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## Trade-offs between blue water use and greenhouse gas emissions related to food systems: An optimization study for French adults

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**Abstract**

Food systems face challenges from both their water and carbon footprints. Data suggest that it is possible to improve these both footprints simultaneously, but their potential conflicts and trade-offs have not been systematically explored. To this end, we here used a compromise programming approach to identify the dietary changes required to improve one and/or the other of these footprints, while ensuring nutritional adequacy and adherence to dietary guidelines, using French data on food consumption (1,456 adults aged 18-64 years from the INCA 3 study) and food environmental impact (Agribalyse® database). A full range of scenarios was identified by prioritizing the two objectives differently, giving weight from 0% to 100%, by 5-% steps, to the improvement in greenhouse gas emissions (GHGe) over the improvement in blue water use (BWU).

Overall, we have shown that it is possible to significantly reduce both BWU and GHGe compared to observed levels. The BWU reduction ranged from 14% to 36% with increasing prioritization, while the simultaneous GHGe reduction varied less, from 52% to 44% with decreasing prioritization. The consumption of some foods varied according to the priority given to BWU over GHGe reduction (namely, vegetables, fruit juice, dairy products, eggs, refined cereal, substitutes, offal and potatoes). In contrast, meat consumption (beef, pork, poultry and processed meat) was systematically removed, while the consumptions of offal and dairy products remained moderate in order to meet nutrient reference values. Fish, whole grains, and fruit also remained relatively constant across scenarios due to the constraints based on dietary guidelines. Whatever the scenario, the modeled diets were more plant-based than the observed diet from which they differed significantly (only 23-31% of common food consumptions), and were therefore healthier (63-76% reduction in the distance to theoretical minimum risk of chronic diseases).

To conclude, while focusing solely on BWU reduction induces a joint GHGe reduction that is near-maximal, the reverse is not true, showing that there is good alignment but also some divergence between these objectives.

**Keywords:** GHGe, Water use, compromise modelling, diet optimization, nutrient adequacy, healthy diet

**Symbols**

$i$  : denotes the 45 food groups

$\lambda$ : increment from 0% to 100%

### **Abbreviations**

BWU: blue water use (m<sup>3</sup> world eq/d)

DALYs: Disability-Adjusted Life Years

DD: diet departure

DSI: diet similarity index

GBD: Global Burden of diseases HRS: health risk score

GHGe: greenhouse gas emissions (CO<sub>2</sub> eq/d)

HRS: Health Risk Score

INSEE: Institut Nationale des Statistiques et des Etudes Economiques

Opt: optimized value

Obs: observed

TMREL: theoretical minimum-risk exposure levels

## 1 Introduction

1 Urbanization and modernization have significantly impacted dietary patterns, especially in developed  
2 countries. These patterns have shifted towards animal-based diets high in salt, fat, and sugar, leading  
3 to increased prevalence of various chronic diseases such as cardiovascular disease, diabetes, and  
4 cancer, particularly colorectal cancer (GBD 2017 Diet Collaborators, 2019; GBD 2019 Risk Factors  
5 Collaborators, 2020), with ~190 million disability-adjusted life years (DALYs) attributed to dietary  
6 risk factors worldwide in 2019 (GBD 2019 Risk Factors Collaborators, 2020). Besides harming health,  
7 the current food system damages the environment by depleting natural resources, such as water and  
8 forests, and contribute climate instability through increased GHGe (Clark et al., 2019; HLPE, 2017).  
9 Planetary boundaries are exceeded for six critical indicators (Campbell et al., 2017; Rockström et al.,  
10 2009; Steffen et al., 2015), including climate change and one component of total water use, namely  
11 green water use (mainly from rainwater and water stored in the soil), the other component being BWU  
12 (groundwater or surface water). Recently, some authors have highlighted the need for integrated  
13 environmental assessments and policy decisions to better understand the trade-offs between different  
14 environmental footprints, which should be studied simultaneously (Vanham et al., 2019).  
15 Today, GHGe from animal-based food production are twice higher than those from plant-based food  
16 (Xu et al., 2021). Specifically, red and processed meat production are the most impactful globally,  
17 regardless of climate change or water use (Clark et al., 2019; Poore and Nemecek, 2018). However,  
18 other productions have a substantial role in GHGe, such as dairy products and some fish, depending on  
19 the production method (captured or farmed). In addition, some of the lower emitting foods (e.g. fresh  
20 fruit, vegetables, and some refined and whole grains) have quite strong water demands (Clark et al.,  
21 2019; Poore and Nemecek, 2018). With regard to diet, there is a large body of scientific literature  
22 documenting its link to environmental pressures, and studies have reported that diets rich in plant-  
23 based foods have lower pressures than animal-based diets, particularly for GHGe (Aleksandrowicz et  
24 al., 2016; Auestad and Fulgoni, 2015; Carey et al., 2023; Perignon et al., 2016; van Dooren, 2018;  
25 Wilson et al., 2019). While studies on diet-related environmental pressures have largely focused on  
26 climate change and the associated GHGe criterion (Jones et al., 2016; Ridoutt et al., 2017), it is not the  
27 only environmental indicator threatened by the food systems, which also significantly impact natural

28 resource depletion and biodiversity loss (Campbell et al., 2017; Eyhorn et al., 2019; Poore and  
29 Nemecek, 2018).

30 Regarding studies on the water footprint of food systems, the available results are not always  
31 consistent and are subject to debate depending on the type of indicator (water footprint, BWU and/or  
32 green water use). In reviews, healthier diets have been associated with higher (Steenson and Buttriss,  
33 2021), similar (Harris et al., 2020) or lower (Aleksandrowicz et al., 2016) BWU. This may be because  
34 some plant-based foods, such as fruits, oils, and nuts, which are essential components of a healthy diet,  
35 are also important contributors to BWU (Clark et al., 2022; Harris et al., 2020; Willett et al., 2019).

36 Most of the diet modeling studies to identify the dietary changes needed to reduce environmental  
37 impact have focused on GHGe, even if some have also considered BWU (Wilson et al., 2019). Only  
38 one diet modeling study has alternately optimized different footprints (Gephart et al., 2016): in this  
39 American study, all optimized diets were rich in plant-based products and low in animal-based  
40 products, regardless of the considered environmental criteria (such as BWU and green water use or  
41 GHGe), and the authors concluded to similarities between diets limiting carbon and water footprints.  
42 However, to date, no study has comprehensively analyzed the similarities or differences between diets  
43 that minimize carbon or water footprints, and little is still known about the alignments and/or conflicts  
44 (with potential trade-offs) between these two objectives.

45 To address this question adequately and systematically, we have used a compromise programming  
46 approach to explore all possible trade-off scenarios between the two extremes of optimizing only one  
47 of the carbon and water footprints. Compromise modelling is a key tool to explore such multi-criteria  
48 optimization problem, which has not yet been applied to this research question. This was done here by  
49 sequentially balancing GHGe and BWU minimization under a set of constraints to ensure nutrient  
50 security, not worsen long-term health risk, and consider cultural acceptability in the French context,  
51 using the average observed diet from the most recent representative dietary survey combined with the  
52 diet-related environmental footprints data from the Agribalyse ® database.

## 2 Methods

## 53 **2.1 Population**

54 This study was conducted using data from the INCA 3 study, a nationally representative French survey  
55 conducted in 2014-2015 by the French Agency for Food, Environmental and Occupational Health  
56 Safety (ANSES). This study initially included 2,121 adult participants who provided food  
57 consumption data using a validated method (Dubuisson et al., 2019). Details of the study design,  
58 recruitment and survey plan (definition of individual weight), and methods used have been described  
59 in detail elsewhere (Dubuisson et al., 2019).

60 Overall, participants were selected according to a three-stage random sampling design (geographic  
61 units, dwellings, and then individuals) drawn by the National Institute of Statistics and Economic  
62 Studies (INSEE). One individual per dwelling was then drawn at random from the eligible individuals  
63 at the time of the household contact. The weight of individuals was calculated according to the INSEE  
64 method to improve representativeness by region, size of the urban area, occupation and socio-  
65 professional category of the household's reference person, household size, education level, gender, and  
66 age (Sautory, 1993).

67 The INCA 3 study protocol was authorized by the National Commission on Informatics and Liberty,  
68 after a favorable opinion from the Advisory Committee on Information Processing in Health Research.  
69 The study also received a favorable opinion from the Conseil National de l'Information Statistique on  
70 15 June 2011 (n°121/D030) and was awarded the label of “general interest” and statistical quality by  
71 the INSEE Label Committee (n°47/Label/D120).

72 The data collected in the INCA 3 survey encompass food and drink consumption and socio-  
73 demographic and lifestyle characteristics. In the present study, we selected adults <65y old (N=1,665)  
74 who were not under-reporter (N=1,456) for energy intake (the procedure for identification of under-  
75 reporters is described in **Method S1**).

## 76 **2.2 Dietary data**

77 Food and beverage consumption data were collected over 3 non-consecutive days (2 weekdays and 1  
78 weekend day) distributed over approximately three weeks, using the 24-hour recall method by phone  
79 conducted by trained interviewers using a standardized validated software for data entry  
80 (GloboDiet)(Aglago et al., 2017). Estimation of portion sizes consumed was performed using a picture

81 booklet of food portions and household measurements, previously sent by post. Mixed dishes were  
 82 decomposed in ingredients using the standardized recipes validated by dieticians.  
 83 Nutrient intakes were calculated using the 2016 food composition database published by the French  
 84 Information Centre on Food Quality (Agence nationale de sécurité sanitaire de l'alimentation de  
 85 l'environnement et du travail (ANSES), 2012). In the modeling procedure, food items consumptions  
 86 were collapsed into 45 broader food groups (the list is provided in **Table S1**). Nutrient composition  
 87 and environmental pressure of each food group were calculated as mean over all items of the group  
 88 weighted by the contribution of the food item consumption to the food group consumption.

### 89 **2.3 Health risk and diet similarity scores**

90 Health risk associated with each observed and modeled diet was assessed using the Health Risk Score  
 91 (HRS) (Fouillet et al., 2023), representing the global normalized distance to the theoretical minimum-  
 92 risk exposure levels (TMREL) for three unhealthy (red meat, processed meat and sweetened  
 93 beverages) and six healthy (whole grains, fruit, vegetables, legumes, nuts and seeds, and milk) food  
 94 groups established by the 2019 Global Burden of diseases (GBD) study (GBD 2019 Risk Factors  
 95 Collaborators, 2020). The HRS measures the distance to each consumption target (TMREL) weighted  
 96 by its relative importance using DALYs attributable to each food group in the French population. By  
 97 construct, HRS varies between 0% and 100%, depending on whether the diet meets all the food group  
 98 targets (i.e., minimum risk) or deviate from them at most (i.e., maximum risk), respectively. The HRS  
 99 calculation is presented and detailed in **Method S2**.

100 For each modeled diet, we also computed DSI (diet similarity index) reflecting the proportion of food  
 101 group consumptions that remained similar to those of the observed diet (Mertens et al., 2020) using the  
 102 following formula:

$$103 \quad \text{DSI} = 100 \times \frac{\sum_{i=1}^{44} \min(\text{Opt}_i, \text{Obs}_i)}{\sum_{i=1}^{44} \text{Obs}_i}$$

104 where  $i$  denotes the 45 food groups except water used in the optimization model. Opt referred to the  
 105 optimized value and Obs to the observed value.

## 106 **2.4 Environmental indicators**

107 Environmental indicators for pressure along with the food chain were estimated using data from the  
108 French database Agribalyse ® 3.1 developed by the French Agency for the Environment and Energy  
109 Management (ADEME). Agribalyse ® 3.1 contains environmental indicators for 2,517 foods  
110 consumed in France. The list was based on the consumption declared in the INCA 3 survey using  
111 common coding (Colomb et al., 2015). The methodology has been extensively explained in *ad hoc*  
112 published reports (ADEME, 2020; Koch and Salou, 2020) and is summarized in **Method S3**. In the  
113 Agribalyse ® database, water footprint has been estimated using the guidelines of The Water Footprint  
114 Network (Hoekstra, 2011) and refers to blue water. The other available indicators are defined in  
115 Method S3. GHGe and BWU for each food group are shown in **Table S1**.

## 116 **2.5 Multicriteria optimization by compromise programming for analyzing GHGe and BWU** 117 **trade-offs**

118 Diet optimization was performed using the procedure SAS/OR ® *optmodel* (version 9.4; SAS  
119 Institute, Inc.) using a non-linear optimization algorithm with multi-start option to minimize the risk of  
120 obtaining only a local minimum. The methodologic approach and used data are summarized in **Figure**  
121 **1**.

122 Starting from the observed food consumptions, we modeled fully nutrient-adequate diets by including  
123 the following constraints in diet optimization:

- 124 - Nutritional constraints on daily energy intake and a set of nutrient intakes were based on the  
125 recently revised ANSES Reference Values (French Agency for Food, Environmental and  
126 Occupational Health Safety (Anses), 2016) according to the 2021 EFSA opinion (“Dietary  
127 Reference Values | DRV Finder,” n.d.). For bioavailable iron and zinc, lower bounds were not  
128 based on current reference values but on lower threshold values ensuring  $\leq 5\%$  deficiency  
129 prevalence, as in our previous study (Fouillet et al., 2023), because we have shown that such  
130 flexibility enables the identification of healthier diets with a better balance in DALYs due to  
131 less cardiometabolic disease, despite a higher prevalence of iron-deficiency anemia (Dussiot et  
132 al., 2021). Nutritional constraints are presented in **Table A.2**. For zinc and iron, bioavailability



133 was considered using reference equations (Armah et al., 2013; Miller et al., 2013). Details of  
 134 computation and formula are presented in **Method S4**.

- 135 - Acceptability constraints were defined by upper bounds set at the weighted 99<sup>th</sup> percentiles  
 136 values of each food group based on the distribution in the INCA3 study (**Table S1**). In the  
 137 lack of specific data on acceptability, these constraints rather represent the overall feasibility  
 138 given current consumption levels, and only an upper threshold was used to limit  
 139 aberrant/unattainable consumption levels while leaving room for change.
- 140 - Epidemiological constraints have been also defined to avoid increasing the health risk (as  
 141 considered by the GBD) beyond its observed level, as follows:
  - 142 ○ at least equal of the observed average (among consumers) consumption for healthy  
 143 food groups (fruits, vegetables, legumes, nuts, whole grains and milk)
  - 144 ○ less than or equal to the observed average intake (among consumers) for the unhealthy  
 145 food groups (red meat, processed meat and sweetened beverages), outweighing the  
 146 corresponding acceptability constraints.

147 Diet optimization was conducted on the mean dietary data for each sex and for both sexes, by  
 148 considering an average individual constituted of 50% male and 25% non-menopausal and 25%  
 149 menopausal female. In the average individual, nutritional references were defined as the weighted  
 150 values of sex specific nutritional references (**Table S1**), and the 99<sup>th</sup> and 95<sup>th</sup> percentiles (see below) of  
 151 food group consumptions were calculated using the same weighting scheme.

152

153 In a preliminary step (**Figure 1**), we applied the nutritional, epidemiological and acceptability  
 154 constraints to identify the modifications needed to comply with the nutritional and epidemiologic  
 155 references only. For this model, the objective function was the minimization of the diet departure  
 156 (DD) from to the initial (observed) situation using a formula accounting for dietary inertia (Kramer et  
 157 al., 2018) as:

$$158 \quad DD = \sum_{i=1}^{45} \left[ \frac{Opt_i - Obs_i}{SD_i} \right]^2 \quad [1]$$

159 Where  $Opt_i$  and  $Obs_i$  denoted the optimized and observed daily consumption of food group (i) and  
 160  $SD_i$  was the standard deviation of the observed daily consumption of food group (i).

161 Then, in a first step (**Figure 1**) we performed a multi-criteria optimization of GHGe and BWU by  
 162 compromise programming.

163 First, we determined the best (minimal) and worst (maximal) values achievable for GHGe and BWU,  
 164 respectively, while satisfying all the model constraints defined above, by mono-criteria optimization as  
 165 following:

$$166 \quad \text{Min GHGe} = \sum_{i=1}^{45} [Opt_i \times GHGe_i], \text{ giving } GHGe_{\text{best}} \text{ and } BWU_{\text{worst}} \quad [2]$$

167 and

$$168 \quad \text{Min BWU} = \sum_{i=1}^{45} [Opt_i \times BWU_i], \text{ giving } BWU_{\text{best}} \text{ and } GHGe_{\text{worst}} \quad [3]$$

169 where  $i$  is the food group,  $Opt_i$  denotes the daily consumption of the food group  $i$  (g/d) in the optimized  
 170 model,  $GHGe_i$  is the greenhouse gas emission for 1 g of the food group  $i$  and  $BWU_i$  is the blue water  
 171 use for 1 g of the food group  $i$ ,  $GHGe_{\text{best}}$  and  $BWU_{\text{best}}$  are the best values and  $GHGe_{\text{worst}}$  and  $BWU_{\text{worst}}$   
 172 the worst values of the corresponding criteria (i.e., the parameters of the pay-off matrix in compromise  
 173 programming) extracted from equations [2] and [3] (Rohmer et al., 2019; Van Mierlo et al., 2017).

174 For purpose of fairness between the GHGe and BWU criteria with different units, the multi-criteria  
 175 optimization was then conducted on the normalized distances to their ideal best values, i.e., on the  
 176 degree of closeness to ideal points  $d_{GHGe}$  and  $d_{BWU}$  defined by:

$$177 \quad d_{GHGe} = \frac{GHGe - GHGe_{\text{best}}}{GHGe_{\text{worst}} - GHGe_{\text{best}}} \quad [4]$$

$$178 \quad d_{BWU} = \frac{BWU - BWU_{\text{best}}}{BWU_{\text{worst}} - BWU_{\text{best}}} \quad [5]$$

179 The compromise programming weighted by  $\lambda$  and  $(100\% - \lambda)$  the  $d_{GHGe}$  and  $d_{BWU}$  terms, respectively,  
 180 using a multi-objective function defined as (Garcia-Launay et al., 2018; Oliveira and Saramago,  
 181 2010):

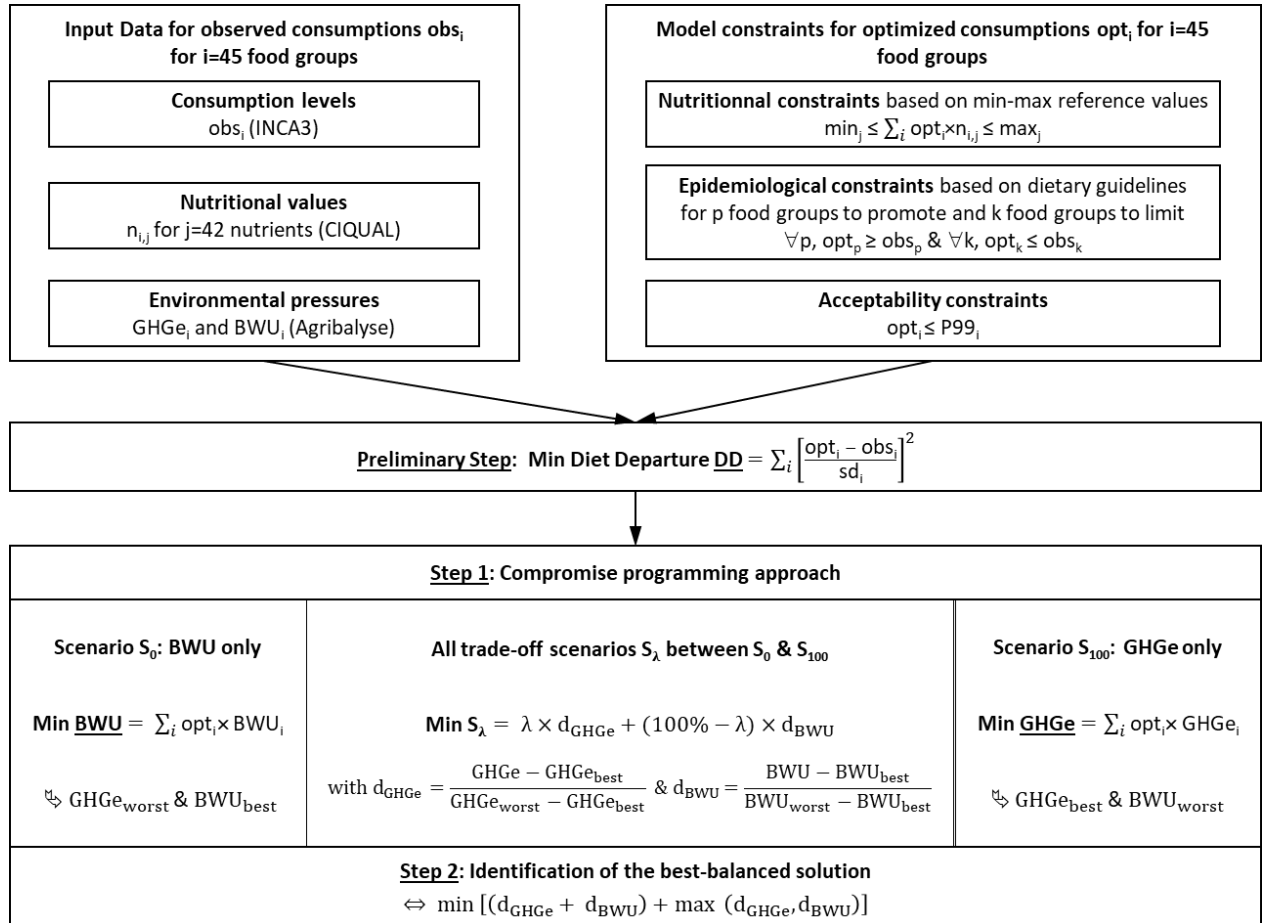
$$182 \quad \text{Min multi-objective function} = \lambda \times d_{GHGe} + (100\% - \lambda) \times d_{BWU} \quad [6]$$

183 with  $\lambda$  ranging from 0% to 100% by increment of 5%, to explore all the compromise solutions  
 184 between minimizing only BWU ( $\lambda = 0\%$ ) or GHGe ( $\lambda = 100\%$ ).

185 Finally, in a last second step, we identified the *best-balanced* solution between the *minsum* (i.e., sum  
 186 of  $d_{GHGe}$  and  $d_{BWU}$ , efficiency metric) and *minmax* (i.e., maximum of  $d_{GHGe}$  and  $d_{BWU}$ , equity  
 187 metric) objectives (Oliveira and Saramago, 2010; Rohmer et al., 2019; Van Mierlo et al., 2017), as:

$$188 \quad \text{best - balanced} \Leftrightarrow \min [(d_{GHGe} + d_{BWU}) + \max (d_{GHGe}, d_{BWU})] \quad [7]$$

189 **Figure 1: Schematic diagram of the optimization phases, models, and parameters<sup>1</sup>**



190

191 Abbreviations: BWU, blue water use; GHGe, greenhouse gas emissions; Obs: observed consumption; Opt;  
 192 optimized consumption  
 193  $Nut_j$  denotes intake of nutrient  $j$  and nutritional constraints are based on revised ANSES Reference Values  
 194 (2016),  $c_i$  denotes consumption of a food item  $i$ . The  $k$  and  $p$  food categories referred to those defined by the  
 195 GBD (red meat, processed meat, sweetened beverages, fruits, vegetables, legumes, nuts, whole grains and milk) ;  
 196 these epidemiologic constraints impose consumption lower/higher than the observed mean for unhealthy/healthy  
 197 food categories.  $S_x$  denotes the scenario  $\lambda=x\%$  in the compromise programming approach, where  $\lambda$  is the  
 198 relative weight given to Greenhouse gas emissions over blue water use in the multi-criteria optimization (i.e.  $S_0$   
 199 and  $S_{100}$  correspond to the minimization of water use only and GHGe only, respectively). The preliminary step  
 200 allows to identify the modifications needed to comply with the nutritional and epidemiologic references only.  
 201

## 202 **2.6 Sensitivity analysis**

203 First, we analyzed the influence of the acceptability constraints: we used stricter acceptability  
 204 constraints by lowering the food group consumption limits to their observed 95<sup>th</sup> percentile, rather than  
 205 their 99<sup>th</sup> percentile as in our main analysis.

206 Second, we analyzed the influence of some nutrient constraints: we used stricter requirements for  
 207 bioavailable iron and zinc by raising their lower bounds to current reference values (1.92 g/d and 3.62  
 208 g/d, respectively), rather than lower threshold values (1.11 g/d and 1.83 g/d, respectively) ensuring  
 209  $\leq 5\%$  deficiency prevalence as in our main analysis.

210 For each modeled diet, we conducted a dual value analysis to identify, among the different constraints  
 211 we used, those that were the most active (compared to the inactive ones that had no effect on the  
 212 results) and that most limited the objective gain (i.e., GHGe and/or WU reduction). In particular, this  
 213 allowed the identification of nutrients requirements that proved to be limiting for GHGe and WU  
 214 reductions. This analysis was conducted as in our previous work (Dussiot et al., 2021), by calculating  
 215 the standardized dual values corresponding to the potential gain in objective (i.e., GHGe and/or WU)  
 216 in the case of a 100% relaxation of the limiting bound of the constraint.

## 217 **2.7 Statistical analysis**

218 The sociodemographic and lifestyle characteristics of the men and women were presented as mean  
 219 (SD) or percentage. The observed and modeled diets were described in terms of food group  
 220 consumptions, environmental pressures, HRS and DSI.

221 All statistical analyses were performed using SAS® (version 9.4; SAS Institute, Inc., Cary, NC, USA)  
 222 and figures were drawn using R version 3.6.

## 223 **3 Results**

### 224 **3.1 Description of the sample**

225 The characteristics of the total study population are presented in **Table 1**. The studied population  
 226 included 1,456 participants (57% women), with a mean age of 42.2 years (SD= 13.5). About 96% of  
 227 the sample completed three 24-hour recalls.

**Table 1: Characteristics of study participants, (INCA 3 study, n=1,456)<sup>1</sup>**

	Men	Women
--	-----	-------

N	621	835
<b>Age (y)</b>	42.05 (13.86)	42.42 (12.45)
<b>Education</b>		
Primary + College	42.82	40.03
High school	17.62	21.59
Undergraduate level	19.39	20.13
Postgraduate level	20.03	16.54
No information	0.14	1.72
<b>Physical activity<sup>2</sup> (%)</b>		
Low	24.4	45.56
Moderate	53.69	46.36
High	21.91	8.08
<b>Living area (%)</b>		
Rural	27.48	23.9
2,000-19,999 inhabitants	13.91	20.04
20,000-99,999 inhabitants	10.39	10.27
≥100,000 inhabitants	35.88	30.97
Paris area	12.35	14.81
<b>Health risk score<sup>3</sup> (%)</b>	92 (29)	83 (26)
<b>Body mass index (kg.m<sup>-2</sup>)</b>	25.08 (3.82)	24.95 (4.94)
<b>Number of 24h recall</b>	2.96 (0.19)	2.95 (0.21)
<b>Energy intake (kcal/d)</b>	2682.64 (747.70)	1963.31 (521.05)
<b>Greenhouse gas emissions (kg CO<sub>2</sub> eq /d)</b>	6.23 (2.68)	4.42 (1.50)
<b>Blue water use (m<sup>3</sup> water eq deprivation /d)</b>	7.07 (3.76)	5.82 (3.26)

228 <sup>1</sup>Values are n, means (SD) or % as appropriate, all data are weighted on the survey design

229 <sup>2</sup>Estimated using the RPAQ questionnaire

230 <sup>3</sup> HRS is the normalized distance to the theoretical minimum-risk exposure levels (TMREL) from the Global  
 231 Burden of Diseases (i.e. HRS= 0% when the diet is at minimal risk by meeting all the TMREL and HRS=100%  
 232 when the diet is at maximal risk by deviating from them at most).

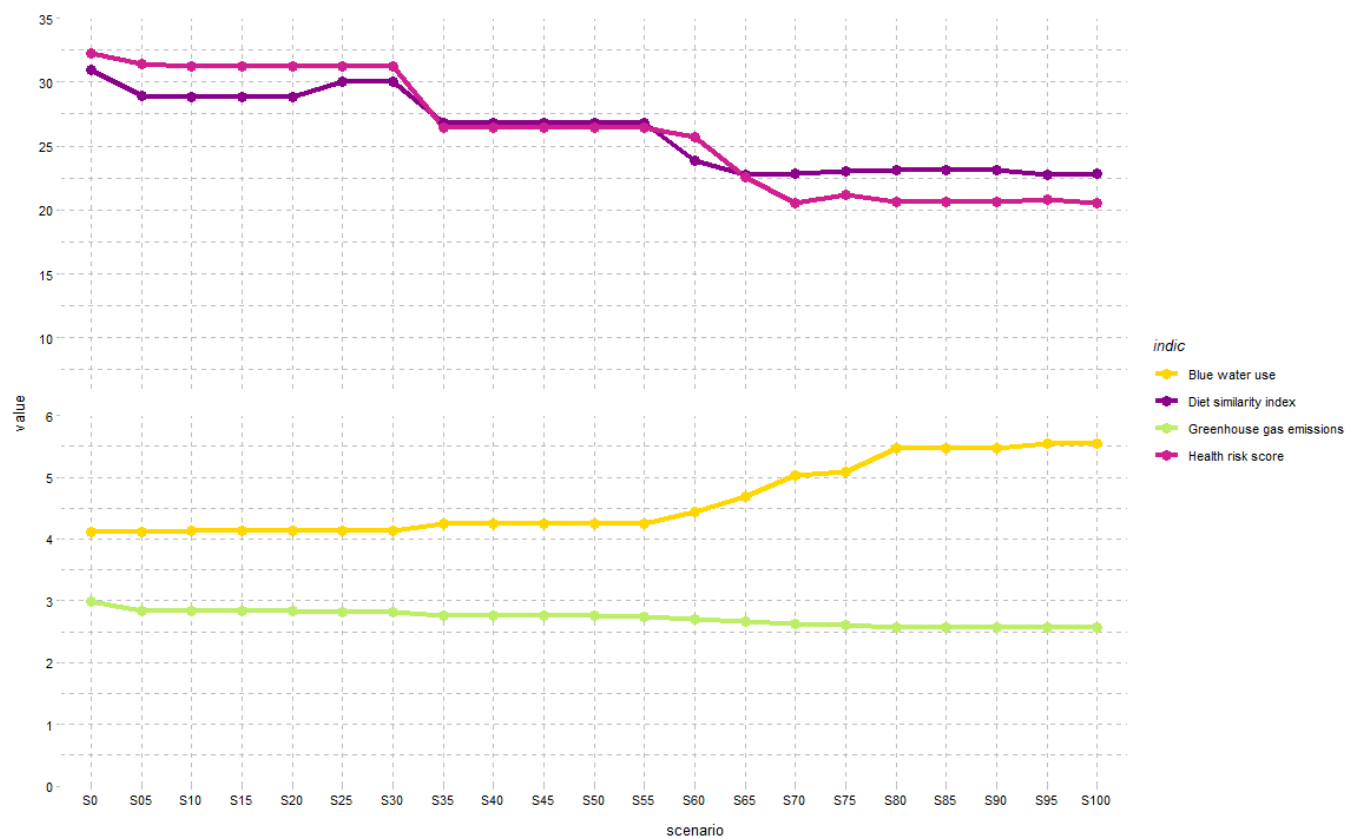
233

### 234 **3.2 Environmental and other key indicators of modeled diets across trade-off scenarios**

235 Compared to the mean observed diet, the modeled diet issued from the first step, i.e., the one closest to  
 236 the observed diet meeting the nutritional and epidemiological constraints, had a Diet Similarity Index  
 237 (DSI, percentage of common food group consumption) of 68%, and 13% higher GHGe (6.05 vs 5.34  
 238 kg CO<sub>2</sub> eq/d), 33% higher BWU (8.56 vs 6.46 m<sup>3</sup> water eq deprivation/d), but 26% lower Health Risk  
 239 Score (HRS, 64 vs 87 %) (**Table S.3**).

240 **Figure 2: Objective functions and diet descriptors in scenarios differentially prioritizing blue water use**  
 241 **over greenhouse gas emissions minimization**

242



243

244 Greenhouse gas emissions in kg CO<sub>2</sub> eq/d and blue water use in m<sup>3</sup> world eq /d).

245 Abbreviations: DSI, diet similarity index (%); HRS, health risk score (%);

246 S<sub>x</sub> denotes the scenario  $\lambda=x\%$  in the compromise programming approach, where  $\lambda$  is the relative weight given to247 Greenhouse gas emissions over blue water use in the multi-criteria optimization (i.e. S<sub>0</sub> and S<sub>100</sub> correspond to

248 the minimization of water use only and GHGe only, respectively). HRS (%) is the normalized distance to the

249 theoretical minimum-risk exposure levels (TMREL) from the Global Burden of Diseases, expressed in % (i.e.,

250 HRS = 0% when the diet is at minimal risk by meeting all the TMREL and HRS=100% when the diet is at

251 maximal risk by deviating from them at most). DSI (%) is the proportion of food group consumptions remaining

252 similar to those observed.

253 **Figure 2** depict the environmental and other key indicators (HRS and DSI) for each modeled diet254 issued from the second step, i.e., from compromise programming by tuning  $\lambda$  from 0% (minimizing

255 BWU only) to 100% (minimizing GHGe only) by steps of 5%, always under the nutritional,

256 epidemiological and acceptability constraints. As regards environmental indicators, whatever the

257 trade-off scenario ( $\lambda$  value), GHGe and BWU were lower in the modeled than observed diets. From  $\lambda$ 

258 = 0% (BWU minimization) to 100% (GHGe minimization), GHGe lowered gradually by -14%, from

259 2.98 to 2.57 kg CO<sub>2</sub> eq/d (i.e., from -44% to -52% of the observed situation, respectively), while BWU260 increased by 35%, from 4.12 to 5.54 m<sup>3</sup> world eq/d (i.e., from -36% to -14% of the observed situation,

261 respectively). Gradually prioritizing GHGe reduction (from  $\lambda=0\%$  to  $100\%$ ) resulted in diets with a  
 262 progressive but slight GHGe decrease, while BWU showed an initially slight and progressive increase  
 263 that became more marked from  $\lambda\approx 60\%$ .

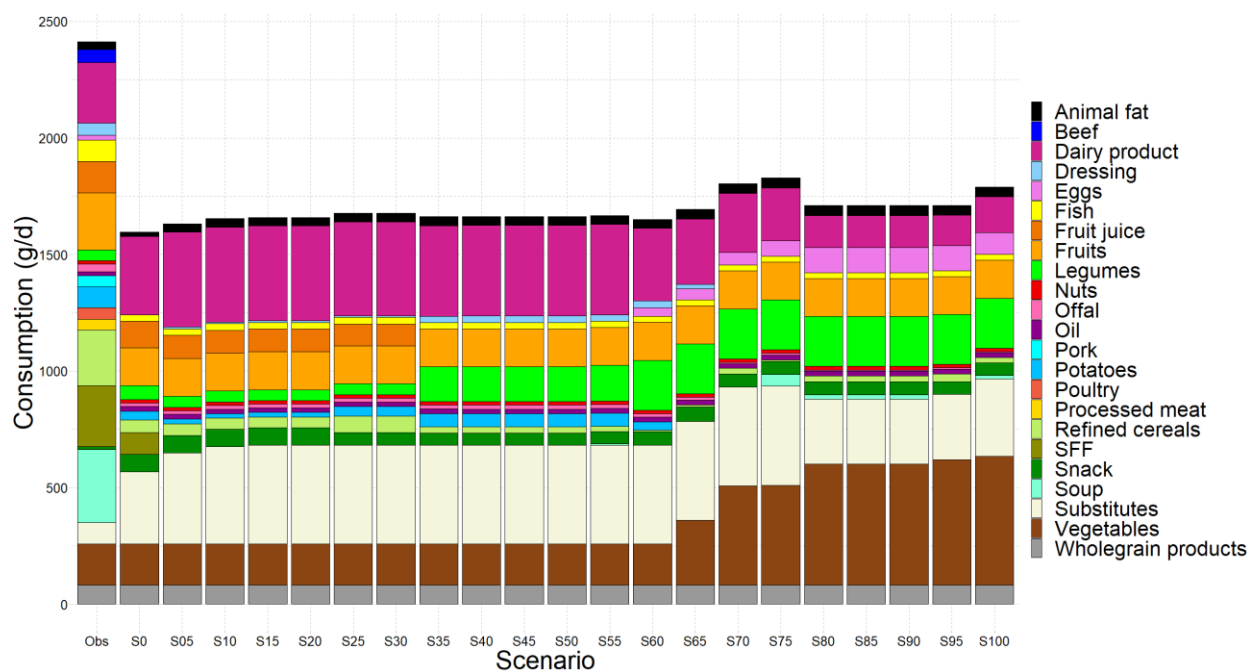
264 Other environmental indicators calculated using the Agribalyse<sup>®</sup> database are presented in the **Table**  
 265 **S.4**. Many indicators (notably land use and energy demand) as well as the single aggregated  
 266 environmental footprint score varied little across the  $\lambda$  range, with similar values at both extreme  $\lambda$   
 267 values, but slightly lower values in the middle range ( $\approx 15\%$  for land use,  $\approx 5\%$  for energy demand  
 268 and  $\approx 7\%$  single environmental footprint score for  $\lambda=35-60\%$  vs  $\lambda=0\%$ ).

269 As regards other key indicators (DSI and HRS, **Figure 2**), all modeled diets were actually greatly  
 270 distant from the observed diet but also greatly healthier. From  $\lambda = 0\%$  to  $100\%$ , the modeled diets had  
 271 no more than  $32\%$  to  $23\%$  common food group consumptions with the observed diet (DSI) and the  
 272 estimated health risk (HRS) was decreased to  $32\%$  to  $21\%$ , respectively (i.e.,  $-63\%$  to  $-76\%$  decrease  
 273 compared to the observed situation). DSI and HSR both reached their minimal values for the scenarios  
 274 with  $\lambda \geq 65\%$  (prioritization of GHGe reduction), but globally they varied only moderately over the  $\lambda$   
 275 range.

### 276 **3.3 Food group consumptions in modeled diets across trade-off scenarios**

277 Compared to the observed diet, some food groups were totally removed from modeled diets, namely  
 278 beef, processed meat, poultry, and pork, alcoholic beverages, sodas, and other beverages (hot drinks)  
 279 (**Figure 3** and **Table S.5**). Across the scenarios gradually prioritizing GHGe reduction (from  $\lambda=0\%$  to  
 280  $100\%$ ), there was an increase in eggs, animal fat, legumes, vegetables, and soup, and inversely a  
 281 decrease in dairy products, offal and snack. Fruit juices were present in the modeled diets up to  
 282  $\lambda=30\%$  then disappeared. Potatoes, refined cereals, substitutes and dressing showed bell-shaped curves  
 283 and the other food groups (fish, oil, fruits, nuts and whole grains) were present in nearly constant  
 284 quantities in most of the modeled diets.

285 **Figure 3: Food group consumptions (g/d) in modeled diets differentially prioritizing blue water use over**  
 286 **greenhouse gas emissions minimization**



287  
 288 Abbreviations: Obs, observed situation; SFF: sweet and fat foods. For clarity purpose, the 45 food groups are  
 289 pooled into 26 broader food categories and beverages are not shown. Sx denotes the scenario  $\lambda=x\%$  in the  
 290 compromise programming approach, where  $\lambda$  is the relative weight given to greenhouse gas emissions over blue  
 291 water use in the multi-criteria optimization (i.e., S0 and S100 correspond to the minimization of blue water use  
 292 only and greenhouse gas emissions only, respectively).  
 293

294 The contributions of food groups to GHGe and BWU for each modeled diet are presented in **Figure**  
 295 **S.1 (Tables S.6 and S.7)**. Overall, as meat (beef, pork and poultry) was removed from all modeled  
 296 diets over the  $\lambda$  range, the main contributors to GHGe were dairy products, offal, eggs, and vegetables  
 297 and substitutes. The main contributors to BWU were vegetables and fruits.

298 The consumption changes explaining the HRS improvement in the modeled compared to the observed  
 299 diets were the lower consumptions of sugar-sweetened beverages, red and processed meat and higher  
 300 consumption of vegetables (**Figure S.2**). Across the scenarios gradually prioritizing GHGe reduction  
 301 (from  $\lambda=0\%$  to 100%), there was a slight progressive improvement in HRS with the vegetables and  
 302 legumes increases and red meat reduction.

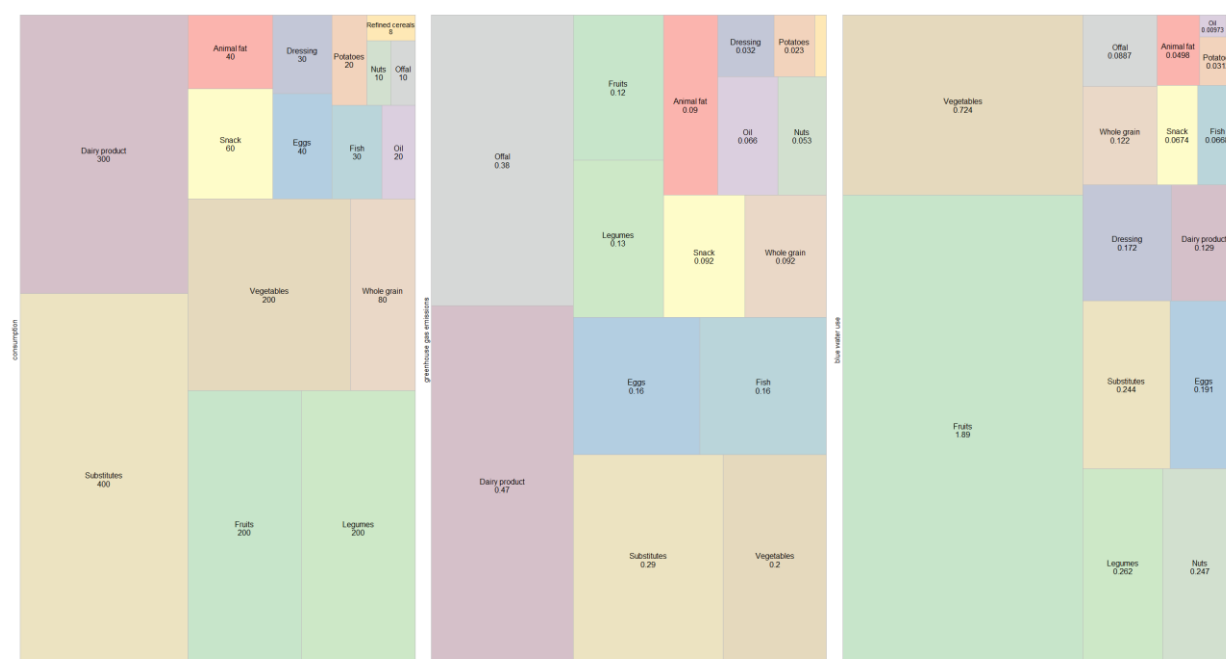
### 303 **3.4 Best trade-off and best-balanced diet**

304 The main characteristics and environmental impacts of this diet are presented in **Figure 4** and **Table**  
 305 **S3**. The best compromise between efficiency and equity in GHGe and BWU improvements (i.e., best-  
 306 balanced solution issued from the last third step) was identified for GHGe of 2.69 kg CO<sub>2</sub> eq/d and



307 BWU of 4.52 m<sup>3</sup> world eq/d (i.e., -50% and -30% of the observed values, respectively). This best-  
 308 balanced diet, corresponding to  $\lambda \approx 65\%$ , had a HRS value of 25% (-71% of the observed value) and a  
 309 DSI of 23%. This diet was composed mainly of dairy products, vegetables, legumes, and substitutes  
 310 and did not contain any animal flesh, except for a little offal.

311 **Figure 4: Food group consumptions, greenhouse gas emissions and blue water use for the best-balanced**  
 312 **modeled diet**



313  
 314 Abbreviation: SFF: sweet and fat foods. For clarity purpose, the 45 food groups are pooled into 28 broader food  
 315 categories.

316 Food group consumptions in g/d, and contributions of food groups to greenhouse gas emissions (GHGe) in kg  
 317 CO<sub>2</sub> eq/d and blue water use (BWU) in m<sup>3</sup> world eq/d for the best-balanced modeled diet regarding efficiency  
 318 and equity, identified by minimizing  $(d_{\text{GHGe}} + d_{\text{BWU}}) + \max(d_{\text{GHGe}}, d_{\text{bBWU}})$ , where  $d_{\text{GHGe}}$  and  $d_{\text{BWU}}$  are the  
 319 normalized distances to the best GHGe and BWU values, respectively.

320  
 321 Some discrepancies were noted between men and women as shown in **Figure S3**. Main differences in  
 322 terms of food group consumptions concerned higher consumption of legumes, offal, nuts, fruits and  
 323 substitutes in men and higher consumption of vegetables in women.

### 324 3.5 Sensitivity analysis

325 The most limiting nutrients in the modeled diets were quite systematically iodine, sodium, vitamin A,  
 326 EPA+DHA, and vitamin B2, vitamin C and the linolenic acid / alpha-linoleic acid ratio (data not  
 327 shown).

328 We tested the influence of being stricter regarding cultural acceptability, by limiting the food group  
329 consumptions to their observed 95<sup>th</sup> rather than 99<sup>th</sup> centile as in our main analysis. The best-balanced  
330 diet identified using these stricter acceptability constraints (**Figure S4**) had quite similar BWU and  
331 GHGe values (4.49 world eq/d and 2.73 kg CO<sub>2</sub> eq/d, respectively, Supplemental Table 1), but was  
332 characterized by lower consumption of legumes, dairy products and substitutes and higher  
333 consumption of refined cereals, potatoes and eggs.

334 We also tested the influence of being stricter regarding nutritional requirements for bioavailable iron  
335 and zinc, by using current reference values rather than lower threshold values ensuring  $\leq 5\%$  deficiency  
336 prevalence. The best-balanced diet identified using these stricter nutritional constraints was  
337 characterized by higher BWU and GHGe values (7.50 m<sup>3</sup> world eq /d and 4.39 kg CO<sub>2</sub> eq/d,  
338 respectively, **Table S3**), less dairy product and snack, more fruit juice, eggs, potatoes, sweet and fat  
339 foods, and, importantly, the reintroduction of processed meat and poultry that had previously been  
340 removed in our main analysis (data not shown).

341

#### 4 Discussion

342 In the present study we have shown that under a set of nutritional and epidemiological constraints  
343 limiting the extent of their potential improvement, it is still possible to reduce BWU up to 36% and  
344 GHGe up to 52%, and also to reduce them jointly by 30% and 50%, respectively. The modeled diets  
345 prioritizing one or the other of these impact reductions share some common main characteristics, such  
346 as the absence of some food groups being important contributors to both GHGe and BWU (beef,  
347 poultry, pork, fatty and sugary products), while they differ in their content in other food groups having  
348 more divergent effects on GHGe and BWU, such as fruit and vegetables which exhibit low GHGe but  
349 elevated BWU.

##### 350 *Nutrition/Health*

351 Regarding the nutritional constraints, based on our previous results (Dussiot et al., 2021), we here  
352 considered lowered threshold values than nutritional references for iron and zinc, allowing for a small  
353 increase in anemia to a 5% prevalence. This should not be considered a limitation, insofar as we have  
354 documented that the current nutritional references for iron and zinc are impediments to identifying

355 healthy and sustainable diets: lowering these thresholds can greatly limit the overall burden of disease  
356 as we have previously shown (Dussiot et al., 2021), and it also allows greater decreases in  
357 environmental pressures as shown here. Besides, as expected, and in line with available data in the  
358 literature, we found that some nutrients typically provided by animal products were limiting in the  
359 modeled diets designed to reduce environmental pressure, usually characterized by reduced  
360 consumption of animal products (Dussiot et al., 2021; Van Mierlo et al., 2017). Furthermore, although  
361 the health risk was greatly improved compared to the observed situation, it was still burdened by low  
362 overall consumption of whole grains.

363 Regarding the nutritional and health dimensions, the modeled diet proposed here as the "best"  
364 compromise, which allowed a substantial reduction in both GHGs (-50%) and BWU (-30%) compared  
365 to the observed situation, also had a reduced human health risk (as estimated by HRS) due to the  
366 suppression of red meat combined with the increase of plant-based foods (except for whole grains,  
367 remained at its lower limit, corresponding to its initial observed value, well below its TMREL value).  
368 The lack of increase in consumption of whole grains seems to be due to their lack of competitive  
369 advantage (or added value) over other food groups: their environmental costs (in terms of GHGe and  
370 BWU) outweigh their nutritional benefits (in terms of nutrients provided).

371 Also, the consumption of substitutes was greatly increased in all the diets modeled, regardless of the  
372 prioritization of BWU over GHGe reduction. Therefore, substitutes could be a good lever for water  
373 use, climate, and food sustainability. However, when we tested stricter dietary requirements for iron  
374 and zinc, substitutes were less introduced and failed to improve GHGe and BWU, arguing for their  
375 fortification as previously reported by us (Salomé et al., 2023). In addition, while substitutes have  
376 lower overall environmental impacts than animal products (Bryant, 2022; Mertens et al., 2017), their  
377 quality may be compromised because they are often ultra-processed (Kraak, 2022). In addition, some  
378 adverse health effects of soy foods have been suspected (Ahsan et al., 2018).

379 *GHGe*

380 The modeled diets prioritizing GHGe reduction ( $\lambda \geq 0.70$ ) were slightly healthier than others, due to  
381 lower consumption of red meat and higher consumption of legumes.

382 Numerous observational studies have documented that Mediterranean, vegetarian diets with a high  
383 contribution of plant-based foods are more favorable to the preservation of natural resources  
384 (Aleksandrowicz et al., 2016; Perignon et al., 2017; Ridoutt et al., 2017; Willett et al., 2019; Wilson et  
385 al., 2019). Diet optimization studies have also confirmed that reducing the share of animal products in  
386 favor of plant-based foods minimizes GHGe (Fouillet et al., 2023; Kesse-Guyot et al., 2021; Wilson et  
387 al., 2019).

388 *BWU*

389 Although the literature has been growing recently on that topic (Harris et al., 2020), a lower number of  
390 studies have focused on water use in food systems compared to those on GHGs (Hatjiathanassiadou et  
391 al., 2023) and the footprint indicators used were not consistent (blue, green, or blue and green).

392 Results can vary greatly depending on the indicator, especially for fruit and rice requiring more blue  
393 water than green water. This is illustrated by the findings reported by Mirzaie-Nodoushan et al. in a  
394 Iranian study aiming at minimizing water use (blue+green and blue) (Mirzaie-Nodoushan et al., 2020).  
395 The dietary changes identified were consistent with our results, regarding the reduced consumptions of  
396 food groups with low “nutritional water productivity” (water demand relative to their nutritional  
397 value) such as red meat and poultry, and the increased consumptions (by up to 80% in the Iranian  
398 study) of food groups with high “nutritional water productivity”, such as milk, fish, vegetables and  
399 legumes, which allowed to comply with certain nutritional references of which animal products are the  
400 principal supports. The intensity of the changes in food group consumptions depended on the type of  
401 water use to be minimized.

402 The findings strongly depend on how the acceptability constraints are defined. A diet modeling study  
403 that aimed at minimizing water use (in the first step of a hierarchical optimization approach) in food  
404 systems was conducted using Hungarian data (Tompa et al., 2022). This study found smaller reductions  
405 in diet-related water footprint (blue and green water use) for women and men compared to our study,  
406 possibly due to the different acceptability constraints used as the Hungarian study more drastically  
407 limited food changes by forcing modeled consumptions to stay between their observed 10<sup>th</sup> and 90<sup>th</sup>  
408 percentiles, which limited the reduction in the animal products being large contributors to green water  
409 use (Harris et al., 2020).

#### 410 *Compromise between GHGe and BWU*

411 In diet modeling studies about environmental footprint reductions, environmental indicators were  
412 generally included as constraints and not as objectives/targets to be minimized (Wilson et al., 2019),  
413 which makes not possible to assess the extent of their maximal improvements and their potential  
414 conflicts or alignments as in our study.

415 In the available scientific literature, the study with the closest design to our study, offering easier  
416 comparison of results, is the study conducted in 2016 by Gephart et al. (Gephart et al., 2016). This  
417 study was conducted on US data to identify diets minimizing several environmental footprints (with  
418 blue+green water use), one by one, under nutritional constraints. The authors concluded that the  
419 modeled diets were relatively similar regardless of the optimized indicator (high in plant foods and  
420 fish and low in other animal products), and interpreted these results as demonstrating synergies rather  
421 than conflicts between environmental indicators.

422 This is partially consistent with our findings. Indeed, we found that the GHGe and BWU values in all  
423 modeled diets, regardless of the relative weight given to their reduction in our compromise  
424 programming approach, were always markedly lower than their initial values in the observed diets.

425 This argues for a kind of synergy between BWU and GHGe in their responses to dietary changes, due  
426 to the fact that their strongest contributors are often the same food groups (e.g. beef, pork, poultry,  
427 offal). However, BWU varied much more widely than GHGe over their differential prioritization  
428 range. Our results show that a reduction in BWU leads to and is accompanied by a reduction in GHGe,  
429 but not necessarily vice versa. This is helpful in explaining why some studies have found a moderate  
430 decrease or even an increase in BWU with low GHGe diets (Fouillet et al., 2023). Indeed, our findings  
431 illustrate that among low emitting foods (e.g. fresh fruit, vegetables, and some refined and whole  
432 grains), some have quite strong water demands.

#### 433 *Best model*

434 In line with the best-balanced solution identified here, a study in India (Milner et al., 2017) that aimed  
435 at reducing BWU and GHGe through dietary changes showed that by reducing wheat, dairy products,  
436 poultry, and nuts, and increasing plants and legumes could achieve BWU reductions of 18% and 30%

437 by 2025 and 2050, respectively, and that there BWU reductions were accompanied by GHGe  
438 reductions of up to 13%.

439 These results seem to concur with ours to indicate that a decrease in water footprint leads to a  
440 concomitant GHGe decrease, while the reverse is not systematic according to our results.

441 This best-balanced diet is also globally coherent with the EAT-Lancet diet (Willett et al., 2019)  
442 proposed to preserve both human and planetary health. It should be noted that dairy products and offal  
443 (in small quantity) are present in this diet to meet the nutrient reference values, including vitamin A,  
444 zinc, iron and calcium.

#### 445 *Transferability, application, and perspectives*

446 Our research was based on data from a French context, covering food supply, eating habits, nutritional  
447 content, and environmental pressures. However, it is uncertain how applicable these findings are to  
448 other geographic locations. Furthermore, our emphasis on BWU, which pertains to reservoirs that can  
449 be used for irrigation, restricts the relevance of our outcomes to areas with differing top resources,  
450 water constraints, and climate conditions. It would be beneficial to examine alterations in  
451 environmental indicators caused by reallocating agricultural land via consequential LCA.

452 Unfortunately, the Agribalyse database does not permit such analysis. Further, climate change may  
453 lead to fluctuations in nutrient content, thereby influencing the nutritional worth of foods (Frumkin  
454 and Haines, 2019) like proteins, iron, zinc, and calcium, which must be kept in mind while designing  
455 prospective diets. These findings provide scientific evidence to inform sustainable food policies and  
456 highlight the importance of taking water usage into greater consideration.

#### 457 *Strengths/limitations*

458 Firstly, the main limitation is that the water footprint only considered the use of blue water.. The  
459 results are necessarily affected by this limitation because the food groups contributing to blue or green  
460 water differ (Harris et al., 2020). For example, animal products strongly contribute to green water  
461 while fruits and cereals strongly contribute to blue water. Second, LCA used herein did not consider  
462 the type of farming system (organic or conventional), limiting the consideration of the variety of  
463 practices and regionality along the food chain. Data on waste were not available, not allowing to focus  
464 on potentially avoidable environmental pressures while some authors have argued that limiting waste

465 throughout the food systems may be an important lever for reducing water use (Jalava et al., 2016).  
466 Furthermore, the models use parameters (nutritional contents, nutritional references, TMREL,  
467 environmental indicators) which are subject to uncertainties. Finally, optimization applied to the diet  
468 has inherent limitations in the methodology, which depends on the options selected in terms of  
469 definitions of constraints (e.g., no lower bound for food group consumption in our study), food  
470 grouping process and objectives (Mariotti et al., 2021). Concerning strengths, the data were based on  
471 LCA according to the standardized guidelines and methodologies environmental data were validated  
472 by several expert entities (ADEME, 2020). We used a food grouping with an appropriately high level  
473 of detail (45 distinct food groups), which provides an averaged picture of the nutrient density and  
474 environmental impact of detailed food categories, in order to identify the rebalancing of food groups  
475 required to achieve each studied objective (BWU and/or GHGe reduction) without the possible  
476 selection and over-representation in the modeled diets of particular food items that are not nutritionally  
477 and/or environmentally representative of their categories. Of note, an innovative compromise  
478 programming approach was used to thoroughly describe and understand the potential conflicts  
479 between environmental indicators which are multiple and not necessarily in alignment.

## 5 Conclusions

480 This study is the first to consider the reduction of GHGe and BWU in the same optimization model  
481 using compromise modelling. While focusing solely on BWU reduction induces a joint GHGe  
482 reduction that is near-maximal, the reverse is not true, showing that there is good alignment but also  
483 some divergence between the two objectives of lowering the carbon and water footprints. Meat is a  
484 major contributor to both indicators and the diets limiting both GHGe and water use are much more  
485 plant-based than actual diets, but if such more plant-based and healthy diets have low GHGe they may  
486 have more or less pronounced water use. Thus, this study suggests that water use in food systems  
487 should be better considered to define healthy and sustainable diets, and that otherwise prioritizing a  
488 lower-emitting diet per se may be counterproductive in terms of water use.

### 489 The authors' contributions

490 EKG conducted the research, implemented the databases, conducted the analyses and wrote the  
491 manuscript; HF provided tools and methodological support; All authors critically helped in the

492 interpretation of results, revised the manuscript and provided relevant intellectual input. They all read  
493 and approved the final manuscript; EKG had primary responsibility for the final content, she is the  
494 guarantor.

## 6 Conflict of Interest

495 The other authors declared no conflict of interest.

## 496 Data availability

497 The data collected in the INCA 3 study are available on the website

498 [https://www.data.gouv.fr/fr/datasets/donnees-de-consommations-et-habitudes-alimentaires-de-letude-  
inca-3/](https://www.data.gouv.fr/fr/datasets/donnees-de-consommations-et-habitudes-alimentaires-de-letude-<br/>499 inca-3/)

500 The data from Agribalyse ® are available on the ADEME website: <https://agribalyse.ademe.fr/>



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