



HAL
open science

Soil ingestion, a key determinant of exposure to environmental contaminants. The case study of chlordecone exposure in free-range pigs in the French West Indies.

Claire Collas, Jean-Luc Gourdine, David Beramice, Pierre-Marie Badot, Cyril Feidt, Stefan Jurjanz

► To cite this version:

Claire Collas, Jean-Luc Gourdine, David Beramice, Pierre-Marie Badot, Cyril Feidt, et al.. Soil ingestion, a key determinant of exposure to environmental contaminants. The case study of chlordecone exposure in free-range pigs in the French West Indies.. *Environmental Pollution*, 2023, 316 (1), pp.120486. 10.1016/j.envpol.2022.120486 . hal-04077964

HAL Id: hal-04077964

<https://hal.inrae.fr/hal-04077964v1>

Submitted on 23 Feb 2024

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 **Soil ingestion, a key determinant of exposure to environmental contaminants. The case**
2 **study of chlordecone exposure in free-range pigs in the French West Indies.**

3

4 Claire Collas ^{a,*}, Jean-Luc Gourdine ^b, David Beramice ^c, Pierre-Marie Badot ^d, Cyril Feidt ^a,
5 Stefan Jurjanz ^a

6

7 ^a *Université de Lorraine, INRAE, URAFPA, 54000 Nancy, France*

8 ^b *INRAE Centre Antilles-Guyane, UR-ASSET, 97170 Petit-Bourg, Guadeloupe, France*

9 ^c *INRAE Centre Antilles-Guyane, UE-PTEA, 97170 Petit-Bourg, Guadeloupe, France*

10 ^d *Université de Bourgogne Franche-Comté-CNRS, Laboratoire Chrono-environnement,*
11 *25030 Besançon Cedex, France*

12

13 * Corresponding author.

14 *E-mail address:* claire.collas@univ-lorraine.fr (C. Collas).

15

16

17 ¹Non-standard abbreviations sorted in alphabetical order

18

¹ BW: body weight, CLD: chlordecone, CR: Creole breed of pigs, DM: dry matter, FM: fresh matter, HP: high pasture paddock, LP: low pasture paddock, LW: Large White breed of pigs, MRL: maximum residue limit, MW: metabolic weight, SP: sweet potato paddock

19 **ABSTRACT**

20

21 Ingested soil may expose free-range animals to environmental pollutants. In pigs, soil
22 ingestion is few described whereas their burrowing behaviour suggests that it could be high.
23 Although highly productive pigs are generally reared indoor, free-range farming is increasing
24 in view of ethical considerations for animal welfare and is a common practice for subsistence
25 agriculture systems. The experiment lasted 8 weeks (2 for adaptation, 6 for measurements)
26 with 24 growing pigs of Guadeloupean Creole (CR) or Large White (LW) breeds. Pigs were
27 assigned to 3 outdoor treatments: high pasture HP (>60 days of regrowth), low pasture LP
28 (35 days of regrowth), and sweet potato SP (sweet potato field). Titanium (soil marker) and
29 chromium (faecal output marker) contents of faeces, vegetation and soil samples were used to
30 estimate individual daily soil ingestions. The average, 10th and 90th percentiles were 440, 200
31 and 726 g of dry soil per 100 kg body weight, respectively, without significant differences
32 between the 3 outdoor treatments or the 2 breeds but with a significant period (i.e. week of
33 measurements) × treatment interaction ($P<0.001$). In the French West Indies, animals may be
34 exposed to chlordecone (CLD), a very persistent organochlorine insecticide. Simulations of
35 CLD tissue contamination due to ingestion of contaminated soil were carried out and
36 compared to the maximum residue limit. These results show that grazing management needs
37 to be adapted to effectively limit soil ingestion by pigs and the impact of a contaminated
38 environment on the sustainability of pig systems.

39

40 **Keywords:** soil intake; feeding behaviour; plant cover; exposure risk; soil-bound pollutants;
41 chlordecone

42

1. Introduction

Literature developed methodologies to show that free-range animals might ingest soil during exploration of outside runs or paddocks (Jurjanz et al., 2014; Mayland et al., 1975; Roberts and Longhurst, 2002) and to quantify it. Even if the ingestion of some invertebrates of the pedofauna can take place (Rose and Williams, 1983), the ingestion of soil is generally of little nutritional value, as the presence of soil in the digestive tract would decrease the digestibility of ingested diet (Jurjanz et al., 2014). In addition, the ingestion of soil has been shown to be a significant exposure pathway of free-range animals to environmental pollutants (Ayrault et al., 2016; Cooke et al., 1996; Johnsen and Aaneby, 2019; Rychen et al., 2013). Indeed, soil can keep over very long-time concentrations of heavy metals, organic pollutants or radionuclides (Comte et al., 2022). Therefore, the studies of soil ingestion are less motivated by nutritional or zootechnical questions but much more by environmental and food safety approaches. To the best of our knowledge, this question has been little studied in pigs where only one study gives quantifications for free-range lactating sows (Jurjanz and Roinsard, 2014).

Although some breeds of pigs are more often reared indoor, societal expectations in favour of animal welfare are encouraging farmers to provide outdoor areas for the animals, and this is a criterion in the specifications for organic farming if the control of African swine fever is met (Martínez Avilés et al., 2019). Moreover, pigs are reputed to their digging activity and to explore easily soil whenever they have an access to (Høøk Presto et al., 2008). Furthermore, pigs are frequently an element of self-catering agricultural systems as they are able to valorize different byproducts of the garden or the household and by consequence would efficiently participate in the food delivery for the family. These agricultural systems are of especial interest as they are more frequent in informal systems not covered by the

68 governmental monitoring programs to control exposure to environmental pollutions (Kagira
69 et al., 2010; Thutwa et al., 2020). Moreover, informal animal rearing would concern more
70 often the less well-off parts of the population, which have generally a more vulnerable health
71 status. The fact that soil may carry environmental pollutants in the food chain especially to
72 these populations strengthen the need to evaluate the degree of soil ingestion of pigs in such
73 informal rearing systems and how it can be limited.

74 In this frame, the Caribbean context has a higher importance as home reared pigs are part of
75 the traditional food habits and were frequently raised outside in a tropical climate
76 characterized by episodes of strong rainfall (Sousa Junior et al., 2014). Therefore, these free-
77 range systems can cumulate all risk factors: species with an elevated digging activity raised
78 on easily soiled vegetation due to humid climate. The presence of a sanitary crisis due to the
79 pollution of the environment by the organochlorine pesticide chlordane (CLD) would
80 reinforce the need to quantify the soil ingestion by free-range pigs and to find out
81 management tools to limit their soil ingestion. In Guadeloupe, as in many tropical and
82 subtropical regions (Robinson et al., 2011), swine production is based on a variety of farming
83 systems, including specialized industrial and landless farms with high pig density, or small
84 family farms with Creole and/or crossbreeds reared in low input conditions (including free-
85 range pigs or tethered to a tree) (Gourdine et al., 2021).

86 Therefore, the present work studies the ingestion of soil in free-range pigs raised on different
87 types of tropical vegetation covers in Caribbean conditions. Besides this main objective, the
88 study target to investigate potential differences in soil ingestion between the local Creole
89 breed in comparison to the European Large White breed.

90

91 **2. Materials and methods**

92

93 *2.1.Ethics statement*

94

95 All animal care handling techniques and procedures as well as the procedures for sampling
96 were approved by the Ethics Committee of French Guyana and West Indies n° 069 (Comité
97 d’Ethique en Matière d’Expérimentation Animale des Antilles et de la Guyane, CEMEAAG)
98 authorized by the French Ministry of Higher Education, Research and Innovation, under the
99 project number APAFIS#6070-2016070721289156v3. The experiment was performed at the
100 INRAE Experimental indoors facilities of PTEA (Plateforme Tropicale d’Expérimentation
101 sur l’Animal) and outdoor facilities of PEYI (Plateforme Expérimentale sur le végétal et les
102 agrosYstèmes innovants en milieu tropical) according to the certificate number A 971-18-02
103 of authorization to experiment on living animals issued by the French Ministry of
104 Agriculture.

105

106 *2.2.Experimental design and animal management*

107

108 A total of 24 growing pigs of two different breeds (12 Creole CR and 12 Large White LW)
109 were used on the experimental facilities of INRAE in Guadeloupe (GPS 16°12’13” N,
110 61°39’24”W). The experimental design consisted of three outdoors treatments: high pasture
111 (HP, pasture with grass more than 60 days of regrowth age); low pasture (LP; pasture with
112 grass at 35 days of regrowth age) and sweet potato (SP; sweet-potato field). The 24 pigs were
113 randomly selected at 13 weeks of age (20.6 ± 3.6 kg and 28.9 ± 5.2 kg BW, for CR and LW
114 pigs, respectively), from the same litters and were affected in each treatment to have half-
115 siblings represented in each treatment (4 CR and 4 LW pigs). All the pigs fed the same before
116 joining the different experimental treatments. The experiment began by 2 weeks of adaptation

117 of the pigs to experimental conditions (wk1-wk2), followed by 6 weeks of measurements on
118 animals, vegetation and soil (wk3-wk8) (Fig. SM1, supplementary material).

119 The total area for each treatment was 880 m² (10 x 88 m) of ferralitic alluvial soils, divided in
120 one plot of 100 m² (10 x 10 m; 12.5 m² per pig) for adaptation (electric fences, pastures for
121 HP and LP, and sweet-potato field for SP) and 6 plots of 130 m² (10 x 13 m²) for the
122 measurement periods. The 8 pigs per treatment grazed a plot for one week, from Monday
123 (day1) to Sunday (day7) (16.25 m² per pig per week). The plot was delimited by a mobile
124 electric fence powered by a solar battery. Every Monday morning (at about 07:00), the pigs
125 were moved to another plot. For each treatment, the pigs had access to two shades (1 m² per
126 pig) and to 2 barrels near the shades, containing 100 L of water each, which were filled twice
127 per day so that the pigs never run out of water, for drinking and for mud wallow. The HP and
128 LP pastures did not contain any cultivated crops, but only natural grasses. The SP field was
129 the results of a planting of sweet-potato cuttings of more than three nodes every 20 cm on a
130 row and 1.60 m between rows (corresponding to 33 000 cuttings per ha). The plots were
131 clean weeded before planting and one month after planting. In LP treatment, the plots were
132 previously mown to 3 cm 35 days before the introduction of the pigs, to ensure the same age
133 of regrowth and the same vegetation stage. The LP and HP diets were based on grass
134 available on the plot, cracked corn (containing <25 µg Ti·g⁻¹ DM and <20 µg Cr·g⁻¹ DM; 750
135 g FM·d⁻¹ from wk1 to wk4, and 1000 g FM·d⁻¹ from wk5 to wk8; DM = 87.8 % FM), and a
136 supplementation of 250 g FM·d⁻¹ of a protein feed (composed by 96% of soya bean meal and
137 0.6 % of chromium oxide; containing 100 µg Ti·g⁻¹ DM and 4215 µg Cr·g⁻¹ DM). The SP
138 diets were based on sweet-potato leaves and tubers available on the plot (the pigs had to dig
139 the soil to catch tubers), a limited amount of non-cultivated grasses available in the sweet-
140 potato field, cracking corn (400 g FM·d⁻¹) and the same supplementation (250 g FM·d⁻¹) of
141 protein feed than LP and HP treatments. The protein feed was given individually in feeding

142 boxes for all treatments in two daily meals: 150 g fresh matter (FM) in the morning and 100 g
143 FM in the afternoon.

144

145 *2.3. Sampling, measurements and analysis*

146

147 Pregrazing surface height of herbage sward of both HP and LP treatments were measured on
148 Monday morning *via* a rising plate herbometer (Fig. SM1, supplementary material). Grass
149 was sampled on the HP treatment plot (outside the experimental paddocks). Potato tubers and
150 aerial parts (leaves and stems) of potatoes were harvested on the SP treatment plot (outside
151 the experimental paddocks) and weighted to estimate the available biomass per surface unit.
152 The potato allowance was estimated in kg dry matter (DM) per period, as well as in g DM per
153 kg body weight (BW). Based on BW of each pig in each period, an average supply per CR
154 pig and per LW pig was estimated (CR pigs being lighter than LW ones, the calculation was
155 done separately for the two breeds). Samples of corn and protein feed were taken weekly. All
156 feedstuffs samples were dried (65°C up to constant weight) and crushed in a ball mill.

157 The behaviour of the pigs was studied twice during the experimental period, at the beginning
158 (the third day of wk3) and in the middle (the fourth day of wk7) of the experiment. During 24
159 continuous hours (from 06:00 to 06:00 of the next day), the physical (rest, feeding, rooting,
160 ...) and feeding (water, leaves, sweet potatoes, concentrate) activities were recorded every
161 five minutes. Animals were weighed each Monday morning during the transfer to the new
162 experimental barn on a scale (PM110, Maréchalle, Chauny). During the distribution of the
163 protein feed, an individual sample of faeces was taken from each animal from Tuesday
164 (day2) to Friday (day5), dried (65°C up to constant weight) and crushed in a ball mill.

165 Each morning (from Monday to Friday, at about 08:00), a standardized picture (from the
166 south-east corner) of each plot (HP, LP and SP) was taken with a digital camera.
167 Consequently, each plot within each treatment had a total of 6 pictures (from Monday to
168 Friday, day1 to day5, and the Monday after pigs had changed plot, called day8) (Fig. SM1,
169 supplementary material). These images were used to quantify the changes in monitor the
170 disappearance of the vegetation cover (see section 2.5).

171 Soil samples were taken once the animals have been moved to the next barn (15 cm, 3
172 elementary samples grouped to one sample of each experimental barn, i.e. each week). After
173 manual separation of gravel (>2 mm) and roots, soil samples were dried and crushed in a ball
174 mill. Titanium (in the form of titanium dioxide TiO_2) and chromium (Cr) were analysed in all
175 samples (vegetation, feedstuffs, faeces and soil) by SARM service of CNRS-CRPG
176 laboratory (Vandœuvre-lès-Nancy, France). The samples were dissolved by melting with
177 lithium metaborate and recovering the melt with diluted HCl. Melting was carried out in a
178 muffle furnace progressively raised to 1000°C (200°C/h). TiO_2 was measured by UV-visible
179 spectrophotometry (measurement at 470 nm of the titanium-chromotropic acid complex at pH
180 3.5) on an AGILENT CARY60 instrument. Cr was measured by flame atomic absorption
181 spectrometry on an AGILENT SPECTRAA 240FS instrument. Regarding Quality Control,
182 all measurements were validated by measuring certified reference materials included in the
183 analysis series. The reference materials analysed are BE-N, UB-N, SO-1, SO-2, BCR-
184 CRM060, GSS-7, CRM-055, BCR-CRM414, GXR-5 and GXR-2 (Jochum et al., 2005). For
185 Quality Assurance, the limit of detection (LOD) was 25 ppm for TiO_2 and 20 ppm for Cr. It
186 was calculated on the measurements of minimum 10 x experimental blank. Uncertainty
187 values vary depending on the content determined in each sample and are based on statistical
188 calculations extrapolated from measurements of reference materials and duplicates of the
189 experimental samples. For Cr the relative uncertainty is <20% for contents below 100 ppm,

190 and <5% for contents above 100 ppm. For TiO₂ the relative uncertainty varies from 10 to
191 20% for contents below 200 ppm, to <5% for contents above 1000 ppm.

192

193 *2.4. Evaluation of soil ingestion*

194

195 The ingestion of soil was estimated at each experimental week based on faecal samples taken
196 between Tuesday (day2) and Friday (day5). The method consisted of estimating the amount
197 of the internal soil marker TiO₂ excreted in faeces, by considering that all faecal TiO₂ was
198 originated from soil according to the very low TiO₂ levels found in plants and crops, 200
199 times lower than those found in soil (Table 1). Therefore, the daily faecal output was
200 estimated *via* the external marker Cr₂O₃. Indeed, the daily applied of chromium, obtained by
201 multiplying the amount of protein feed daily ingested (refusals were considered) by the
202 chromium content of the protein feed, was divided by the chromium concentration in the
203 faeces:

$$204 \text{ Daily faecal output [kg DM}\cdot\text{d}^{-1}] = \text{protein feed ingestion [g DM]} \times \text{Cr in protein feed } [\mu\text{g} \\ 205 \text{ Cr}\cdot\text{g}^{-1} \text{ DM}] / \text{Cr in faeces } [\mu\text{g Cr}\cdot\text{g}^{-1} \text{ DM}] \times 1000$$

206 Then, the daily soil ingestion was estimated *via* the concentration of the soil-specific internal
207 marker TiO₂ in faecal output and in soil:

$$208 \text{ Daily soil ingestion [g}\cdot\text{d}^{-1}] = \text{TiO}_2 \text{ in faeces } [\mu\text{g TiO}_2\cdot\text{g}^{-1} \text{ DM}] \times \text{faecal output [g DM}\cdot\text{d}^{-1}] / \\ 209 \text{TiO}_2 \text{ in soil } [\mu\text{g TiO}_2\cdot\text{g}^{-1} \text{ dry soil}].$$

210

211 *2.5. Statistical analyses*

212

213 Statistical analyses were performed using R software (version 4.1.1) (R Development Core
214 Team, 2020). Pregrazing sward surface heights were analysed in a linear model to compare
215 HP and LP treatments, period and treatment \times period interaction. The latter was not
216 significant and remove from final model. Tukey's *post-hoc* test was used for multiple
217 comparisons between the 6 experimental periods.

218 For the behaviour data, the data collected every five minutes were converted in time spent in
219 physical of feeding activities. Each behaviour trait (eating, rooting, ...) was analyzed
220 separately by robust analysis of variance (ANOVA) with the WRS2 package, using firstly a
221 three-way robust ANOVA: the effects of treatment (HP, LP vs. SP), breed (CR vs. LW) and
222 period (wk3 vs. wk7) and within-animal effects, due to repeated measurements (wk3 and
223 wk7). Pairwise group comparisons were performed within each breed \times treatment interactions
224 and when interactions were found not significant ($P>0.05$), two-way ANOVA was performed
225 and pairwise group comparisons were performed within each treatment (HP, LP and SP).
226 Pearson correlations between estimated ingestion levels and the time spent for rooting
227 activities were calculated to check the linear relationship between soil ingestion and rooting
228 activities.

229 The daily pictures of plots were analyzed using the method proposed by Mahieu et al. (2019).
230 The aim of this method is to discriminate defoliation patterns based on a pixel index, the
231 visible atmospherically resistant index for green matter (VARIgreen). As reported by Mahieu
232 et al. (2019), the pixel index values were distributed in three categories: i) category A
233 contained pixels associated with bare soil, mud and/or soil litter; ii) category B contained
234 pixels associated with stems and other senescent plant parts with low chlorophyll levels; and
235 iii) category C contained pixels associated with green matter. The percentage of pixels were
236 calculated for each category and were normalized with the arcsine of the square root of the
237 proportions. The transformed data were analyzed using ANOVA of the R package agricolae,

238 with the effects of treatment (HP, LP vs. SP), day (1 to 5 and 8) and the interaction between
239 day and category (A, B and C).

240 Animal variables (BW, faecal production, soil ingestion) were tested using linear mixed
241 models with type of paddock (HP, LP, SP), period, breed (CR, LW) and their two-by-two
242 interactions as fixed effects; and individual as a random effect (R package nlme). Individual
243 animals were used as the experimental unit. Type of paddock \times breed interaction was never
244 significant and remove from final models. Only the interactions of period with the type of
245 paddock, or with the breed, were retained in the final models when they were significant
246 ($P < 0.05$) or tended to be ($0.05 < P < 0.1$). The 'lsmeans' function (R package lsmeans) was
247 used for multiple comparisons (Tukey method for adjusting p.values).

248

249 **3. Results**

250

251 *3.1. Paddock characteristics*

252

253 Herbage and potato quality, as well as stocking rate, expressed in kg BW per m² per period,
254 are detailed in Table 1. Pregrazing sward surface heights were significantly higher on HP
255 than LP paddocks with an average gap of 12.9 cm ($P < 0.001$; Table 1). For both LP and HP
256 treatments, the sward heights in wk8 (27.0 cm) were significantly higher than the sward
257 heights in weeks 3 to 6 (19.3 to 21.9 cm; $P < 0.001$). Sward heights ranged from 14.2 to 21.2
258 cm and from 23.9 to 32.9 cm for LP and HP paddocks respectively. Treatment \times period
259 interaction was not significant.

260 The potato biomass on SP treatment was estimated to 38.4 and 18.7 kg DM per period for
261 tubers and leaves/stems respectively. This corresponded to a daily potato allowance of
262 approximately 20.1 and 10.1 g DM \cdot kg⁻¹ BW for tubers and leaves/stems respectively. In

263 relation to the BW of the pigs this would be equivalent to a daily allowance of 615 and 756 g
264 DM tubers per pig, and 300 and 368 g DM leaves/stems per pig, for CR and LW breeds
265 respectively.

266

267 *3.2. Pig behaviour and impacts on paddocks*

268

269 *3.2.1. Pig behaviour*

270

271 The interaction effects of breed with treatment and period on rooting, drinking and resting
272 activities were not significant ($P>0.20$). Rooting activities significantly increased from wk3
273 to wk7 in HP and LP treatment but remained the same in SP treatment (Fig. 1). Eating
274 activities were found to be different between breeds ($P <0.01$). Irrespective of the treatment,
275 CR pigs spent much more time eating than LW pigs. Eating (Fig. 2) and drinking (Fig. SM2,
276 supplementary material) activities in outdoor conditions significantly increased or trended to
277 increase from wk3 to wk7 (except for CR pigs in HP conditions and for pigs in SP conditions,
278 for eating and drinking activities, respectively). Consequently, time dedicated for resting
279 decreased from wk3 to wk7 (Fig. SM3, supplementary material).

280

281 *3.2.2. Soiling and depletion of the vegetation*

282

283 The ANOVA analysis of the percentage of pixels showed that the latest was not affected by
284 either treatment ($P>0.8$) or week ($P>0.6$). At the opposite, the percentage of pixels was
285 affected by the interaction between days of pasture and VARIgreen category ($P<0.001$).
286 Figure 3 illustrates that the proportion of pixels corresponding to green matter (category C)
287 but also to stems and senescent plant parts (category B) decreased with the day of pasture and

288 it is closed to zero after pasture (day 8). At the opposite, the percentage of pixels
289 corresponded to bare soil, mud and/or soil litter (category A) increased and the value was
290 closed to 100 % after moving pigs to the next plot (day 8). The percentage of pixels of C
291 category (green matter) dropped dramatically from the first to the second day of pasture and
292 inversely the percentage of pixels of A category increased. These results of the picture
293 analysis were in line with our observations (Fig. SM4, supplementary material).

294

295 *3.3. Soil ingestion*

296

297 *3.3.1. Influence of the type of paddock and the period*

298

299 There was a significant type of paddock \times period interaction for the four variables: BW
300 ($P<0.05$), daily faecal output ($P<0.05$), faecal TiO_2 content ($P<0.001$) and daily soil
301 ingestions ($P<0.001$; Fig. 4; Table 2). Expressed in g DM per pig per day, the means per
302 treatment and period ranged from 91.9 (wk4) to 130.6 (wk5), from 131.2 (wk7) to 262.1
303 (wk4), and from 80.3 (wk5) to 178.4 (wk8), for HP, LP and SP treatments respectively.
304 Regarding the individual values, daily soil ingestions expressed in g DM per pig ranged from
305 38.5 to 258.6, from 58.4 to 514.5, and from 35.2 to 310.8, for HP, LP and SP treatments
306 respectively. In LP treatment, two values were particularly high for one pig due to its high
307 faecal output and high TiO_2 contents in faeces. The latest were kept in the dataset as they did
308 not appear abnormal, but they pull the average of the LP treatment upwards (Fig. 4).
309 Irrespective of the period (wk3 or wk7), no significant correlations were found between
310 rooting behaviour and the level of soil ingestion (first behaviour data set during wk3: $r =$
311 0.18, $P=0.40$; second behaviour data set during wk7: $r = -0.34$, $P=0.10$).

312

313 3.3.2. *Influence of BW and breed*

314

315 Pig's BW was not different between the three treatments (i.e. the three types of paddock: HP,
316 LP and SP) (33.0 ± 1.4 kg, mean \pm SE) but was significantly higher for LW than CR pigs
317 ($P < .001$; Table 1). Pigs' BW increased during the six periods from 26.9 to 39.3 kg (on
318 average + 12.4 kg in 5 weeks). The effect of breed was only significant on daily soil ingestion
319 when the latter was expressed per animal and not per unit BW. The breed effect was therefore
320 confounded with the BW effect since CR pigs were lighter than LW ones. Indeed, BW was
321 on average 29.3 kg for pigs of CR breed and 36.7 kg for pigs of LW breed. During the
322 experiment, CR pigs have grown from 23.8 to 34.9 kg and LW pigs have grown from 30.0 to
323 43.6 kg. The BW difference between CR and LW pigs was on average 6.2 kg at the start and
324 8.6 kg at the end of the experiment. Daily soil ingestions were on average 116.2 and 162.9 g
325 DM per pig, and 410.5 and 469.2 g DM per 100 kg BW, for CR and LW breeds respectively
326 (Fig. 5). The 10th and 90th percentiles were 181.8 and 662.6 g DM per 100 kg BW for CR,
327 and 218.1 and 789.8 g DM per 100 kg BW for LW. Period \times breed and treatment \times breed
328 interactions were not significant for all variables tested ($P > 0.05$; Fig. 5; Table 3).

329

330 **4. Discussion**

331

332 *4.1. Soil ingestion in pigs and possible drivers*

333

334 We hypothesized that soil ingestion would increase from HP, LP to SP due to i) the
335 difference in pasture allowance between HP and LP; ii) the higher rooting activities in SP
336 than LP and HP since pigs had to root to find sweet-potato tubers. In contrast to our
337 expectations, soil ingestion levels did not differ between treatments, with a daily average of

338 440 g DM per 100 kg BW, while we observed higher rooting activities in SP than LP and HP
339 and higher pregrazing sward heights in HP than LP. A high variability in soil ingestions
340 between pigs and between periods was observed with average, 10th and 90th percentiles at
341 439.9, 200.2 and 726.3 g DM per 100 kg BW, respectively.

342 The BW of the animals remained similar between the three types of paddocks and the pigs
343 grew by about 353 g per day (i.e. +12.4 kg between the 1st and 6th week of measurements).
344 This growth is consistent with the references for these breeds fed with these types of diets in
345 our conditions (Gourdine et al., 2018), so it appears that the three types of paddocks allowed
346 the pigs to meet their nutritional requirements for maintenance and growth. Therefore, we
347 consider that the animals on the LP treatment were not nutritionally limited.

348 Although the 24-hour behavioural observations in weeks 3 and 7 showed that pigs in the SP
349 paddock spent more time for rooting activities, this did not result into higher soil ingestion.
350 The higher rooting activities on the SP treatment may be explained by the expression of
351 natural foraging behaviour reinforced here by the search for underground food in contrast to
352 the aerial grass available on the LP and HP treatments. Several hypotheses can be put forward
353 to try to explain the rooting and/or soil ingestion behaviour, such as the natural exploratory
354 behaviour of pigs (Studnitz et al., 2007), the search for nutritional elements as pedofauna,
355 minerals, trace elements or fibres. However, even if the nutrient intakes from the different
356 diets of the three treatments have not been quantified, the possibility of a restriction in
357 minerals or trace elements is unlikely.

358 According to the picture analysis of the dirty grass or bare soil evolution over a period, the
359 condition of the paddocks remained fairly similar between the three treatments but with a
360 clear degradation from one day to the next. From day 5 (i.e. the last day of faces collection),
361 the paddocks had a fairly high proportion of bare soil (85.1 %). On the morning of the 8th day

362 (i.e. the day the pigs were moved to a new paddock for a next period), the bare soil, mud
363 and/or soil litter represented almost the entire surface. Faeces were collected from day 2 to
364 day 5 of each period, so soil ingestion was assessed in the first half of each period when
365 sufficient vegetation was still available to observe differences between treatments if there had
366 been any.

367 The period effect on soil ingestion was not related to the increase in stocking rate over time,
368 although the increase in grazing pressure from one period to another may increase the impact
369 of trampling and soiling of the grass, but the highest soil ingestions not being concentrated in
370 the last periods. The first experimental period was the rainiest with 49 mm the day the
371 animals entered the first experimental paddock, and 45 mm the day before. These conditions
372 could explain at least in part the higher soil uptake observed in wk3, particularly in the LP
373 treatment. Such an effect of humidity on soil ingestion has been shown previously in sheep
374 (Abrahams and Steigmajer, 2003) and cattle (Collas et al., 2019). The consequences of the
375 heavy rainfall may be more impactful on the short grass of LP than on the HP and SP
376 treatments where the tall grass, or the aerial parts of the potatoes, can more easily attenuate
377 the "splash" effect linked to the soil projections on the grass (Dreicer et al., 1983; Hinton et
378 al., 1995). The results of this study suggest that the amplitude of soil ingestion is more related
379 to the natural rooting behaviour of pigs in comparison with other species, the variability in
380 climatic conditions and the time spent in each plot.

381

382 *4.2. Soil ingestion of pigs in comparison to other species*

383

384 This very first quantification in growing pigs showed average daily ingestion of soil between
385 116 and 171 g per animal. That is quite similar to these reported for cattle, also in Caribbean

386 systems (Collas et al., 2019), but the large different BW change their significations. Indeed,
387 the comparison of ingestions are generally carried out for growing animals by putting the
388 ingested amount in ratio to 100 kg of BW or for comparison between species relative to the
389 metabolic weight (MW, correspond to $BW^{0.75}$). In such a frame, the measured soil ingestion
390 of these free-range pigs of 368 to 548 g per 100 kg of BW are much higher than ingestions
391 reported in growing cattle in Caribbean systems of less than 100 g per 100 kg BW (Collas et
392 al., 2020, 2019; Jurjanz et al., 2017), or less than 200 g per 100 kg BW in growing broilers
393 (Jurjanz et al., 2015). By expressing per kg MW, cattle and broilers ingested generally not
394 more than 4 g DM per kg MW (Collas et al., 2019; Jurjanz et al., 2017, 2015), grazing horses
395 and lactating sows around 5 g $DM \cdot kg^{-1}$ MW (Jurjanz et al., 2021; Jurjanz and Roinsard,
396 2014), whereas these free-range pigs in tropical conditions ingest much more with 9 to 13 g
397 $DM \cdot kg^{-1}$ MW (Table 2). These comparisons confirm clearly that the natural digging
398 behaviour of pigs would result in a significantly higher soil ingestion than in other species
399 what would, by consequence, expose such rearing systems much more to pollutants in a
400 context of contaminated areas.

401

402 *4.3. Implications for animal exposure to soil-bound pollutants and human health – the* 403 *case of chlordecone*

404

405 During grazing, pigs ingest soil and this is enhanced by rooting activity. Although this natural
406 behaviour may allow ingestion of pedofauna and roots or mineral supplementation (Edwards,
407 2003), it also presents risks as pigs are exposed to the full range of contaminants that may be
408 present in the soil, which may have serious repercussions on their health and/or the safety of
409 the products. In the French West Indies, nearly one fifth of the agricultural soil in
410 Guadeloupe and two fifths of the agricultural soil in Martinique are at risk of CLD

411 contamination (Comte et al., 2022; Ministère de la Transition Ecologique, 2017). Due to its
412 strong persistence, this insecticide, used against the banana weevil (*Cosmopolites sordidus*)
413 until its ban in the 1990s, caused soil pollution and the contamination of water and
414 ecosystems. A major challenge for this territory, as for any territory affected by such a health
415 crisis, is to maintain ecosystem services, in particular the food produced, and the economic
416 activities that depend on them (Perrette et al., 2020).

417 The concentration of CLD in perirenal adipose tissue (reference at the slaughterhouse for
418 comparison with the maximum residue limit MRL) was simulated as a function of the daily
419 soil ingestion and the concentration of CLD in the soil (CLD_{soil}). The CLD concentration in
420 serum ($\mu\text{g/mL}$) was first estimated by dividing the maintenance dose by the clearance. The
421 maintenance dose is the amount of CLD daily ingested ($\mu\text{g CLD}\cdot\text{kg}^{-1}\text{ BW}$) obtained by
422 multiplying daily soil ingestion by CLD_{soil} . Daily soil ingestions (CR: 164.2, 72.7 and 265.0 g
423 DM per pig; LW: 246.3, 114.5 and 414.6 g DM per pig; average, 10th and 90th percentiles,
424 respectively) were calculated for each breed using values expressed per 100 kg BW, obtained
425 for CR and LW pigs in this study, and the average BW of free-range growing pigs (CR: 40
426 kg; LW: 52,5 kg; average between post-weaning and slaughtering BW considering linear
427 relation, with 20 and 60 kg for CR pigs, 25 and 80 kg for LW pigs, respectively). Different
428 CLD_{soil} from 0.01 to 0.1 $\text{mg}\cdot\text{kg}^{-1}$ were considered in these simulations. The clearance ($\text{mg}\cdot\text{kg}^{-1}$
429 1 per day) was obtained from Fourcot et al. (2020) for both breeds (average of the two breeds
430 and separately). Finally, the CLD concentrations in perirenal adipose tissue ($\mu\text{g per g of fat}$)
431 were obtained by dividing the serum concentration by the tissue partition coefficient (K_p).
432 The K_p in tissue is obtained by the ratio of concentrations between serum and tissue (Fourcot,
433 2020). Simulated results show that when CLD_{soil} is 0.1 $\text{mg}\cdot\text{kg}^{-1}$, CLD concentrations in
434 perirenal adipose tissue exceed the MRL set at 0.02 mg CLD per kg of fat regardless of breed
435 and soil ingestion. When CLD_{soil} is 0.05 $\text{mg}\cdot\text{kg}^{-1}$, only the 10% of animals ingesting the least

436 amount of soil complied for both breeds (with a concentration 2 times lower than the MRL).
437 When CLD_{soil} is $0.03 \text{ mg}\cdot\text{kg}^{-1}$, only the 10% of animals that ingest the most soil do not
438 comply for both breeds. For CLD_{soil} of 0.01 and $0.02 \text{ mg}\cdot\text{kg}^{-1}$, all animals comply (Table
439 SM5, supplementary material). These simulations on the compliance of animals at slaughter,
440 in relation to their ingestion of soil and the level of soil contamination, may contribute to
441 establishing thresholds for regulating soil use in agriculture (Li, 2020). Our results therefore
442 raise the question of the values to be considered in the risk assessment. Should we consider
443 the mean or median ingestion levels, the 90 or 95 percentiles or outliers? Since self-
444 consumption is a widespread practice in the French West Indies, this point is of crucial
445 importance for self-consumers, who can therefore be markedly overexposed. Free-range
446 livestock farming is a relatively common practice in the tropics in self-sufficient animal
447 husbandry and has many advantages in terms of economy, environment, animal welfare and
448 legislation (European organic regulation), so it is important to be able to secure these farming
449 systems against health risks. A precise determination of daily soil ingestion is therefore
450 necessary for the assessment of the health risk.

451 Our study shows that a significant amount of soil can be ingested by pigs on a daily basis. In
452 order to limit the exposure of pigs to environmental contaminants, especially persistent
453 organic pollutants, it is necessary to adapt grazing practices (i.e. stocking rate, optimal time
454 step for changing plots, to limit soil ingestion. Thus, soil quality (texture, structure and
455 composition), particularly the amount of soil pollutants, and climatic parameters has to be
456 considered in grazing management. Even if grass is initially available in the paddock, the
457 foraging behaviour of the pig can quickly remove the grass, leaving bare or even turned over
458 ground for the remainder of the time spent in the paddock (Edwards, 2003). Thus,
459 maintaining sufficient grass availability and/or increasing the supply of concentrate feed may
460 help to limit soil ingestion associated with foraging and rooting on pasture (Jakobsen et al.,

461 2015; Stern and Andresen, 2003). Rivera Ferre et al. (2000) observed a high frequency of
462 rooting in unring sows compared to nose-ring sows, but animal welfare implications limit
463 this practice as a recommendation to reduce soil ingestion (Horrell et al., 2001).

464 The present study focused on soil ingestion, which is an important route of exposure of
465 animals to environmental contaminants. However, other matrices can also contribute to
466 animal exposure such as food and water consumption. The biomass of sweet potato tubers on
467 the SP treatment plot was assessed which provides information on the amount offered to the
468 animals (i.e. 38.4 kg DM per paddock of 130 m² offered per week, which corresponds to 20.1
469 g DM/kg BW⁻¹ per day). From the pictures of the paddocks (Fig SM4, supplementary
470 material), it can be considered that the pigs consumed almost all the available potatoes, which
471 allows to roughly estimate the daily intake of potato tubers (i.e. 615 and 756 g DM per pig for
472 CR and LW breeds respectively) based on the average BW of the animals of both breeds. The
473 CLD concentration in potato tubers was estimated according to Achard et al. (2007) using a
474 proportional relationship with the CLD content of the soil and a coefficient adapted for
475 ferralitic soils such as the study plots. These data make it possible to evaluate the intake of
476 CLD via the ingestion of sweet potato tubers. Using the average soil ingestions obtained in
477 this study for the CR and LW pigs in the SP treatment, together with the CLD contamination
478 of the soil to be simulated, it is possible to estimate the intake of CLD via soil ingestion. For
479 soil at 0.05 mg CLD.kg⁻¹ for example, this would correspond to a CLD intake of 9.95 (CR)
480 and 12.17 (LW) µg, with 5.92 (CR) and 7.21 (LW) µg via soil, and 4.03 (CR) and 4.96 (LW)
481 µg via sweet potato tubers. The soil contribution to exposure would be 60% with this kind of
482 diet.

483 Drinking water can also expose animals to pollutants as CLD, especially for unmonitored
484 spring water or ponds. However, this route of exposure is easier to control if farmers can use
485 potable water drinking supply as a limit has been set at 0.1 µg CLD.L⁻¹. Considering a water

486 consumption of 10% BW (i.e. on average 0.29 and 0.37 L per day for CR and LW pigs
487 respectively under the conditions of this study) (Chambre d'Agriculture de Bretagne, 2021),
488 this would represent an intake of CLD up to 0.29 and 0.37 μg for CR and LW pigs
489 respectively in the case of water reaching the potability threshold of $0.1 \mu\text{g CLD.L}^{-1}$.

490

491 **5. Conclusions**

492

493 This study provided references on soil ingestion by free-range pigs by comparing two breeds
494 (Creole and Large White) and three types of outdoor paddocks (pasture with two different
495 grass availabilities and sweet potato field). The results show that daily soil ingestions are not
496 influenced by breed or paddock type in this experiment, but this type of management exposes
497 pigs to substantial soil ingestions about half a kg of dry soil per animal per day. When free-
498 range farming is carried out in an area where the soil is contaminated, soil ingestion levels
499 such as those observed in this study can result in contamination of pigs, and their tissues, at
500 values sometimes exceeding the MRL. In this context, adapting farming practices to limit pig
501 exposure to soil-bound pollutants, in particular persistent organic pollutants as some
502 phytosanitary molecules, appears inevitable to ensure the safety of animal products.

503

504 **Acknowledgements**

505

506 This work was supported by The French National Research Agency (ANR, project
507 INSSICCA; grant number: 16.CE21.0008.01) and the Region Guadeloupe (AGROECODIV
508 project and CPER CRB-PSA project). The authors thank warmly Bruno Bocage, Mélain
509 Bructer, Katia Benony and Laurent Dantec from PTEA Experimental Unit (INRAE), Harry
510 Archimède, Matthieu Desfonds, Dalila Feuillet and Valérieuse Calif from UR-ASSET

511 (INRAE), Pamela Hartmeyer and Agnès Fournier from URAFPA (Université de Lorraine,
512 INRAE), and Nadia Crini from Chrono-Environment Laboratory (Université de Franche-
513 Comté-CNRS).

514

515 **References**

516

517 Abrahams, P.W., Steigmajer, J., 2003. Soil Ingestion By Sheep Grazing the Metal Enriched
518 Floodplain Soils of Mid-Wales. *Environ. Geochem. Health* 25, 17–24.
519 <https://doi.org/10.1023/A:1021217402950>

520 Achard, R., Cabidoche, Y.M., Caron, A., Nelson, R., Dufeal, D., Lafont, A., Jannoyer, M.,
521 2007. Contamination des racines et tubercules cultivés sur sol pollué par la
522 chlordécone aux Antilles.

523 Ayrault, S., Catinon, M., Boudouma, O., Bordier, L., Agnello, G., Reynaud, S., Tissut, M.,
524 2016. Metal exposure in cows grazing pasture contaminated by iron industry: Insights
525 from magnetic particles used as tracers. *Environmental Pollution* 212, 565–573.
526 <https://doi.org/10.1016/j.envpol.2016.03.006>

527 Chambre d'Agriculture de Bretagne, 2021. Réglementation Fiche 1 - Réglementation et
528 besoins en eau. [https://opera-connaissances.chambres-
529 agriculture.fr/doc_num.php?explnum_id=170037](https://opera-connaissances.chambres-agriculture.fr/doc_num.php?explnum_id=170037)

530 Collas, C., Mahieu, M., Badot, P.-M., Crini, N., Rychen, G., Feidt, C., Jurjanz, S., 2020.
531 Dynamics of soil ingestion by growing bulls during grazing on a high sward height in
532 the French West Indies. *Sci. Rep.* 10, 17231. [https://doi.org/10.1038/s41598-020-
533 74317-0](https://doi.org/10.1038/s41598-020-74317-0)

534 Collas, C., Mahieu, M., Tricheur, A., Crini, N., Badot, P.-M., Archimède, H., Rychen, G.,
535 Feidt, C., Jurjanz, S., 2019. Cattle exposure to chlordecone through soil intake. *The*

536 case-study of tropical grazing practices in the French West Indies. *Sci. Total Environ.*
537 668, 161–170. <https://doi.org/10.1016/j.scitotenv.2019.02.384>

538 Comte, I., Pradel, A., Crabit, A., Mottes, C., Pak, L.T., Cattan, P., 2022. Long-term pollution
539 by chlordecone of tropical volcanic soils in the French West Indies: New insights and
540 improvement of previous predictions. *Environmental Pollution* 303, 119091.
541 <https://doi.org/10.1016/j.envpol.2022.119091>

542 Cooke, A.I., Green, N., Rimmer, D.L., Weekes, T.E.C., Wilkins, B.T., Beresford, N.A.,
543 Fenwick, J.D., 1996. Absorption of radiocaesium by sheep after ingestion of
544 contaminated soils. *Sci. Total Environ.* 192, 21–29.

545 Dreicer, M., Hakonson, T.E., Whicker, F.W., White, G.C., 1983. Investigation of the pathway
546 of contaminated soil transported to plant surfaces by raindrop splash. (UCRL--
547 88450). United States. Research Org.: Lawrence Livermore National Lab., CA
548 (USA); Los Alamos National Lab., NM (USA); Colorado State Univ., Fort Collins
549 (USA).

550 Edwards, S.A., 2003. Intake of nutrients from pasture by pigs. *Proc. Nutr. Soc.* 62, 257–265.
551 <https://doi.org/10.1079/PNS2002199>

552 Fourcot, A., 2020. Distribution et élimination de la chlordécone chez les animaux d'élevage –
553 modélisation des processus. Université de Lorraine.

554 Fourcot, A., Feidt, C., Bousquet-Mélou, A., Ferran, A.A., Gourdine, J.L., Bructer, M.,
555 Joaquim-Justo, C., Rychen, G., Fournier, A., 2020. Modeling chlordecone
556 toxicokinetics data in growing pigs using a nonlinear mixed-effects approach.
557 *Chemosphere* 250, 126-151.

558 Gourdine, J.-L., Bambou, J.-C., Giorgi, M., Loranger-Merciris, G., Archimède, H., 2018.
559 Performance of growing pigs reared indoors or outdoors in sweet-potato fields. *Rev.*

560 D'élevage Médecine Vét. Pays Trop. 71, 41–46. <https://doi.org/10.19182/remvt.31347>
561 70

562 Gourdine, J.-L., Fourcot, A., Lefloch, C., Naves, M., Alexandre, G., 2021. Assessment of
563 ecosystem services provided by livestock agroecosystems in the tropics: a case study
564 of tropical island environment of Guadeloupe. *Trop. Anim. Health Prod.* 53(4), 435.
565 <https://doi.org/10.1007/s11250-021-02880-3>

566 Hinton, T.G., Stoll, J.M., Tobler, L., 1995. Soil contamination of plant surfaces from grazing
567 and rainfall interactions. *J. Environ. Radioact.* 29, 11–26.
568 [http://dx.doi.org/10.1016/0265-931X\(95\)00008-X](http://dx.doi.org/10.1016/0265-931X(95)00008-X)

569 Høøk Presto, M., Andersson, H.K., Folestam, S., Lindberg, J.E., 2008. Activity behavior and
570 social interactions of pigs raised outdoors and indoors. *Arch. Anim. Breed.* 51(4),
571 338–350.

572 Horrell, R.I., A'Ness, P.J.A., Edwards, S.A., Eddison, J.C., 2001. The use of nose-rings in
573 pigs: consequences for rooting, other functional activities, and welfare. *Anim. Health*
574 10, 3–22.

575 Jakobsen, M., Kongsted, A.G., Hermansen, J.E., 2015. Foraging behaviour, nutrient intake
576 from pasture and performance of free-range growing pigs in relation to feed CP level
577 in two organic cropping systems. *Animal* 9, 2006–2016.
578 <https://doi.org/10.1017/S1751731115001585>

579 Jochum, K.P., Nohl, U., Herwig, K., Lammel, E., Stoll, B., Hofmann, A.W., 2005. GeoReM:
580 A New Geochemical Database for Reference Materials and Isotopic Standards.
581 *Geostand. Geoanalytical Res.* 29, 333–338. [https://doi.org/10.1111/j.1751-](https://doi.org/10.1111/j.1751-908X.2005.tb00904.x)
582 [908X.2005.tb00904.x](https://doi.org/10.1111/j.1751-908X.2005.tb00904.x)

583 Johnsen, I.V., Aaneby, J., 2019. Soil intake in ruminants grazing on heavy-metal
584 contaminated shooting ranges. *Sci. Total Environ.* 687, 41–49.
585 <https://doi.org/10.1016/j.scitotenv.2019.06.086>

586 Jurjanz, S., Collas, C., Lastel, M.L., Godard, X., Archimède, H., Rychen, G., Mahieu, M.,
587 Feidt, C., 2017. Evaluation of soil intake by growing Creole young bulls in common
588 grazing systems in humid tropical conditions. *animal* 11, 1363–1371.
589 <https://doi.org/10.1017/S1751731116002755>

590 Jurjanz, S., Collas, C., Quish, C., Younge, B., Feidt, C., 2021. Ingestion of Soil by Grazing
591 Sport Horses. *Animals* 11(7), 2109. <https://doi.org/10.3390/ani11072109>

592 Jurjanz, S., Germain, K., Dziurla, M.A., Juin, H., Jondreville, C., 2014. Use of acid-insoluble
593 ash and n-alkanes as markers of soil and plant ingestion by chickens. *Anim. Feed Sci.*
594 *Technol.* 188, 92–101. <http://dx.doi.org/10.1016/j.anifeedsci.2013.11.004>

595 Jurjanz, S., Germain, K., Juin, H., Jondreville, C., 2015. Plant and soil intake by organic
596 broilers reared in tree- or grass-covered plots as determined by means of n-alkanes
597 and of acid-insoluble ash. *animal* 9, 888–898.
598 <https://doi.org/10.1017/S1751731114002870>

599 Jurjanz, S., Roinsard, A., 2014. Valorisation de l’herbe par des truies plein-air. *ALTERAGRI*
600 125, 25–28.

601 Kagira, J.M., Kanyari, P.W.N., Maingi, N., Githigia, S.M., Ng’ang’a, J.C., Karuga, J.W.,
602 2010. Characteristics of the smallholder free-range pig production system in western
603 Kenya. *Trop. Anim. Health Prod.* 42, 865–873. [https://doi.org/10.1007/s11250-009-](https://doi.org/10.1007/s11250-009-9500-y)
604 [9500-y](https://doi.org/10.1007/s11250-009-9500-y)

605 Li, Z., 2020. PBCLM: A top-down causal modeling framework for soil standards and global
606 sustainable agriculture. *Environ. Pollut.* 263, 114404.
607 <https://doi.org/10.1016/j.envpol.2020.114404>

608 Mahieu, M., Arquet, R., Tricheur, A., Collas, C., Jurjanz, S., 2019. A method for monitoring
609 grazing patterns using a non-professional digital camera and free software. *Fourrages*
610 240, 335–340.

611 Martínez Avilés, M., de la Torre, A., Prodanov-Radulović, J., Bellini, S., 2019.
612 Characterising outdoor pig production in Europe [WWW Document]. URL
613 <https://www.thepigsite.com/articles/characterising-outdoor-pig-production-in-europe>
614 (accessed 3.14.22).

615 Mayland, H.F., Florence, A.R., Rosenau, R.C., Lazar, V.A., Turner, H.A., 1975. Soil
616 Ingestion by Cattle on Semiarid Range as Reflected by Titanium Analysis of Feces. *J.*
617 *Range Manag.* 28, 448–452.

618 Ministère de la Transition Ecologique, 2017. La contamination chronique des sols antillais
619 par la chlordécone [WWW Document]. Données Études Stat. URL
620 [http://www.donnees.statistiques.developpement-](http://www.donnees.statistiques.developpement-durable.gouv.fr/lesessentiels/essentiels/sol-contamination-antille.htm)
621 [durable.gouv.fr/lesessentiels/essentiels/sol-contamination-antille.htm](http://www.donnees.statistiques.developpement-durable.gouv.fr/lesessentiels/essentiels/sol-contamination-antille.htm) (accessed
622 3.14.22).

623 Perrette, J., Le Floch, C., Naves, M., Gourdine, J.-L., Alexandre, G.G., 2020. Intérêts des
624 secteurs formels et informels de l'élevage pour une fourniture variée de services
625 écosystémiques : le cas de la Guadeloupe. *Ethnozootechnie* 107, 95–108.

626 R Development Core Team, 2020. *R: A language and environment for statistical computing.*
627 R Foundation for Statistical Computing, Vienna, Austria.

628 Rivera Ferre, M.G., Edwards, S.A., Mayes, R.W., Riddoch, I., DeB. Hovell, F.D., 2000.
629 Grass utilisation by outdoor sows in different seasons measured by the n-alkane
630 technique. Presented at the Ecological Animal Husbandry in the Nordic Countries:
631 Proceedings from NJF-seminar No. 303, Danish Research Centre for Organic
632 Farming, pp. 87–92.

633 Roberts, A.H.C., Longhurst, R.D., 2002. Cadmium cycling in sheep- grazed hill- country
634 pastures. *N. Z. J. Agric. Res.* 45, 103–112.

635 Robinson, T.P., Thornton, P.K., Franceschini, G., Kruska, R.L., Chiozza, F., Notenbaert, A.,
636 Cecchi, G., Herrero, M., Epprecht, M., Fritz, S., You, L., Conchedda, G., See, L.,
637 2011. *Global livestock production systems*, Rome, Food and Agriculture Organization
638 of the United Nations (FAO) and International Livestock Research Institute (ILRI).
639 ed.

640 Rose, C.J., Williams, W.T., 1983. Ingestion of earthworms, *Pontoscolex corethrurus*, by
641 village pigs, *Sus scrofa papuensis*, in the highlands of Papua New Guinea. *Appl.*
642 *Anim. Ethol.* 11, 131–139.

643 Rychen, G., Jurjanz, S., Fournier, A., Toussaint, H., Cyril, F., 2013. Exposure of ruminants to
644 persistent organic pollutants and potential of decontamination. *Environ. Sci. Pollut.*
645 *Res. Int.* 21(10), 6440–6447. <https://doi.org/10.1007/s11356-013-1882-8>

646 Sousa Junior, A.A.O., Freitas, A.B., Mariante, A., Sierra Vásquez, A., Vadell, A., Barba
647 Capote, C., Abeledo, C.M., Lorenzo Machorro, C.R., Campagna, D., Leite, D.M.G.,
648 Sponenberg, D.P., Pimenta Filho, E.C., Costa, E.P., Diéguez, F.J., Castro, G.,
649 Rodríguez-Galván, G., Silva, H.T., Santana, I., Beranger, I., Ly, J., León Jurado, J.M.,
650 Ortiz Ortiz, J., Sereno, J., Hernández Zepeda, J.S., Delgado Bermejo, J.V., Álvarez
651 Franco, L.A., Sarmiento Franco, L., Vásquez Chegüen, L., Zaragoza-Martínez, L.,
652 Montenegro, M.C., Canul Solís, M., Bender, M.F., Rosado, M.M., Macías, M.,
653 Barlocco, M., Silva Filha, O.L., Gámiz Ramírez, P., Lopes, P.S., Silva, P., Gagliardi,
654 R., Jáuregui Jiménez, R., Perezgrovas Garza, R., Llambí, S., Piovezan, U., López,
655 V.T., 2014. *Las razas porcinas Iberoamericanas: un enfoque etnozootécnico*,
656 Salvador, BA (Brasil): IF Baiano. ed.

657 Stern, S., Andresen, N., 2003. Performance, site preferences, foraging and excretory
658 behaviour in relation to feed allowance of growing pigs on pasture. *Livest. Prod. Sci.*
659 79, 257–265. [https://doi.org/10.1016/S0301-6226\(02\)00171-9](https://doi.org/10.1016/S0301-6226(02)00171-9)

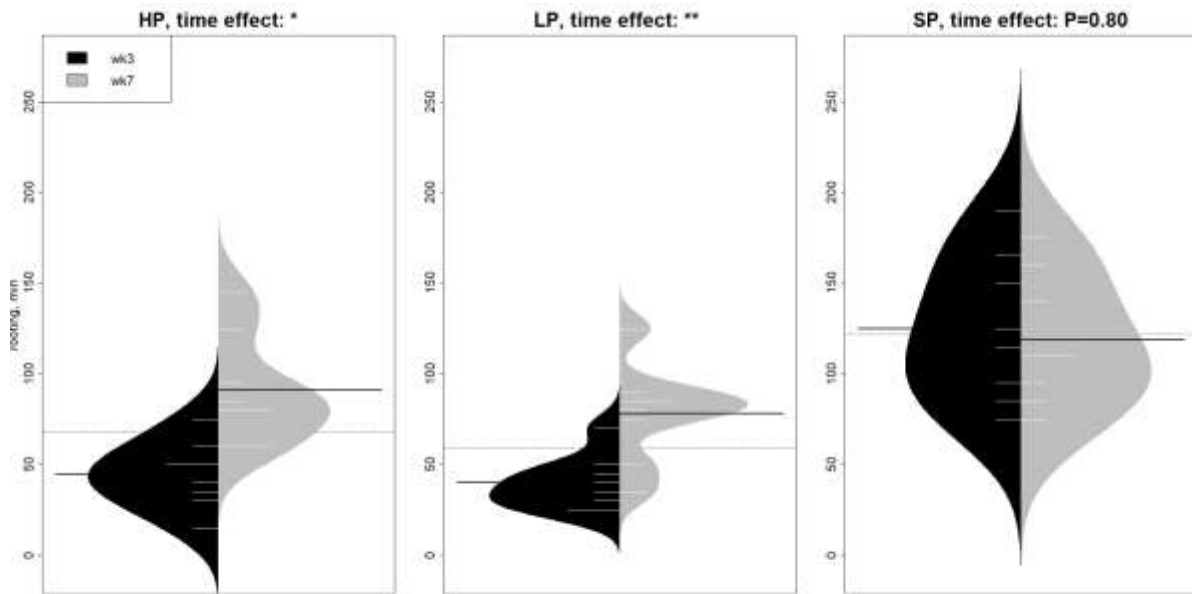
660 Studnitz, M., Jensen, M.B., Pedersen, L.J., 2007. Why do pigs root and in what will they
661 root? A review on the exploratory behaviour of pigs in relation to environmental
662 enrichment. *Appl. Anim. Behav. Sci.* 107, 183–197.
663 <https://doi.org/10.1016/j.applanim.2006.11.013>

664 Thutwa, K., Chabo, R., Nsoso, S.J., Mareko, M., Kgwatalala, P.M., Owusu-Sekyere, E.,
665 2020. Indigenous Tswana pig production characteristics and management practices in
666 southern districts of Botswana. *Trop. Anim. Health Prod.* 52, 517–524.
667 <https://doi.org/10.1007/s11250-019-02037-3>

668

669 **Figures**

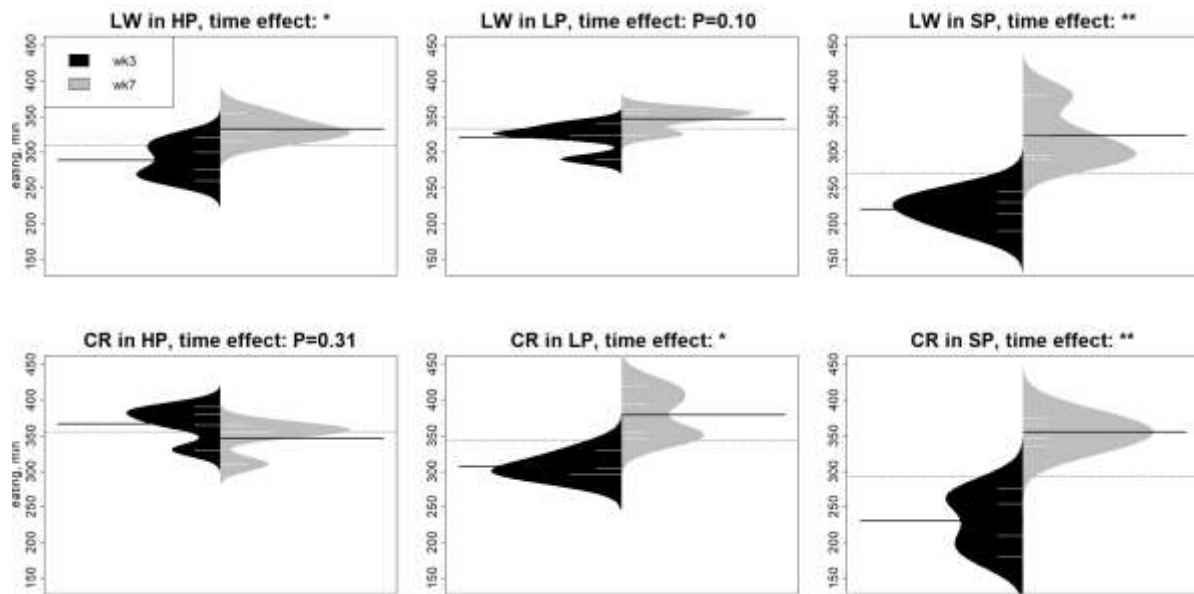
670



671

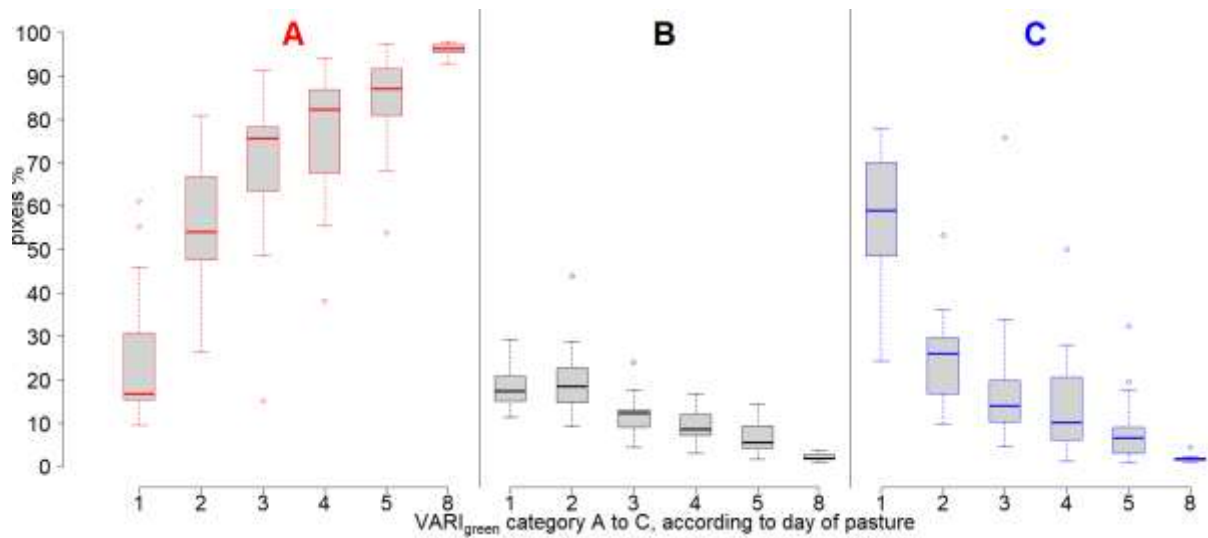
672 **Fig. 1.** Beanplot of rooting activities (time spent within 24-h, in min) of outdoor pigs reared
673 in HP (pasture with grass more than 60 days of regrowth age), LP (pasture with grass at 35
674 days of regrowth age) or SP (sweet-potato field) conditions. The horizontal bars correspond
675 to the means of the distributions (if they were normal): one mean for the black distribution,
676 one mean for the grey and one mean for the joint distribution. Asteriks indicate significant
677 difference (*: $P < 0.05$; **: $P < 0.01$; ***: $P < 0.001$) between wk3 and wk7 within treatment.

678



679
 680 **Fig. 2.** Beanplot of eating activities (time spent within 24-h, in min) of outdoor pigs reared in
 681 HP (pasture with grass more than 60 days of regrowth age), LP (pasture with grass at 35 days
 682 of regrowth age) or SP (sweet-potato field) conditions. Asteriks indicate significant
 683 difference (*: $P < 0.05$; **: $P < 0.01$; ***: $P < 0.001$) between wk3 (in black) and wk7 (in grey)
 684 within breed \times treatment interactions (LW: Large White, CR: Creole).

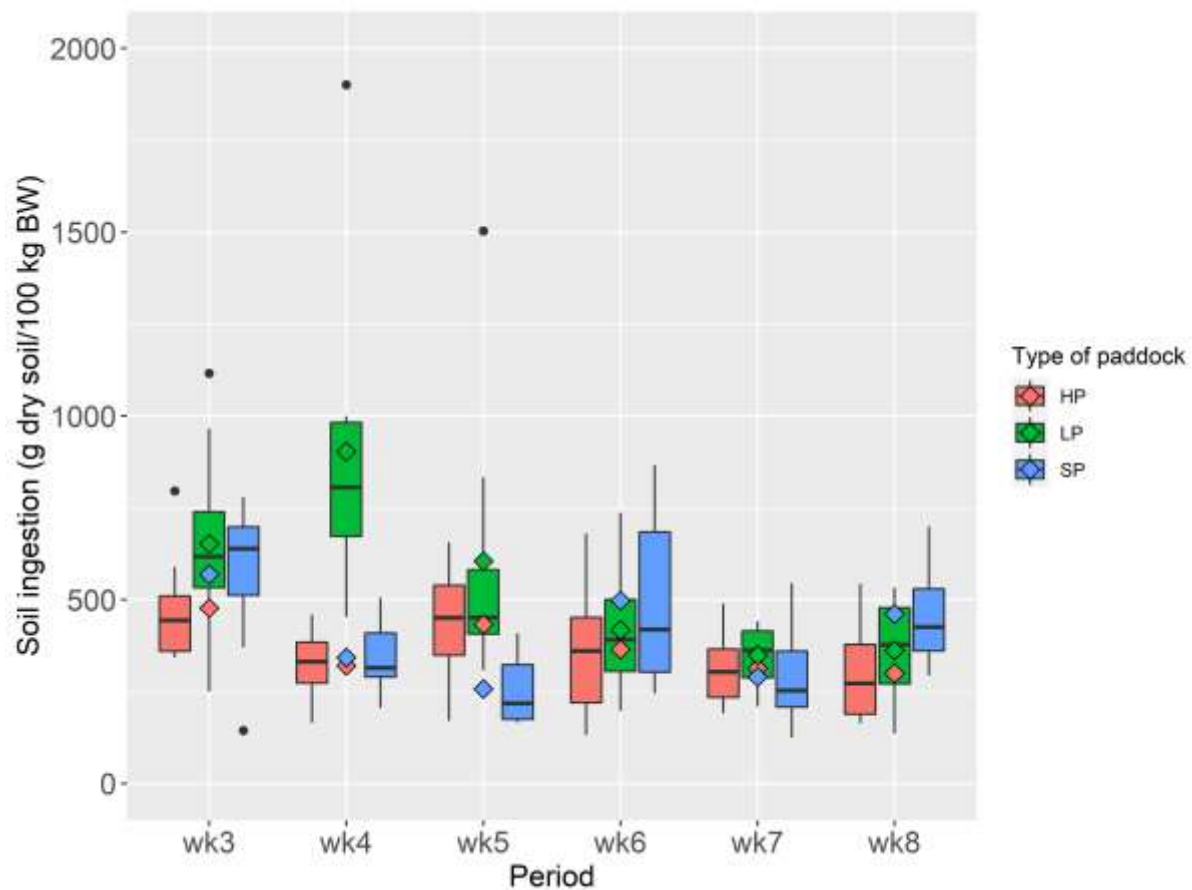
685



686

687 **Fig. 3.** Boxplot of the percentage of pixels in each VARIgreen category (A: bare soil, mud,
 688 and/or soil litter; B: stems and senescent plant parts; and C: green matter: leaves and sheaths)
 689 according to the day of pasture (1 to 5: first to fifth day of pasture; 8: the first day after
 690 pasture).

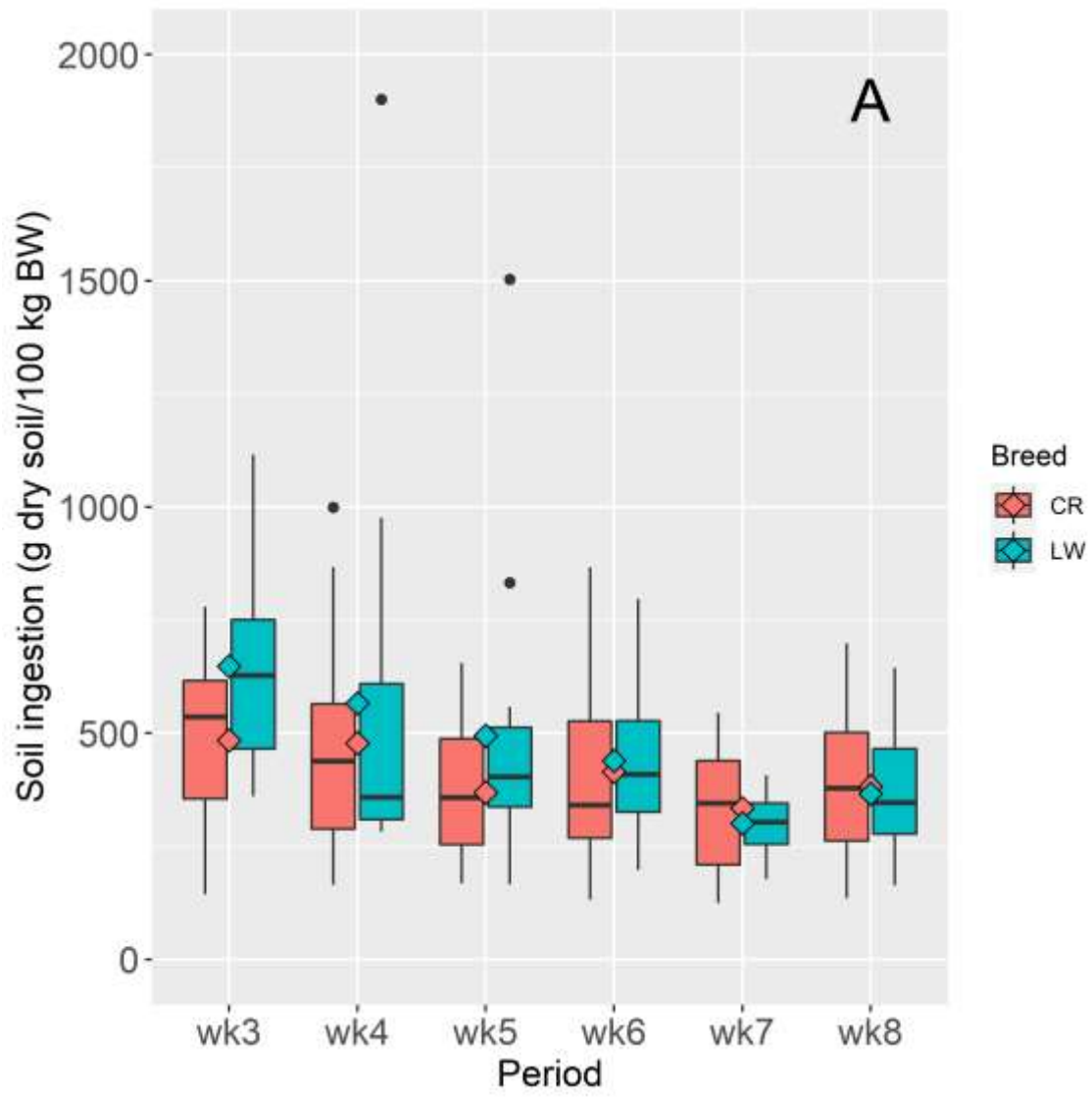
691



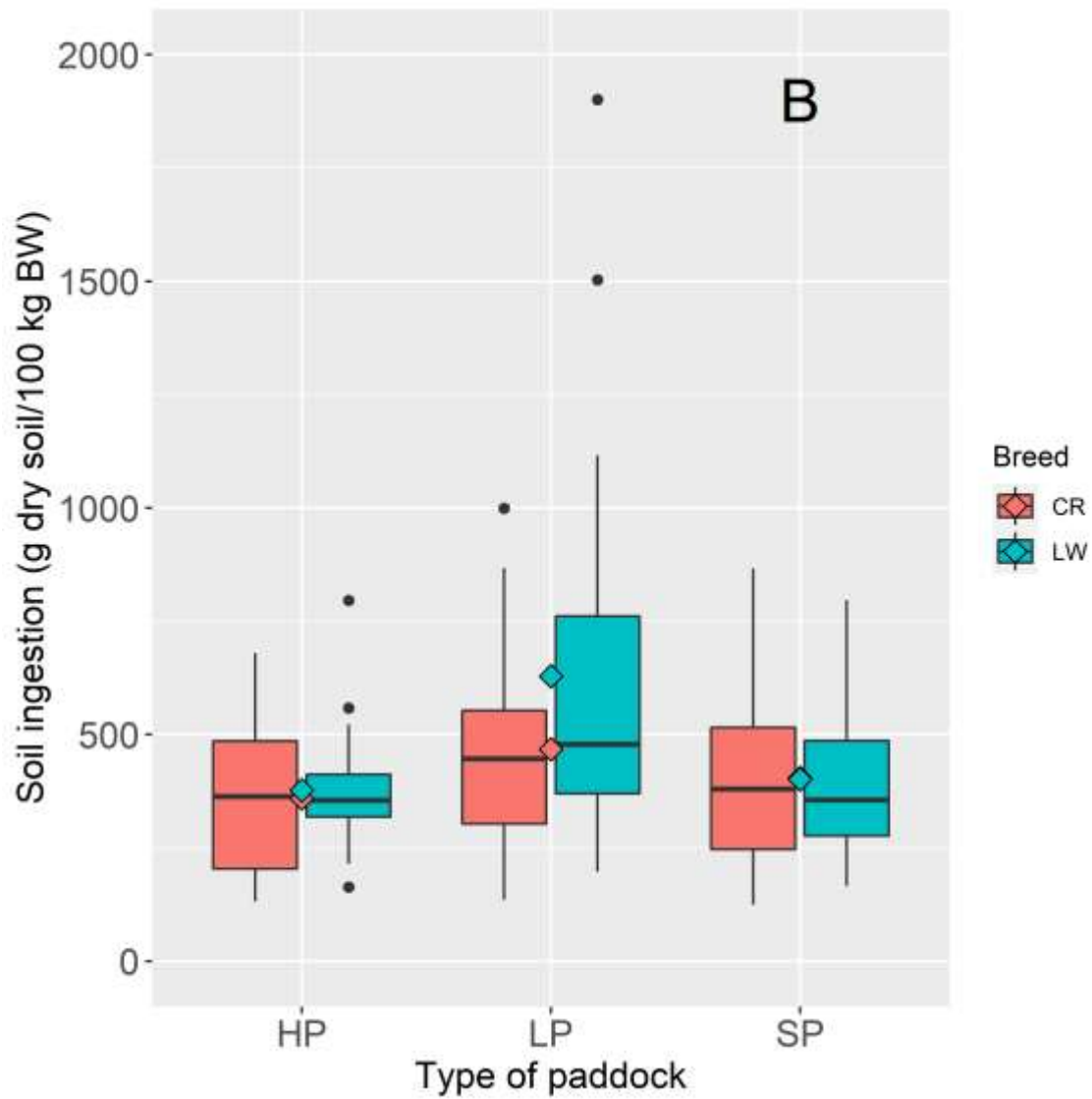
692

693 **Fig 4.** Effect of type of paddock (HP: pasture with grass more than 60 days of regrowth age;
 694 LP: pasture with grass at 35 days of regrowth age; SP: sweet-potato field) and experimental
 695 week (wk3 to wk8) on the ingestion of soil (the diamonds represent the averages of each
 696 treatment for each period).

697



698



699

700 **Fig. 5.** Effect of breed (CR: Creole; LW: Large White) and experimental period (A) (wk3 to
 701 wk8) or type of paddock (B) (HP: pasture with grass more than 60 days of regrowth age; LP:
 702 pasture with grass at 35 days of regrowth age; SP: sweet-potato field) on the relative
 703 ingestion of dry soil, i.e. per 100 kg of BW.

704

705 **Tables**

706

707 **Table 1.** Grazing management, vegetation and soil characteristics according to the paddock
 708 type (HP: pasture with grass more than 60 days of regrowth age; LP: pasture with grass at 35
 709 days of regrowth age; SP: sweet-potato field) (average value of the 6 weeks for each type of
 710 paddock).

Item	Type of paddock			
	HP	LP	SP	
			Tubers	Leaves/stems
Grazing management				
Daily surface area, m ² per animal per period	16.25	16.25	16.25	
Stocking rate, kg BW per m ²	2.00	2.05	2.04	
Stocking rate (week 3), kg BW per m ²	1.64	1.66	1.67	
Stocking rate (week 8), kg BW per m ²	2.39	2.44	2.42	
Sward height, cm	28.6	15.7	-	
Vegetation characteristics				
Dry matter content, g·kg ⁻¹	24.8	23.1	28.2	16.2
Organic matter content, g·kg ⁻¹ DM	779	-	881	801
Crude protein content, g·kg ⁻¹ DM	62	-	56	122
Neutral detergent fiber content, g·kg ⁻¹ DM	669	-	60	302
Cr ¹ content, µg·g ⁻¹ DM	< 20	< 20	< 20	< 20
Ti ² content, µg·g ⁻¹ DM	36	59	< 25	69
Ti and Cr levels in soils				
Cr content, µg·g ⁻¹ DM	<20	<20	<20	
Ti content, µg·g ⁻¹ DM	14223	14082	13282	

711 ${}^1\text{Cr}$ = chromium.

712 ${}^2\text{Ti}$ = titanium.

713

714 **Table 2.** Soil ingestion by pigs according to the type of paddock (HP: pasture with grass
 715 more than 60 days of regrowth age; LP: pasture with grass at 35 days of regrowth age; SP:
 716 sweet-potato field), period and their interaction.

Item	Type of paddock				<i>P</i> -value		
	HP	LP	SP	RSD ¹	Type of paddock	Period	Type of paddock × Period
Body weight (BW), kg	32.5	33.3	33.2	0.84	NS ²	<.0001	0.043
Faecal production							
Faecal Cr ³ content, µg·g ⁻¹	3118	3050	3109	498.3	NS	0.004	0.068
Daily faecal output, g	357	387	343	66.5	NS	0.058	0.015
DM per pig							
Daily faecal output, g							
DM per 100 kg BW							
Faecal Ti ⁴ content, µg·g ⁻¹	4585 _b	6052 _a	4924 _{ab}	800.6	0.036	<.0001	<.0001
Soil ingestion							
Soil ingestion, g DM per pig	116 _b	171 _a	131 _{ab}	41.1	0.050	0.024	<.0001
Soil ingestion, g DM per 100 kg BW	368	548	403	145.0	0.081	<.0001	<.0001
Soil ingestion, g DM per kg MW ⁵	8.7	12.9	9.6	3.26	0.069	NS	<.0001

717 ^{a-b}Means within a row with different superscripts differ (*P* <0.05).

718 ¹RSD = residual standard deviation.

719 ²NS = non-significant ($P > 0.1$).

720 ³Cr = chromium.

721 ⁴Ti = titanium.

722 ⁵MW = metabolic weight.

723

724 **Table 3.** Soil ingestion by pigs according to the breed (CR: Creole; LW: Large White) and its
 725 interaction with the period (same model and continuation of the results presented in Table 2).

Item	Breed			<i>P</i> -value	
	CR	LW	RSD ¹	Breed	Breed × Period
Body weight (BW), kg	29.3	36.7	0.84	<.001	<.0001
Faecal production					
Faecal Cr ² content, µg·g ⁻¹	3676	2509	498.3	<.0001	0.009
Daily faecal output, g DM per pig	284	440	66.5	<.0001	0.019
Daily faecal output, g DM per 100 kg					
BW					
Faecal Ti ³ content, µg·g ⁻¹	5442	4932	800.6	NS ⁴	NS
Soil ingestion					
Soil ingestion, g DM per pig	116	163	41.1	0.015	0.089
Soil ingestion, g DM per 100 kg BW	411	469	145.0	NS	NS
Soil ingestion, g DM per kg MW ⁵	9.4	11.4	3.26	NS	0.077

726 ¹RSD = residual standard deviation.

727 ²Cr = chromium.

728 ³Ti = titanium.

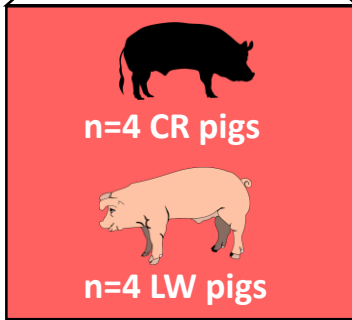
729 ⁴NS = non-significant (*P*>0.1).

730 ⁵MW = metabolic weight.

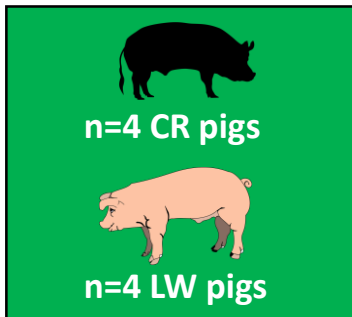
731

wk3 → wk4 → wk5 → wk6 → wk7 → wk8

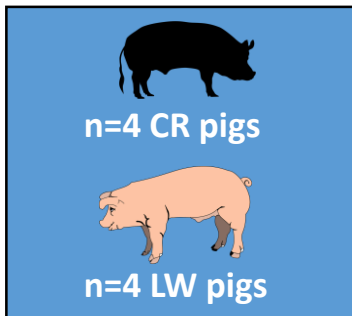
6 experimental weeks wk3 to wk8
(preceded by 2 adaptation weeks wk1 and wk2)
Pigs moved to a new paddock each week



High Pasture HP

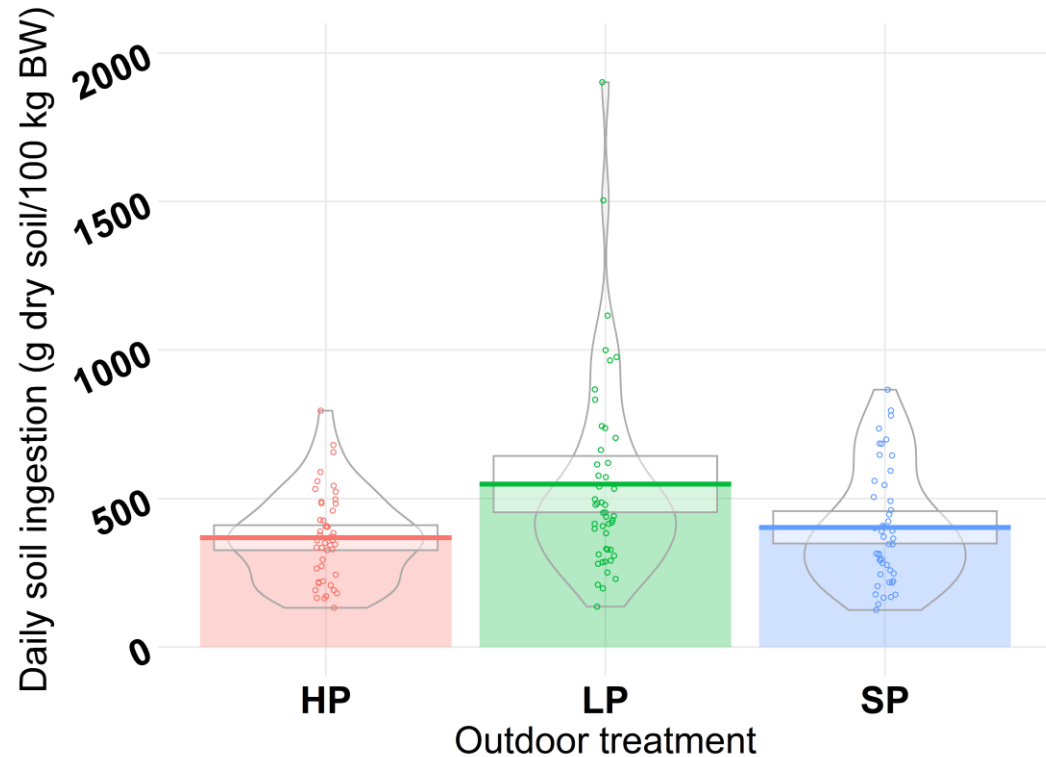


Low Pasture LP



Sweet Potato SP

3 outdoor treatments (HP, LP, SP), 24 free-range pigs, 2 breeds
(CR: Creole of Guadeloupe, LW: Large White)



Individual daily soil ingestion of pigs
No difference between the 3 treatments and the 2 breeds

Reference data for predictive models of
pig exposure to soil-bound pollutants