured in each population, $reps$: number of
 377 repeated measures per individuals). A represents a scenario where only the among-
 378 individual variance (V_I) varies between environments, B represents a case where the
 379 within-individual variance (V_W) varies between environments, and both among and within-
 380 individual variance vary between environments in scenario C. In scenario D, all variance
 381 components are equal while in scenario E, variance components are different but variance
 382 ratios are equal across environments. Dashed lines correspond to 80 % threshold similar to
 383 recommendations for power analyses.

384 In scenarios B and D, the correct differences among variance components were
385 identified > 80 % of the time, even at low sample sizes (Fig. 3). In all other scenarios this
386 threshold was only reached with high sample sizes and a high number of repeated
387 measures. For scenarios C and E—which correspond to cases where the variance ratio
388 differs as a result of among-individual variance (C) or when the variance ratio remains the
389 same despite changes to both among- and within-individual variance (E)—datasets with
390 only 2 repeated measures per individual never achieved a probability of identifying the
391 generating model above 0.8, even with sample sizes above 200 units per environment (i.e. a
392 minimum of 800 total measurements, Fig. 3). Increasing the number of repeated measures
393 only marginally alleviated the problem. For example, in scenario C, only datasets with 4 or
394 more repeated measures per individual reached statistical power above 0.8 with sample
395 sizes above 120 individuals per environment, which is higher than many ecological or
396 evolutionary studies can provide under realistic scenarios.

397 Note that for AIC model comparison, we calculated power as the number of times
398 the best model corresponded to the generating model. A more conservative approach is to
399 calculate the proportion of times the best model is at least 2 AIC units lower than the
400 second model. This method corresponds to a common threshold to detect statistically
401 distinct models (Burnham and Anderson 1998). When using this more conservative
402 threshold (Fig. S1), datasets generated according to scenarios A and D were never
403 statistically distinguishable from non-generating models, although the correct model was
404 consistently ranked as the best model. This discrepancy is likely because when the
405 generating model does not include differences in the within-individual variability
406 (scenarios A and D), sampling error is erroneously identified as heterogeneity. At smaller

407 sample sizes this error is greater on average, and thus detectable. At larger sample sizes
408 this sampling error is smaller but more easily detected and therefore manifests as a
409 difference between groups. To address this, in addition to measures of variance differences
410 like the described ΔV statistic, researchers should also compare mean-standardized
411 variance estimates like the coefficient of variation or Houle's evolvability between groups
412 (Houle 1992; Hansen et al. 2011; Dochtermann and Royauté 2019).

413 The comparison of relative bias, precision, and accuracy among statistical methods
414 produced mixed results. On average, Bayesian LMMs consistently underestimated the
415 among-individual variance for scenarios in which the among-individual variance differed
416 between environments (scenarios A, C, and E) resulting in a bias at small sample sizes (Fig.
417 S2). However, Bayesian LMMs also had higher precision and accuracy compared to
418 maximum likelihood (Fig. S3, S4). This means that Bayesian estimates tend to be
419 consistently more conservative than maximum likelihood regarding the magnitude of the
420 among-individual variance but that these estimates nonetheless more closely matched
421 simulation conditions.

422 **DISCUSSION**

423 Comparing variability across datasets is important for many questions in evolutionary
424 ecology (e.g. Table 1). However, variance ratios are not sufficient to address questions
425 about how variation is expressed across environments, populations, or sexes. The inability
426 to determine why groups differ based on ratios is in addition to the numerous conceptual
427 and theoretical problems inherent to the estimation of variance ratios (Houle 1992; Hansen
428 et al. 2011). Instead, many questions require the direct comparison of variances.

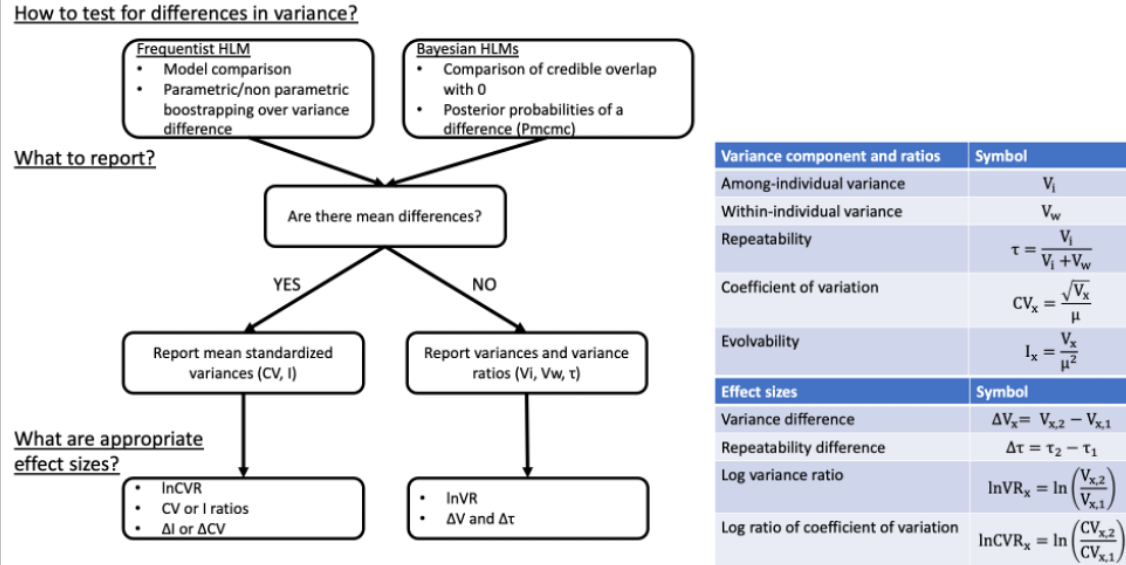
429 *What are appropriate sample sizes for detecting differences in variance?*

430 Our simulations show that regardless of the statistical methods used, comparing
431 variance components across groups is a “data hungry” question. Scenarios where the
432 among-individual variance differed between environments were particularly hard to detect
433 at low sample sizes. Note that our objective was not to provide a full exploration of
434 parameter space. Instead, we focused on a subset of scenarios that are likely to be common
435 in ecology and evolution (Fig. 2). Based on our simulations, the probability to detect
436 differences in variance components will depend in large part on the ability to estimate the
437 among-individual variance component (V_I). In the most complex case where differences
438 occur among and within-individuals (scenario E), researchers would require a minimum of
439 1,600 observations to correctly detect differences (i.e, 200 individuals measured 4 times in
440 each environment). This is far higher than sample sizes needed for single populations,
441 where moderate repeatabilities only need ~ 100 observations to be estimated with > 0.8
442 power (at least 25 individuals measured 4 times to detect a repeatability of 0.3; see
443 (Dingemanse and Dochtermann 2013).

444 Given these challenges, we recommend that researchers conduct power calculations
445 prior to the experiment whenever possible (see R code for *a priori* power analyses in ESM2
446 and an R Markdown tutorial in ESM3). If not, a simple rule for sampling can be to estimate
447 the sample size needed to detect the lowest among-individual variance value of interest
448 (see, for example, (Martin et al. 2011; van de Pol 2012; Dingemanse and Dochtermann
449 2013) and multiplying that sample size by the number of experimental groups involved.

450

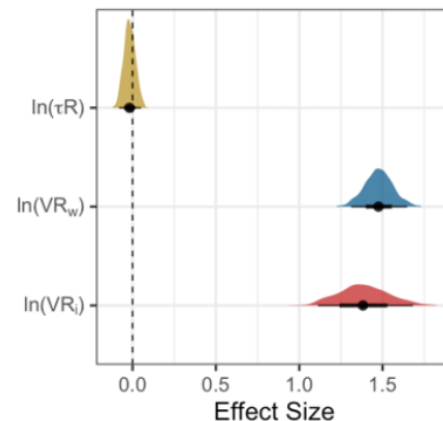
A



B

Model	Comparison	k	Log-Likelihood	AIC	ΔAIC
Model 1	V_i & $V_w =$	4	-6538.02	13084.04	383.84
Model 2	$V_i \neq$ & $V_w =$	6	-6493.45	12998.90	298.70
Model 3	$V_i =$ & $V_w \neq$	5	-6383.17	12776.34	76.14
Model 4	V_i & $V_w \neq$	7	-6343.10	12700.20	0.00

Variance and ratio	Environment	
	Low elevation Median [95% CI]	High elevation Median [95% CI]
V_i	165.96 [132.39; 199.65]	662.29 [540.31; 815.68]
V_w	38.26 [33.71; 42.64]	167.07 [148.77; 187.21]
τ	0.81 [0.77; 0.84]	0.80 [0.76; 0.84]
Difference	Δ (High - Low) Median [95% CI]	Pmcmc
V_i	493.89 [370.25; 648.69]	1.00
V_w	128.76 [107.57; 148.4]	1.00
τ	-0.01 [-0.06; 0.04]	0.69



451

452 **Fig. 4** A) Flowchart showing decision rules regarding how to test for differences in variance
 453 components, which metrics to report and which effect sizes can be calculated, along with their
 454 definitions in table format. B) Reporting example based on the simulated case study in Fig. 1B, C.
 455 The first Table used REML model selection with AIC to compare the support for different
 456 hypotheses for how variance components of aggression may differ between the low and high
 457 elevation populations. The best model is one where among and within-individual variances are
 458 higher in the high elevation population. The second Table compares all components by
 459 environment (posterior medians and 95 % credible intervals estimated from a Bayesian mixed
 460 model with model 4, note that frequentist confidence interval can also be reported using non-
 461 parametric bootstrapping as shown in ESM3). Finally, because aggression does not differ on
 462 average between populations, InVR is an appropriate metric to report the effect size for the
 463 difference in variance between populations.

465 Given the issues discussed above, how should researchers interested in ecological
466 and evolutionary variation design their studies and report their findings? We suggest that
467 researchers report their results in a manner that focuses on the magnitude of the difference
468 in variability between experimental groups rather than solely focus on statistical
469 significance.

470 To this effect, we believe that reporting the results of the full model rather than just
471 the most parsimonious model will be most appropriate in most cases (i.e. model 4 in our
472 conceptual example). This is because model selection only gives information on whether
473 differences among groups are statistically detectable. In contrast, questions regarding the
474 magnitude and precision of the estimated differences are answerable only with
475 interpretation of the most complete statistical model (see tutorial in ESM3).

476 In addition to presenting results of the full model, we suggest that measures of effect
477 sizes for the differences in variance component also be presented. As reported above, ΔV
478 provides a simple metric to estimate the magnitude of these differences, but it is by no
479 mean the only one. In our theoretical example, the mean trait value did not differ by
480 environments, but in many cases mean and variance are related. In such cases, using
481 comparisons based on Houle's (1992) I^2 value or coefficients of variation for each
482 component as opposed to variance component themselves can be preferable (Hansen et al.
483 2011; Dochtermann and Royauté 2019). Effect sizes based on the coefficient of variation
484 can also be calculated within an LMM framework as described by (Nakagawa et al. 2015)
485 (see also (Carmona et al. 2016; Fontana et al. 2018) for approaches relevant to functional
486 trait diversity).

487 We provide a synthetic guide for which statistical tests and effect sizes are most
488 appropriate depending on the nature of the dataset in Fig. 4A. Returning to our dahu
489 example, an appropriate analysis of the difference in aggression variance would follow the
490 tables and figures from Fig. 4B. Here the repeatability is unchanged between environments
491 (posterior median [95 % credible interval]; $\Delta\tau = -0.01 [-0.06; 0.04]$, probability of
492 difference: $P_{mcmc} = 0.68$). However, the high-elevation population shows significantly
493 higher variation among and within-individuals ($\Delta VR_I = 493.89 [370.25; 648.69]$, $P_{mcmc} =$
494 1.00 ; $\Delta VR_W = 128.76 [107.57; 148.40]$, $P_{mcmc} = 1.00$). This difference is also biologically
495 relevant since the effect sizes are also > 1 ($\ln VR_I = 1.38 [1.10, 1.66]$; $\Delta VR_W = 1.48 [1.31,$
496 $1.64]$). Biologically, this means that the high elevation population is composed of
497 individuals that are more distinct in behavior compared to the low elevation population.

498 While we limited our conceptual example to comparisons between two
499 environments, the LMM approach we propose is by no mean restricted to two-groups
500 comparisons. For example, Jenkins (2011) used model comparison to tease apart the
501 relative influence of sex, species and their interaction on the expression of behavioral
502 variation in kangaroo rats. Similarly, (Coblentz et al. 2017) show how model selection
503 combined with Bayesian GLMM can allow the comparison of indices of diet specialization
504 within and among species. In both cases, model selection can provide a first pass at
505 whether differences in variance components are detectable among groups, while specific
506 pairwise comparisons of effect sizes (using ΔV or other metrics) will allow discernment of
507 the most pronounced differences in variance component. Regardless of the statistical
508 approach used, we suggest it is important that researchers clearly outline the direction and,
509 when possible, magnitude of the expected effects in their predictions.

510 Finally, our conceptual examples focus exclusively on the case of “well-behaved”
511 data with normal error distributions. While these comparisons can be made with
512 generalized extensions to LMMS (i.e. GLMMs), researchers must take extra precautions
513 when calculating and comparing the within-individual variances (i.e. the residual variance).
514 Indeed, in the case of non-Gaussian data, the residual variance depends on both the link
515 function used and how the software deals with overdispersion (additive vs. multiplicative
516 overdispersion). (Nakagawa and Schielzeth 2010)) provides a very useful and extensive
517 guide explaining how the correct residual variation can be calculated.

518 **CONCLUSIONS**

519 Variance ratios are straightforward metrics to describe how various ecological and
520 evolutionary processes occur. However, comparing these ratios across studies or group can
521 be misleading if poor attention is given to the specific variance components making up
522 those ratios. More importantly, a lack of difference in these ratios does not mean that
523 variation is expressed equally among groups. Given these limitations, we advocate for
524 techniques allowing the estimation of differences in each variance components rather than
525 focusing solely on variance ratios. The statistical tools allowing comparison of trait
526 variation have become increasingly sophisticated and now allow asking very precise
527 questions. Specifically, we can now ask how trait variation is generated and how variation
528 differs among groups. However, despite the availability of these tools, researchers
529 interested in ecological and evolutionary variation must remain careful in their study
530 designs. As our simulations show, scenarios involving differences in among-individual
531 variance are particularly difficult to detect without substantial sample sizes. Finally, we

532 hope the statistical approaches and tools for power analysis presented here will allow for
533 appropriate comparisons of trait variation in ecological and evolutionary studies.

534

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664

665 **Supporting Information (SI) and Electronic Supplementary Materials (ESM)**

666 **ESM 1** Raw data from simulations along with R code for data analysis and figures

667 (<https://osf.io/5aw42/>)

668 **ESM 2** R code for conducting *a priori* power analysis (<https://osf.io/5aw42/>)

669 **ESM 3** R tutorial for comparing variance components using *nlme*, *MCMCglmm* and *brms*

670 packages (<https://osf.io/5aw42/>)

671 **Table S1** Scenarios tested in simulations to estimate the power to detect differences in

672 variance components of varying magnitude

673 **Fig. S1** Effect of sampling design on the probability to detect differences in variance

674 components by scenario type and statistical modeling approach with $\Delta AIC > 2$ threshold for

675 model comparison

676 **Fig. S2** Effect of sampling design on relative bias by scenario type and statistical modeling

677 approach

678 **Fig. S3** Effect of sampling design on estimate precision (width of the interquartile interval)

679 by scenario type and statistical modeling approach

680 **Fig. S4** Effect of sampling design on model accuracy (estimated as the root mean square of

681 error, RMSE) by scenario type and statistical modeling approach