

Comparison of enrichment methods for efficient nitrogen fixation on a biocathode

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14	
15	Abstract

16 The production of nitrogen fertilizers in modern agriculture is mostly based on the Haber-Bosch process, representing nearly 2% of the total energy consumed in the world. Low-energy 17 18 bioelectrochemical fixation of N₂ to microbial biomass was previously observed but the mechanisms 19 of microbial interactions in N₂-fixing electroactive biofilms are still poorly understood. The present 20 study aims to develop a new method of enrichment of autotrophic and diazotrophic bacteria from soil 21 samples with a better electron source availability than using H_2 alone. The enrichment method was 22 based on a multi-stage procedure. The first enrichment step was specifically designed for the selection 23 of N₂-fixing bacteria from soil samples with organic C as electron and carbon source. Then, a polarized 24 cathode was used for the enrichment of autotrophic bacteria using H_2 (hydrogenotrophic) or the 25 cathode as electron source. This enrichment was compared with an enrichment culture of pure 26 diazotrophic hydrogenotrophic bacteria without the use of a microbial electrochemical system. 27 Interestingly, both methods showed comparable results for N₂ fixation rates at day 340 of the 28 enrichment with an estimated average of approximately 0.2 mgN_{fixed}/L.d. Current densities up to -15 29 A/m² were observed in the polarized cathode enrichments and a significant increase of the microbial 30 biomass on the cathode was shown between 132 and 214 days of enrichment. These results confirmed 31 an enrichment in autotrophic and diazotrophic bacteria on the polarized cathode. It was hypothesied 32 that autotrophic bacteria were able to use either the H_2 produced at the cathode or directly the 33 cathode through direct electron transfer (DET) as more biomass was produced than with H_2 alone. 34 Finally, the analysis of the enriched communities suggested that Desulforamulus ruminis mediated 35 microbial interactions between autotrophic anaerobic and heterotrophic aerobic bacteria in polarized cathode enrichment. These interactions could play a key role in the development of biomass in these 36 37 systems and on N₂ fixation. Based on these findings, a conceptual model on the functioning of mixed 38 cultures N₂-fixing electroactive biofilms was proposed.

39 *Keywords:* Nitrogen fixation, Microbial electrochemical system, Biomass electrostimulation,

- 40 Enrichment method
- 41

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Introduction

45 Nitrogen is one of the essential elements for the growth of all living organisms, especially for cellular 46 protein synthesis. In modern agriculture, ammonia is often used as a nitrogen source for plants (Bagali, 47 2012; Burris & Roberts, 1993; Masclaux-Daubresse et al., 2010). This compound is produced at industrial 48 scale by the Haber-Bosch process which allows the reduction of N₂ to NH₃ at the expanse of large quantities 49 of H₂ and energy (Kandemir et al., 2013; Martín et al., 2019). This process is associated with significant CO₂ 50 emissions, due to the source of H_2 obtained either by methane steam reforming or coal gasification (Martín 51 et al., 2019). Alternatives for green H_2 production, such as water electrolysis, are therefore nowadays 52 considered to feed the Haber-Bosch process that is also contributing to a high energy demand of the 53 process (Cherkasov et al., 2015). A more direct alternative is to reduce N₂ directly on a cathode by chemical 54 catalysts. However, these catalysts are not renewable and are currently not sufficiently selective regarding 55 the hydrogen evolution reaction at ambient conditions (Deng et al., 2018; A. Liu et al., 2020).

56 Recently some authors proposed to use N₂-fixing bacteria in association with electrochemical systems 57 for N₂ reduction at low energy cost (C. Liu et al., 2017; Rago et al., 2019). This idea was inspired by previous 58 observations of microbial CO₂ fixation on microbial cathodes in a process known as microbial 59 electrosynthesis (A. Liu et al., 2020; Logan et al., 2019). The microbial fixation of both N₂ and CO₂ with a 60 polarized cathode was demonstrated in two recent studies (C. Liu et al., 2017; Rago et al., 2019). First, Liu 61 et al. (2017) demonstrated the growth of Xanthobacter autotrophicus in a hybrid organic-inorganic 62 electrochemical system in the absence of nitrogen sources other than N₂. Then, Rago et al. (2019) demonstrated N₂ fixation through microbial electrosynthesis (MES) with a mixed microbial community (C. 63 64 Liu et al., 2017; Rago et al., 2019).

65 Other works investigated the mechanisms with pure bacterial strain like Soundararajan et al. (2019) 66 and Chen et al. (2020) with Rhodopseudomonas palustris and Pseudomonas stutzeri(Chen et al., 2020; 67 Soundararajan et al., 2019). Yadav et al. (2022) demonstrated the possible use of N₂-fixing bacteria as 68 nitrogen source in a microbial electrosynthesis process of acetate(Yadav et al., 2022). Zhang et al. (2022) 69 also worked on a system of N₂ fixation in microbial electrolysis cell (MEC)(Zhang et al., 2022). In these 70 works, the authors investigated the interactions existing between CO_2 and N_2 fixation microbial processes. 71 Coupling capabilities of N_2 fixation in bioelectrochemical systems as possible source of nitrogen for other 72 biological systems was subject of interest. Indeed, such coupling can lead to the production of molecules 73 of interest such as acetate by reducing the environmental impact of the use of reactive nitrogen often in 74 the form of NH₄Cl (Yadav et al., 2022). The work of Li et al. (2022) in a single-chamber system and 75 highlighted an importance of synergy within an N₂-fixing community in a microbial bioremediation system 76 (Li et al., 2022).

77 All this work has demonstrated that it is possible to use a cathode as an electron source for biomass 78 growth by fixing N₂ and CO₂. This biomass could then be used as a fertilizer with a low impact on the 79 environment (Chakraborty & Akhtar, 2021; Rago et al., 2019; y. Hafeez et al., 2006). Although the proof of 80 concept was made for this process, the microbial interactions supporting N_2 fixation in this microbial 81 electrochemical systems are still poorly understood. Different N₂ fixation scenarios are indeed possible, 82 such as: (i) fixation by a single population capable of fixing N_2 and CO_2 using the electrode as sole electron 83 source, (ii) fixation by heterotrophic diazotrophic bacteria that can utilize the organic carbon produced by 84 electro-autotrophic bacteria, (iii) fixation through an interaction between methanogenic archaea and 85 methanotrophs that could use CH₄ as an energy source for N₂ fixation and (iv) fixation through an 86 interaction mediated by direct interspecies electron transfer (DIET) between electro-autotrophic bacteria 87 and diazotrophic bacteria (Rago et al., 2019). A better understanding of these interactions is essential to 88 optimize N₂ fixation in microbial electrochemical system.

89 In nature, biological N₂ fixation is a key mechanism of the nitrogen cycle where atmospheric nitrogen 90 is uptaken by living organisms (Bagali, 2012). It is carried out by so-called diazotrophic bacteria responsible 91 for the transformation of N₂ into NH₃ (Kim & Rees, 1994). Some of these bacteria can be found on the roots 92 of plants where they are living in symbiosis (Burris & Roberts, 1993; Franche et al., 2009). These 93 microorganisms are able to fix N_2 from the air, making it assimilable in the form of NH_4^+ or amino acids (L-94 glutamine, L-glutamate) which are further used in plants for protein or DNA synthesis (Burris & Roberts, 95 1993). In exchange, the bacteria use the organic matter of root exudates produced by the plants as carbon 96 and energy sources. Among the diazotrophic bacteria, the genera Frankia and Rhizobium spp. are often

97 associated with leguminous plant roots (Burris & Roberts, 1993; Peoples & Craswell, 1992). In contrast,
98 other free-living N₂-fixing bacteria such as *Azospirillum*, are able to fix N₂ with or without interacting with
99 plants and can use organic or inorganic materials to produce their own energy (Tilak et al., 1986).

100 In order to better understand the microbial mechanisms supporting N₂ fixation in polarized cathode 101 enrichment and produce biomass, this work aims at developing an enrichment method of cathodic biofilms 102 for direct fixation of CO₂ and N₂. Rago et al. (2019) demonstrated the capacity of producing biomass from 103 air, CO₂ and a solid electrode polarized negatively (Rago et al., 2019). Here a specific enrichment of 104 microorganisms capable of N₂ fixation was developed. It was hypothesized that a multi-step enrichment 105 with specific medium could select a microbial community able to fix N₂ and CO₂ supported by electrons 106 brought by a cathode. It was assumed that these communities enriched by this procedure could use a large 107 number of interactions leading to N₂ fixation and biomass growth. For that, it was assumed that the 108 enrichment in autotrophic diazotrophic bacteria in the presence of a cathode with pre-enrichment steps 109 in presence of several electron donors (organic C and cathode) would be more efficient than enrichment 110 cultures of hydrogenotrophic and diazotrophic bacteria. For this, soil samples were used as sources of N₂-111 fixing bacteria, and successive enrichments in autotrophic bacteria in polarized cathode enrichment (PCE) 112 were performed to select a electroactive biofilm capable of fixing N_2 and CO_2 with a cathode as sole 113 electron source. The enriched biofilm was compared with a classical enrichment of N2-fixing hydrogen-

- oxidizing bacteria (HOB) in flasks (named H₂ enrichment, H₂E) (X. Hu et al., 2020; C. Liu et al., 2017).
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Methods

116 Inoculum

Soil samples from a forest, agricultural crop field and a commercial compost were used as microbial inoculum sources. Samples were collected from forest and agricultural soils in the Haute Vallée de l'Aude, France. These sources were selected based on their assumed abundance of N₂-fixing bacteria and their theoretical C/N ratio (Khan et al., 2016). One to two mg of each samples were used as inoculum in 50mL of medium for preliminary enrichment culture.

122 Culture media

123 The culture media were both formulated on the basis of H3 medium (81 DSMZ) used for enrichment of 124 soil autotrophic bacteria. The medium consisted of 2.3g KH₂PO₄ and 2.9g Na₂HPO₄ 2H₂O per liter as buffer, 125 0.5g MgSO₄ 7H₂O, 0.01g CaCl₂ 2H₂O, 0.005g MnCl₂ 4H₂O, 0.005g NaVO₃ H₂O, and 5mL of SL-6 trace element 126 solution per liter of medium, with 5mL of vitamin solution. The SL-6 trace element solution consisted of 127 0.1g ZnSO₄ 7H₂O, 0.03g MnCl₂ 4H₂O, 0.3g H₃BO₃ 0.2 CoCl₂ 6H₂O, 0.01g CuCl₂ 2H₂O, 0.02g NiCl₂ 6H₂O, and 128 0.03g Na₂MoO₄ 2H₂O per liter of solution. The vitamine solution consisted of 10mg Riboflavin, 50mg 129 Thiamine-HCl 2H₂O, 50mg Nicotinic acid, 50mg Pyridoxine-HCl, 50mg Ca-Pantothenate, 0.1mg Biotin, 130 0.2mg Folic acid and 1mg Vitamin B₁₂ for 100 mL of distilled water. Iron citrate was added to the enrichment 131 bottles at a concentration of 0.05 g/L but not to the microbial bioelectrochemical systems in which the 132 cathode was used as sole electron source when only CO₂ was suplied. An organic carbon solution (organic 133 C) was composed of 2g/L D-glucose, 1g/L yeast extract, 1g/L Na-acetate, 1g/L DL-malic acid, 1g/L Na-134 lactate, 1g/L Na-pyruvate, and 1g/L D-mannitol and used when indicated. NH₄Cl was added at 1 g/L only 135 when indicated. All enrichment procedures were maintained at 30°C and the pH was adjusted to 6.8 with 136 NaHCO₃ in the microbial bioelectrochemical systems and the enrichment cultures without organic C 137 addition. When the organic C solution was used, the pH was adjusted between 6.3 and 6.5.

138 Design of microbial electrochemical system

139 The electrochemical system used for our enrichment were composed of two chambers separated by 140 an anion exchange membrane (AEM) (fumasep ® FAB-PK-130). The AEM was used to avoid the migration 141 of NH_4^+ ions to the anodic chamber. Each chamber had a total volume of one liter. The pH of the reactors 142 was adjusted to 6.8 at the beginning of each batch experiment. Each system had a 25 cm² square carbon 143 felt working electrode with a thickness of 0.7 cm and a 16 cm² square Pt-Ir grid as a counter electrode. The 144 carbon felt electrodes were conditioned using chemical treatment with HCl, a flush with ethanol and a 145 heat treatment at +400°C as described elsewhere by Paul et al (Paul et al., 2018). The systems were 146 inoculated with the flask enrichments used for N₂ fixation in presence of organic C (N-free medium). After

147 inoculation, the organic C supply was reduced to 10% of the initial supply in all reactors. The organic C 148 supply was then totally replaced by CO₂ supply after 60 days. Two of the systems were polarized and two 149 other reactors were used as controls with open current voltage (non-polarized cathode enrichment, nPCE). 150 The working electrodes of the polarized cathode enrichment (PCE) were poised at a potential of -0.940 V 151 vs. saturated calomel electrode (SCE) used as a reference. The system was connected to a VMP3.0 152 potentiostat (BioLogic). The current was measured over time by a chronoamperometry method. The 153 current intensity was used to monitor the availability of electrons in the electroactive biofilm. An increase 154 of the current intensity was representative of an increase of the reduction reactions at the cathode. The 155 increase of these reduction reactions, abiotic or not, was assumed to bring more electrons to the bacterial 156 community. It was therefore assumed that an increase in current intensity was related to the enrichment 157 of bacteria able to use the cathode as electron source (Zaybak et al., 2013). The current density J was 158 calculated using the surface of the working electrode, i.e. 25 cm².

Before inoculation of the reactors, an initial chronoamperometry measurement was performed along the first four days of operation with only organic carbon in the medium to determine the basal current density in absence of bacteria. Two other abiotic reactors for 15 days were then implemented to measure the current density in a medium without organic C. In order to validate the role of the cathode as electron source, two OCV (nPCE) reactors were carried out. The current densities of the abiotics reactors over a short period were therefore used as a reference to be compared with the measurements made after inoculation and monitor the increase of the activity of reduction at the cathode.

166 Electrons required for the production of microbial metabolites and for biomass growth was then used167 to calculate the Coulombic efficiency of the PCE according to the equations 1-5:

169 Eq. 1 $2.1 H_2 + CO_2 + 0.2NH_4^+ \rightarrow CH_{1.8}O_{0.5}N_{0.2} + 1.5H_2O + 0.2H^+$ for biomass production 170 (21 mol_e/mol_{Nbiomass}) (Wresta et al., 2021)

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Eq. 2 $N_2 + 3H_2 \rightarrow 2NH_3$ for nitrogen fixation (3 mol_{e-}/mol_{Nfixé}) (Zhang et al., 2022)

174 Eq. 3 $2CO_2 + 4H_2 \rightarrow CH_3COO^- + H^+ + 2H_2O$ for acetate production (8 mol_{e-}/mol_{CH3COO-}) 175 (Wresta et al., 2021)

176 177 Eq. 4 $2H^+ + 2e^- \rightarrow H_2$ for H₂ production (2 mol_e/mol_{H2})

Eq. 5

178 179

$$\eta_{CE} = \frac{n_e \times F \times n_{produc}}{\int i \, dt}$$

180 With η_{CE} the Coulombic efficiency in percentage of electron recovery in products, n_e moles of electrons 181 per moles of product (mol_e-/mol_{product}) calculated from the stoichiometric equations, *F* the Faraday 182 constant (96485 C.mol⁻¹), $n_{product}$ the number of mol_{product} and *i* the current intensity.

183 Enrichment procedures

184 Two enrichment procedures based on a sequential procedure by enriching separately N₂ fixation and 185 the use of an inorganic electron source were carried out. Each soil sample (Forest, Leguminous and 186 Compost) and a mix of all were used as inoculum in one batch of pre-enrichment. Pre-enrichment cultures 187 were then carried out in the same medium as first enrichment step for each procedure (i.e. supplied with 188 organic C for the first procedure and inorganic medium for the second) and for 20 days prior to be used as 189 inoculum in the enrichment procedures. Samples of the pre-enrichment were considered as initial 190 microbial community of the enrichments. Thus, at the end of pre-enrichment, the mixed culture was used 191 to inoculate two bottles of each enrichment and all unmixed soil samples were mixed to inoculate four 192 bottles of each enrichment.

193The first enrichment cultures, polarized cathode and non-polarized cathode enrichments (PCE and194nPCE) were performed in three steps:

195 - The first step was performed in a 120 mL bottle with N-free medium supplemented with organic C 196 source. This first step was used to select N₂-fixing bacteria using organic compounds as electron donors. 197 The headspace was composed of an $Ar/O_2/N_2$ mixture (80/5/15) at 0.5 bar (absolute pressure). Cultures 198 were carried out in bottles containing 50 mL of liquid and 70 mL of headspace. Subcultures of these

enrichments were performed every 7 days for 6 weeks. The time between 2 subcultures was then reduced
 to 3-4 days using 10% of the volume of the previous culture (5mL/50mL).

201 - After 55 days of enrichment in 10 successive batches, the N₂-fixing bacteria enriched cultures were 202 used to inoculate the cathodic chambers of the polarized microbial electrochemical system (Polarized 203 Cathode Enrichment, PCE) and the non-polarized microbial electrochemical system (non-Polarized Cathode 204 Enrichment nPCE). The same inorganic medium supplemented with 10% of the organic C source was fed 205 each week to start the enrichment of autotrophic bacteria. Inoculation of the cathode in presence of 206 organic C sources was made to favor bacterial growth. 80% of the medium was renewed every second 207 week to promote biofilm growth on the cathode. Headspace composition was monitored by GC and flushed 208 with N_2 if pO_2 exceeded 10% of the gas volume. Organic C supply was stopped when a significant current 209 density was measured in the polarized systems with regard to the controls.

210 - In the third and final enrichment step, CO₂ was used as sole carbon source as presented in Figure 1. 211 Organic C sources were then removed and the cathode was used as sole electron donor in reactor. Here, 212 only bacteria able to use cathode as electron source by direct interaction or indirect with H₂ were able to 213 grow. A 80/20 (v:v) CO_2/N_2 atmosphere was set up in the headspace with trace amounts of O_2 (< 5%). The 214 medium was replaced every 15-30 days. The gas recycling vessel was filled with CO₂ and was replaced with 215 a new one when O₂ exceeded 5% of the volume due to gas volume depletion. The nPCE controls were 216 operated in the same conditions as the polarized systems but without monitoring the current density. The 217 only available electron source in the nPCE controls was the organic C fed at the beginning of the 218 enrichment.

219 In the second enrichment method, the H2-fed enrichment (H2E), autotrophic bacteria were enriched in 220 inorganic medium supplemented with H₂ as sole electron source. These enrichments on H₂ were obtained 221 by pre-enriching in strict autotrophic bacteria using 50 ml of inorganic medium in a 120 ml bottle. The 222 headspace consisted of a mixture of 75/15/8/2 (v:v:v) H₂/CO₂/N₂/O₂ at an initial pressure of 1.5 bar 223 (absolute). Two two-weeks batches (30 days) were performed with NH₄Cl (1g/L) as nitrogen source in the 224 first stage of this enrichment as seen in Figure 1. This nitrogen source was then replaced by N₂ as only N 225 source to enrich N₂-fixing bacteria in the second stage of enrichment. Centrifugation (10min, 7500 RPM, ~7500g) of 80% of the initial medium (40mL) was done at each subculture every 15-20 days. The pellets 226 227 obtained after centrifugation were suspended in 5mL of sterile medium before being used for subculturing.

228 Table 1 presents a summary of the different procedures used.

229

Table 1 - Comparison between each enrichments procedure used in this work.

	PCE	nPCE	H ₂ E	
Enrichment reactor	Bottle then cathodic chamber	Bottle then cathodic chamber	Bottle	
Electrode Yes		Yes	No	
Applied potential -0,7 V vs SHE		Open circuit voltage	NA	
Enrichment step	3 steps	3 steps	2 steps	
	(1) soil diazotrophic bacteria in	(1) soil diazotrophic bacteria in	(1) soil autotrophic bacteria in	
	bottle	bottle	bottle	
	(2) diazotrophic bacteria in	(2) diazotrophic bacteria in	(2) autotrophic diazotrophic	
	MEC	MEC in OCV	bacteria in bottle	
	(3) autotrophic diazotrophic	(3) control without electron		
	bacteria in MEC	sources		
Number of Batch	(1) 10	(1) 10		
	(2) 4 medium replacement	(2) 4 medium replacement	(1) 2	
	(3) 14	(3) 14	(2) 25	
Batch duration	3-7 days (Organic C)	3-7 days (Organic C)	15-30 days	
	15-30 days (CO ₂)	15-30 days (CO ₂)		
Enrichment duration	>300 days	> 300 days	> 300 days	
Electron donor	(1) (2) Organic C	(1) (2) Organic C	(1) NH ₄ Cl and H ₂	
	(2) (3) Cathode (direct or	(3) none	(2) H ₂	
	indirect)			
Electron acceptor	O ₂ /N ₂ /CO ₂	O ₂ /N ₂ /CO ₂	O ₂ /N ₂ /CO ₂	

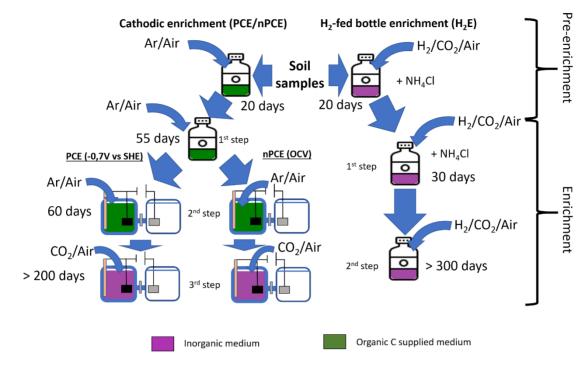


Figure 1 - Diagram of the enrichment process

234 Medium analyses

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 $\begin{array}{lll} 235 & {\sf NH_4^+, NO_2^- and NO_3^- concentrations were measured using a Gallery+ sequential analyzer (Thermo Fisher Scientific). Two mL of samples were centrifuged at 13500 RPM (~12300g) and then filtered to 0.2 \mum with nylon membranes before being stored at 4°C. The remaining pellets were stored at -20°C and used for community analysis. VFAs and other carbon compounds were measured on Clarus 580 GC equipped with FID and a Dionex UltiMate 3000 HPLC as described elsewhere (Carmona-Martínez et al., 2015; Moscoviz et al., 2019) \\ \end{array}$

Total nitrogen was measured using a CHNS Flashsmart elemental analyzer (Thermo Fisher Scientific). The sample (approximatively 2.5 mg) was weighed and was introduced into the oxidation/reduction chamber of the analyzer. 200 mL of medium were sampled at each medium change. These samples were dried for 4-5 days at 60°C. The samples were then freeze-dried and then grounded with a mortar. Two to four mg of each sample was used in the CHNS analyzer. The nitrogen content was then compared to the dry weight measured before freeze-drying to determine the mass of nitrogen in the medium. No CHNS analysis was performed on H₂ –based enrichment due to a low culture volume (50mL).

248 Nitrogen present in the biomass (biofilm and planktonic) was estimated from quantification of 16S 249 rDNA with qPCR. The rrnDB-5.7 database was used to estimate the actual bacterial amount from 16S rDNA 250 qPCR using theoretical 16S rDNA copies/genome per strain, genus or family given by the database and 251 sequencing results from our communities (Stoddard et al., 2015). Then, the theoretical calculated number 252 of bacteria was used to determine the nitrogen content in the biomass using the theoretical average dry 253 mass of an *Escherichia coli* cell of 216×10⁻¹⁵ g/bacterium and with a theoretical relative mass of nitrogen in 254 microbial biomass, ie. 11.4% according to the average biomass formula CH_{1.8}O_{0.5}N_{0.2} (Heldal et al., 1985; 255 Loferer-Krößbacher et al., 1998). The nitrogen present in the biomass was therefore estimated using the 256 formula below:

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Eq. 6 $N_{(in biomass)} = theoretical bacteria count \times mass E. coli \times relative mass of N in biomass$

260 With $N_{in \ biomass}$ the nitrogen concentration in the medium from biomass in mg_N/L, the *theoretical* 261 *bacteria count* based on the calcul of concentration in bacteria with results of qPCR of 16S rDNAfor each 262 enrichment (number of bacteria/L), mass E. coli a constant of 2.16 10⁻¹⁰ mg/cell_{E. coli} and relative mass of N 263 *in bacteria* is 11.4% of the dry mass.

264 Gas analysis and acetylene reduction assay (ARA)

CO₂, H₂ and N₂ used in the headspace of the enrichments were provided at laboratory grade. Pure
 ethylene was also supplied for calibration of the gas chromatography ethylene measurement.

Headspace compositions and pressures were analyzed every 1-2 days. The pressure was manually measured with a Keller LEO 2 manometer (KELLER AG, Wintherthur, Switzerland). Gas analyses were carried out on a Perkin Elmer Clarus 580 GC equipped with RT-Q-Bond and RT-Msieve 5Å columns with a TCD allowing the quantification of H₂, CO₂, N₂, O₂ and CH₄ with Ar as carrier gas as described by A. Carmona-Martínez (Carmona-Martínez et al., 2015). Acetylene and ethylene were measured using a Perkin Elmer Clarus 480 GC equipped with RT-U-Bond and RT-Msieve 5Å columns with TCD and He as carrier gas as described in a previous work (Carmona-Martínez et al., 2015).

The acetylene reduction assay (ARA) was performed to quantify the rate of N₂ fixation using the ability of nitrogenases to reduce acetylene to ethylene. This reaction occurs at a rate proportional to the rate of N₂ fixation according to the theoretical ratio C₂H₂:N₂ (3:1) (Bergersen, 1970). The ARA was performed only after 18 batch cycles for H₂-fed enrichment (H₂E) and 11 batch cycles in polarized cathode enrichment (PCE) fed with CO₂ (340 days). Acetylene concentration were calculated as follow :

279 Eq. 7
$$n_{C2H4} = \frac{P_{NaOH}V_{C2H4}}{RT} = \frac{(P_{total} - P_{CO2}) \times V \times \%_{C2H4}}{RT}$$

280 With $(P_{total}$ the pressure in headspace, P_{CO2} the partial pressure of CO₂ measured without CO₂ trap, V281 the volume of gas in headspace, $\%_{C2H4}$ the part of C₂H₄ measured by GC-TCD in headspace, R the perfect 282 gas constant and T, the temperature of the reactor.

283 And Acetylene production rate were calculated following next equation:

284 Eq. 8
$$\Delta[C_2H_4] = \frac{[C_2H_4]_{t1} - [C_2H_4]_{t0}}{t1 - t0} = \frac{\frac{n_{C2H_4t_1}}{V} - \frac{n_{C2H_4t_0}}{V}}{t1 - t0}$$

285 With $\Delta[C_2H_4]$ the rate of C₂H₄ production in µmol/L.d, *t0* the last measure without C₂H₄ observed, 286 *t1* the first measure of C₂H₄ in headspace and V the volume of liquid which is constant between two ARA 287 measurement.

288 A specific N₂-fixing activity was then calculated with the N₂-fixing bacteria measured by quantification of the nifH gene used as a marker for these bacteria. This specific activity corresponds to the rate of C₂H₄ 289 290 produced per bacteria capable of fix N₂ measured by qPCR of *nifH* gene. The acetylene used for the 291 acetylene reduction assay (ARA) was obtained by adding calcium carbide (CaC₂) in water and recovered in 292 a bag. The acetylene concentration in the bag was then measured by gas chromatography. Gas from the 293 bag was added to each enrichment to reach a composition of 10% V/V acetylene and an equivalent amount 294 of gas was removed from the headspaces. Ethylene production was then daily monitored for 7 days in the 295 microbial electrolysis systems (PCE and nPCE) and 15 in the H₂E bottles by the Perkin Elmer Clarus 480 GC 296 with TCD. To ensure separation of ethylene from CO₂ on the RT-U-Bond column, a CO₂ trap with sodium 297 hydroxide (6M NaOH) was used at the time of sampling and is taken into account in the calculations. After 298 the ARA method was completed, the headspaces were flushed with N₂ and the gas recycling system was 299 changed.

300 Community sequencing and biomass quantification

301 The microbial communities were quantified using the 16S rDNA qPCR to determine the total bacterial 302 concentration and *nifH* gene qPCR for N₂-fixing bacteria. In parallel, 16S rDNA sequencing was performed 303 to identify major members of each community. This sequencing was also necessary to convert the amount 304 of 16S rDNA to total bacteria using the rrnDB-5.7 database. To analyze the communities present in 305 suspension, 1.8 mL of sample were collected for qPCR. For the cathodes, 1 cm² was recovered at several 306 times. The piece of carbon felt was then chopped with a sterile scalpel before being immersed in 20 mL of 307 sterile inorganic media. 1.8 mL was then recovered after shaking 20 mL of medium to resuspend as much 308 biomass as possible. These samples were then centrifuged 10 min 13500 RPM (12340g). The supernatant 309 was discarded and the pellets retained for DNA extractions. After qPCR, the concentrations measured in 310 the electrode samples were expressed considering the volume of medium.

311 Genomic DNA was extracted using the PowerSoil™ DNA Isolation Sampling Kit (MoBio Laboratories, 312 Inc., Carlsbad, CA, USA) according to the manufacturer's instructions. The qPCR amplification program was 313 performed in a BioRad CFX96 Real-Time Systems C1000 Touch thermal cycler (Bio-Rad Laboratories, USA). 314 For the analysis of total bacteria, primers 330F (ACGGTCCAGACTCCTACGGG) and 500R 315 (TTACCGCGGCTGCTGGCAC) were used. For the bacteria qPCR mix: SsoAdvanced™ Universal SYBR Green 316 Supermix (Bio-rad Laboratories, USA), primer 330F (200 nM), primer 500R (200 nM), 2 µL of DNA, and 317 water were used to a volume of 12 µL. The qPCR cycle was as follows: incubation for 2 min at 95°C and 40 318 cycles of dissociation (95°C, 10 s) and elongation (61°C, 20 s) steps. The results were then compared to a 319 standard curve to obtain the copy number of the target in the sample. Both the 16S rDNA concentration 320 of the PCE media and the cathodes are considered in the calculation of the total 16S rDNA concentration 321 of the PCE. These concentrations are used as an indicator of the biomass present and the use of a database 322 of the number of 16S operons per bacterial genome was used to estimate the actual amount of bacteria.

323 The presence of N_2 -fixing bacteria was monitored by qPCR of the *nifH* gene of the Fe-Fe subunit of 324 nitrogenases (Dos Santos et al., 2012; Gaby & Buckley, 2012). The *nifH* gene is known as marker of N₂ fixing 325 bacteria, common to all nitrogenases and is used for their quantification because of its necessary presence 326 for the fixation of N₂ (Dos Santos et al., 2012; Gaby & Buckley, 2012). All qPCR amplification programs were 327 performed in a BioRad CFX96 Real-Time Systems C1000 Touch thermal cycler (Bio-Rad Laboratories, USA). 328 The primers PoIF-TGCGAYCCSAARGCBGACTC and PoIRmodify reverse-AGSGCCATCATYTCRCCGGA were 329 used (Poly et al., 2001). The mixture: 6µl SsoAdvanced[™] Universal SYBR Green Supermix (Bio-rad 330 Laboratories, USA), F primer (500 nM), R primer (500 nM), 2 µL of DNA and water was used up to a volume 331 of 12 µL. The qPCR cycle was as follows: incubation for 2 min at 95°C and 40 cycles of dissociation (95°C, 332 30 s) and elongation (60°C, 30 s) steps. Then, the results were compared to a standard curve to obtain the 333 number of copies of the target in the sample. These two quantifications allow us to calculate the ratios of 334 N₂-fixing bacteria per total bacteria of the enrichments at different points to track the enrichment of N₂-335 fixing bacteria. This ratio can also help us to derive hypotheses on the functioning of our communities that 336 can be completed by the analysis of the communities during sequencing.

337 After quantification, our enriched communities were sequenced according to their 16S rDNA and the 338 results are available on NCBI repository PRJNA976100, Biosample SAMN28447998-SAMN28448066. The V3-V4 region of the 16S rDNA was amplified using universal primers as reported elsewhere (Carmona-339 340 Martínez et al., 2015). The PCR mixture consisted of MTP Taq DNA Polymerase (Sigma-Aldrich, Germany) 341 $(0.05 \text{ u/}\mu\text{L})$ with enzyme buffer, forward and reverse primers (0.5 mM), dNTPs (0.2 mM), sample DNA (5-342 10 ng/ μ L), and water to a final volume of 60 μ L. 30 cycles of denaturation (95°C, 1 min), annealing (65°C, 1 343 min), and elongation (72°C, 1 min) were performed in a Mastercycler thermal cycler (Eppendorf, Germany). 344 A final extension step was added for 10 min at 72 °C at the end of the 30th amplification cycle. PCR 345 amplifications were verified by the 2100 Bioanalyzer (Agilent, USA). The GenoToul platform (Toulouse, 346 France http://www.genotoul.fr) used an Illumina Miseq sequencer (2 x 340 bp pair-end run) for the 347 sequencing reaction. The raw sequences obtained were analyzed using bioinfomatic tools. Mothur version 348 1.39.5 was used for cleaning, assembly and quality control of the reads. Alignment was performed with 349 SILVA version 128 (the latter was also used as a taxonomic contour).

Communities sequenced on pre-enrichment bottles were used as initial community of the enrichment cultures. For PCE and nPCE, sequenced communities came from the biofilm formed on electrodes. Two replicates per potential were used. For H₂E bottles, 6 bottles were used for sequence analysis and qPCR. For pre-enrichment, sequences corresponded to each soil samples (leguminous, forest, compost) and a mix of them.

355 Data analysis

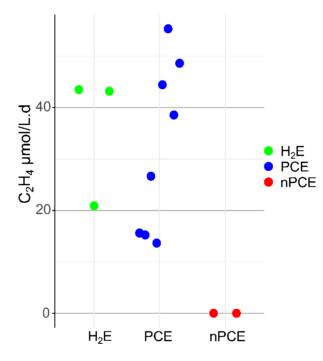
All results were analyzed using R (4.2.0) and Rstudio (2022.07.1) for calculations and graphics. The Tidyverse package was used for data manipulation(*Tidyverse*, n.d.). The packages ggplot2, ggpubr, scales, cowplot, corrplot and palettetown were used for the graphical representations. Visual representation of bacterial relative abundances was performed with the phyloseq package (McMurdie, 2011/2023). Inkscape software was also used to edit the graphs when necessary. The uncertainties shown for the values presented are standard deviations. All data and scripts used here are available online (Rous, 2023).

363

Results and discussion

364 Nitrogen fixation after 340 days of enrichment

365 N₂ fixation was quantified at the day 340 of the enrichment using acetylene reduction assays (ARA). 366 This assay was performed in H₂-fed bottle enrichments (named 'H₂E'), in polarized cathode enrichment 367 (named 'PCE') and in the non-polarized cathode enrichments as controls (named 'nPCE'). As shown in Figure 2, the ARA results confirmed the N₂ fixation capacity of the enriched communities (Bergersen, 1970). 368 369 This indicates that the cathode and/or H₂ was used as electron sources for N₂ fixation both in PCE and in 370 H_2E bottles. The average rates were similar in both enrichment methods with $32\pm17 \mu molC_2H_4/L.d$ in PCE 371 and 36±13 μ molC₂H₄/L.d in H₂E bottles. The PCE corresponding N₂ fixation rates ranged from 0.12 372 mg_{Nfixed}/L.d (minimum) to 0.51 mg_{Nfixed}/L.d (maximum), which is consistent with the rate of 0.2 mg_{Nfixed}/L.d 373 estimated by Rago et al. (2019) and also with N_2 fixation rates reported for soil bacteria (Hardy et al., 1973; 374 Kifle & Laing, 2016; Rago et al., 2019). Despite these significant N₂ fixation rates and the long duration of 375 the experiments, the rate of ammonium production in solution remained lower than 0.07 mg_N/Ld at the 376 day 340 (Table 2), indicating that most of the fixed N₂ was probably rapidly used by bacteria. 377



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Figure 2 - Reduction rate of acetylene in μ mol C₂H₄/L.d in the different reactors after 340 days of operation . or H₂ condition, 3 bottles were used for the acetylene reduction assay with one injection for each bottle, i.e. 3 measurements. For PCE, 2 reactors were used and 4 injections were made to validate the repeatability of the measurement when C₂H₄ 382 which gives 8 measurements for each of these two conditions. For nPCE, 2 reactors were used which gives 2 383 measurements for each conditions.

384 The ability of the microbial communities to fix N_2 was also assessed by gPCR of the *nifH* gene (Dos Santos et al., 2012; Pogoreutz et al., 2017). The amounts of N₂-fixing bacteria after 340 days of enrichment 385 386 are reported in Table 2. The average *nifH* gene concentration in PCE was estimated at 7.8 10^7 copies_{*nifH*}/mL, 387 two orders of magnitude higher than the average concentration of 8.3 10⁵ copies_{nift}/mL measured in the 388 H_2E enrichment bottles. This observation was surprising since N_2 -fixation rates were similar in both 389 configurations (Figure 2). This suggests that the fixation rate per *nifH* copy was much higher in H_2E than in PCE. The estimated specific activities per *nifH* copy were indeed of $0.2\pm0.3 \mu mol_{C2H4}/10^8$ copies_{*nifH*} d in the 390 391 PCE and $2.1\pm0.7 \mu mol_{C2H4}/10^8$ copies_{nifH}.d in H₂E bottles (Table 2). Furthermore, the comparison between 392 *nifH* gene and 16S rDNA copy numbers gives an idea of the proportions of N_2 fixing bacteria in each 393 microbial community. Interestingly, this proportion was 18% in H₂E bottles that was four times higher than 394 the value of 4.6% in PCE (Table 2). Therefore, N_2 -fixing bacteria constituted a smaller proportion of the

395 bacterial populations in PCE than in H_2E and only a small proportion of the bacteria participated to N_2 396 fixation in the PCE. In nPCE controls, the biomass was higher than in H₂E and the *nifH*/16S ratio lower (Table 397 2). This higher biomass concentration was likely due to the enrichment period in presence of organic C (day 398 0 to 115 including 60 days in BES) for the PCE and nPCE enrichments. NH₄⁺ in nPCE was also observed at a 399 rate of 0.07 mg/Ld as presented in Table 2 but without acetylene accumulation, meaning that no N_2 400 fixation occurred. In absence of electron source, nPCE enrichment communities could have been 401 maintained through cryptic growth. The presence of NH4⁺ in the nPCE was likely related to cell lysis since 402 no measurable fixation was detected by ARA even though significant biomass production was observed.

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406

Table 2 - The average ammonium production rates, the number of *nifH* gene copies, the number of 16S gene copies, 404 and the nifH/16S ratio were assessed after 340 days for the three experimental configurations. Average values were 405 measured on the last batch of 21 days, before 340 days of enrichment for the two polarized cathode enrichment (PCE), the

two non-polarized Cathode Enrichments (nPCE) and the six H₂ enrichment bottles (H₂E)

	PCE	nPCE	H ₂ E
N-NH4 ⁺ mg/L.d	0.07±0.01	0.07±0.01	0.04±0.03
nifH gene copies copies/mL	7.8±9.5 10 ⁷	2.5±2.8 10 ⁶	8.3±2.2 10 ⁵
16S rDNA gene copies copies/mL	1.7±1.9 10 ⁹	1.4±0.2 10 ⁸	4.6±1.5 10 ⁶
nifH/16S	3.8 %	1.7 %	19.0 %

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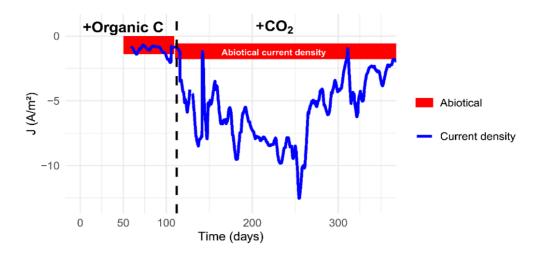
408 Current density and Autotrophic enrichment in polarized cathode enrichment

409 The average current density for the two PCE over experimental time is shown in Figure 3. As the current 410 measured at the cathode was negative by convention, a more negative corresponded to a higher reduction 411 activity. Regarding the current densities in the abiotic systems, the average current densities were 412 measured at -0.75 A/m² for two times four days with an organic C source and -1.1 A/m² for 16 days with 413 only CO₂ as the carbon source.

414 The average current density measured in the PCE was not different from the current density measured 415 in the abiotic control (approximately -1 A/m^2 with organic C supply) for the first 45 days of operation (days 416 55 to 100 of the enrichment). An increase up to -2 A/m² appeared between 100 and 115 days of enrichment 417 in PCE. CO_2 was then used as sole carbon source after this increase appeared. Following this change in 418 carbon source, a sharp increase in current density to -5 A/m² was observed in the PCE with regards to the 419 range of current densities in the abiotic controls (approximately -2 A/m² on CO₂)(Figure 3). The higher 420 current density was assumed to be associated to the electroactive activity of electron uptake by enriched 421 bacteria. The current density in the PCE then continuously increased until day 250 of the enrichment to 422 reach a value of -15 A/m^2 .

423 The high current density observed after 250 days of enrichment indicated a high redox activity linked 424 either to hydrogen evolution, oxygen reduction, or possibly direct electron transfer. As proposed by Z. 425 Zaybak et al. (2013), the high activity was probably resulting from a high metabolic activity in the biofilm 426 with significant microbial catabolic process (Zaybak et al., 2013). Compared to the current densities 427 obtained Rago et al. (2019) in the order of magnitude of -10 mA/m² at the same potential (-0.7 vs SHE), the 428 current densities observed here (-5 to -10 A/m²) were about 1000 times higher. These current density levels 429 are close to those measured by Zhang et al. (2022) who reported a maximum of -10 A/m^2 at the same 430 applied potential (Zhang et al., 2022).

431 After 230 days, power failures occurred, interrupting temporarily the current supply to the cathodes. 432 An important decrease of the current density was observed afterwards, down to -5 A/m² after 260 days 433 and -3 A/m² after 320 days. The lower current density reflected a change in the functioning of the microbial 434 communities, leading to less electron exchange with the cathode.



436

437 Figure 3 - Mean current density measured for the two PCE (blue line). Levels shown in red correspond to theoretical 438 mean current density and standard deviation estimated from two abiotic electrochemical systems current densities. For 439 abiotic electrochemical system, one batch of 2 days were made with Organic C and one batch of 16 days with CO₂. The 440 peaks observed are due to the batch operation of the PCE with disturbances each time the medium was renewed. Power 441 failures occurred at 230 days and 260 days. The time indicated on the x axis corresponds to the experimental time starting 442 at day 0 of the enrichment where soil samples were first introduced in the bottles with a medium containing organic C. Day 443 55 corresponds to the start of microbial electrochemical system with the precultured communities. The dashed line on day 444 110 corresponds to the passage on CO_2 as the sole carbon source in PCE. A 5-day curve smoothing was applied on PCE 445 current denstity curve.

446 **Biomass quantification**

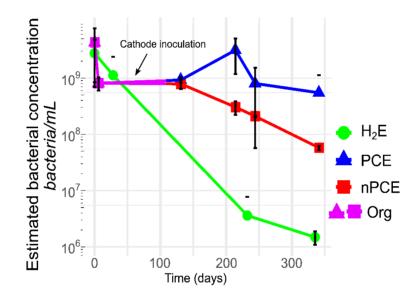
447 Bacterial biomass was monitored in the enrichments by measuring 16S rDNA concentrations by gPCR 448 (Figure 4). At day 131 (18 days after switching from organic C to CO₂), the average 16S rDNA concentrations 449 measured in the polarized cathode enrichment (PCE) and in the non-polarized cathode enrichment (nPCE) controls were 4.6 ± 0.4 10⁹ and 4.2 ± 1.2 10⁹ copies16SrDNA/mL, respectively. These concentrations 450 451 corresponded to 9.4±1.0 10⁸ and 8.0±1.5 10⁸ bacteria/mL, respectively, as presented in Figure 4. These 452 bacteria concentrations resulted from the first enrichment phase with organic C. During this phase, organic 453 substrates were used as carbon and electron sources for biomass growth in both configurations (PCE and 454 nPCE). At day 214, the 16S rDNA concentration in the PCE was used to calculate a concentration of 3.3±2.1 455 10⁹ bacteria/mL, corresponding to a biomass increase by a factor of 3.5 between 131 and 214 days (Figure 456 4). At the same time, the bacterial concentration dropped in nPCE from 8.0±1.5 10⁸ bacteria/mL to 3.0±0.8 457 10⁸ bacteria/mL. This drop was explained by the lack of available energy source for growth, which led to a 458 sharp decrease of the bacterial populations. At day 214, the 16S rDNA concentration in PCE was therefore 459 11-fold higher than in nPCE controls. This difference is consistent with the difference reported by Rago et 460 al. between polarized and non-polarized conditions, with electroactive biocathodes enriched in 461 autotrophic diazotrophic bacteria (Rago et al., 2019). These results suggested that the enriched microbial 462 communities were able to use the electrodes polarized at -0.7 V vs. SHE as sole electron sources to grow 463 while fixing N₂.

464 In the H_2 -fed enrichment (H_2E) bottles, the 16S rDNA concentration steadily decreased along the 465 experiment. The concentration decreased from 1.1±1.3 10⁹ bacteria/ml at the beginning of the enrichment 466 down to 1.3±0.3 10⁶ bacteria/ml after 340 days (Figure 4). These concentrations appear lower than the 467 biomass concentrations observed in the PCE medium at the same time of enrichment, ie. 4.5±5.9 10⁶ 468 bacteria/mL at day 244 and 8.6±8.8 10⁶ bacteria/mL at day 340. These results confirm that bacterial growth 469 was higher on the cathodes than in an H₂ supplied environment. It was therefore concluded that the PCE 470 provided more favorable environment for biomass growth than H₂ fed bottles as the surface provided by 471 the electrode was likely favorable for biofilm growth.

472 We also calculated the *nifH*/16S ratio representing the part of bacteria able to fix N₂ among the total 473 bacteria. A ratio of 0.0006 of *nifH* gene copies per 16S rDNA copy was measured for samples at the very 474 beginning of enrichment, both for H₂E and PCE. After 131 days, corresponding to the switch to CO_2 as sole 475 C-source, this level increased to 0.03 in PCE and 0.02 in nPCE control. These results are consistent with an 476 enrichment in diazotrophic bacteria during the enrichment phase on organic carbon (Bowers et al., 2008). 477 The bacterial enrichment in nPCE was likely possible as the organic C was used by the bacteria as energy 478 source. After 214 days, the level decreased to 0.02 in PCE but remained higher than the level at the 479 beginning of enrichment. This variation suggests interactions within the community that favored the 480 growth of non-N₂ fixing bacteria after the shift to CO₂ as sole C-source. After 340 days, the ratio of *nifH* to 481 16S rDNA was 0.04as presented in Table 2. In parallel, a ratio of nifH to 16S rDNA of 0.90 was measured for 482 H₂E at 244 days. Therefore, most of the bacteria were able to fix N₂ in H₂E bottles, confirming the efficient 483 enrichment in diazotrophic bacteria (Bowers et al., 2008). Given the loss of biomass observed in H₂E during 484 the experiment (Figure 4), this high ratio corresponded likely to the surviving bacteria that were selected 485 on their ability to fix N₂. The ratio measured in these H₂E then decreased down to 0.19, suggesting a 486 decrease in N₂-fixing bacteria in biomass.

487 As previously mentioned, after 230 days, power failures occurred and interrupted the polarization of 488 the electrodes. These interruptions impacted the microbial communities with a decrease in biomass 489 concentration to $8.1\pm7.6\ 10^8$ bacteria/mL at 244 days and $5.5\pm6.0\ 10^8$ bacteria/mL after 340 days compared 490 to the concentration of $3.3\ 10^9$ bacteria/mL measured at 214 days. At the same time, the *nifH*/16S rDNA 491 ratio increased up to 5%, indicating that N₂-fixing bacteria were more resistant. Nevertheless, a decrease 492 was observed in *nifH* quantities, from $2.3\ 10^8$ copies_{*nifH*}/mL after 214 days to $7.8\ 10^7$ copies_{*nifH*}/mL after 340 493 days.

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495

496 Figure 4 - Bacteria concentrations over time in the different enrichments calculated from 16S rDNA qPCR
 497 quantifications in bulk and biofilm. Green disks correspond to H₂ enrichments in bottles, blue triangles correspond to
 498 polarized cathode enrichment (PCE), red squares correspond to controls in non-polarized cathode enrichment (nPCE). The
 499 partially purple symbols marked Org correspond to the first phases of enrichment with organic C for the PCE and nPCE. The
 500 arrow indicates the transition from bottle enrichments to cathode enrichments in microbial electrochemical systems for
 501 PCE and nPCE. Error bars correspond to the calculated standard deviation.

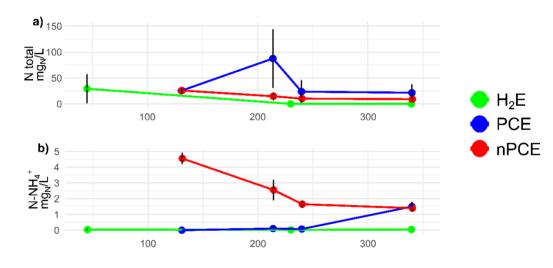
502 N quantification and coulombic efficiency

503 Total N contents of the different experiments are shown in Figure 5a. The total N corresponded to the 504 sum of the nitrogen measured in the liquid phase by N-ion concentration analysis (N-NH₄⁺, N-NO₃⁻, N-NO₂⁻ 505), in the medium by CHNS elemental analysis for PCE and nPCE, in the suspended biomass from qPCR results 506 only for H_2E where the dry mass was not measured, and on the electrode based on the bacterial 507 concentrations. The total N concentration was estimated after 131 days of enrichment (ie. before the shift 508 to CO₂ as sole C source) at 25.5 \pm 0.4 mg_N/L and 25.8 \pm 4.7 mg_N/L in the PCE and in the nPCE controls, 509 respectively. After 214 days of enrichment, total N increased up to 87.6 ± 56.1 mg_N/L in the PCE with regards 510 to the low value of 14.8 ± 7.0 mg_N/L in the nPCE controls. Ammonium represented only a small fraction of 511 the total N in the PCE. The maximum ammonium concentration observed at a batch end in the PCE was 512 $1.5\pm0.3 \text{ mg}_{N}/\text{L}$ at the day 340 of the enrichment in comparison with the maximum value of $4.5\pm0.4 \text{ mg}_{N}/\text{L}$ 513 found in nPCE control after 131 days (1.4 mg_N/L at the day 340)(Figure 5b). The average N fraction in the

form of ammonium was therefore of 12% in nPCE and only 1.6% in PCE. It was thus assumed that the higher level of ammonium in nPCE was related to the decay of biomass in absence of electron sources. In counterpart, the ammonium produced by N₂ fixation in the PCE was likely directly used for protein synthesis as suggested elsewhere (Bueno Batista & Dixon, 2019; Temple et al., 1998). In addition, nPCE N-NO₃⁻ concentrations were 5 to 15 times lower than N-NH₄⁺ concentrations with maximum values of 0.9 mg_N/L N-NO₃⁻ in the nPCE control and 0.2 mg_N/L N-NO₃⁻ in PCE. An average concentration of 0.1±0.2 mg_N/L N-NO₃⁻ in H₂E bottles was also observed. NO₂⁻ concentrations were negligible.

Using our method of N mass estimation on biomass lost as presented in Figure 4 for nPCE, the total concentration of N lost by biomass would be estimated around 130 to 200 mgN/L depending on the community (2 nPCE). This loss would then be equivalent to rates of 0.6 to 0.9 mgN/L.d released by this the biomass on average along the enrichment. Assuming a loss of a constant portion of biomass, a release rate of 0.1 mgN/L.d was estimated as average for the batch ending at 340 days, close to the 0.07 mg/L.d presented in Table 2. This result supported our hypothesis that NH₄⁺ release was linked to biomass loss in nPCE enrichments after organic C addition was stopped.

The H₂ enrichment bottles (H₂E) showed an average N-NH₄⁺ accumulation of 35.6±36 μ g_N/L after 131 days and an average of 0.1±0.2 mg_N/L over the duration of enrichment. This concentration represented a small fraction of the total N at the very beginning of the enrichment, with an estimated concentration of 29.4±28 mg_N/L. The N concentration decreased during the enrichment, which is consistent with the biomass loss as shown in Figure 4. H₂E were therefore less efficient for N₂ accumulation than polarized cathode enrichment, with a lower microbial biomass production.



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Figure 5 - (a) Total N concentration based on the sum of (1) N estimated from biomass measurement (suspended biomass and biofilm), (2) N content in ionic forms (N-NH₄⁺, N-NO₃, N-NO₂) and (3) N measured in the dry weight of the medium of the polarized cathode enrichment (PCE) and non-polarized cathode enrichment (nPCE) and (b) N-NH₄⁺ concentration in H₂ enrichment (H₂E), PCE and nPCE.

540 Current densities and rates of acetate production, N₂ fixation and biomass growth are shown in Table 541 3. Coulombic efficiencies associated with each reaction were estimated based on these results (Table 3). 542 During the first 214 days of enrichment, 0.6 to 3.3% of the electrons were used for N₂ fixation in the two 543 PCE. In comparison, efficiencies of 0.5% and 20% for NH₄⁺ synthesis was reported in two recent works 544 carried out under similar conditions (Yadav et al., 2022; Zhang et al., 2022). As the amount of fixed N was 545 highly dependent on biomass accumulation, negative results were obtained at day 244 when biomass 546 started to decrease. The electrons used at the cathode for biomass synthesis during the first period (131 547 to 214 days) accounted for 2.8% and 17.3% in the two PCE. These high coulombic efficiencies were probably 548 also associated to acetotrophic and acetogenic bacteria. Indeed, acetate produced using electron from the 549 cathode could have been used by acetotrophic bacteria for growth, lowering coulombic efficiency of 550 acetogenesis and increasing CE for bacterial growth. As acetogenic bacteria do not tolerate the presence

of oxygen, dissolved O₂ was very probably consumed in some part of the biofilm, leaving other parts in strict anaerobic conditions more favorable for acetogenic bacteria growth.

553 The H₂ recovered in the headspace of the PCE accounted for 12 to 22% of the electrons supplied to the 554 cathode as presented in Table 3. Therefore, H_2 was not related to the biological activity and mostly resulted 555 from an abiotic reaction at the cathode. In addition, in presence of O_2 in the cathodic chamber, oxygen 556 reduction reactions were expected with regards to the potential used in this study. Indeed, a two-electron 557 reduction could have occurred, resulting in the production of hydrogen peroxide (H_2O_2), which can then 558 undergo further reduction to form water (H₂O) (Rozendal et al., 2009; Sim et al., 2015). Hydrogen peroxide 559 was not measured, however, the amount of biomass found on the cathodes (Figure 4) suggested that the 560 concentrations of hydrogen peroxide were sufficiently low to have minimal to no impact on the microbial 561 community during the enrichment process. Nonetheless, a fraction of the electrons may have been lost 562 through these oxygen reduction reactions, which could partially account for the low coulombic efficiencies 563 observed in this study. Interestingly, a significant production of acetate was also observed. An average rate 564 of 149.1 µmol/L.d and 421.6 µmol/L.d were measured in both PCE for the period from day 131 to day 214, 565 as presented in Table 3. Acetate production almost stopped with the power failures with acetate measured 566 only on one to two batches per PCE. This decrease correspond to rate of 61.0 µmol/L.d and 57.4 µmol/L.d 567 of acetate in both PCE. Acetate production accounted for 7.9 and 39% of the cathodic electrons during the 568 first period (up to 214 days) and for less than 5% after power failures. To explain the decrease in acetate 569 production, biomass growth and power consumption, it was hypothesized that acetate production might 570 have been due to a specific loss of members able to fix CO₂, and more especially autotrophic bacteria using 571 cathode as electron source (H₂ or DET), within the enriched community. Thus, with acetate no longer being 572 produced, heterotrophic bacteria did not have enough organic C to sustain their growth, causing their 573 decrease. Therefore, autototrophic bacteria responsible for CO₂ fixation and heterotrophic bacteria that 574 could also be H_2 dependent for N_2 fixation were greatly affected, leading to a decrease in the reduction 575 reactions at the cathode and subsequently in current density. In addition, acetate was not found in the H₂E 576 bottles, indicating the absence of acetogenesis and an important difference in microbial pathways and/or 577 communities. As seen in the Table 3, electrons were retrieved in biomass production, N₂ fixation products, 578 H₂ found in headspace and in CH₃COOH product from CO₂. These products were not sufficient to close the 579 electron mass balance. The loss of electrons and the differences between cathodes of PCE 1 and PCE 2 was 580 explained by side reactions, such as O_2 reduction or biological reaction such as exopolysaccharide (EPS) 581 production.

Table 3 - Current densities, rates and coulombic efficiencies for the two polarized cathode enrichment (PCE) over two
 different periods of current density with CO₂ as sole carbon source. During the first period (131-214 days) current density
 increased, whereas during the second period (>215) current density decreased after several power outages (see Figure 3).

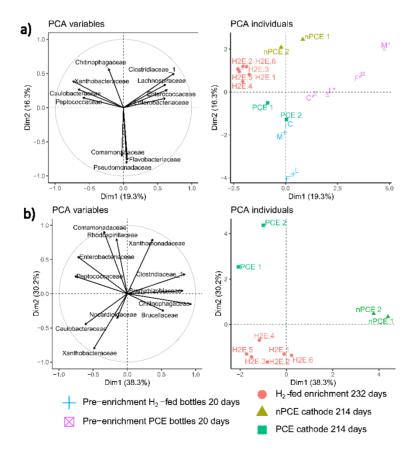
		131-214 days		> 215 days	
		PCE 1	PCE 2	PCE 1	PCE 2
J A/m²	Mean	7.7±3.1	5.9±2.4	6.2±4.1	3.8±2.2
	Max	28.2	11.3	19.5	9.9
Acetate µmol/L.d	Mean	149.1±203.1	421.6±216.7	57.4±151.8	61.0±106.5
N μmol/L.d	N-NH4 ⁺	0.108±0.225	0.155±0.249	2.0±1.4	1.5±1.9
	N Bulk Dry weigth	10.3±6.0	27.5±13.4	9.0±6.4	12.1±6.4
	N Biomass theoretical (16S/bact)	16.7	80.4	-3.0	-7
Biomass bact/L.j	Bulk	0.4 ± 0.5 10 ⁹	3.5 ± 2.0 10 ⁹	0.4±0.5 10 ⁹	0.2±0.1 10 ⁹
	Cathode	0.9 10 ¹⁰	4.6 10 ¹⁰	-2.6±3.5 10 ¹⁰	-5.9±7.8 10 ¹⁰
Coulombic efficiency %	CO ₂ to Acetate	7.9±9.5	30.9±8.9	2.1±5.6	5.0±6.6
	H ⁺ to H ₂	22.3±15.4	12.1±8.8	9.4±13.5	14.4±14.4
	N ₂ fixation	0.6±0.2	3.3±0.8	0.2±1.0	-0.5±2.2
	Biomass growth	2.8±0.5	17.3±5.3	-2.5±4.0	-8.3±14.2
	Total	33.6±14.4	63.6±9.0	9.2±16.5	10.5±18.8

In comparison with the other works dealing with N₂-fixing cathodic biofilms, Zhang et al (2022) showed a maximum of 40.5 mg/L of NH₄⁺ in 4 days with mixed communities, in regard to a maximum of 0.8 mg/L NH₄⁺ observed in Yadav et al (2022) and 6.31 mg/L NH₄⁺ in 10 days for Li et al. (Li et al., 2022; Yadav et al., 2022; Zhang et al., 2022). When demonstrating the N₂ fixation in MES, Rago et al (2019) showed a N₂ fixation of 0.2 mgN/L.d in biomass and 5 10⁹ bacteria/L.d (Rago et al., 2019). In the present study, biomass production in biofilms was 2 to 10 times higher than in Rago et al. (2019), as was the nitrogen found in the biomass varied between 0.2 and 1 mg/L.d before 214 days (Rago et al., 2019).

593 Microbial communities

16S rDNA sequencing was performed at the end of pre-enrichment , and at 214 or 232 days of enrichment in polarized cathode enrichment (PCE) and in H₂-fed enrichment bottles (H₂E), respectively. The sampling days were selected because they were associated to a high microbial activity (high current densities and high biomass concentrations). In H₂E, the *nifH*/16S abundance ratio was also maximum (0.9) at day 232. Principal Component Analysis (PCA) was used to present the communities for each enrichment. Each reactors and bottles are presented as individuals and major families as variables in PCA presented in the Figure 6 and relative abundances are presented in the Figure 7.

601 For communities at the end of the pre-enrichment, the principal component analysis (PCA) showed an 602 important link between the families of the Clostridiaceae, Enterobacteriaceae, Enterococcaceae and 603 Lachnospiraceae and with the PCE pre-enrichment (Figure 6a). Indeed, the group of communities of pre-604 enrichment with organic C is well separated from the others groups and follow the same direction as these 605 four families. A dominance of the Enterobacteriaceae family (mainly of the Citrobacter genus) was 606 observed for each pre-enriched sample expect for the pre-mixed sample where the three others families are highly present. These families are therefore absent or very weakly represented in the other sequenced 607 608 communities as shown in Figure 7.



609

Figure 6 - Results of the principal component analysis (PCA) performed on the microbial communities of a) pre enrichments, H₂-fed enrichment bottles (H₂E) after 232 days and cathodes of PCE and nPCE enrichments after 214 days and
 b) H₂-fed enrichment bottles (H₂E) after 232 days and cathodes of PCE and nPCE enrichments after 214 days. Only families
 of the five major bacterial OTU in each sampled community were used for the analysis. The microbial communities in the

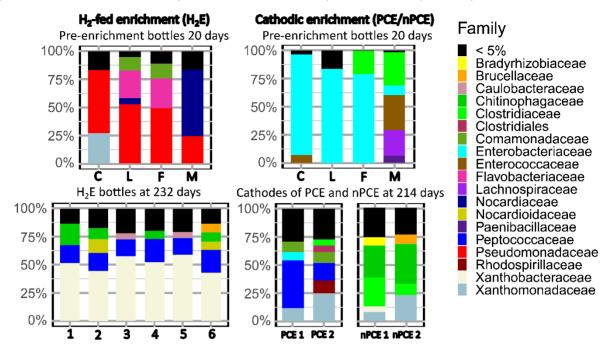
614pre-enrichment bottles are represented by the following abbreviations: F for forest soil, C for compost, L for the615rhizosphere of leguminous plants, and M for a mix of all. Variables least close to the correlation circle are not displayed616(cos2 < 0.2).</td>

617 For the H₂-fed bottle pre-enrichments (H₂E), the group is also seprated of the 232-days enriched 618 community H₂E. Pseudomonadaceae (45% of average relative abundance) family was mostly dominant at 619 the end of pre-enrichment. Nocardiaceae, Flavobacteriales, Xanthomonadaceae and Comamonadaceae 620 were also present as seen in Figure 7. Flavobacteriales and Comamonadaceae families are also higly linked 621 with the group of pre-enrichment in PCA of Figure 6a. These families, with the exception of 622 Flavobacteriales, are also known to have members possessing the set of genes necessary for N₂ fixation 623 (Dos Santos et al., 2012; Ghodhbane-Gtari et al., 2019; Huda et al., 2022). These families accounted for 624 77% of the sequences which is high compared to the nifH/16S rDNA ratio of less than 0.01 at the same time 625 point. This suggests that either the *nifH* primers were not adapted to these specific species or that the 626 species found at this point did not possess the genes for nitrogenases. As the H₂E pre-enrichment cultures 627 started on a medium containing NH₄Cl, the presence of this source of nitrogen was likely favorable to the 628 growth of non-diazotrophic bacteria.

629 After 214 days of enrichment, PCE communities were affiliated to Peptococcaceae (29% in average), 630 Xanthomonadaceae (18% in average), Rhodospirillaceae (11% in PCE 2, Azospirillum) and 631 Comamonadaceae (10% in average) as presented in Figure 7. As seen in the PCA presented Figure 6b, 632 Rhodospirillaceae, Comamonadaceae, Enterobacteriaceae and Xanthomonadaceae families are 633 representative of the PCE cathode communities. *Peptococaccaceae* appear to be shared with communities 634 of H₂-fed enrichment (H₂E) bottles. As seen in Figure 6a and Figure 7, a clear shift in microbial communities 635 from the end of pre-enrichment was therefore observed as the difference between PCE communities at 636 214 days and at the end of pre-enrichment.

637 Only Enterobacteriaceae family was maintained although at minor relative abundance. Members of 638 Clostridiales incertae sedis absent from original inoculum also appeared on polarized cathode. These 639 families are known to exhibit the role of plant growth promoting bacteria (PGPB). These communities could 640 thus be beneficial when used as living fertilizers (Cassan & García de Salamone, 2008; Rojas-Tapias et al., 641 2012; Singh et al., n.d.). The Comamonadaceae as well as the Enterobacteriaceae families mostly include 642 heterotrophic species, which would be consistent with our hypotheses about the existence of interactions 643 between heterotrophic and autotrophic populations (F. Liu et al., 2011; Wu et al., 2018). More precisely, 644 the Peptococcaceae sequences were affiliated to species Desulforamulus ruminis (>98%). This species was 645 already described for their ability to fix N₂ (Postgate, 1970). The Desulforamulus and Desulfotomaculum 646 genera have also several species able to grow with H_2 and CO_2 as electron and C sources (Aullo et al., 2013; 647 Klemps et al., 1985; Zaybak et al., 2013). They were previously reported to be able to produce acetate by 648 CO₂ reduction through the Calvin cycle (Klemps et al., 1985), and some were already found in microbial 649 electrochemical system on a biocathode producing acetate (Zaybak et al., 2013). The other main family, 650 Xanthomonadaceae, was represented by several genera with a majority of Pseudoxanthomonas. In this 651 genus, some members were identified as N₂ fixers with a need of external organic C source, exhibiting a 652 mixotrophic metabolism depending on the environmental conditions (J. Hu et al., 2022; Ryan et al., 2009). 653 Sequences associated to the Rhodospirillaceae family were mainly affiliated to the species Azospirillum 654 lipoferum which is able to grow in autotrophy with H₂, CO₂ and N₂ (Tilak et al., 1986). This soil bacterium is 655 also known for its role as a PGPB with a capacity to solubilize phosphates, making it a good candidate as a 656 fertilizer (Cassan & García de Salamone, 2008; Tilak et al., 1986). Interestingly, many of the identified 657 bacteria in the polarized cathode enrichment were previously described to possess the N₂-fixing genes and 658 capability. This supports the fact that the primers were not able to amplify the full diversity of nifH genes 659 from these communities.

660 The Xanthobacteraceae (51% in average), Peptococcaceae (17%, identified as Desulforamulus), 661 Chitinophagaceae (8%) and Nocardioidaceae (5%) families were found to be dominant in H₂E bottles at day 662 214 as presented in Figure 7. The Xanthobacteraceae family, highly linked to H₂E communities as seen in 663 Figure 6b, was mostly represented by the species Xanthobacter autotrophicus which is known as N₂-fixing 664 HOB (Wiegel, 2005). This species was already been used for N₂ fixation by Liu et al. (C. Liu et al., 2017) in 665 an hybrid system using the H₂ produced by a cathode. Xanthobacter autotrophicus was also found in the 666 medium of the polarized cathode enrichment but in lower abundance (< 5%). Therefore, the community 667 enriched with H₂ in bottles was mostly composed of N₂-fixing bacteria, as also supported by the high 668 *nifH*/16S ratio (0.9). After 214 days of enrichment, diazotrophic HOB were specifically selected.



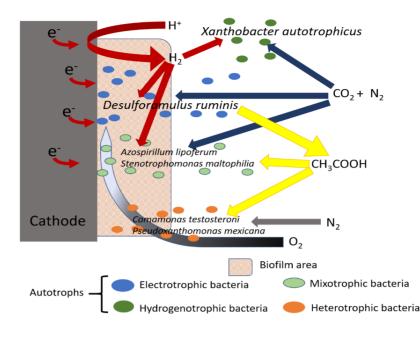
669

Figure 7 - Barplots of relative abundances of major bacterial families of pre-enrichments, of H₂-fed enrichment (H₂E
 after 232 days) and cathodic enrichment (PCE and nPCE after 214 days). The microbial communities in the pre-enrichment
 bottles are represented by the following abbreviations: F for forest soil, C for compost, L for the rhizosphere of leguminous
 plants, and M for a mix of all. Only families with a relative abundance ≥ 5% are shown for each sample.

674

The presence of mixotrophic and heterotrophic bacteria in the PCE suggested that carbon-based interactions could have occurred. Acetate was the only abundant soluble carbon metabolite found in these enrichments (see Table 3). Therefore, acetate was assumed to be used as intermediate for carbon and electron transfer between autotrophic homoacetogens, e.g. *Desulforamulus rumnis*, and heterotrophic bacteria such as *Comamonas sp.*).

680 Furthermore the low concentration of N-NH4⁺ in the PCE before day 210 (Table 3) was probably due to 681 its rapid consumption for bacteria growth. Considering these hypothesis, a conceptual scheme of microbial interactions between the main bacterial families in the PCE was proposed and is presented in Figure 8. The 682 683 presence of heterotrophic bacteria and their potential use of O₂ as a final electron acceptor was also 684 considered. The concentration of dissolved O_2 would have decreased in a deep layer of the biofilm due to 685 its use by heterotrophic bacteria. A structure of the biofilm in two layers could then be proposed with a 686 first layer composed mainly of homoacetogens fixed on the cathode and reducing CO₂ to acetate, and a 687 second layer composed mainly of heterotrophic bacteria using acetate and dissolved O₂ to sustain their growth. It was assumed that bacteria in the first layer would not access to N₂ that would be mostly fixed 688 689 by organisms of the second layer.



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Figure 8 - Conceptual scheme of microbial interactions occurring in polarized cathode enrichment (PCE) after enrichment for N₂ fixation with inorganic energy and carbon sources

Conclusion

695 Enrichment cultures of N_2 -fixing bacteria were successfully carried out in H_2 -fed bottles (H_2E) and in 696 polarized cathode enrichment (PCE). Both methods showed significant N₂ fixation after 340 days of 697 enrichment. The microbial communities selected were able to fix N_2 with CO_2 as sole carbon source and H_2 698 or cathodic electrons as sole electron sources. Biomass growth on the cathode up to 4.6 10¹⁰ bacteria/L.d 699 is another evidence of autotrophic growth in the PCE while bacterial growth was much lower in the H_2E . 700 Current density suggests the activity of autototrophic bacteria in the PCE and the availability of electron 701 sources. As the coulombic efficiency of N₂ fixation was low with a maximum of 3.3% and considering the 702 low concentrations of NH4⁺, it was concluded that the major part of the nitrogen was incorporated into 703 microbial biomass during the enrichment procedure. Interestingly, acetate was also produced in the PCE 704 corresponding to a coulombic efficiency of 27%. The related microbial communities found in both 705 enrichments had some bacterial families in common, but the communities found in the PCE appeared 706 metabolically more diverse, suggesting probable rich microbial interactions with exchanges of electrons, 707 carbon and nitrogen between autotrophic, heterotrophic and mixotrophic populations. Several members 708 of the enriched communities were furthermore reported as plant growth promoting bacteria (PGPB) which 709 could be interesting for the production of environment friendly fertilizers. To summarize, a conceptual 710 model of microbial interactions between the main bacterial families found in the bioelectrochemical 711 system was proposed suggesting a key role of each autotrophic, heterotrophic and mixotrophic 712 populations in the process of N_2 fixing by cathodic biofilms. In order to focus on the enriched microbial 713 community and avoid to disrupt the enrichments, a comprehensive screening with different potential 714 electron acceptors was not performed here. However, such screening would be interesting to be 715 investigated and could be the subject of further studies on synthetic community to further explore the 716 impact of electron acceptors on microbial communities, current densities and coulombic efficiencies.

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724	Data and scripts are available online on: https://doi.org/10.57745/ONNGWZ (Recherche Data gouv)			
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726 727	The authors declare that they comply with the PCI rule of having no financial conflicts of interest in relation to the content of the article.			
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730	References			
731 732 733	Aullo, T., Ranchou-Peyruse, A., Ollivier, B., & Magot, M. (2013). Desulfotomaculum spp. And related gram-positive sulfate-reducing bacteria in deep subsurface environments. <i>Frontiers in Microbiology</i> , 4. <u>https://www.frontiersin.org/article/10.3389/fmicb.2013.00362</u>			
734 735	Bagali, S. (2012). Review: Nitrogen fixing microorganisms. <i>Int. J. Microbiol. Res., 3</i> , 46–52. https://doi.org/10.5829/idosi.ijmr.2012.3.1.61103			
736 737 738	Bergersen, F. J. (1970). The Quantitative Relationship between Nitrogen Fixation and the Acetylene-Reduction Assay. <i>Australian Journal of Biological Sciences</i> , 23(4), Article 4. <u>https://doi.org/10.1071/bi9701015</u>			
739 740 741	Bowers, T. H., Reid, N. M., & Lloyd-Jones, G. (2008). Composition of nifH in a wastewater treatment system reliant on N2 fixation. <i>Applied Microbiology and Biotechnology</i> , <i>79</i> (5), 811–818. https://doi.org/10.1007/s00253-008-1486-2			
742 743 744	Bueno Batista, M., & Dixon, R. (2019). Manipulating nitrogen regulation in diazotrophic bacteria for agronomic benefit. <i>Biochemical Society Transactions</i> , <i>47</i> (2), 603–614. <u>https://doi.org/10.1042/BST20180342</u>			
745 746	Burris, R. H., & Roberts, G. P. (1993). Biological Nitrogen Fixation. <i>Annual Review of Nutrition, 13</i> (1), Article 1. <u>https://doi.org/10.1146/annurev.nu.13.070193.001533</u>			
747 748 749	Carmona-Martínez, A. A., Trably, E., Milferstedt, K., Lacroix, R., Etcheverry, L., & Bernet, N. (2015). Long-term continuous production of H2 in a microbial electrolysis cell (MEC) treating saline wastewater. <i>Water Research, 81</i> , 149–156. <u>https://doi.org/10.1016/j.watres.2015.05.041</u>			
750 751	Cassan, F., & García de Salamone, I. (2008). Azospirillum: Cell physiology, plant response, agronomic and environmental research in Argentina. ISBN 978-987-98475-8-9			
752 753	Chakraborty, T., & Akhtar, N. (2021). Biofertilizers: Prospects and Challenges for Future. In <i>Biofertilizers</i> (pp. 575–590). John Wiley & Sons, Ltd. <u>https://doi.org/10.1002/9781119724995.ch20</u>			

- 754 Chen, S., Jing, X., Yan, Y., Huang, S., Liu, X., Chen, P., & Zhou, S. (2020). Bioelectrochemical nitrogen
- fixation to extracellular ammonium by Pseudomonas stutzeri. *Applied and Environmental Microbiology*, 87(5). <u>https://doi.org/10.1128/AEM.01998-20</u>
- 757 Cherkasov, N., Ibhadon, A. O., & Fitzpatrick, P. (2015). A review of the existing and alternative
- 758 methods for greener nitrogen fixation. *Chemical Engineering and Processing: Process Intensification*,
 759 90, 24–33. https://doi.org/10.1016/j.cep.2015.02.004
- Deng, J., Iñiguez, J. A., & Liu, C. (2018). Electrocatalytic Nitrogen Reduction at Low Temperature. *Joule*, *2*(5), 846–856. https://doi.org/10.1016/j.joule.2018.04.014
- Dos Santos, P. C., Fang, Z., Mason, S. W., Setubal, J. C., & Dixon, R. (2012). Distribution of nitrogen
 fixation and nitrogenase-like sequences amongst microbial genomes. *BMC Genomics*, *13*(1), 162.
 https://doi.org/10.1186/1471-2164-13-162
- 765 Franche, C., Lindström, K., & Elmerich, C. (2009). Nitrogen-fixing bacteria associated with leguminous
- and non-leguminous plants. *Plant and Soil*, *321*(1), 35–59. <u>https://doi.org/10.1007/s11104-008-</u>
 9833-8
- Gaby, J. C., & Buckley, D. H. (2012). A Comprehensive Evaluation of PCR Primers to Amplify the nifH
 Gene of Nitrogenase. *PLoS ONE*, *7*(7), e42149. <u>https://doi.org/10.1371/journal.pone.0042149</u>
- 770 Ghodhbane-Gtari, F., Nouioui, I., Hezbri, K., Lundstedt, E., D'Angelo, T., McNutt, Z., Laplaze, L.,
- Gherbi, H., Vaissayre, V., Svistoonoff, S., Ahmed, H. ben, Boudabous, A., & Tisa, L. S. (2019). The
- plant-growth-promoting actinobacteria of the genus Nocardia induces root nodule formation in
- 773 Casuarina glauca. *Antonie van Leeuwenhoek*, *112*(1), 75–90. <u>https://doi.org/10.1007/s10482-018-</u>
 774 1147-0
- Hardy, R. W. F., Burns, R. C., & Holsten, R. D. (1973). Applications of the acetylene-ethylene assay for
 measurement of nitrogen fixation. *Soil Biology and Biochemistry*, 5(1), 47–81.
- 777 <u>https://doi.org/10.1016/0038-0717(73)90093-X</u>
- Heldal, M., Norland, S., & Tumyr, O. (1985). X-ray microanalytic method for measurement of dry
- matter and elemental content of individual bacteria. *Applied and Environmental Microbiology*, *50*(5),
 1251–1257. <u>https://doi.org/10.1128/aem.50.5.1251-1257.1985</u>
- Hu, J., Tang, H., Wang, Y. Z., Yang, C., Gao, M., Tsang, Y. F., & Li, J. (2022). Effect of dissolved solids
- released from biochar on soil microbial metabolism. *Environmental Science: Processes & Impacts*,
 24(4), 598–608. <u>https://doi.org/10.1039/D2EM00036A</u>
- Hu, X., Kerckhof, F.-M., Ghesquière, J., Bernaerts, K., Boeckx, P., Clauwaert, P., & Boon, N. (2020).
- 785 Microbial Protein out of Thin Air: Fixation of Nitrogen Gas by an Autotrophic Hydrogen-Oxidizing
- 786Bacterial Enrichment. Environmental Science & Technology, 54(6), Article 6.
- 787 <u>https://doi.org/10.1021/acs.est.9b06755</u>
- Huda, N. ul, Tanvir, R., Badar, J., Ali, I., & Rehman, Y. (2022). Arsenic-Resistant Plant Growth
- 789 Promoting Pseudoxanthomonas mexicana S254 and Stenotrophomonas maltophilia S255 Isolated
- from Agriculture Soil Contaminated by Industrial Effluent. *Sustainability*, *14*(17), Article 17.
 <u>https://doi.org/10.3390/su141710697</u>
- 792 Kandemir, T., Schuster, M. E., Senyshyn, A., Behrens, M., & Schlögl, R. (2013). The Haber–Bosch
- 793 Process Revisited: On the Real Structure and Stability of "Ammonia Iron" under Working Conditions.

- 794 Angewandte Chemie International Edition, 52(48), Article 48.
- 795 <u>https://doi.org/10.1002/anie.201305812</u>
- 796 Khan, K. S., Mack, R., Castillo, X., Kaiser, M., & Joergensen, R. G. (2016). Microbial biomass, fungal
- and bacterial residues, and their relationships to the soil organic matter C/N/P/S ratios. *Geoderma*,
 271, 115–123. <u>https://doi.org/10.1016/j.geoderma.2016.02.019</u>
- 799 Kifle, M. H., & Laing, M. D. (2016). Isolation and Screening of Bacteria for Their Diazotrophic
- 800 Potential and Their Influence on Growth Promotion of Maize Seedlings in Greenhouses. *Frontiers in*
- 801 *Plant Science*, 6. <u>https://doi.org/10.3389/fpls.2015.01225</u>
- Kim, J., & Rees, D. C. (1994). Nitrogenase and biological nitrogen fixation. *Biochemistry*, 33(2), Article
 <u>https://doi.org/10.1021/bi00168a001</u>
- Klemps, R., Cypionka, H., Widdel, F., & Pfennig, N. (1985). Growth with hydrogen, and further
 physiological characteristics of Desulfotomaculum species. *Archives of Microbiology*, *143*(2), 203–
 208. <u>https://doi.org/10.1007/BF00411048</u>
- 807 Li, F., Li, F., Lin, Y., Guo, L., Zhang, L., Li, R., Tian, Q., Wang, Y., Wang, Y., Zhang, X., Liu, J., & Fan, C.
- 808 (2022). Investigating the performance and mechanism of nitrogen gas fixation and conversion to
- ammonia based on biocathode bioelectrochemistry system. *Journal of Chemical Technology &*
- 810 *Biotechnology*, *97*(8), 2163–2170. <u>https://doi.org/10.1002/jctb.7092</u>
- Liu, A., Yang, Y., Ren, X., Zhao, Q., Gao, M., Guan, W., Meng, F., Gao, L., Yang, Q., Liang, X., & Ma, T.
- 812 (2020). Current Progress of Electrocatalysts for Ammonia Synthesis Through Electrochemical
- 813 Nitrogen Reduction Under Ambient Conditions. *ChemSusChem*, *13*(15), 3766–3788.
- 814 <u>https://doi.org/10.1002/cssc.202000487</u>
- Liu, C., Sakimoto, K. K., Colón, B. C., Silver, P. A., & Nocera, D. G. (2017). Ambient nitrogen reduction
- cycle using a hybrid inorganic–biological system. *Proceedings of the National Academy of Sciences*,
 114(25), Article 25. <u>https://doi.org/10.1073/pnas.1706371114</u>
- Liu, F., Zhao, C., Xia, L., Yang, F., Chang, X., & Wang, Y. (2011). Biofouling characteristics and
- 819 identification of preponderant bacteria at different nutrient levels in batch tests of a recirculating
- 820 cooling water system. *Environmental Technology*, *32*(8), 901–910.
- 821 https://doi.org/10.1080/09593330.2010.517220
- Loferer-Krößbacher, M., Klima, J., & Psenner, R. (1998). Determination of Bacterial Cell Dry Mass by
- 823 Transmission Electron Microscopy and Densitometric Image Analysis. *Applied and Environmental*
- 824 Microbiology, 64(2), 688–694. https://doi.org/10.1128/AEM.64.2.688-694.1998
- Logan, B. E., Rossi, R., Ragab, A., & Saikaly, P. E. (2019). Electroactive microorganisms in
- bioelectrochemical systems. *Nature Reviews Microbiology*, *17*(5), Article 5.
- 827 <u>https://doi.org/10.1038/s41579-019-0173-x</u>
- 828 Martín, A. J., Shinagawa, T., & Pérez-Ramírez, J. (2019). Electrocatalytic Reduction of Nitrogen: From
- Haber-Bosch to Ammonia Artificial Leaf. *Chem*, 5(2), Article 2.
- 830 https://doi.org/10.1016/j.chempr.2018.10.010
- 831 Masclaux-Daubresse, C., Daniel-Vedele, F., Dechorgnat, J., Chardon, F., Gaufichon, L., & Suzuki, A.
- 832 (2010). Nitrogen uptake, assimilation and remobilization in plants: Challenges for sustainable and
- 833 productive agriculture. Annals of Botany, 105(7), 1141–1157. https://doi.org/10.1093/aob/mcq028

- McMurdie, P. J. (2023). *Phyloseq* [R]. <u>https://github.com/joey711/phyloseq</u> (Original work published
 2011)
- 836 Moscoviz, R., Desmond-Le Quéméner, E., Trably, E., & Bernet, N. (2019). Bioelectrochemical Systems
- 837 for the Valorization of Organic Residues. In J.-R. Bastidas-Oyanedel & J. E. Schmidt (Eds.), *Biorefinery:*
- 838 Integrated Sustainable Processes for Biomass Conversion to Biomaterials, Biofuels, and Fertilizers
- 839 (pp. 511–534). Springer International Publishing. <u>https://doi.org/10.1007/978-3-030-10961-5_21</u>
- Paul, D., Noori, M. T., Rajesh, P. P., Ghangrekar, M. M., & Mitra, A. (2018). Modification of carbon
- 841 felt anode with graphene oxide-zeolite composite for enhancing the performance of microbial fuel
- 842 cell. Sustainable Energy Technologies and Assessments, 26, 77–82.
- 843 https://doi.org/10.1016/j.seta.2017.10.001
- 844 Peoples, M. B., & Craswell, E. T. (1992). Biological nitrogen fixation: Investments, expectations and
- actual contributions to agriculture. *Plant and Soil*, 141(1), 13–39.
- 846 <u>https://doi.org/10.1007/BF00011308</u>
- 847 Pogoreutz, C., Rädecker, N., Cárdenas, A., Gärdes, A., Wild, C., & Voolstra, C. R. (2017). Nitrogen
- 848 Fixation Aligns with nifH Abundance and Expression in Two Coral Trophic Functional Groups.
- 849 Frontiers in Microbiology, 8. https://www.frontiersin.org/articles/10.3389/fmicb.2017.01187
- Poly, F., Ranjard, L., Nazaret, S., Gourbière, F., & Monrozier, L. J. (2001). Comparison of nifH Gene
- 851 Pools in Soils and Soil Microenvironments with Contrasting Properties. Applied and Environmental
- 852 Microbiology, 67(5), 2255–2262. https://doi.org/10.1128/AEM.67.5.2255-2262.2001
- Postgate, J. R. (1970). Nitrogen Fixation by Sporulating Sulphate-reducing Bacteria Including Rumen
 Strains. *Microbiology*, 63(1), 137–139. https://doi.org/10.1099/00221287-63-1-137
- 855 Rago, L., Zecchin, S., Villa, F., Goglio, A., Corsini, A., Cavalca, L., & Schievano, A. (2019).
- 856 Bioelectrochemical Nitrogen fixation (e-BNF): Electro-stimulation of enriched biofilm communities
- drives autotrophic nitrogen and carbon fixation. *Bioelectrochemistry*, *125*, 105–115.
- 858 <u>https://doi.org/10.1016/j.bioelechem.2018.10.002</u>
- 859 Rojas-Tapias, D., Moreno-Galván, A., Pardo-Díaz, S., Obando, M., Rivera, D., & Bonilla, R. (2012).
- 860 Effect of inoculation with plant growth-promoting bacteria (PGPB) on amelioration of saline stress in 861 maize (Zea mays). *Applied Soil Ecology*, *61*, 264–272. https://doi.org/10.1016/j.apsoil.2012.01.006
- 862 Rozendal, R. A., Leone, E., Keller, J., & Rabaey, K. (2009). Efficient hydrogen peroxide generation
- from organic matter in a bioelectrochemical system. *Electrochemistry Communications*, 11(9), 1752–
- 864 1755. https://doi.org/10.1016/j.elecom.2009.07.008
- 865 Ryan, R. P., Monchy, S., Cardinale, M., Taghavi, S., Crossman, L., Avison, M. B., Berg, G., van der Lelie,
- 866 D., & Dow, J. M. (2009). The versatility and adaptation of bacteria from the genus
- 867 Stenotrophomonas. *Nature Reviews Microbiology*, 7(7), Article 7.
- 868 <u>https://doi.org/10.1038/nrmicro2163</u>
- 869 Sim, J., An, J., Elbeshbishy, E., Ryu, H., & Lee, H.-S. (2015). Characterization and optimization of
- 870 cathodic conditions for H2O2 synthesis in microbial electrochemical cells. *Bioresource Technology*,
 871 *195*, 31–36. <u>https://doi.org/10.1016/j.biortech.2015.06.076</u>
- Singh, R. K., Singh, P., Li, H.-B., Guo, D.-J., Song, Q.-Q., Yang, T., Malviya, M. K., Song, X.-P., & Li, Y.-R.
- 873 (2020). Plant-PGPR interaction study of plant growth-promoting diazotrophs Kosakonia radicincitans
- 874 BA1 and Stenotrophomonas maltophilia COA2 to enhance growth and stress-related gene

- 875 expression in Saccharum spp. *Journal of Plant Interactions*, 15,427-445.
 876 <u>https://doi.org/10.1080/17429145.2020.1857857</u>
- 877 Soundararajan, M., Ledbetter, R., Kusuma, P., Zhen, S., Ludden, P., Bugbee, B., Ensign, S. A., &
- 878 Seefeldt, L. C. (2019). Phototrophic N2 and CO2 Fixation Using a Rhodopseudomonas palustris-H2
- 879 Mediated Electrochemical System With Infrared Photons. *Frontiers in Microbiology*, *10*, 1817.
- 880 <u>https://doi.org/10.3389/fmicb.2019.01817</u>
- Stoddard, S. F., Smith, B. J., Hein, R., Roller, B. R. K., & Schmidt, T. M. (2015). rrnDB: Improved tools
- 882 for interpreting rRNA gene abundance in bacteria and archaea and a new foundation for future
- 883 development. *Nucleic Acids Research*, *43*(D1), D593–D598. <u>https://doi.org/10.1093/nar/gku1201</u>
- Temple, S. J., Vance, C. P., & Stephen Gantt, J. (1998). Glutamate synthase and nitrogen assimilation.
 Trends in Plant Science, *3*(2), 51–56. <u>https://doi.org/10.1016/S1360-1385(97)01159-X</u>
- 886 *Tidyverse*. (n.d.). Retrieved 9 February 2023, from <u>https://www.tidyverse.org/</u>
- Tilak, K. V. B. R., Schneider, K., & Schlegel, H. G. (1986). Autotrophic growth of nitrogen-
- fixingAzospirillum species and partial characterization of hydrogenase from strain CC. *Current Microbiology*, *13*(6), 291–297. https://doi.org/10.1007/BF01577194
- Wiegel, J. (2005). *Xanthobacter Wiegel, Wilke, Baumgarten, Opitz and Schlegel 1978, 573 AL*.
 https://doi.org/10.1007/0-387-29298-5 135
- 892 Wresta, A., Widyarani, R., Boopathy, R., & Setiadi, T. (2021). Thermodynamic approach to estimating
- reactions and stoichiometric coefficients of anaerobic glucose and hydrogen utilization. *Engineering Reports*, *3*(6), e12347. <u>https://doi.org/10.1002/eng2.12347</u>
- 895 Wu, Y., Zaiden, N., & Cao, B. (2018). The Core- and Pan-Genomic Analyses of the Genus Comamonas:
- From Environmental Adaptation to Potential Virulence. *Frontiers in Microbiology*, *9*, 3096.
 <u>https://doi.org/10.3389/fmicb.2018.03096</u>
- y. Hafeez, F., yasmin, S., Ariani, D., Renseigné, N., Zafar, yusuf, & A. Malik, K. (2006). Plant growthpromoting bacteria as biofertilizer. *Agronomy for Sustainable Development*, *26*(2), 143–150.
 https://doi.org/10.1051/agro:2006007
- 901 Yadav, R., Chiranjeevi, P., Yadav, S., Singh, R., & Patil, S. A. (2022). Electricity-driven bioproduction
- from CO2 and N2 feedstocks using enriched mixed microbial culture. *Journal of CO2 Utilization*, 60,
 101997. <u>https://doi.org/10.1016/j.jcou.2022.101997</u>
- 204 Zaybak, Z., Pisciotta, J. M., Tokash, J. C., & Logan, B. E. (2013). Enhanced start-up of anaerobic
- 905 facultatively autotrophic biocathodes in bioelectrochemical systems. *Journal of Biotechnology*,
- 906 *168*(4), 478–485. <u>https://doi.org/10.1016/j.jbiotec.2013.10.001</u>
- 907 Zhang, L., Tian, C., Wang, H., Gu, W., Zheng, D., Cui, M., Wang, X., He, X., Zhan, G., & Li, D. (2022).
- 908 Improving electroautotrophic ammonium production from nitrogen gas by simultaneous carbon
- 909 dioxide fixation in a dual-chamber microbial electrolysis cell. *Bioelectrochemistry (Amsterdam,*
- 910 Netherlands), 144, 108044. https://doi.org/10.1016/j.bioelechem.2021.108044
- 911