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1 **Trapped microplastics within vertical redeposited sediment:**
2 **Experimental study simulating lake and channeled river systems**
3 **during resuspension events**

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14 **Keywords:** Microplastics, river sediments, low-velocity flow, deposition, settling velocity.

15 **Abstract**

16 Plastic waste and its fragments (microplastics; <5 mm) have been observed in almost all types of
17 environments. However, the mechanisms underlying the flow and transport processes of plastics are
18 unknown. This is particularly valid for river sediments, where complex interactions occur between
19 particles and influence their vertical and horizontal distribution patterns. In this study, we investigated
20 the vertical redistribution of 14 pristine microplastics (MPs) with different densities, sizes, and shapes
21 within disturbed sediment without lateral transport (i.e., low-velocity flow). MPs were spiked into
22 sediments (height: 8 cm) in a column with a height of 1 m (diameter: 6 cm) filled to the top with
23 water. The sediment was perturbed by turning the column upside-down to simulate remobilization
24 and the subsequent deposition of sediment. After the complete sedimentation of the particles, the
25 water column was filtered and the sediment was cut into vertical sections. MPs were then extracted
26 from the sediment using sieves and a density separation method, and were counted under a
27 stereomicroscope. Low-density polymers were mainly recovered in the water column and at the
28 surface of the sediment, whereas high-density polymers were found within all sediment sections. The
29 vertical distribution of high-density polymers changes primarily with the sediment grain size. The
30 distribution of each polymer type changes depending on the size and/or shape of the particles with
31 complex interactions. The observed distributions were compared with the expected distributions
32 based only on the vertical velocity formulas. Overall, the formulas used did not explain the
33 sedimentation of a portion of low-density polymers and predicted a lower distribution in the sediment
34 than those observed in the experiment. In conclusion, this study highlights the importance of
35 considering MPs as multi-dimensional particles and provides clues to understand their fate in low-
36 velocity flow systems, considering that they undergo scavenging in sediments.

37 **1. Introduction**

38 Attention to microplastics (MPs) and their economic, social, and environmental issues has increased
39 over the last few decades (Horton, 2021). MPs are found everywhere, from mountains to central open
40 oceans, and are transported by winds, adrift in water bodies, deposited or trapped in sediment, and
41 captured inside organisms (*e.g.*, Allen et al., 2019; Constant et al., 2021a; Miller et al., 2017; Santos
42 et al., 2021). Plastics have a slow degradation rate under most environmental conditions and can
43 therefore persist for long periods of time (Andrady, 2015). Moreover, plastic materials can contain or
44 adsorb contaminants (Rochman, 2015). MPs are of particular concern because both the probability of
45 entering the bodies of organisms (through ingestion or inhalation) and the reaction/adsorption surface
46 increase with decreasing size (GESAMP, 2015).

47 The majority of marine litter is known to have a land-based origin (UNEP and GRID-Arendal, 2016)
48 and rivers are considered as the main conveyors transporting MPs from terrestrial to marine
49 ecosystems (Lebreton and Andrady, 2019). Observations also suggest that rivers can act as a sink for
50 MPs trapped in sediment (Constant et al., 2021a), as is evident for other pollutants, including trace
51 metals and organic substances. MPs can accumulate or be remobilized at higher flow velocities. The
52 quantitative and temporal extent of this sequestration is poorly understood, but it is of paramount
53 importance, both for the exposure of benthic organisms and for the global budget of MPs. Despite an
54 increasing number of studies to identify and quantify MPs in aquatic environments, their sources,
55 transport, and fate remain unclear (Petersen and Hubbart, 2021; Rochman, 2018; Zhang et al., 2020;
56 Constant et al., 2017; Weiss et al., 2021). This is particularly true for the behavior of MPs in
57 freshwater, including sediments (Waldschläger et al., 2022; He et al., 2021). This lack of
58 understanding restricts our ability to properly estimate and understand the distribution of MPs in
59 rivers and their export to the oceans (Waldschläger and Schüttrumpf, 2019a).

60 In the water column, the movement of particles depends on their properties, hydrodynamics, and
61 interactions with other particles (Waldschläger and Schüttrumpf, 2019b). Sediment and/or
62 contaminant transport models generally integrate the density and size as parameters of particle

63 properties (Nizzetto et al., 2016). The first theoretical assessment of MP transport assumed that the
64 behavior of MPs is similar to that of sediments (Kooi et al., 2018). However, MPs are characterized
65 by a broad range of densities, sizes, and shapes (Rochman et al., 2019). Therefore, plastics and
66 sediments may behave differently (Waldschläger and Schüttrumpf, 2019). In particular, sediment
67 shapes range from nearly spherical (e.g., mature siliciclastic sediment) to nearly flat (e.g., muscovite
68 and biotite), and MPs mainly comprise fibers (elongated and thin cylinders; Constant et al., 2021a).
69 Understanding the similarities and differences between the behaviors of MPs and sediment, as well
70 as their interactions, is of particular interest for the accurate estimation of MP deposition in river
71 sediments.

72 This study aims to better understand the distribution of several MP types in sediment that were
73 recently deposited after perturbation under low-velocity flow conditions (e.g., barge traffic in canals
74 or dredging) through column experiments and theoretical approaches. Low-velocity flows are
75 common in lakes and anthropic waterways such as canals, which result in favorable conditions for
76 the deposition of particles. The objectives of our study were to: (1) observe the influence of MP and
77 sediment parameters on the redistribution of MPs; (2) predict the distribution of MPs based only on
78 vertical velocity formulas using MP and sediment features (size, shape, and density); and (3) compare
79 the experimental and theoretical data to identify the key parameters explaining the differences
80 between observations and a simple theoretical sedimentation situation.

81 Several field studies has investigated the link between sediment grain size and MP (e.g. Constant et
82 al., 2021a; Corcoran et al., 2020; Dhivert et al., 2022; Vermaire et al., 2017). However, experiments
83 on the sedimentation of particles and MPs have been carried out separately, while laboratory tests are
84 crucial for understanding this relationship, highly complex in the nature. This study investigates them
85 together. The results are expected to help future studies to better estimate the complex sedimentation
86 of MPs, improve global budgets, and predict future trends. Notably, we demonstrate the extent to
87 which sediments can be a sink for the densest MPs, redistributed within the sediment column, and
88 also how sediments can be a source of MPs when they are resuspended. These experiments and

89 simulations will integrate in the future the lateral transport of particles to improve models and to study
90 the export of these pollutants from terrestrial to marine environments.

91 2. Materials and methods

92 2.1. Column experiments: observed distribution

93 2.1.1. Spiked pristine microplastics and sediments

94 In each experiment, we used a mixture of low- and high-density polymers with different sizes and
95 shapes (Table 1 and Fig. A.1). We used polyethylene (PE; 0.9 g/cm^3) to represent low-density
96 polymers, and polyethylene terephthalate (PET; 1.4 g/cm^3) and polyester (PES; density: 1.4 g/cm^3) to
97 represent high-density polymers. PE, PES, and PET are among the most commonly produced
98 polymers (PlasticsEurope, 2020). The pellets were purchased from a plastic manufacturer (Acordis).
99 Fragments were obtained by cryo-crushing the pellets. Films were obtained by cutting pieces of
100 packing films and plastic bottles. Fibers were obtained by cutting the bobbin threads. The added MPs
101 could easily be distinguished from other particles (Fig. A.1) already present in the samples
102 (contamination). Each pristine MP type was previously identified by Fourier transform infrared
103 spectroscopy-attenuated total reflectance (FTIR-ATR) spectroscopy to check its polymeric
104 composition (Constant et al., 2021b; see method details in Appendix A.).

105 Calibrated natural siliceous sand sediments (data from the manufacturer: >98% silica with a density
106 of 2.65 ± 0.05 ; SNL) were used in the experiments. The features and granulometry are presented in
107 Appendix A (Table A.1 and Fig. A.2). Sediment and MPs were dried at $40 \text{ }^\circ\text{C}$ overnight and sieved
108 using a sieve column (2, 1.25 mm; 800, 630, 500, 400, 315, 250, 200, 125, 80, and $40 \text{ }\mu\text{m}$) placed in
109 a mechanical shaker. Sediment and MP size classes corresponded to the size classes obtained by
110 sieving, except for fibers, which were measured when cut.

111 **Table 1.** Main features of the virgin MPs spiked in sediment matrices. LD: low-density;
 112 HD: high-density.

Polymer	PE			PET			PES
Shape (dimension)	Pellet (3D)	Fragment (3D)	Film (2D)	Pellet (3D)	Fragment (3D)	Film (2D)	Fiber (1D)
Density (g/cm ³)	0.9 (low-density)			1.4 (high-density)			
Type	LD Fragment/Pellet		LD Film	HD Fragment/Pellet		HD Film	HD Fiber
Size classe (mm)	3-4	1-2 0.5-0.63	3-4 1-2 0.5-0.63	3-4	1-2 0.5-0.63	3-4 1-2 0.5-0.63	3-4 1-2

113

114 **2.1.2. Experimental setup**

115 For all tests, approximately 350 g of dry sediment (corresponding to a height of approximately 8 cm
 116 of dry sediment in the column) was transferred to a plastic column with a height and length of 1 m
 117 and 6 cm, respectively (Fig. A.3). A total of 140 MPs (10 per size and type) were added to the top of
 118 the sediment layer. The top of the column was then filled with tap water (~10 L). The column was
 119 turned upside-down six times (we observed, during preliminary tests, that sediment did not sink
 120 homogeneously until 3-4 upside-down and we selected 6 reversals to keep a margin), with a
 121 sedimentation period of 10 min. After the sixth reversal and complete sedimentation of the particles
 122 (2 h), the floating particles were collected using a pipette with a large opening. Water was siphoned
 123 off and filtered through a 40- μ m metal sieve. Particles collected on the sieve were pooled with those
 124 collected at the water surface and designated in the following text, figures, and tables as “floating
 125 MPs”. The sediments were pushed toward the upper part of the column, and five sections were cut
 126 with a metal trowel using a cylindrical guide and transferred to an aluminum container. The first
 127 section had a height of 0.5 cm (“surface” layer) and contained the MPs deposited on the sediment.
 128 The sediment was cut every 2 cm. The resulting five sediment sections and floating particles were
 129 dried at 40 °C overnight and then sieved through the same set of 12 sieves as used for sediment and
 130 MP preparation. The contents of the sieves were transferred to glass petri dishes. The experiment was
 131 repeated three times with 11 sediment size classes from 0.04 to 2 mm (i.e., 33 tests; Table A.1).

132 **2.1.3. MP separation and counting**

133 For most sieve contents, the amount of sediment was low, and MPs could directly be observed on
134 petri dishes under a stereomicroscope. When the amount of sediment covered the MPs (> 5 g) and
135 made the observation too laborious, the MPs were separated from the sediment using density-based
136 NaI (1.6 g/mL) extraction according to Constant et al. (2021a, 2021b). In particular, the sediments
137 were transferred to a glass container, and depending on the amount of sediment, 100-150 mL of
138 extracting solution was added. After 30 s of manual agitation, the first centrifugation was conducted
139 (5 min, 500 rpm), and the resulting supernatant was filtered through a metal sieve (40 µm). This
140 extraction procedure (centrifugation and sieving) was repeated thrice. The collected materials were
141 filtered through filter papers (Whatman©; 47 mm diameter; porosity of 2 µm). Finally, the filters and
142 petri dishes were observed under a Leica MZ12 dissecting stereomicroscope (magnifications of 6 ×,
143 12 ×, and 25 ×). For all MP types, more than 90% of spiked MPs were recovered.

144 **2.2. Estimation of vertical velocities and deposition: theoretical distribution**

145 The vertical velocities (w_s) of the seven types of MPs (Table 1) were estimated based on two empirical
146 equations (Equations 1 and 3). To infer the deposition pattern of the MPs in relation to the sediments
147 in the absence of lateral transport (i.e., simple sedimentation), the vertical velocities of sediments
148 were also calculated for two contrasting density intervals: between 1.5 and 1.7 (“soil”), representing
149 clays, soils, or river sediments with a high organic matter content, and between 2.5 and 2.7 (“sand”),
150 representing inorganic sediments.

151 Pellets and sediments were assimilated to spheres, whereas fragments and films were assimilated to
152 flat pieces of variable proportions. The formula published by Zhiyao et al. (2008) was used for these
153 shapes (Equation 1):

$$w_s = \frac{v}{d} d_*^3 [38.1 + 0.93 d_*^{12/7}]^{-7/8} \quad (1)$$

156 where d^* is the dimensionless particle diameter (Equation 2), d is the diameter (mm), and ν is the
157 kinematic viscosity of the fluid ($\text{mm}^2 \cdot \text{s}^{-1}$). For fragments, d corresponds to a characteristic size equal
158 to the cube root of the longest axis multiplied by the intermediate axis and the shortest axis.

$$d_* = \left(\frac{1}{\nu^2} \frac{g (\rho_w - \rho_p)}{\rho_w} \right)^{1/3} d \quad (2)$$

161 where g is the gravitational acceleration ($\text{mm} \cdot \text{s}^{-2}$), ρ_p is the particle density, and ρ_w is the fluid density
162 ($\text{kg} \cdot \text{m}^{-3}$).

163 For fibers, we used the semi-empirical formula (Equation 3) published by Khatmullina and Isachenko
164 (2017):

$$w_s = \frac{\pi}{2} \frac{1}{\nu} \frac{g (\rho_w - \rho_p)}{\rho_w} \frac{D L}{c1 L + c2} \quad (3)$$

167 where D is the diameter (mm), L is the characteristic length (mm), ρ_p is the particle density ($\text{kg} \cdot \text{m}^{-3}$),
168 and $c1$ (mm^{-1}) and $c2$ (unit-less) are empirical coefficients.

169 For each type of particle, 10,000 estimations of w_s were calculated. The sizes of the particles were
170 randomly retrieved over the size intervals used in the column experiment (MPs: 0.5-4 mm; sediments:
171 0.04-2 mm). Each size was associated with a density, which was also randomly retrieved over the
172 density intervals of the polymers (Table 1) and the density intervals of the two contrasting sediments
173 selected for deposit pattern comparison. Additional details regarding the used parameters are
174 presented in Appendix A. (Table A.2).

175 w_s values were then partitioned by MP and sediment size classes (Tables 1, A.1, and A.2). For each
176 class, the 2.5% ($Q_{0.025}$) and 97.5% ($Q_{0.975}$) quantiles, representing 95% of the particles, were selected
177 to describe the vertical velocities. The resulting vertical velocity ranges ($Q_{0.025}$ - $Q_{0.975}$) of MPs were
178 then compared with those of the sediment. The deposition patterns were inferred by considering three
179 situations (Fig. A.4):

- 180 • $w_{s[\text{MP}]} > w_{s[\text{sediment}]}$: MPs sink slower than sediments and settle at the top (surface) of the
181 deposit.
- 182 • $w_{s[\text{MP}]} < w_{s[\text{sediment}]}$: MP sink faster than sediments and settle at the base (bottom) of the deposit.
- 183 • $w_{s[\text{MP}]} = w_{s[\text{sediment}]}$: MP sink as fast as the sediments and settle homogeneously between the
184 top and bottom of the deposit (middle).

185 **2.3. Data analysis**

186 The percentage of MPs (%) was calculated for each pristine MP based on the number of MPs
187 recovered within each sediment section compared to the total number of MPs recovered in all
188 sections. Statistical analyses were conducted using R software (R Core Team, 2018), including the
189 data manipulation packages “dplyr” (Wickham et al., 2017), “tidyr” (Wickham, 2021) and “purrr”
190 (Henry and Wickham, 2020), and graphical package “ggpubr” (Kassambara, 2017) and colorblind-
191 friendly color scale “viridis” (Garnier et al., 2021). As the normality of distribution was not observed
192 (Shapiro-Wilk test), four non-parametric tests were conducted: the Wilcoxon-Mann-Whitney test to
193 compare two groups; Kruskal-Wallis test to compare more than two groups; Scheirer-Ray-Hare test
194 to investigate the influence of two different factors and the interaction among factors; and post-hoc
195 Dunn’s test to compare the differences of all possible pairs and pinpoint specific medians that are
196 significantly different from the others.

197 **3. Results**

198 **3.1. Observed distribution from the column experiments**

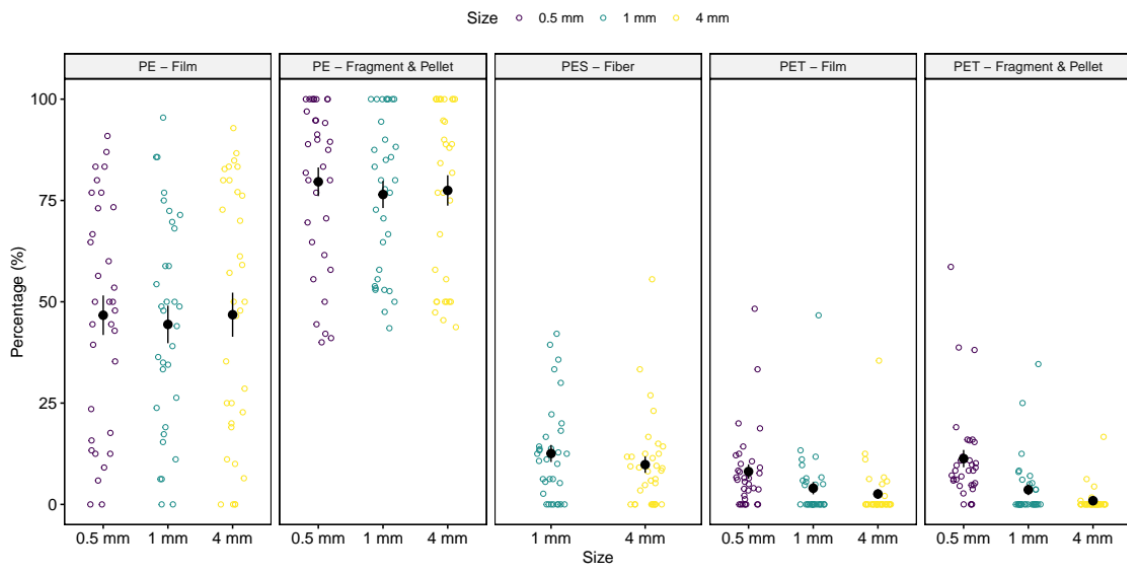
199 **3.1.1. Floating MPs**

200 Low-density MPs were significantly more abundant ($62 \pm 29\%$) within the water column than high-
201 density polymers ($7 \pm 10\%$) (Wilcoxon-Mann-Whitney test, $p < 0.01$; Table A.3).

202 The percentages of floating low-density polymers were very heterogeneous between tests, ranging
203 from 29 to 87% (Fig. 1). The percentage of fragments/pellets was significantly higher ($80 \pm 20\%$)
204 than those of films ($46 \pm 30\%$) (Wilcoxon-Mann-Whitney test, $p < 0.01$). The percentages were not
205 significantly different between each size class (Kruskal-Wallis test, $p = 0.80$). The interaction
206 between the shape and size was not significant (SHR test, $p = 0.93$).

207 The percentages of floating high-density polymers varied moderately between tests, ranging from 0
208 to 33%. Fibers were significantly more abundant within the water column ($11 \pm 12\%$) than films and
209 fragments/pellets (Wilcoxon-Mann-Whitney test, $p < 0.01$). The percentages of fragments/pellets ($5 \pm$
210 9%) and films did not differ significantly (Wilcoxon-Mann-Whitney test, $p = 0.89$). The percentages
211 of floating MPs were not significantly different between each size class of fiber (Wilcoxon-Mann-
212 Whitney test, $p = 0.23$). The percentages of small (0.5 mm) films and fragments/pellets (8 ± 10 and
213 $11 \pm 12\%$, respectively) were significantly higher than those of larger ones (1 and 4 mm) (3 ± 8 and
214 $2 \pm 5\%$, respectively) (Kruskal-Wallis test, $p < 0.01$). The interaction between the shape and size was
215 not significant ($p = 0.31$).

216 PE fragments/pellets, PES fibers, and PET films were significantly less abundant within the water
217 column with very fine sand (and fine sand for PE fragments/pellets) than with coarser sediments
218 (Kruskal-Wallis test, $p < 0.01$, $p < 0.01$, and $p = 0.04$, respectively). The distribution of PE films and
219 PET fragments/pellets was not significantly different across the sediment size classes (Kruskal-Wallis
220 test, $p = 0.32$ and $p = 0.08$, respectively).



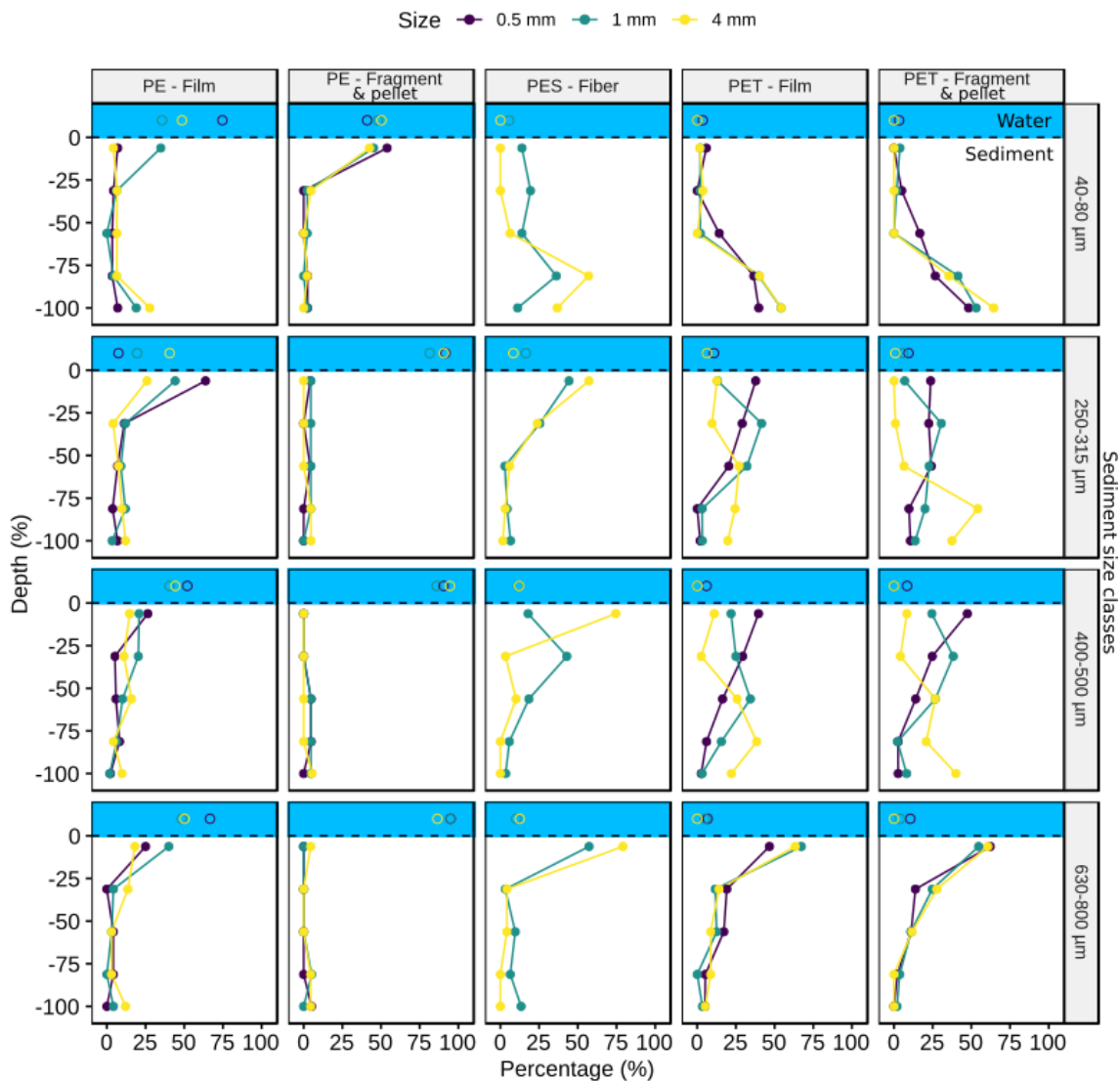
221 **Fig. 1.** Percentage of floating MPs (%) for each size, polymer, and shape. Small empty
 222 circles show the results of each test (i.e., 11 sediments \times 3 replicates). Large solid circles
 223 and bars represent means and standard deviations, respectively.

224 3.1.2. Sinking MPs

225 The percentages of sinking low-density polymers were very heterogeneous between tests, ranging
 226 from 12 to 71% (Fig. 2). Within the sediment column, the distribution of low-density polymers
 227 changed between the films and fragments/pellets (Fig. 2). PE films were more heterogeneously
 228 distributed and mostly present at the surface of all sediment size classes, except for the finest sediment
 229 (40-80 μ m), which was an important part of the bottom sediment column. PE fragments/pellets were
 230 homogeneously distributed along the vertical axis, with low percentages for each section, except for
 231 the finest sediment (40-80 μ m), where most of them were present at the surface of the sediment
 232 column. The distributions were relatively similar between the sizes for both the film and
 233 fragments/pellets.

234 The percentages of sinking high-density polymers varied moderately between tests, ranging from 67
 235 to 100% (Fig. 2). Inside the sediment column, the distribution changed with the sediment size class.
 236 At the smallest sediment size classes (40-80 μ m), the majority of high-density MPs were concentrated
 237 in the bottom part of the sediment column. For larger sediment sizes, the MPs were less deeply
 238 distributed. In the largest size classes (> 630 μ m), most of the high-density MPs were concentrated
 239 at the surface of the sediment column. The maximum concentrations of fibers were less deep than

240 those of fragments/pellets and films, except for the largest size classes ($> 630 \mu\text{m}$). Film and
 241 fragments/pellets had a relatively similar distribution. Fine and medium sand ($125\text{-}500 \mu\text{m}$) exhibited
 242 a unimodal distribution, skewed toward the bottom with fine sediments and toward the top with
 243 coarser sediment. Smallest fibers were deposited slightly deeper than larger ones (except at $40\text{-}80$
 244 μm), contrary to films and fragments/pellets, whose distributions became deeper with the increase in
 245 their sizes.



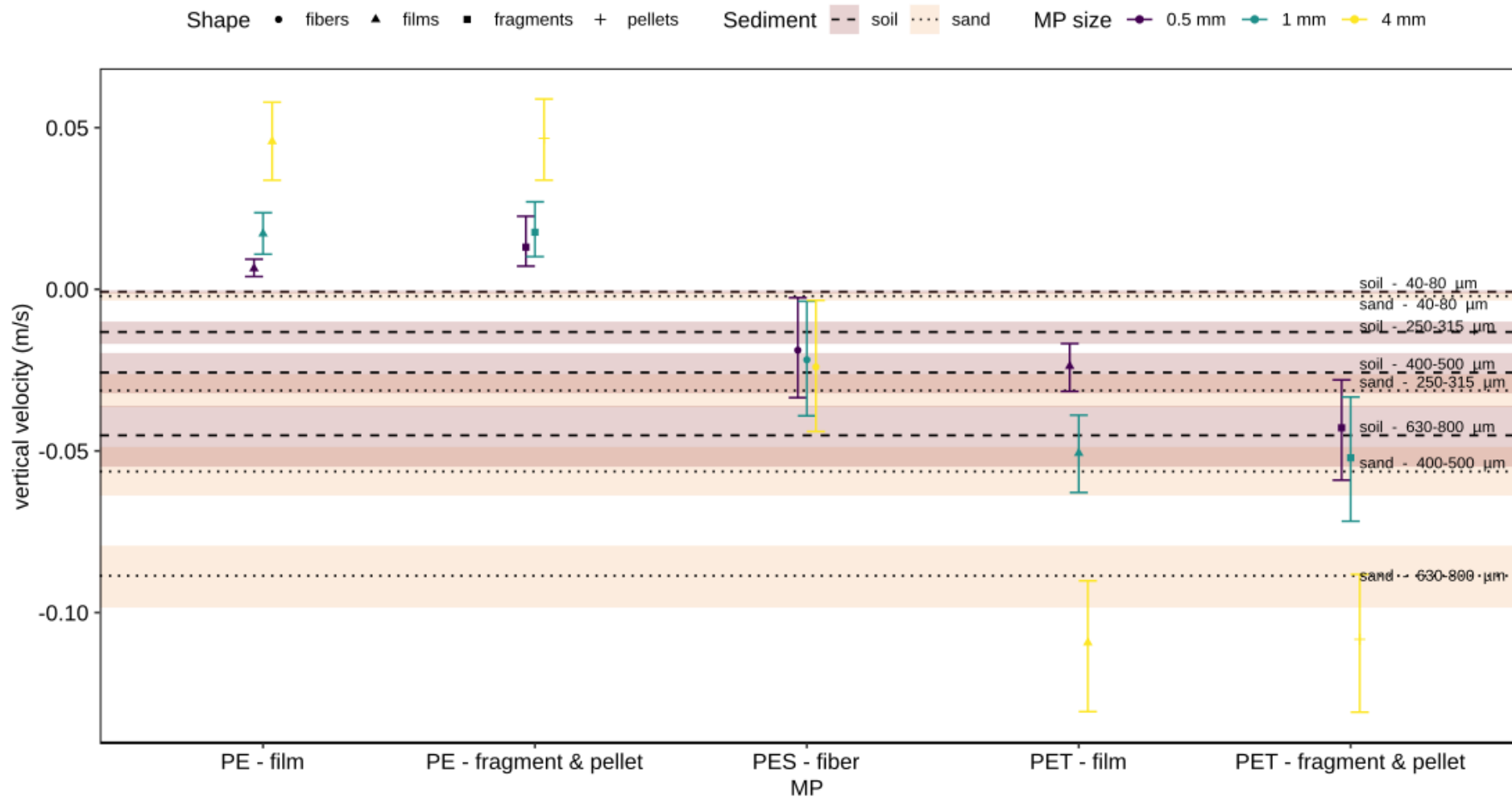
246 **Fig. 2.** Mean percentage (averages of three replicates) of MPs (%) within the water
 247 column and each sediment depth depending on four sediment grain size classes and
 248 sorting based on the polymer and shape. Empty circles represent the results within the
 249 water column (blue rectangle), whereas solid circles and lines represent the results within
 250 the sediment column. Fragment & pellet (PE and PET): 0.5 and 1 mm represent
 251 fragments, and 4 mm represent pellets. These four sediment size classes were selected
 252 because they show clearly visible changes from one to another. However, the entire set
 253 of results is displayed in Fig. A.5.

254 **3.2. Theoretical distribution based on the simulated vertical velocity**

255 The vertical velocities (w_s) were estimated (Table A.4) and compared for MPs and sediment particles
256 over the size intervals of the column experiment (Figs. 3, A.6, and A.7). Low- and high-density
257 polymers have a positive (rising) and negative (sinking) w_s values, respectively.

258 w_s values of the low-density polymers increase from smaller to larger particles and range between 0
259 and 0.05 m/s. Low-density films and fragments/pellets have similar w_s , excepted for those with a size
260 of 0.5 mm, wherein the w_s values of fragments/pellets are slightly more positive.

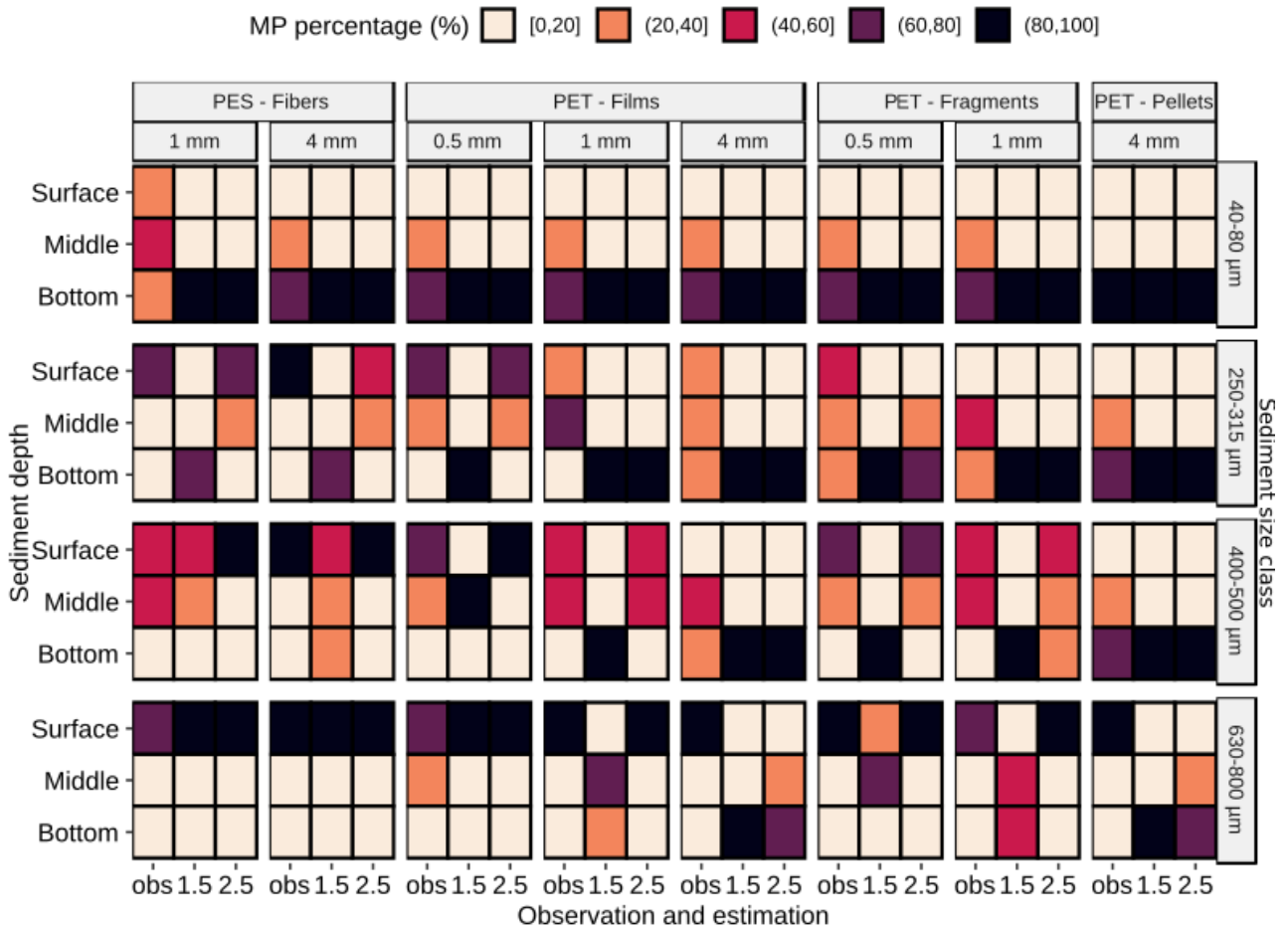
261 High-density films and fragments/pellets show more negative w_s than fibers, excepted for films with
262 a size of 0.5 mm. The w_s values of the fibers are roughly the same regardless of size, with a low
263 negative vertical velocity ranging between 0 and -0.05 m/s. For films and fragments/ pellets, the w_s
264 values decrease from smaller to larger and range from 0 to -0.15 m/s. Fragments with a size of 0.5
265 mm had slightly more negative w_s than the films with a size of 0.5 mm, but both their w_s values were
266 similar at sizes of 1 and 4 mm. $w_{s,\text{fiber}}$ was more negative than w_s of the finest sediment (40-80 μm)
267 and less negative than that of the coarsest sediments (soil: 630-800 μm ; sand >400 μm). The w_s values
268 of films and fragments/pellets are more negative than those of fine sediments (soil: <315 μm ; sand:
269 40-80 μm), and the w_s values of the largest particles (4 mm) are more negative than those of the
270 coarsest sediments (630-800 μm).



271 **Fig. 3.** Theoretical vertical velocities (m/s) of MPs and four sediment size classes. Icons and bars represent the means and interquartile ranges
 272 (95%) of the vertical velocity of MPs, respectively. Horizontal dashed lines and rectangles represent the means and interquartile ranges (95%)
 273 of the vertical velocity of soil (density: 1.5-1.7) and sand (density: 2.5-2.7), respectively. Fragments and pellets (PE and PET): 0.5 and 1 mm
 274 represent fragments (square) and 4 mm represent pellets (cross). Results for the 11 sediment size classes are shown in Fig. A.7. The details
 275 of the vertical velocity estimation in presented in section 2.2.

276 3.3. Comparison between the experiment and theory

277 We summarized the results acquired from the column experiments (“observed” distribution; Fig. 2)
278 over three simplified depths (surface: top 0-0.5 cm; middle: between 0.5-6.5 cm; bottom: bottom 6.5-
279 8 cm), and compared them with the expected deposition patterns (“theoretical” distribution) obtained
280 using the vertical velocity results based on the hypothesis of simple sedimentation (Fig. 3; see method
281 details in section 2.2). The observed distribution using sand was closer to the theoretical distribution
282 estimated (density: 2.5-2.7) than that using with soil (density: 1.5-1.7) (Fig. 4 and A.9). To a large
283 extent, the theoretical distribution of high-density MPs within soil is deeper than the observed
284 distribution. Within the finest sediment size class (40-80 μm), high-density MPs should theoretically
285 be below the sediment (“2.5” distribution). In the experiment (“observed” distribution), most of the
286 MPs (60-80%) were expected to be effectively within the bottom section, but a significant portion of
287 MPs (20-40%) were present in the middle part. PES fibers with a size of 1 mm were mainly (40-60%)
288 present in the middle part. For the coarser sediments, the fiber-observed distributions are fairly similar
289 to the 2.5-density theoretical distributions, slightly below the distributions of particles with sizes of
290 400-500 μm and 600-830 μm . For the 250-315 μm fraction, the PET was mainly above the theoretical
291 distribution. For the 400-500 μm fraction, the observed distribution was similar to the theoretical
292 distribution for the films with a size of 1 mm and fragments with a size of 0.5 mm, and slightly above
293 the theoretical distribution for the other films and fragments/pellets. Finally, for the largest sediment
294 size class (600-830 μm), high-density MPs were mainly present at the surface, whereas simulated
295 vertical velocities predicted a deeper distribution for large films and fragments/pellets (4 mm).



296 **Fig. 4.** Observed and theoretical distributions of high-density polymers (PES and PET)
 297 within three simplified sediment depths (surface, middle, and bottom) depending on four
 298 sediment size classes sorted according to the polymer and shape. Observed (obs) MP
 299 percentages are the average of three replicates. Observation and estimation: “observed
 300 distribution” (obs), theoretical distribution with soil of “density 1.5-1.7” (1.5), and sand
 301 of “density 2.5-2.7” (2.5). Details about the theoretical distributions are presented in
 302 section 2.2. Sediment depth: “top 0-0.5 cm” (surface), “between 0.5-6.5 cm” (middle),
 303 “bottom 6.5-8 cm” (bottom). Low-density polymers (PE) have a positive vertical velocity
 304 and are theoretically not present within the sediment. Results for the 11 sediment size
 305 classes are shown in Fig. A.8. Differences between the observation and estimation are
 306 shown in Figs. A.9 and A.10.

307 **4. Discussion**

308 Theoretically, two major processes occur after the resuspension of the sediment: the settling or the
309 rise of the particles in the water column, followed by the superposition of the sinking particles on top
310 of each other (Baba and Komar, 1981). In the water column, the movement of particles also depends
311 on their properties and interactions with each other (Kooi et al., 2018). Denser, larger, and more
312 spherical particles settle faster; however, some processes (e.g., particle aggregation) can modify this
313 pattern (Alimi et al., 2017; Andersen et al., 2021; Leiser et al., 2021; Li et al., 2022, 2019). The
314 superposition pattern of the particles is related to their settling velocity, as well as their capacity to
315 slip into the interstices between other particles during deposition (infiltration behavior). In this
316 experiment, infiltration was presumed to be negligible, as the large size of the MPs drastically reduced
317 the probability of infiltration (Waldschläger and Schüttrumpf, 2020).

318 In the column experiments, the distribution of MPs varied greatly within and among the sediments.
319 The MP type also plays a key role in this distribution. Nevertheless, some patterns were clearly
320 demonstrated. The features of the sediment (density and size) and MPs (density, size, and shape)
321 influence the deposition of MPs, but to different degrees. Additionally, the characteristics interact
322 with each other, resulting in opposing trends. We observed that the distribution of MPs was directly
323 related to their density, wherein the majority of low-density polymers remained in the water column
324 after resuspension, whereas the majority of high-density polymers were trapped in the bulk sediment
325 (i.e., throughout the sediment column). The vertical distribution of MPs within the sediment also
326 changes with sediment features. Overall, the percentage of MPs reaching the deepest sections
327 increased with decreasing sediment grain size. The vertical distribution of high-density polymers also
328 changes according to their shape. Fibers were found to be more abundant than pellets, fragments, and
329 films. Finally, the distributions of fragments/pellets and films changes with their sizes, in contrast to
330 that of fibers. Thus, with the sediments tested, the distribution of MPs was roughly consistent with
331 their theoretical order of sinking and appeared to be preliminarily related to their settling velocities.

332 The most important deviation from theory is the significant fraction of low-density polymer that settle
333 at the top of the sediment (Fig. 1). Low-density polymers have positive vertical velocities, which
334 force them to rise within the water column. Particle adhesion, aggregation, or biofouling may decrease
335 the vertical velocity of low-density polymers (Chubarenko et al., 2016; Leiser et al., 2021). According
336 to the used equations, the rising velocities of the PE change with size but not with shape (Fig. 3).
337 However, the observed distribution of low-density MPs is affected by the particle shape, as mentioned
338 by Waldschläger et al. (2020), they observed that the rising velocity of films was lower than that of
339 fragments and highlighted the deformation behavior of films. For the same size and density, the
340 irregular shape of a particle, compared to a sphere, increases its secondary movements (i.e., not
341 vertical), which retards its sinking or rising behavior (Waldschläger and Schüttrumpf, 2019a).
342 Additionally, films might be particularly sensitive to particle adhesion, aggregation, or biofouling (in
343 nature but not in these experiments) because of their high surface-area-to-volume ratio (Chubarenko
344 et al., 2016). High-density polymers also constitute a small but significant fraction that reaches the
345 water column. Moreover, for low-density polymers, interactions with “slower” particles (e.g., fine
346 sediments or organic matters) may decrease the settling velocity of particles (Chubarenko et al.,
347 2016). However, this result has an important probability of being an artifact. Notably, virgin MPs can
348 have a significant hydrophobicity ((Al Harraq et al., 2022); more details are presented in the last
349 paragraph of the discussion). To a lesser extent, the distribution of high-density fibers in the sediment
350 also deviates from expectations. Theoretically, the influence of the fiber size is low. Notably, fibers
351 align and sink horizontally at the same velocity, regardless of their length (Waldschläger and
352 Schüttrumpf, 2019a). Moreover, thinner fibers can infiltrate deeper (Waldschläger and Schüttrumpf,
353 2020), but the fibers used in the experiment had the same diameter, and as mentioned above, such
354 phenomena have a low probability of occurrence. In the experiment, small fibers tended to be
355 deposited slightly deeper in fine and coarse sediment than in large fibers, and the opposite was true
356 in very fine sediment (namely, large fibers were deposited above small fibers). As the fibers are
357 regular cylinders, the surface of the contact increases with the length (at a constant diameter), which

358 may explain the slight difference observed in our study. Finally, the sedimentation of low-density
359 MPs increases with very fine sediments for fragments/ pellets, but not for films. Li et al. (2022)
360 observed that the settling rate of small polystyrene microspheres (<300 μm) increased in fine
361 suspended sediment (median: 16.3 μm) compared to that in no sediment.

362 Xia et al. (2021) studied the resuspension of MPs within columns of sediment samples using a
363 customized particle entrapment simulator (PES) device. Compared to that in the control samples (i.e.,
364 undisturbed), a higher proportion of small-sized MPs (50–500 μm) was observed within the overlying
365 water or the deep sediment and a lower proportion settled at the sediment surface, suggesting a
366 disturbance-induced transport from the sediment surface. No change was observed in the distribution
367 of large-sized MPs (0.5-5 mm). Unfortunately, their results cannot be compared with those in our
368 study, as the authors did not indicate whether density and shape affected these patterns.

369 Overall, the observed distribution of virgin MPs was in accordance with the expected distribution
370 based on the vertical velocity formulas (Fig. 4). The formulas logically did not predict the
371 sedimentation of low-density polymers but tended to position high-density polymers lower in the
372 sediment. Excluding the potential biases of the experiment (see details below), these results may be
373 explained either by a misconfiguration of some parameters in the formulas or by the absence of
374 parameters on the effect of particle interactions. The vertical velocity formulas used here are based
375 on Stokes' formula and include empirical parameters determined in previous experimental studies
376 (section 2.2). These parameters should be corrected in further experiments considering the high
377 diversity of MPs (Kooi et al., 2018). Some equations accounting for microbial colonization have been
378 proposed (Nguyen et al., 2020), but parameters accounting for hetero-aggregation with inorganic
379 particles are still missing (Alimi et al., 2017). This study highlights the importance of including the
380 latter in future theoretical studies. Finally, the vertical velocity formulas can be used in the first
381 approximation, but still need to be strengthened.

382 Before presenting some assumptions regarding the environmental implications of these observations,
383 it is necessary to mention the limitations of this experiment. First, MPs were placed at the surface of
384 the sediment, which may have induced different results than if the MPs and sediment particles were
385 mixed in advance. However, as column reversal was conducted six times, this bias was probably
386 limited. Second, the experiments displayed a simplified version of the conditions occurring in natural
387 systems. In natural environments, sediment movements are not horizontally restricted, and flow
388 velocity is rarely null; therefore, sedimentation is probably more diffuse. This may also result in an
389 amplified trapping effect due to a high concentration of particles, which transport some MPs in the
390 sedimentary flux independently of their sedimentation behavior. In contrast, the column may
391 engender the wall effect (Baba and Komar, 1981) and generate air bubbles (Waldschläger and
392 Schüttrumpf, 2019a), both of which may modify the settling or rising velocity. Finally, we used virgin
393 MPs to increase the repeatability of the experiments and facilitate the counting procedure; however,
394 they are more prone to have a significant hydrophobicity (Al Harraq et al., 2022)) and are not
395 associated with other solids, such as biofilms, for a long time. In addition, composition and surface
396 coating may influence the interaction with other particles, including homo- and hetero-aggregation
397 (Wang et al., 2021). Considering these limitations, some hypotheses on the fate of MPs are proposed:
398 (i) the physical perturbation of the sediment may modify the distribution and pool of MPs; (ii) low-
399 density polymers trapped in the sediment may be released to the water column and be dragged along
400 by the current; (iii) they may settle subsequently, especially films, but will be concentrated at the
401 surface of the sediments; (iv) high-density polymers have a higher probability of being trapped again
402 in the sediment, and their global vertical distribution will depend on the sediment grain size, but also
403 on the nature of the MPs themselves (size and shape). This last point explains, in to a certain extent,
404 why no relevant correlation has been observed between sediment grain size and MP distribution in
405 field studies (Constant et al., 2021a; Corcoran et al., 2020; Vermaire et al., 2017).

406 All these hypotheses examine the consequences of sediment resuspension in aquatic systems, where
407 lateral transfers (currents) are limited (lakes and canalized rivers in particular). This may cause

408 bioturbation phenomena at a scale of millimeters or centimeters, wave action at scales of several
409 centimeters or even tens of centimeters when propellers touch the sediment surface, and dredging at
410 scales of several tens of centimeters to a meter because these processes systematically lead to
411 important sediment remobilization.

412 **5. Conclusion**

413 This study investigated key parameters to understand the sdeposition and sinking of MPs by
414 experimentally and theoretically comparing the distributions of MPs in water and sediment after the
415 remobilization of sediments without lateral transport (i.e., low-velocity flow). Fourteen combinations
416 of density, size, and shape of MPs were tested for light and dense pristine polymers in 11 sediment
417 grain sizes. The main conclusions based on the 33 experimental tests and simple sedimentation
418 simulations are as follows:

419 (1) The observed distribution largely follows the deposition pattern induced by differential vertical
420 velocities between particles: most the low-density polymers float; most the high-density polymers
421 settle deeper in fine sediments than in coarse ones; high-density pellets, fragments, and films settle
422 deeper than fibers and pellet, and large fragments and films settle deeper than small ones.

423 (2) Unexpectedly, an important part of the low-density polymers settled at the surface of the sediment.
424 This pattern was more important for films than fragments and pellets, probably because of the specific
425 vertical behavior of the films. The sedimentation of low-density fragments increased with very fine
426 sediments, thereby highlighting the importance of particle interactions. The effect of shape was not
427 the same for high- and low-density polymers.

428 (3) Vertical velocity formulas roughly predict the distribution of MPs but tend to position high-
429 density polymers lower in fine sediment and upper in coarse sediment. Vertical velocity formulas
430 could therefore be useful as a preliminary approach but still need to be improved.

431 This study can be regarded as a first step in the investigation of MP deposition in disturbed sediments,
432 as many questions still need to be answered. Further experiments should be conducted on the effects
433 of organic matter and heterogeneous-sized sediments, as well as the effects of the MP degradation
434 state and their associations with other particles, such as biofilms, on the MP deposition behavior.
435 Finally, from the initial conclusions of vertical particle sedimentation, interesting results can be

436 obtained by integrating the lateral transport of particles during the experiments and theoretical
437 models, as is the case in rivers, estuaries, and seawater.

438 **CRedit authorship contribution statement**

439 Mel Constant: Methodology, Investigation, Writing - Original Draft. Claire Alary: Conceptualization,
440 Writing - Review & Editing, Supervision. Lisa Weiss: Formal analysis, Writing - Review & Editing.
441 Alix Constant: Investigation. Gabriel Billon: Conceptualization, Writing - Review & Editing,
442 Supervision.

443 **Declaration of competing interest**

444 The authors declare that they have no known competing financial interests or personal relationships
445 that could have appeared to influence the work reported in this paper.

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452 plastic pellets.

453 **Appendix A. Supplementary data**

454 Methods: spectroscopy FTIR.

455 Figures: Photos of MPs spiked (A.1) and the column (A.3), grain size distribution (A.2), explanatory
456 diagram of section 2.2 (A.4), percentage of floating MPs (A.5), theoretical vertical velocities (A.6
457 and A.7), observed and theoretical distributions (A.8), and difference between them (A.9 and A.10).

458 Tables: Grain size distribution (A.1), parameters and results of w_s (A.2 and A.4), and statistical
459 analysis (A.3).

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