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# Evolutionary Epidemiology Consequences of Trait-Dependent Control of Heterogeneous Parasites

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**ABSTRACT:** Disease control can induce both demographic and evolutionary responses in host-parasite systems. Foreseeing the outcome of control therefore requires knowledge of the eco-evolutionary feedback between control and system. Previous work has assumed that control strategies have a homogeneous effect on the parasite population. However, this is not true when control targets those traits that confer to the parasite heterogeneous levels of resistance, which can additionally be related to other key parasite traits through evolutionary trade-offs. In this work, we develop a minimal model coupling epidemiological and evolutionary dynamics to explore possible trait-dependent effects of control strategies. In particular, we consider a parasite expressing continuous levels of a trait-determining resource exploitation and a control treatment that can be either positively or negatively correlated with that trait. We demonstrate the potential of trait-dependent control by considering that the decision maker may want to minimize both the damage caused by the disease and the use of treatment, due to possible environmental or economic costs. We identify efficient strategies showing that the optimal type of treatment depends on the amount applied. Our results pave the way for the study of control strategies based on evolutionary constraints, such as collateral sensitivity and resistance costs, which are receiving increasing attention for both public health and agricultural purposes.

**Keywords:** trade-offs, mathematical modeling, evolutionary epidemiology, disease management, heterogeneous pathogens, trait-structured populations.

## Introduction

Disease and pest control strategies aim to eradicate or mitigate the exploitation of a parasite population on a host

population of economic (agriculture) or public health (humans) interest (Gilligan 2002). By altering the ecological host-parasite interactions, a control strategy can induce both demographic effects (Gilligan and van den Bosch 2008) and evolutionary responses (Day et al. 2020). Predicting the outcome of control strategies in a system therefore depends on our understanding of its eco-evolutionary feedbacks (Day and Gandon 2006), that is, on the evolutionary epidemiology behavior (Galvani 2003).

In addition to experimental studies, theoretical work based on evolutionary epidemiology has provided insights into control strategies and on their (often counterintuitive) consequences. For instance, host reduction (e.g., via culling) may increase disease abundance and prevalence (Bolzoni and De Leo 2013); altering parasite growth, such as in vaccination campaigns, can lead to selection for more virulent parasites, thus increasing disease severity (Gandon et al. 2001, 2003; Zurita-Gutiérrez and Lion 2015). Studies of the emergence and spread of multidrug resistance have also shown that the efficacy of a control strategy may depend on the structure of the host population (e.g., age or spatial distributions) as well as on parasite heterogeneity (Blanquart et al. 2018; Lehtinen et al. 2019; McLeod and Gandon 2021). Furthermore, models also provide a low-cost tool to explore and optimize large-scale agricultural practices (Rimbaud et al. 2018), such as the deployment of disease-resistant cultivars (Taylor and Cunniffe 2022) and crop rotation (Bargués-Ribera and Gokhale 2020), whose in-field implementation can consume considerable time and resources.

Most of the experimental and theoretical work has assumed that control strategies have a homogeneous effect on parasites or that parasites are endowed with a genetically

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encoded resistance that is either present or absent (qualitative resistance; REX Consortium 2010). However, this picture neglects the cases where the efficacy of a treatment depends on quantitative traits that can be heterogeneously expressed in the parasite population (quantitative resistance; Corwin and Kliebenstein 2017). In fact, parasites have developed several defense mechanisms that can quantitatively affect drug uptake, leading to variable levels of resistance (Munita and Arias 2016; El Meouche and Dunlop 2018). Widespread examples of such mechanisms are metabolic regulators (Chebotar et al. 2021), biofilms (Costerton et al. 1999; Fanning and Mitchell 2012), efflux pumps (Martinez et al. 2009), flagella (Lyu et al. 2021), and capsules (Song et al. 2021). Thus, a parasite population will often exhibit heterogeneity in the expression of key traits (Hewitt et al. 2016; González et al. 2019; Perrier et al. 2019; Schröter and Dersch 2019; Dutta et al. 2020), potentially leading to heterogeneous trait-dependent treatment effects (Porco et al. 2005; Laine and Barrès 2013; Martínez et al. 2019; Alizon 2020) and ultimately threatening the control strategy's overall efficacy (Gefen and Balaban 2009; Patyka et al. 2016; Weigel and Dersch 2018; Dewachter et al. 2019).

The occurrence of trait-dependent treatment effects is expected to be exacerbated in the future, as many of the new promising strategies to counter resistance escalation are based on the exploitation of trait-specific evolutionary constraints (Palmer and Kishony 2013; Lässig et al. 2017; Furusawa et al. 2018), such as fitness costs of resistance (Lenski 1998; Andersson and Hughes 2010; Vincent et al. 2013; Hawkins and Fraaije 2018), life history (Shoval et al. 2012), and metabolic (Weißé et al. 2015; Pinheiro et al. 2021) trade-offs and collateral sensitivity (Lázár et al. 2018; Barbosa et al. 2019; Maltas and Wood 2019; Maeda et al. 2020; Roemhild and Andersson 2021). For instance, bacterial efflux pumps rely on proton motive force both to import some toxic compounds (e.g., aminoglycosides; Taber et al. 1987; Alekshun and Levy 2007) and to expel others (e.g.,  $\beta$ -lactams; Okusu et al. 1996; Lázár et al. 2013; Suzuki et al. 2014); therefore, strains with a reduced proton motive force will be more resistant to one antibiotic but more sensitive to the other—and vice versa (Pál et al. 2015; Roemhild and Andersson 2021). This phenomenon is a textbook example of antibiotic collateral sensitivity, but it also affects other types of control strategies, such as fungicides, copper use, and phage therapy, and it can potentially involve other fields of application, such as biocontrol methods in agriculture.

Phage therapy employs viruses (phages) that selectively attack bacteria and ultimately kill them. It has been observed that resistance to both phages and antibiotics is often costly for bacteria (Meade et al. 2015; Mangalea and Duerkop 2020; Laure and Ahn 2022): selection pressure often leads to the evolution of bacterial strains that are resis-

tant to either phage or antibiotic therapy, which can therefore be applied in combination to obtain synergistic antimicrobial effects (Torres-Barceló and Hochberg 2016; Coyne et al. 2022; Kebriaei et al. 2022). In addition, recently discovered phages targeting mechanisms involved in both antibiotic resistance and virulence are smart tools to restore antibiotic treatment efficacy and co-select for avirulent strains (Chan et al. 2018; Gurney et al. 2020). Phages attach to specific bacterial external structures (e.g., flagella, capsules, or efflux pumps) that are involved in important biological processes, such as antibiotic resistance and pathogenicity (Chan et al. 2016; Song et al. 2021; Esteves and Scharf 2022). Modifications in these components make phage infection more difficult and are therefore selected for in bacterial populations exposed to phages. These changes inhibit the bacteria's previous ability to cause disease and to resist antibiotics, thereby restoring their sensitivity to the treatment (Chan et al. 2016; Chiarelli et al. 2020; Gurney et al. 2020). Specifically, bacteria with a downregulated production of efflux pumps would, on the one hand, avoid phage infection. On the other hand, however, they would be more sensitive to antibiotics or toxic heavy metals that are detoxified by these pumps.

A similar mechanism underlies the behavior observed in *Myzus persicae*, an aphid considered a major threat to agriculture (Van Emden and Harrington 2017) and an important living model for the study of insecticide resistance (Bass et al. 2014): mutations in the metabolic activity can lead to the emergence of clones that are more resistant to insecticides due to reduced uptake rates but also more vulnerable to natural enemies (Foster et al. 2007), with non-trivial consequences on their demography (Foster et al. 2011). When exposed to both insecticides and biocontrol (natural enemies or pathogens), insects may face fitness trade-offs that prevent them from maintaining the same level of resistance (Lacey et al. 2015). Therefore, by exerting different selection pressures on a pest, a synergistic use of chemical and biocontrol has the potential to contain resistance development and maintain crop productivity while minimizing the negative environmental impacts by potentially reducing chemical doses (Ons et al. 2020), akin to the phage-antibiotics case.

Since traits that lead to resistance are often also involved in the parasite's ability to exploit and harm the host (Beceiro et al. 2013; Alcalde-Rico et al. 2016; Giraud et al. 2017; Copin et al. 2019), the outcome of a trait-specific treatment may be complicated by the presence of trade-offs between resistance mechanisms and other life history traits of the parasite (Boots and Haraguchi 1999; Boots and Bowers 2004), such as those involving host exploitation and disease-induced mortality (known as the transmission-virulence trade-off) in the case of microparasites (Bull 1994; van den Bosch and Gilligan 2008; Alizon et al. 2009) and those

involving foraging activity and life span in the case of macroparasites (Werner and Anholt 1993; Anholt et al. 2000; Gotthard 2000; Brodin and Johansson 2004; Stoks et al. 2005; Strobbe et al. 2011). Knowledge of novel trait-specific mechanisms is broadening the spectrum of possible selection pressures we can exert on parasites (Allen et al. 2014), and a full exploitation of the potential of such new strategies depends on our understanding of the eco-evolutionary feedbacks between the treatment and the biological system, at various levels of description (Burmeister et al. 2021; Perry 2021).

Accounting for a detailed description of the therapy-parasite interactions have provided insightful information on the *in vitro* behavior of specific systems (Bull et al. 2014; Mattei et al. 2018; Nichol et al. 2019; Rodriguez-Gonzalez et al. 2020; Aulin et al. 2021). However, their results can be hardly generalizable, and the corresponding population-level information can be tricky to obtain. Here, we are interested in developing a general framework, shared in principle by any control strategy, where (i) treatment efficacy depends on a target trait; (ii) different treatment types correlate differently with the target trait; (iii) target traits are heterogeneously expressed across the parasite population, leading to heterogeneous treatment effects; and (iv) target traits may be related to other parasites' traits through trade-offs.

We tackle this issue by means of a minimal ecological model describing the dynamics of a generic, valuable resource and of a generic parasite population (Lafferty et al. 2015). Following a well-established evolutionary epidemiological approach (Day and Proulx 2004), we model parasites as a trait-structured population, characterized by heterogeneous levels of expression of key traits. We consider a single proxy trait variable accounting for the possible trade-offs between the parasite's level of exploitation and mortality and its resistance to treatment. Crucially and differently from previous work (Frank 1996; Alizon and van Baalen 2005; Porco et al. 2005), we allow treatment efficacy to either correlate or anticorrelate with the proxy trait to reproduce potentially different control strategies as well as combinations of them. We focus our analysis on the implications of trait-dependent treatment from an agricultural perspective, where maximizing resources and reducing treatment use are two major objectives (WHO 2014; Medina-Pastor and Triacchini 2020). Using a simple multicriteria analysis, we show that the ability to tune trait-dependent control can harness parasite heterogeneity to our advantage: in particular, we show the existence of optimal treatment types and the emergence of saturation effects on resource production gains. By focusing on a quantity of general economic interest (healthy resource at equilibrium), our results can be applied to a variety of agricultural practices and are potentially extendable to different scenarios.

## Models and Methods

### *The Model*

In the spirit of Lafferty et al. (2015), we start with a minimal ecological model describing the interaction between resource ( $R$ ) and parasite ( $P$ ) biomass, potentially representative of both microparasite and macroparasite scenarios. The former scenario neglects the explicit dynamics of within-host abundance (as is commonly done for pathogens; Anderson and May 1979, Keeling and Rohani 2011):  $R$  represents the biomass of healthy, susceptible hosts, while  $P$  represents the biomass of hosts infected by the pathogen. In the latter scenario (typical of crop-pest and plant-herbivore systems),  $R$  represents the resource biomass, and  $P$  represents the biomass of the parasite that consumes it. In both scenarios,  $R$  is treated as a renewable resource exploited by  $P$ . The treatment has the effect of killing the parasite, thus removing  $P$  biomass. In what follows, we first present the underlying assumptions of the minimal model, from which we then develop the eco-evolutionary formulation.

*Homogeneous RP Formulation with Treatment.* Our assumptions are summarized as follows. First, resource biomass is renewed at a constant rate. Second, resource biomass is converted to parasite biomass upon exploitation: in the microparasite scenario, it corresponds to the transmission of the infection from an infected to a susceptible host; in the macroparasite scenario, it corresponds to the consumption of the resource by the consumer. Third, resource and parasite biomass can be removed from the system by various possible mechanisms acting on both (e.g., natural mortality) or either (e.g., disease-induced mortality). Fourth, parasite biomass is eradicated at a rate proportional to treatment application and efficacy: in the microparasite scenario, eradication is intended in the sense of removing the pathogen from the host (Hall et al. 2004; Castle and Gilligan 2012); in the macroparasite scenario, eradication is intended in the sense of killing the parasite (van den Bosch and Gilligan 2008).

Under the above assumptions, the dynamics of  $R$  and  $P$  biomass is given by the following system of equations:

$$\frac{dR}{dt} = \theta - \delta^R R - \beta RP + \zeta \gamma \phi P, \quad (1a)$$

$$\frac{dP}{dt} = \epsilon \beta RP - (\delta^P + \gamma \phi) P, \quad (1b)$$

where  $\theta$  is the resource renewal rate;  $\delta^{R,P}$  are the mortality rates;  $\beta$  is the exploitation rate;  $\epsilon$  is the coefficient accounting for biomass conversion;  $\gamma \phi$  is the total eradication rate per unit of parasite biomass, where parameter  $\gamma$  is the treatment application rate and  $\phi$  is the treatment efficacy; and  $\zeta$  is the treated parasite's fate parameter.

In the microparasite scenario,  $\beta$  corresponds to the transmission rate of the infection, and  $\epsilon = 1$ . The fate of the treated infected host is determined by  $\zeta$  as follows: for  $\zeta = 1$ , the treated host is restored to  $R$  (i.e., treatment provides recovery without immunity, as may be the case with animal antibiotics; Hethcote 2000; Forster and Gilligan 2007); for  $\zeta = 0$ , the treated host is removed from the system (recovery with immunity or permanent removal, as is generally the case with plant diseases; Hall et al. 2004). Intermediate values ( $0 < \zeta < 1$ ) can model in-between cases. As will be shown, for our purposes it is not necessary to specify the fate of the treated host, since the results presented here are independent of the value of  $\zeta$ .

In the macroparasite scenario,  $\beta$  corresponds to the consumption rate of the resource,  $0 < \epsilon < 1$  (biomass conversion from resource to consumer) and  $\zeta = 0$  (treatment removes consumers from the system).

In the standard ecological formulation of the model, the population is considered to be homogeneous, so that treatment efficacy, exploitation, and mortality rates have constant values across individuals. The behavior of this basic model is well known, and we refer the reader to Korobeinikov and Wake (2002) for details.

*Trait-Structured RPx Formulation with Trait-Dependent Treatment.* Following previous work (Day and Proulx 2004; Korobeinikov 2018; Day et al. 2020; Sasaki et al. 2022), we elaborate the formulation equivalent to model equations (1) in the case of a parasite heterogeneously expressing a trait. The parameters of the model and their description are summarized in table 1. The following assumptions are introduced. First, the parasite has a continuum of strains, mathe-

**Table 1:** Variables and parameters used in the model

|               | Description  |
|---------------|--|
| $R$           | Resource biomass   |
| $P$           | Parasite biomass   |
| $x$           | Trait variable   |
| $p(x)$        | Parasite trait distribution                              |
| $\bar{x}$     | Average trait value                                      |
| $\mu$         | Mutations diffusion coefficient                          |
| $\theta$      | Resource renewal rate                                    |
| $\delta^R$    | Resource mortality rate                                  |
| $\gamma$      | Treatment application rate                               |
| $\zeta$       | Treatment fate parameter                                 |
| $\delta^P(x)$ | Parasite mortality rate                                  |
| $\delta_0^P$  | Parasite baseline natural mortality rate                 |
| $\delta_1^P$  | Trait-dependent mortality contribution                   |
| $\beta(x)$    | Parasite exploitation rate                               |
| $\beta_0$     | Baseline exploitation rate                               |
| $\beta_1$     | Trait-dependent exploitation contribution                |
| $\phi(x)$     | Trait-dependent treatment efficacy                       |
| $\phi_1$      | Type of treatment ( $ \phi_1 $ degree of specialization) |

matically described by a single continuous proxy variable  $x \in \mathcal{T}$ , where  $\mathcal{T}$  is the trait space. For the sake of simplicity, we will also consider the unit interval  $\mathcal{T} = [0, 1]$  as trait space. The generalization to any positive interval  $[x_{\min}, x_{\max}]$  is straightforward and can be mapped to the interval  $[0, 1]$  by rescaling of the parameters. Heterogeneity is described by the trait distribution  $p(x)$ , that is, the density of parasites carrying trait  $x$ . Second, the trait determines the parasite's levels of resource exploitation, mortality, and treatment efficacy, which are now represented by the functions  $\beta(x)$ ,  $\delta^P(x)$ , and  $\phi(x)$ . Third, parasites undergo mutations that induce small changes in their traits and maintain heterogeneity within the population; mutation rates are high compared with the ecological timescale, they are unbiased, and no preferred direction is assumed.

In the heterogeneous formulation, the system is described by the ecological and evolutionary states. The ecological state is given by the amount of  $R$  and  $P$  biomass. The evolutionary state is given by the trait distribution  $p(x)$ . Mathematically speaking,  $p(x)$  is a probability distribution over which it is possible to compute average quantities with respect to the parasite population. Given the assumptions described above, the dynamics of the heterogeneous system is provided by the following system of equations (details are provided in the supplemental PDF):

$$\frac{dR}{dt} = \theta - \delta^R R - \overline{\beta(t)} RP + \zeta \gamma \overline{\phi(t)} P, \quad (2a)$$

$$\frac{dP}{dt} = \epsilon \overline{\beta(t)} RP - [\overline{\delta^P(t)} + \gamma \overline{\phi(t)}] P, \quad (2b)$$

$$\frac{dp(x)}{dt} = \mu \frac{\partial^2 p(x)}{\partial x^2} + p(x)[F(x) - \overline{F(t)}], \quad (2c)$$

with

$$F(x) = \epsilon \beta(x) R - \delta^P(x) - \gamma \phi(x). \quad (3)$$

The bar notation indicates the average over the trait distribution; thus,  $\overline{\beta(t)}$  is the average exploitation rate of the population:

$$\overline{\beta(t)} = \int_{\mathcal{T}} \beta(x) p(x) dx. \quad (4)$$

In the equation above, the time dependence is retained in order to recognize that such averages are not fixed but change over time with the time variation of the trait distribution  $p(x)$ . Equivalent definitions apply to average mortality  $\overline{\delta^P(t)}$ , the average treatment efficacy  $\overline{\phi(t)}$ , and the function  $\overline{F(t)}$ .

Equations (2a) and (2b) describe the population dynamics at the demographic level. They are equivalent to the classical formulation equations (1) upon replacing

the single-strain parameters  $\beta$ ,  $\delta^p$ , and  $\phi$  with their population-average counterparts  $\overline{\beta(t)}$ ,  $\overline{\delta^p(t)}$ , and  $\overline{\phi(t)}$ . Equation (2c) instead describes the population dynamics at the evolutionary level, as it governs the changes in the parasite trait distribution due to mutations and competition for limited resource.

Phenotypic mutations are captured by the diffusion operator  $\partial^2/\partial x^2$  over the trait space,  $\mu$  being the diffusion coefficient related to mutations; this choice assumes that mutations induce small perturbations in the quantitative trait, that is, a parasite mutates into a “phenotypically close” variant (possible biases in the direction of mutations may be accounted for by introducing a gradient term in the last equation; Kimura 1965; Chisholm et al. 2016; Lorenzi et al. 2016).

Concomitantly, parasites compete between each other for access to a limited amount of resource (i.e., infection of a limited number of hosts or consumption of a limited resource), according to the trait-dependent function  $F(x)$ : the exploitation term  $\epsilon\beta(x)R$  contributes to increasing the density of trait  $x$ , whereas mortality  $\delta^p(x)$  and efficacy  $\phi(x)$  contribute to decreasing it.

The overall success of parasites with trait  $x$  depends on the difference between its value of  $F(x)$  and the population average  $\overline{F}$ , as in a replicator dynamics (Schuster and Sigmund 1983). Thus,  $F(x)$  represents the fitness landscape structuring the parasite’s competition for exploitation. Unlike purely theoretical work, this fitness landscape is not assumed. Rather, it emerges from the ecological interactions (Day et al. 2020). Note that the evolutionary equation takes the same form regardless of the parasite’s nature and of the fate parameter  $\zeta$ .

The system of equations (2) shows neatly the intertwining of ecological and evolutionary levels of description that is typical of eco-evolutionary dynamics: on the one hand, the demography of the population (given by  $R$  and  $P$ ) depends on the trait distribution  $p(x)$  via the average quantities  $\overline{\beta(t)}$ ,  $\overline{\delta^p(t)}$ , and  $\overline{\phi(t)}$ ; on the other hand, the trait distribution depends on the demography via the ecological interactions (as exploitation depends on  $R$ ). The solution of the heterogeneous problem (and the methods needed to obtain it) depends on the choice of the trait-dependent functions, which are detailed below.

#### *Trait-Dependent Trade-Offs*

We are interested in cases where the proxy trait  $x$  provides the parasite with different levels of resource exploitation and, consequently, mortality due to possible life history trade-offs. This trait will also provide a quantitative response to treatment, depending on the type of control strategy employed, so as to model different possible trait-specific treatments and their consequent heterogeneous efficacy. Therefore, in the following we will refer to exploitation,

mortality, and efficacy as functions of the proxy trait variable  $x$ . Our mathematical choices aim to capture the basic biological features of the trade-offs of interest while maintaining mathematical tractability. Relaxation of such choices does not alter the qualitative features of our model (details are provided in the supplemental PDF). References to the examples presented in the introduction should be taken as qualitative connections underpinning our approach rather than exact, detailed descriptions of the resistance, exploitation, and mortality mechanisms at play.

*Exploitation.* We assume exploitation rate  $\beta$  to be linearly affected by the trait variable  $x$ ; that is,

$$\beta(x) = \beta_0 + \beta_1 x, \quad (5)$$

where  $\beta_0$  is the baseline rate and  $\beta_1$  the trait-dependent contribution. We recall that in the microparasite scenario equation (5) corresponds to the transmission rate of the infection, and in the macroparasite scenario it corresponds to the consumption rate.

*Mortality.* In the microparasite scenario, increased exploitation (i.e., transmission rate) is often associated with increased harm to the infected host (i.e., the transmission-virulence trade-off; Montarry et al. 2006; Sacristán and García-Arenal 2008; Laine and Barrés 2013; Zhan et al. 2015; Nelson and May 2020); in the macroparasite scenario, increased exploitation (i.e., consumption rate) is often associated with a reduced parasite life span due to, for example, increased respiration or risk exposure (Werner and Anholt 1993; Anholt et al. 2000; Gotthard 2000; Brodin and Johansson 2004; Stoks et al. 2005; Strobbe et al. 2011). In both scenarios, the respective trade-offs lead to an increase in parasite mortality. Accordingly, we assume that parasite mortality  $\delta^p$  can be linearly dependent on trait  $x$ :

$$\delta^p(x) = \delta_0^p + \delta_1^p x, \quad (6)$$

where the parameter  $\delta_0^p$  is the baseline natural mortality and  $\delta_1^p$  is the trait-dependent contribution (Day and Proulx 2004; Porco et al. 2005; Bolzoni and De Leo 2013).

For values of both  $\delta_1^p$  and  $\beta_1 > 0$ , parasites with higher exploitation will also have higher mortality, consistent with the trade-off hypothesis described above.

*Treatment Efficacy.* We assume that efficacy  $\phi(x)$  is maximal for one of the extreme values of the trait ( $x = 0$  and  $x = 1$ ). Consistent with Porco et al. (2005), we choose a linear functional dependence:

$$\phi(x) = C_{\phi_1} + \phi_1 x, \quad (7)$$

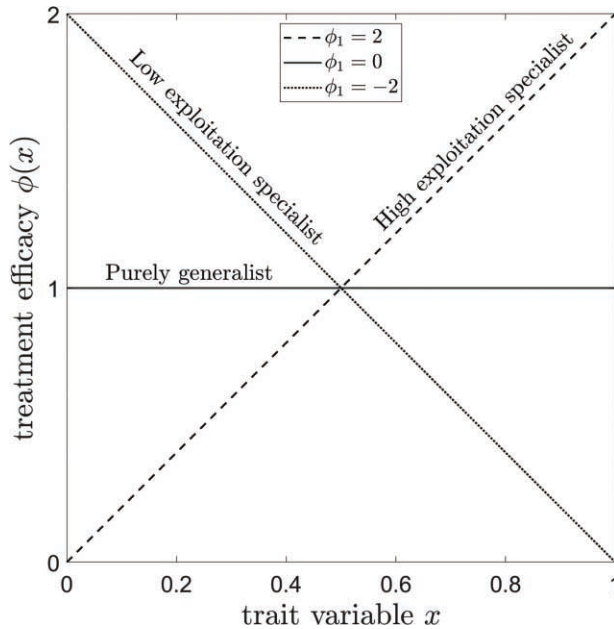
where the parameter  $\phi_1$  represents the degree of correlation between the treatment and the trait (as explained in the next subsection) and  $C_{\phi_1}$  is a normalization factor. A

graphical summary of the treatment spectrum is shown in figure 1B (note that the bound  $|\phi_1| < 2$  ensures a positive  $\phi(x)$ ). The linear assumption is made for the sake of mathematical tractability, although nonlinear saturating choices are considered more realistic (Alizon 2020). The effects of nonlinearities can be explored semianalytically, but they do not affect the quality of the results presented herein (details are provided in the supplemental PDF).

Combination of equations (5), (6), and (7) captures the possible trade-offs occurring between trait-dependent treatment efficacy and the parasite's exploitation levels. Ultimately, the eco-evolutionary dynamics of the system depends on the values of the parameters of the functions described above and thus on how a control strategy interacts with the proxy trait value  $x$ .

### Modeling Treatment Spectrum

We assume that the environmental and economic costs of the use of treatment are proportional to its application rate  $\gamma$ . We also assume that trait  $x$  determines the level of expression of a target trait, such as efflux pump expression, metabolic activity, or proton motive force. The type of



**Figure 1:** Spectrum of treatments with trait-dependent efficacy. The type of treatment is determined by the sign and the magnitude of the parameter  $\phi_1 \in [-2, 2]$ : positive  $\phi_1$  models types with maximal effect on strains with higher exploitation levels ( $x = 1$ ), and negative  $\phi_1$  models types with maximal effect on strains with lower exploitation levels ( $x = 0$ ). Large  $|\phi_1|$  are specialized in targeting extreme values of the trait, and small  $|\phi_1|$  are generalist types with more uniform action.

treatment then depends on how it correlates with the target trait, which is specified by the slope parameter  $\phi_1$ .

For instance, efflux pumps provide bacteria with resistance to chemical compounds but make them vulnerable to attack by certain phages. Therefore, a standard antibiotic or pesticide treatment is more efficient on strains with lower levels of efflux pump expression (small  $x$ ), and it is represented by a negative  $\phi_1$ ; instead, phage therapy benefits from higher levels of efflux pump expression (large  $x$ ), and it is represented by a positive  $\phi_1$ .

Metabolic activity provides aphids an increased ability to consume and reproduce, but it also makes them also more sensitive to pesticides, as they will tend to take up more toxic compounds. Therefore, the pesticide is more efficient on fast-exploiting strains (large  $x$ ) and less efficient on slow-exploiting strains (small  $x$ ); its slope  $\phi_1$  is then positive. At the same time, reduced metabolic activity makes the aphids more vulnerable to natural enemies, so that predator-based biocontrol will have a negative  $\phi_1$ .

Proton motive force reduces the import of aminoglycosides but also the export of  $\beta$ -lactams. Therefore, aminoglycoside is more efficient against strains expressing less proton motive force (small  $x$ ) and less efficient against strains with more proton motive force (large  $x$ ); it is thus represented by a negative  $\phi_1$ . However,  $\beta$ -lactam has the opposite effect, so it is represented by a positive  $\phi_1$ .

In any case, maximal efficacy is obtained at the extremes of the trait space (either  $x = 0$  or  $x = 1$ ), so it is assumed that the correlation of efficacy with traits is unimodal. In the absence of more precise data, we consider this assumption to be a reasonable starting point and leave other possible cases for discussion. The normalization factor  $C_{\phi_1} = 1 - \phi_1/2$  ensures that  $\int_T \phi(x) dx$  is always normalized to 1, for any value of  $\phi_1$ . In addition to removing an arbitrary degree of freedom, this condition also imposes a plausible evolutionary constraint consistent with the notions of costs of resistance and of collateral sensitivity: resistance with respect to a particular treatment is paid for by high sensitivity with respect to others, as has been observed in the examples illustrated above.

Recognizing that a treatment may be either positively or negatively correlated with the proxy trait variable, we assume the existence of a continuous spectrum from which it is, in principle, possible to choose. This continuous spectrum mimics the possibility of synergistically combining different types of treatments, as has been documented for phages and antibiotics (Gu Liu et al. 2020; Kebriaei et al. 2022), antibiotics and antivirulents (Rezzoagli et al. 2020), fungicide mixtures (van den Bosch et al. 2014), antibiotic mixtures (Cokol et al. 2011; Nichol et al. 2019), fungicide biocontrol (Lima et al. 2006), and pesticide biocontrol (Foster et al. 2007). Crucially, we assume that the overall efficacy

is the result of the sum of the single different treatment types. Therefore, any slope can be obtained by adjusting the relative proportions of antibiotic and phage therapy. Although we expect a full spectrum to be practically unavailable, it allows us to fully explore the eco-evolutionary behavior of the system. Henceforth, the term “specialist” will refer to treatments with a large degree of correlation (large  $|\phi_1|$ ), and the term “generalist” will refer to treatments with a small degree of correlation (small  $|\phi_1|$ ). A control strategy is thus characterized by one value of application rate  $\gamma$  and one value of treatment type  $\phi_1$ .

#### *Agent-Based Numerical Simulations*

We compared the deterministic dynamics presented in our article with numerical simulations of the equivalent agent-based dynamics. The Python code is available from Zenodo (<https://zenodo.org/record/7874696#.ZFu4AnZBybh>; Miele et al. 2023). The agent-based dynamics simulates the stochastic events of renewal, mortality, exploitation, treatment, and mutation, each occurring at a rate consistent with the deterministic equations (2). The deterministic theory is expected to be consistent with our agent-based simulations as long as large populations are involved. Our aim is to provide a numerical validation of the existence and stability of the endemic equilibrium solution predicted by the theory. The analysis of finite size effects on the system is beyond the scope of this article, although it is straightforward to perform once the agent-based codes are set up (Ardaševa et al. 2020). Details of the numerical simulations can be found in the supplemental PDF.

### **Results**

#### *Equilibrium Trait Distribution and Evolutionary States*

The state of the parasite population at equilibrium is described by the steady-state trait distribution  $p(x)$ , the solution of

$$\mu \frac{\partial^2 \widehat{p(x)}}{\partial x^2} + \widehat{p(x)} [F(x) - \bar{F}] = 0, \quad (8)$$

where the average quantities  $\bar{\beta}$ ,  $\bar{\delta}^p$ , and  $\bar{\phi}$ , appearing in  $\bar{F}$ , are not known a priori, as they depend on the solution  $p(x)$  itself. Note that their time dependence has been dropped, since they reach a constant value at equilibrium. The equation given above has a trivial solution  $p_0(x) = 0$ , corresponding to the parasite-free equilibrium, and a nontrivial solution describing the endemic equilibrium, where resource and parasites coexist (mathematical details are provided in the supplemental PDF).

The coexistence equilibrium exists (and is stable) as long as the following condition is satisfied:

$$R_0 = \frac{\theta}{\delta^R} \frac{\epsilon \bar{\beta}}{\delta^p + \gamma \bar{\phi}} \geq 1, \quad (9)$$

where  $R_0$  is the expected offspring produced by a parasite encountering an unexploited resource, per unit of biomass (also known as the basic reproduction number), for the heterogeneous system. Contrary to the classical formulation, the condition described above cannot be calculated directly in terms of the ecological parameters, since it depends on average quantities that are not known a priori. Therefore, one must first solve equation (8) for a given set of parameters, then calculate the average quantities and check with condition equation (9) the existence of the endemic equilibrium. In the supplemental PDF, we show that the behavior of the solution to equation (8), with linear trait-dependent functions, is entirely determined by the following compound parameter  $\Omega$ :

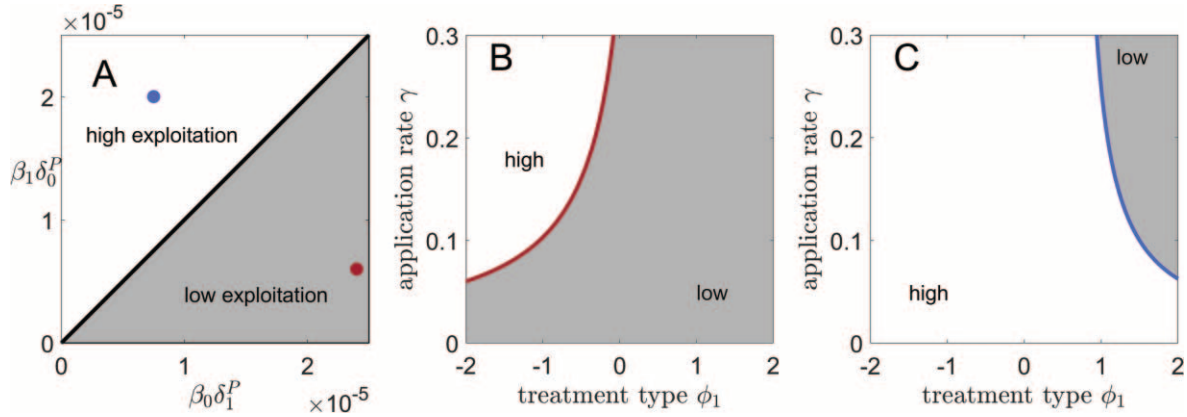
$$\Omega = \beta_1 \delta_0^p - \beta_0 \delta_1^p + \gamma(\beta_1 - \beta_1 \phi_1 / 2 - \beta_0 \phi_1). \quad (10)$$

In particular, if  $\Omega < 0$ , the solution is monotonically decreasing and the trait distribution is mostly distributed close to the trait  $x = 0$ . We will refer to this as the “low-exploitation” state because the corresponding trait has a minimum value for  $x = 0$ . On the other hand, if  $\Omega > 0$ , the solution is monotonically increasing and the trait distribution is mostly distributed close to the trait  $x = 1$ . Likewise, we will refer to this as the “high-exploitation” state.

Below we characterize the phase diagram corresponding to a set of parameters, both in the presence and in the absence of treatment. In this case, the proxy trait variable  $x$  simultaneously determines the levels of transmission and mortality of the parasite as well as the efficacy of the treatment. In the absence of treatment (i.e.,  $\gamma = 0$ ), the parasite population can be found in either the high- or low-exploitation state, depending on the value of the ecological parameters. With reference to figure 2A, low states will be favored for large baseline exploitation  $\beta_0$  and trait-dependent mortality contribution  $\delta_1^p$ ; instead, high states will be favored for large trait-dependent exploitation  $\beta_1$  and baseline mortality  $\delta_0^p$ .

The introduction of treatment (i.e.,  $\gamma \neq 0$ ) can lead to a change in state, depending on the control strategy  $(\phi_1, \gamma)$  employed. In figure 2B, we show how a system initially in the low state (parameters corresponding to the red point in fig. 2A) adapts after treatment application, as a function of the control parameters  $\gamma$  and  $\phi_1$ . At low doses (i.e., low  $\gamma$ ), the parasite population will remain in the low state, regardless of the type employed. However, increasing the application rate will eventually bring the system to the high state if negative  $\phi_1$  or generalist types are employed. If the system is initially in the high state in the absence of treatment (fig. 2C; blue point of fig. 2A), the complementary behavior is observed.

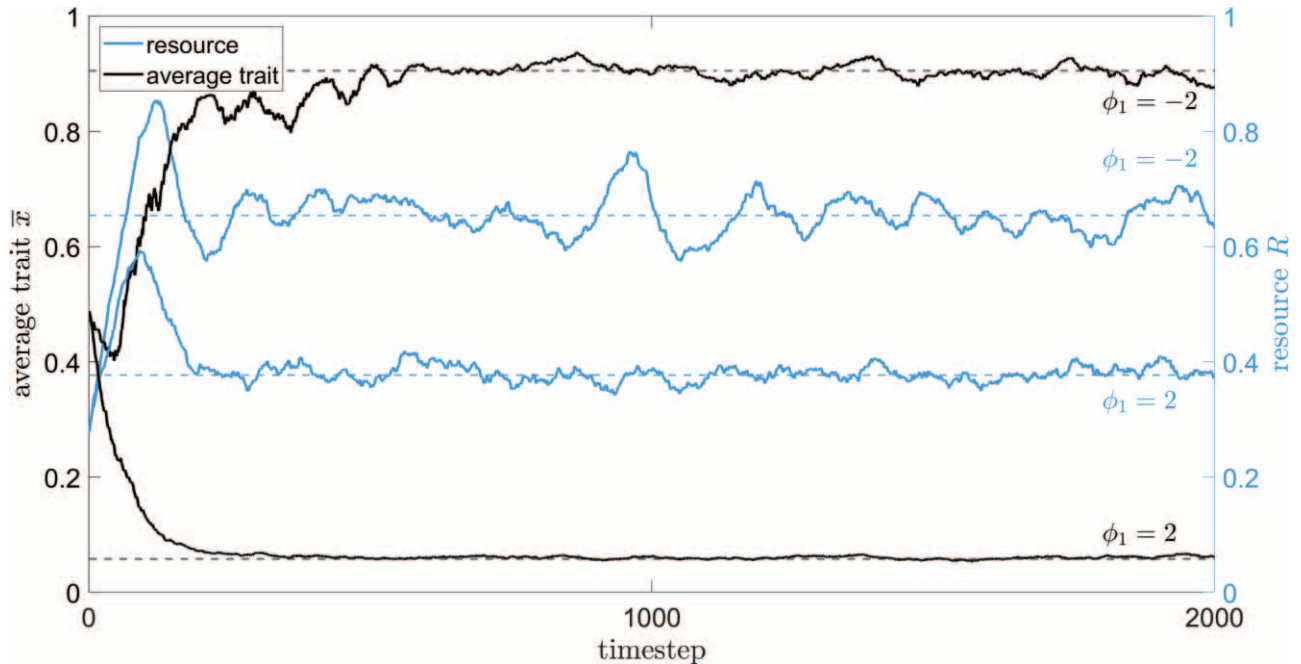




**Figure 2:** Trait distribution states. *A*, State diagram of the trait distribution in the absence of treatment. The parasite population can adapt toward either low-exploitation (gray regions) or high-exploitation (white regions) states. *B*, A system initially in the low state (red point in *A*) can switch toward the high state under a range of control strategies. The red curve separates the two regions of the control parameters. *C*, Likewise, a system initially in the high state (blue point in *A*) can switch toward the low state under a range of control strategies. The blue curve separates the two regions of the control parameters. Parameters: for the red point,  $\beta_0 = 0.0001$ ,  $\beta_1 = \beta_0/2$ ,  $\delta_0^p = 0.12$ ,  $\delta_1^p = 0.24$ ; for the blue point,  $\beta_0 = \beta_1 = 0.0001$ ,  $\delta_0^p = 0.2$ ,  $\delta_1^p = 0.075$ .

In addition to affecting the evolutionary state of the system at endemic equilibrium, the control strategy also affects the amount of equilibrium resource. The adaptation toward the two possible states is shown in figure 3, where we plot the trajectories of the simulated agent-based dynamics (solid

lines), differing for the treatment type employed. The other parameters (indicated in the figure caption) and initial conditions are identical. For  $\phi_1 = 2$  the system adapts toward the low state, and for  $\phi_1 = -2$  it adapts toward the high state. Consequently, the resource reached at equilibrium is



**Figure 3:** Simulated temporal trajectories. Solid lines: temporal trajectories of the resource  $R$  (rescaled with respect to  $R_0$ ) and of the average trait  $\bar{x}$ , obtained from agent-based numerical simulations of the dynamics. Dashed lines: analytical equilibrium values predicted by the deterministic theory. The system is initialized with identical initial conditions and the same application rate  $\gamma = 0.1$  but a different treatment type  $\phi_1$ . For  $\phi_1 = \pm 2$ , resource and parasite average trait attain different equilibrium values. Other parameters:  $\theta = 200$ ,  $\delta^r = 0.04$ ,  $\beta_0 = 0.0001$ ,  $\beta_1 = \beta_0/2$ ,  $\delta_0^p = 0.12$ ,  $\delta_1^p = 2\delta_0^p$ ,  $\mu = 7 \times 10^{-5}$ .

different in the two cases. The figure also shows the agreement between the analytical predictions for the equilibrium values (dashed lines) and the agent-based trajectories (solid lines), which holds for all sets of parameters considered in our analysis.

*Equilibrium Resource and Treatment Effects*

In the following, we will focus on the equilibrium resource  $R$ , which can be considered as the amount of harvest with economic value (Cunniffe et al. 2015; Vyska et al. 2016). The population equilibria are obtained by setting equations (2) to zero, and they are equivalent to the classical formulation. If  $R_0 < 1$ , we have a trivial parasite-free equilibrium:

$$\left(\widehat{R}_0 = \frac{\theta}{\delta^R}; \widehat{P}_0 = 0\right), \tag{11}$$

corresponding to the extinction of the parasite. If  $R_0 \geq 1$ , we have the following stable endemic equilibrium:

$$\left(\widehat{R} = \frac{\overline{\delta^P} + \gamma\overline{\phi}}{\epsilon\overline{\beta}}; \widehat{P} = \frac{\epsilon\overline{\beta}\theta - \delta^R(\overline{\delta^P} + \gamma\overline{\phi})}{\overline{\beta}[\overline{\delta^P} + \gamma\overline{\phi}(1 - \zeta)]}\right). \tag{12}$$

Note that the equilibrium averages are function of the control strategy  $(\phi_1, \gamma)$ , since the equilibrium trait distribution  $p(x)$  depends on such parameters. Therefore, equation (12) provides all the information about the complex relationship between resource production and control strategy, and it allows the systematic exploration of the whole parameter space. Note that the value of resource at the endemic equilibrium  $R$  is independent of  $\zeta$ . Therefore, a control strategy with such a quantity as the objective will have the same outcome regardless of the fate of the treated parasite. In the following, we will focus on a particular set of parameters as an example of the relevant behavior of the system.

In figure 4, we plot the equilibrium resource  $R$  (rescaled with respect to the parasite-free resource  $R_0$ ) as a function of application rate  $\gamma$  and compare the effect of five different types of treatment:  $\phi_1 = -2, -1, 0, 1, \text{ and } 2$ . We present two sets of parameters corresponding to the two opposite states of the parasite trait distribution in the absence of treatment: the left panel corresponds to the red point in figure 2A, which is a low-exploitation state; the right panel corresponds to the blue point in figure 2A, which is a high-exploitation state. Note that employing a treatment type that is inconsistent with the state of the trait distribution in the absence of control (e.g.,  $\phi_1 = 2$  for the left panel,  $\phi_1 = -2$  for the right panel) leads to a small increase in resource as the dose increases (green and blue curves, respectively). Instead, employing a type extremely specialized in the trait that dominates in the absence of control ( $\phi_1 = -2$  for the left panel,

$\phi_1 = 2$  for the right panel) leads to a more significant increase in resource, at least for low application rates (blue and green curves, respectively). However, as increasing the rate of application tends to push the system toward the opposite state, extremely specialized treatments can quickly become less effective, and the resource gain will eventually saturate. At this point, switching to a more moderate type (smaller  $\phi_1$ ) rather than further increasing the application rate  $\gamma$  will provide more resource gain. Depending on the value of the renewal rate  $\theta$ , the system may eventually reach the parasite-free equilibrium (where  $\widehat{R} \equiv \widehat{R}_0$ ). Overall, when comparing the two panels, we find that the outcome of the control depends on the state of the parasite population in the absence of treatment, so that very efficient treatments in one case may be very inefficient in the other. We also find that an increase in  $\gamma$ , regardless of the choice, always corresponds to an increase in the resource. Therefore, maximizing the resource and minimizing the treatment application are conflicting objectives. Nevertheless, it is possible to identify efficient strategies  $(\phi_1, \gamma)$ , as explained below.

*Pareto-Efficient Strategies*

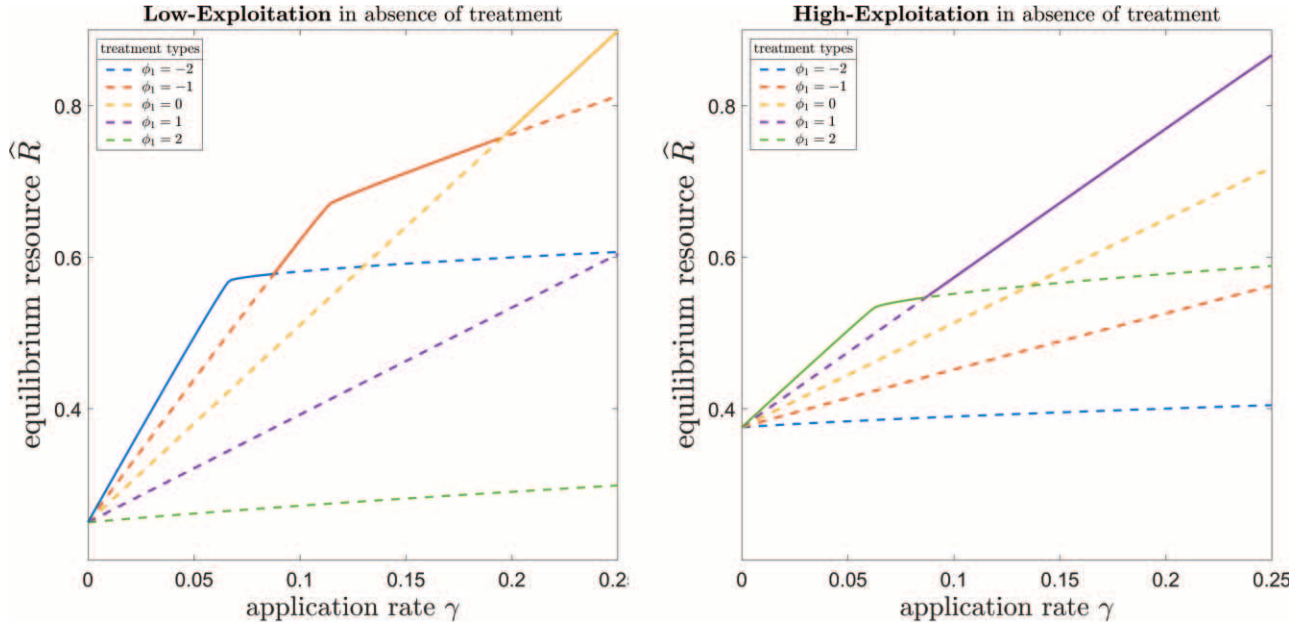
Figure 4 shows that resource maximization and treatment application minimization are conflicting objectives. In the presence of conflicting objectives, multicriteria analysis highlights the best compromises in the form of Pareto-efficient solutions (Kennedy et al. 2008). Among all of the possible choices of our control strategy  $(\phi_1, \gamma)$ , the Pareto-efficient solutions are those for which it is not possible to improve one objective without worsening the other. As such, they provide the decision maker with a smaller set of privileged alternatives to choose from, depending on the different management scenarios and on the decision maker's priorities.

The resulting Pareto-efficient solutions to our control strategy are identified by the solid curves in figure 4: choosing a control strategy  $(\phi_1, \gamma)$  different from the Pareto-efficient ones will inevitably worsen the outcome (dashed curves) either by reducing the amount of resource or by increasing the costs associated with the treatment application.

We note that when moving along the same type  $\phi_1$ , the resource shows a decrease in the incremental gain for the following threshold application rate:

$$\gamma^{\text{th}}(\phi_1) = \frac{2(\beta_1\delta_0^p - \beta_0\delta_1^p)}{(2\beta_0 + \beta_1)\phi_1 - 2\beta_1}. \tag{13}$$

In the presence of a constraint on the application rate, the problem collapses to a unique objective function, which is optimized by (mathematical details in the supplemental PDF)



**Figure 4:** Equilibrium resource as function of control strategy. Shown is the resource at equilibrium  $R$  obtained using five different treatment types, as a function of application rate  $\gamma$  (the  $y$ -axis is normalized with respect to the disease-free resource  $R_0$ ). *Left*, in the absence of treatment, the system is in the low-exploitation state; parameters correspond to the red point of figure 2. *Right*, in the absence of treatment, the system is in the high-exploitation state; parameters correspond to the blue point of figure 2. Very specialized types ( $\phi_1 = \pm 2$ ) are efficient for low  $\gamma$ , but their correspondent gain in resource saturates as the application rate is increased. Therefore, if the application rate can be increased, more generalist treatment should be privileged. The Pareto-efficient strategies of the control strategy are highlighted with solid curves. Other parameters:  $\mu = 7 \times 10^{-5}$ ,  $\theta = 150$ ,  $\delta^R = 0.04$ .

$$\phi_1^{\text{opt}}(\gamma) = \frac{2}{2\beta_0 + \beta_1} \left( \frac{\beta_1 \delta_0^p - \beta_0 \delta_1^p}{\gamma} + \beta_1 \right). \quad (14)$$

With respect to figure 2, the equation given above corresponds to the curves separating the two states in the  $(\phi_1, \gamma)$  phase diagrams, for which  $\Omega = 0$ . The effect of each parameter on  $\phi_1^{\text{opt}}(\gamma)$  is summarized in table 2. In particular, we note that the optimal degree of specialization is a decreasing function of application rate  $\gamma$ , so that extremely specialized treatments types will perform better at low  $\gamma$ ; the optimal choice is independent of biomass conversion  $\epsilon$ ; and exploitation and mortality rates (both baseline and trait-dependent contributions) play a nontrivial role in shaping the optimal choice.

Overall, equations (13) and (14) provide qualitative insight into the role played by each ecological interaction in shaping the control strategy behavior.

### Discussion

We have developed a mathematical model to explore the implications of possible correlations between treatment efficacy and key traits of the parasite. We have considered a general parasite that may express continuous levels of exploitation and mortality (eqq. [5], [6]) and a treatment

that exerts an eradicator action, which may be either positively or negatively correlated with the levels given above (eq. [7]; fig. 1), depending on the type of treatment. As a result of eco-evolutionary feedback, the parasite population can adapt toward evolutionary states dominated by either high or low exploitation levels (fig. 2), and the final resource will depend on the control strategy employed (fig. 4). The transition between these two possible states triggers several implications, depending on the management scenarios, which we discuss below.

*Scenario 1: both the application rate and treatment type are freely tunable.* In this case, the efficient strategies are represented by a Pareto front (solid lines in fig. 4). The Pareto front does not identify a single best strategy. Rather, it highlights a collection of best compromises between resource production and treatment use: whether to favor economic, environmental, or ethical objectives will therefore depend on the priorities of the decision maker, as well as on how the resource and the application rate will map into a real cost benefit.

*Scenario 2: the type of treatment may be constrained by the unavailability of alternatives or the inability to play with synergistic effects.* In this case, there will be a threshold  $\gamma^{\text{th}}$ , above which the decision maker should begin to question the benefit of further increasing the application rate.

**Table 2:** Effect of the parameters on the Pareto-efficient strategies

| Parameter    | Effect on $\phi_1^{\text{opt}}(\gamma)$  | Effect on $\gamma^{\text{th}}(\phi_1)$ |
|--------------|--|--|
| $\gamma$     | $\uparrow$ if $(\beta_0\delta_1^p - \beta_1\delta_0^p) < 0$ and $\downarrow$ otherwise | ...                                    |
| $ \phi_1 $   | ...  | $\downarrow$ always                    |
| $\beta_0$    | $\uparrow$ if $(\delta_0^p + \delta_1^p/2 - \gamma) > 0$ and $\downarrow$ otherwise    | $\downarrow$ always                    |
| $\beta_1$    | $\uparrow$ if $(\delta_0^p + \delta_1^p/2 - \gamma) < 0$ and $\downarrow$ otherwise    | $\downarrow$ always                    |
| $\delta_0^p$ | $\uparrow$ always  | $\uparrow$ always                      |
| $\delta_1^p$ | $\downarrow$ always  | $\downarrow$ always                    |

Note: The symbol  $\uparrow$  ( $\downarrow$ ) indicates that an increase in the parameter in question leads to an increase (decrease) of  $\phi_1^{\text{opt}}(\gamma)$  and  $\gamma^{\text{th}}(\phi_1)$ .

The economic impact of these saturation effects can be further assessed by including such information in economic evaluations of agricultural systems (Paveley et al. 2001; Ney et al. 2013; Day et al. 2021). Note that these saturation effects differ from those typically reported in the literature with treatment dose-response curves (Elderfield et al. 2018), which are accounted for by using nonlinear, saturating functions. Here, saturation is due to a transition between opposite evolutionary states. This kind of saturation is then dynamic, and it is inherent to the eco-evolutionary nature of the system.

*Scenario 3: the application rate may be constrained by limits on the use of antibiotics for safety reasons or by limits on the spread of pesticides/copper/fungicides for legislative agricultural constraints.* In this case, there will be a unique optimal type  $\phi_1^{\text{opt}}$ . As a general rule, very specialized treatment types should be employed at low application rates; instead, generalist types, with a more uniform action over the trait space, are likely to perform better at high application rates. This value represents the optimal choice from an ideal continuous spectrum of possibilities. We do not expect this spectrum to be fully available or even possible to design in practice. Nevertheless, it should provide qualitative guidance to the decision maker when calibrating synergistic treatments.

These scenarios assume that the control parameters  $\phi_1$  and  $\gamma$  are independent, which may not always be the case. However, a possible relation between application rate and treatment could be considered if the function  $\gamma(\phi_1)$  is known. Similarly, treatment application may be related to aspects of the host-parasite system (e.g., virulence or severity of the symptoms; Porco et al. 2005). In such cases, it may be possible to derive optimal treatment conditions in terms of the epidemiological parameters of the parasite, provided that the corresponding functions are known.

Overall, our results can contribute to the ubiquitous call to reduce the use of chemicals in public health (WHO 2014) and agriculture (Medina-Pastor and Triacchini 2020). Specifically, they point toward many of the European Union's principles (P) for sustainable

farm management (Barzman et al. 2015): valuable synergistic use of alternative control methods (P4), minimizing environmental impact (P5), reducing the use of chemicals (P6), and antiresistance strategies (P7). A concrete and urgent issue is the widespread use of copper in agriculture to combat plant diseases (Nunes et al. 2016). Because of its questionable efficacy and toxic side effects, there is an urgent call for its reduction (Tamm et al. 2018). In bacteria, resistance to heavy metals such as copper is mediated by efflux pumps, which are also involved in virulence to plants (Ryan et al. 2007; Martinez et al. 2009). Therefore, our theoretical framework could be used to support experimental studies of phage-copper synergy, which remain to be tested.

The assumption of a broad spectrum of treatment effects was motivated by the growing interest in developing therapies targeting specific traits and the possibility of combining them synergistically with traditional synthetic treatments (Lima et al. 2006; Allen et al. 2014; Baym et al. 2016). So far, we have referred to collateral sensitivity, phage therapy, and biocontrol as examples of control strategies that interfere with heterogeneous traits involved in both resistance and resource exploitation. We argue that the phenomenon may be of interest to other treatments based on heterogeneity and evolutionary constraints, such as antibiofilm, photodynamic activation, and more, on phage therapy.

Antibiofilm and photodynamic activation rely on a planktonic-versus-sessile evolutionary trade-off (Almeida et al. 2014; Tits et al. 2020; Feng et al. 2021): some bacteria can either live and move individually (planktonic phase) or can aggregate together into immobile structures called biofilms (sessile phase). Targeting such structures is a desirable strategy because they are involved in virulence and resistance to treatment. However, the efficacy of an antibiofilm treatment would depend on the trait composition of the target bacterial population, which in turn depends on the trade-off between the two phases, similar to the phenomenon considered in this work.

A promising application of phage therapy is the use of phages that have bacterial capsules as receptors. Capsules

are external polysaccharide layers that protect bacteria and facilitate attachment to host tissue. Capsules are therefore involved in host colonization and in evasion of the immune response or treatment. Phage selection for reduced capsule production will impose trade-offs between virulence and antibiotic sensitivity (Chiarelli et al. 2020; Song et al. 2021).

A proper experimental investigation and measurement of specific trait-dependent effects would likely require two stages. The first would be a single-cell stage to detect heterogeneity in the level of expression of the trait of interest and to measure the resulting trait-dependent interactions with the different treatments (Fernandes et al. 2011); this task could be performed using *in vitro* setups, such as microscopy, flow cytometry, or RNA sequencing (Avraham et al. 2015; Mohiuddin et al. 2020). The second would be a population stage where the overall effects on the demography and on the trait distributions can be monitored and measured; this task could be performed in controlled *in vivo* environments, such as bioreactors (Levin and Udekwu 2010), greenhouses, field plots, and animal facilities (Band et al. 2016).

Our analysis has assumed a linear, monotonic, and unimodal dose-response curve for treatment efficacy so as to keep the model as simple as possible and to favor mathematical tractability. In the supplement PDF, we show that the addition of nonlinearities does not affect the qualitative behavior of the system as long as an evolutionary constraint is considered. Therefore, we conjecture that the emergence of optimal and threshold behaviors are inherent to the system and that they are due to evolutionary constraints, rather than to its exact functional form. From a theoretical perspective, it would be interesting to prove this conjecture formally, for any kind of evolutionary constraint imposed on the treatment.

Instead, we expect that breaking the unimodal assumption will have nontrivial implications. On the one hand, a bimodal efficacy function would likely trigger multiple peaks in the trait distributions, therefore leading to possible branching phenomena where the parasite population splits into two separate subclasses of, for example, mid-low and mid-high levels of exploitation; however, we are currently unaware of any evidence for bimodal treatment efficacy, and it would need to be motivated by (at least) phenomenological arguments. On the other hand, it is reasonable to imagine a treatment that would have maximal efficacy at an intermediate value of the proxy trait (rather than at the extremes), which would likely lead to nonmonotonous trait distributions; accordingly, the variance of this putative bell-shaped efficacy could function as a tunable control parameter, governing possible intermediate scenarios between the homogeneous (rather flat, large variance) and the heterogeneous (rather peaked, small variance) extremes.

Although nonlinear, concave-down functions are typically considered to relate exploitation and mortality, a sim-

pler linear choice has allowed us to take full advantage of the mathematical analysis while preserving the possibility to model exploitation-mortality trade-offs; it also provides the baseline results against which to compare nonlinear functions, thus disentangling the role of nonlinearities from the role of the trade-off alone. The relaxation to nonlinear functions is discussed in the supplement PDF, where it is again shown that the qualitative behavior of the linear case is preserved.

Although our analytical derivation of the optimal treatment type relies on many simplifying assumptions, our work highlights the qualitative role of the various epidemiological interactions, and it provides a starting point for introducing further elements of complexity. To conclude our discussion, we highlight some potentially interesting issues.

Although multidimensional trait spaces are rarely considered, the simultaneous presence of multiple traits encoding different features of the parasite would improve the realism of the model. In particular, it would be interesting to explore the case where the treatment correlates with a subset of them, in order to mimic intervention policies with imperfect coverage (Walter and Lion 2021). For instance, one could consider a pathogen endowed with a trait defining its transmission capacity and a trait defining its disease-induced mortality; then one could imagine the existence of an intervention affecting the transmission trait (e.g., protectant effects of a pesticide, quarantine policy, vaccination campaigns) and an intervention affecting the mortality trait. Performing a similar analysis, one might be able to compute the optimal combination of the two actions and relate it to the geometry of the trait space (Miele et al. 2021), as well as to the possible evolutionary and economic constraints.

Our minimal model used simple demography for both resource and parasite dynamics. The introduction of more complex demographic functions (Cunniffe and Gilligan 2010) could lead to oscillating regimes around the endemic equilibrium. Such a maintained out-of-equilibrium demography results in a time-varying fitness landscape that could trigger out-of-equilibrium evolutionary responses, characterized by the alternation of low- and high-exploitation regimes of the trait distribution.

The expanding knowledge of the ecological, evolutionary, and molecular interactions between parasites and treatments, coupled with the theoretical feedback, should continue to provide opportunities to effectively address the challenge of disease management. Combining the practical development of trait-specific treatments with our theoretical methods of investigation (Saubin et al. 2023) may allow us to exploit heterogeneity of parasite populations—almost always seen as a key difficulty by allowing the evolution of resistance to human intervention—to

our advantage. Ultimately, evolutionary epidemiology is an instantiation of a more general theory of evolutionary ecology (Lion 2018). As such, the potential of the approach presented here can be exploited to investigate trait-dependent intervention in other domains, such as public health (Stearns 2012) and cancer dynamics (Gatenby et al. 2009).

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### Statement of Authorship

Conceptualization: L.M., R.M.L.E., D.B. Funding acquisition: R.M.L.E. Methods development: L.M. Model analysis: L.M. Coding simulation: L.M. Supervision: R.M.L.E., D.B. Writing—original draft: L.M. Writing—review and editing: L.M., R.M.L.E., N.J.C., C.T.-B., D.B.

### Data and Code Availability

Code is available from Zenodo (<https://zenodo.org/record/7874696#.ZFu4AnZBybh>; Miele et al. 2023).

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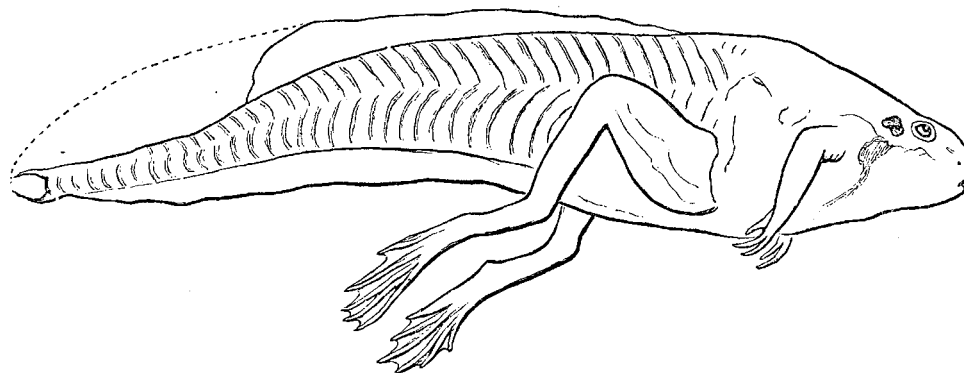
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“Then we should notice the beginning of segmentation, its progress, and the successful changes of form in the embryo, until it tears the shell, and with great, wondering eyes stares out upon its watery world a tadpole.” From “Pseudis, ‘The Paradoxical Frog,’” by S. W. Garman (*The American Naturalist*, 1877, 11:587–591).