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Climate anomalies and neighbourhood crowding interact

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Abstract

 1. Climate extremes and biotic interactions at the neighbourhood scale affect tropical forest dy- namics with long-term consequences for biodiversity, global carbon cycling and climate change mitigation. However, forest disturbance may change crowding intensity, and thus the relative con- tribution of climate extremes and neighbourhood interactions on tree growth, thereby influencing tropical forest resistance and resilience to climate change. Here, we aim to evaluate the sepa-³⁴ rate and interactive effects of climate and neighbours on tree growth in old-growth and disturbed tropical forests.

 2. We used 30 years of growth measurements for over 300 tropical tree species from 15 forest plots in French Guiana to investigate the separate and interactive effects of climate anomalies (in solar radiation, maximum temperature, vapor pressure deficit and climatic water deficit) and neighbourhood crowding on individual tree growth. Contrasting old-growth and selectively-logged forests, we also examined how disturbance history affects tree growth sensitivity to climate and neighbours. Finally, for the most abundant 100 species, we evaluated the role of 12 functional traits pertaining to water relations, light and carbon use in mediating tree growth sensitivity to climate anomalies, neighbourhood crowding, and their interactions.

 3. Climate anomalies tied to heat and drought stress and neighbourhood crowding independently reduced tree growth, and showed positive interactive effects which attenuated their separate effects on tree growth. Their separate and interactive effects were stronger in disturbed than undisturbed forests. Fast-growing species (i.e. higher intrinsic growth rates) were more abundant in disturbed forests and more sensitive to climate anomalies and neighbourhood crowding. Traits related to water relations, light and carbon use captured species sensitivities to different climate anomalies and neighbourhood crowding levels, but were weak predictors of their interactions.

 4. **Synthesis:** Our results demonstrate that climate anomalies and neighbourhood crowding can interact to shape tropical tree growth, suggesting that considering the biotic context may improve predictions of tropical forest dynamics facing altered climate regimes. Furthermore, species traits ₅₄ can capture tree growth sensitivity to the separate effects of climate and neighbours, suggesting that better representing leading functional dimensions in tropical tree strategies offers a promising way towards a better understanding of the underlying ecological mechanisms that govern tropical forest dynamics.

 Keywords: climate change, drought stress, functional traits, tropical forest, tree growth, water relations

Résumé

 1. Les extrêmes climatiques ainsi que les interactions biotiques à l'échelle du voisinage, affectent la dynamique des forêts tropicales, avec des conséquences à long terme pour la biodiversité, le cycle global du carbone et l'atténuation du changement climatique. Cependant, les perturbations forestières peuvent faire varier localement les voisinages, et ainsi modifier la contribution relative des extrêmes climatiques et des interactions de voisinage sur la croissance des arbres, ce qui peut impacter la résistance et la résilience des forêts tropicales au changement climatique. Nos travaux visent à évaluer les effets individuels et interactifs du climat et des voisins sur la croissance des arbres dans les forêts tropicales non-pertubées et perturbées.

 2. Nous avons utilisé 30 ans de mesures de croissance pour plus de 300 espèces d'arbres tropicaux provenant de 15 parcelles forestières en Guyane française pour étudier les effets individuels et interactifs des anomalies climatiques (en termes de rayonnement solaire, de température maximale, de déficit de pression de vapeur, et de déficit hydrique climatique) et des interactions de voisinage sur la croissance individuelle des arbres. En comparant les forêts exploitées sélectivement aux forêtsnon-pertubées, nous avons également examiné comment l'historique des perturbations peut influencer la sensibilité de la croissance des arbres au climat et aux voisins. Enfin, pour les 100 espèces les plus abondantes, nous avons évalué le rôle de 12 traits fonctionnels reflétant les relations hydriques, l'utilisation de la lumière et du carbone sur la réponse de la croissance des arbres aux anomalies climatiques, aux interactions de voisinage et à leurs interactions.

 3. Les anomalies climatiques liées aux stress thermique et hydrique, ainsi que les interactions de voisinage, réduisent la croissance des arbres. De plus, elles peuvent interagir positivement, ce qui atténue leurs effets individuels sur la croissance des arbres. Leurs effets individuels et interactifs sont plus forts dans les forêts perturbées que dans les forêts non-perturbées. Les espèces à croissance rapide (c'est-à-dire à taux de croissance intrinsèque plus élevé) sont plus abondantes dans les forêts perturbées, et aussi plus sensibles aux anomalies climatiques et aux interactions de voisinage. Les traits liées aux relations hydriques, à la lumière et à l'utilisation du carbone captent la sensibilité des espèces aux différentes anomalies climatiques et aux différents niveaux d'interactions de voisinage, mais ne prédisent pas leurs interactions.

 4. **Synthèse:** Nos résultats montrent que les anomalies climatiques et les interactions de voisinage peuvent interagir pour façonner la croissance des arbres tropicaux, ce qui suggère que prendre en compte le contexte biotique peut améliorer les prévisions de la dynamique des forêts tropicales face aux changements climatiques. En outre, les traits des espèces peuvent capter la sensibilité de la croissance des arbres aux effets individuels du climat et des voisins, ce qui suggère que mieux représenter les principales dimensions des stratégies fonctionnelles des arbres tropicaux ouvre sur une meilleure compréhension des mécanismes écologiques qui régissent la dynamique des forêts tropicales.

 Mots-clés : changement climatique, stress hydriqyue, traits fonctionnels, forêt tropicale, crois-sance des arbres, relations hydriques

Introduction

 Tropical forests are biodiversity hotspots (Gatti et al. 2022, Pillay et al. 2022) and major contrib- utors to global carbon and water cycles (Needham et al. 2018, Pan et al. 2011), but the services they provide to human populations are threatened by climate change and anthropogenic distur- bance (Brodribb et al. 2020, Malhi et al. 2008, Nunes et al. 2022, Vancutsem et al. 2021). In particular, extreme droughts are linked to higher solar radiation, maximum temperatures, atmo- spheric evaporative demands and lower soil water availability, which have been shown to negatively impact tropical forest dynamics by reducing tree growth and increasing tree mortality (Bauman et al. 2022b,a, Rifai et al. 2018). As climate change is predicted to result in higher frequencies, intensities, and duration of extreme droughts in the tropics (Shukla et al. 2022), this may have important long-term implications for tropical tree dynamics and diversity, and therefore global carbon sequestration and residence time (Sullivan et al. 2020).

 Tropical forest response to climate extremes is the compounded outcome of individual tree re- sponses, which may depend on local biotic interactions, such as interactions between neighbour- ing trees. While crowding by neighbours generally reduces tropical tree growth (Dawkins 1959, Gourlet-Fleury and Houllier 2000, Fortunel et al. 2018, Uriarte et al. 2004, 2010), the net effects of neighbourhood interactions may change from negative to positive under specific environmental conditions (Grossiord 2020, Fichtner et al. 2020) such as with water levels (Aakala et al. 2018) and disturbance history. For instance, competition for water may accentuate the negative effects of water limitations on tree growth, as shown in temperate and Mediterranean forests (Bottero et al. 2017, Gómez-Aparicio et al. 2011, Astigarraga et al. 2020). Conversely, species interactions can alter the depth of soil water uptake (Grossiord et al. 2014a) or transpiration response to increasing drought stress (Grossiord et al. 2014b, Haberstroh and Werner 2022). Yet, the effects of climate on tree growth have rarely been evaluated together with the effects of neighbourhood crowding in tropical forests (but see: Uriarte et al. 2016a), and we still know little about the extent to which the interaction between climate extremes and neighbourhood interactions can attenuate or exac-erbate their separate effects on tree growth, especially in tropical forests with different disturbance

histories.

 More than half of the world's tropical forests are designated production forests (Blaser et al. 2011), and forest management such as selective logging and thinning has long-lasting effects on forest composition, diversity and structure (Clark and Covey 2012, Cannon et al. 1994, Finegan and Camacho 1999, Mirabel et al. 2021, Osazuwa-Peters et al. 2015), which may alter forest resistance and resilience to extreme climate events (Hiltner et al. 2021). Specifically, disturbed forests host more light-demanding tree species than old-growth forests (Carreño-Rocabado et al. 2012, Baraloto et al. 2012, Kuusipalo et al. 1996, Mirabel et al. 2020, 2021, Whitmore 1989), and these fast-growing species may be more sensitive to climate extremes (Bauman et al. 2022a, Uriarte et al. 2016b) and neighbourhood crowding (Fortunel et al. 2016). Moreover, selective logging and thinning creates gaps, thereby increasing the openness of forest canopies and changing micro-climatic conditions (Ghuman and Lal 1987, Peña-Claros et al. 2008). These logging gaps influence tree growth rates (Baraloto et al. 2005, Hérault et al. 2010, Peña-Claros et al. 2008, Silva et al. 1995, Finegan et al. 1999), but it remains unclear how disturbance history may influence tree growth responses to climate and neighbours.

 Species differences in growth sensitivity to climate and neighbours likely depend on their functional strategies. While commonly measured traits pertaining to light capture, nutrient and carbon use such as specific leaf area and wood density can capture species differences in tree growth response to neighbourhood crowding (Fortunel et al. 2016, Kunstler et al. 2016), they are not always good proxies of species sensitivity to drought (Anderegg et al. 2018, Powell et al. 2017, Powers et al. 2020, Wagner et al. 2014b, but see: Uriarte et al. 2016a, Greenwood et al. 2017, Serra-Maluquer et al. 2022). A promising way forward is to integrate leaf traits that directly link to plant carbon- water relations or physiological responses to heat and water stress (Anderegg et al. 2018, Barros et al. 2019, McGregor et al. 2021, Powell et al. 2017, Powers et al. 2020, Rowland et al. 2021, Serra-Maluquer et al. 2022) (see: Table 1). For instance, stomatal density (Drake et al. 2013, Machado et al. 2021) and oxygen and carbon isotope composition (Farquhar et al. 1989, 2007, Moreno-Gutiérrez et al. 2012) can determine trade-offs between carbon assimilation and water loss at the leaf level under water-limited conditions (Guerrieri et al. 2019, Prieto et al. 2018, Scheidegger et al. 2000, Wang and Wen 2022b). Leaf water potential at turgor loss point is linked to species abilities to maintain key plant functions and growth during drought (Bartlett et al. 2012b, 2016, Maréchaux et al. 2018). Leaf minimum conductance (Blackman et al. 2019, Duursma et al. 2019, Machado et al. 2021) and leaf saturated water content (Blackman et al. 2019, Luo et al. 2021) may indicate species ability to resist or avoid heat- and drought-induced desiccation of plant tissue, and hence the potential for rapid post-drought growth. These leaf traits likely influence competition for water, but no study so far has evaluated their role in mediating tree growth response in the context of neighbourhood effects.

 Here, we tested the separate and interactive effects of climate and neighbours by leveraging 30 years of individual tree diameter growth data for more than 300 tropical tree species from 15 tropical forest plots, including six old-growth forest plots and nine post-logging forest plots, located at the CIRAD (Centre de coopération Internationale en Recherche Agronomique pour le Développement) experimental site of Paracou in French Guiana (Gourlet-Fleury et al. 2004). For the most abundant species (ca. 100), we collected 12 traits pertaining to water relations, light capture and carbon use (Table 1) to examine their role in mediating tree growth response to climate anomalies and neighbourhood crowding. Specifically we asked:

 1) How do climate anomalies and neighbourhood crowding influence tree growth? We expect tree growth to decline with positive anomalies in maximum temperature (Tmax), atmospheric evaporative demand (i.e. vapor pressure deficit, VPD) and climatic water deficit (CWD). As higher levels of solar radiation (SRad) lead to increased light availability, but also higher temperature and vapour pressure deficit, positive anomalies in SRad could either enhance or reduce tree growth. We further expect tree growth to decrease with higher neighbourhood crowding. Finally, we expect that climate anomalies and neighbourhood crowding can interact, thereby accentuating or attenuating their separate effects on tree growth.

 2) Does tree growth response to climate and neighbours vary with forest disturbance history? We expect stronger climate-induced growth declines in disturbed forests as a more open canopy offers less buffer against climate anomalies. We also expect stronger sensitivity to climate anomalies and neighbourhood crowding in disturbed forests as they host more fast-growing species that may be more vulnerable to both abiotic and biotic stressors.

 3) Do leaf water-related traits and commonly-measured leaf and stem traits pertaining to light and carbon use capture species differences in tree growth response to climate anomalies and neigh- bourhood crowding? We expect greater tree growth sensitivity to heat and drought stress and neighbourhood crowding in fast-growing, light-demanding, water-spender or drought intolerant species.

Materials and Methods

Study site

 Our analyses used the permanent forest plot system of the CIRAD experimental site at the Paracou Tropical Forest Research Station (5°18'N, 52°53'W) in French Guiana. Mean annual precipitation at Paracou is 3041 mm yr*−*¹ , with a pronounced dry season (*<* 100 mm month*−*¹) from August through November. Paracou is characterised as a tropical lowland forest site consisting of old- growth terra firme forest. Fifteen permanent forest plots of 6*.*25 ha each were established from 1984 to 1990, of which nine plots were subjected to different types (selective logging and thinning) and intensities of sylvicultural treatments (Fig. 1), resulting in above-ground biomass (AGB) loss (12–33% AGB loss in the three plots subjected to selective logging and 33-56% AGB loss in the six plots that were subjected to selective logging and thinning) (Gourlet-Fleury et al. 2004, Hérault and Piponiot 2018) and differences in species composition (Mirabel et al. 2020, 2021). In the 2021 inventory, there were 70,238 trees belonging to 591 species (including subspecies), 227 genera and 63 families.

Tree growth and neighbourhood data

202 In each of the 15 forest plots, all trees $\geq 10 \text{ cm}$ DBH (diameter at breast height, i.e. 1.3 m) were mapped to a precision of 0*.*5 m and were botanically identified. Tree inventories took place at

Figure 1. Spatial dimension of the 15 permanent forest plots and stand density in $10 \text{ m} \times 10 \text{ m}$ quadrats, given as basal area $(m^2 \text{ ha}^{-1})$, in a) 1991 and b) 2019. Disturbed plots are denoted by black contours. Box plots show differences in basal area between the 6 undisturbed (U) and the 9 disturbed (D) forest plots in c) 1991 and d) 2019. t value of Welch's two sample t-test, performed on log transformed basal area, and significance level (*** = $p<0.001$) are given in the left hand corner of the box plots.

 one or two year intervals during which the status (alive/dead) and circumference to a precision of 0*.*5 cm of every tree were recorded, from which we calculated DBH. New recruits (trees reaching the threshold DBH of 10 cm) were mapped and identified at each census (Gourlet-Fleury et al. 2004, Derroire et al. 2022a,b,c,d,e). To reduce noise in annual growth data (potentially caused by measurement/rounding errors), for each individual *i* of species *s*, we calculated the annualised absolute diameter growth rate (AGR, $cm yr^{-1}$) from DBH values between censuses *t* and $t-2$ as:

$$
AGR_{i,s,t} = \frac{DBH_{i,s,t} - DBH_{i,s,t-2}}{t - (t-2)}
$$
 (eqn 1)

 For each individual tree *i*, we thus calculated AGR for 15 two-year census intervals between 1991 and 2021. We excluded the following stems from the focal tree data: (a) Palm species due to their lack of secondary growth, (b) individuals with human-induced damage (e.g. through logging or thinning treatments), (c) missing or uncertain botanical determination, (d) with estimated circumference (e.g. due to buttresses), and (e) abnormal tree growth values (*≤ −*2 cm yr*−*¹ , resulting f_{215} from a shift in the point of measurement, and $\geq 5 \,\mathrm{cm \,yr^{-1}}$. The upper and lower thresholds for abnormal growth values were based on expert knowledge for Paracou site.

 $_{217}$ The growth of an individual tree *i* can be influenced by its $N(i)$ neighbours within a given radius ²¹⁸ via their size (DBH²) and the inverse of their distance to the focal tree (*d*) (Canham et al. 2004, Fortunel et al. 2016, Uriarte et al. 2010). For each focal tree individual *i* at the start of the growth 220 census interval $t - 2$, we calculated its neighbourhood crowding index (NCI_{i,t^{*-*}}) as the weighted $_{221}$ sum of all neighbours $N(i)$ within a 10 m radius of individual *i* (Fortunel et al. 2018, Lasky et al. 2014, Uriarte et al. 2016a), as:

$$
NCI_{i,t-2} = \sum_{j \in N(i)} \frac{DBH_{j,t-2}^2}{d_{i,j}}
$$
 (eqn 2)

 We chose a 10 m radius to be able to compare with previous studies conducted in tropical forests (Fortunel et al. 2018, Lasky et al. 2014, Uriarte et al. 2016a), and well capture neighbourhood effects while keeping edge effects low (Zambrano et al. 2020). To avoid incomplete neighbourhoods, all focal tree individuals closer than 10 m to the plot edge were excluded. We kept as neighbours individuals with human-induced damage, missing or uncertain botanical determination, estimated circumferences, and corrected abnormal diameter increments (see: Supplementary Methods S1 for details).

Climate data

 We considered four climate variables that have been previously shown to influence tropical tree growth: solar radiation (SRad) representing light availability, maximum temperature (Tmax) cap turing heat stress, vapour pressure deficit (VPD) relating to atmospheric evaporative demands and climatic water deficit (CWD) capturing soil water availability by relating precipitation to evapo- transpiration (Bauman et al. 2022a, Choat et al. 2018, Cunningham 2005, Doughty and Goulden 2008, Grossiord et al. 2020, Wagner et al. 2012). We extracted climate variables from TerraCli- mate, a high-resolution global data set of monthly climate variables and climatic water balance (Abatzoglou et al. 2018).

²³⁹ We identified anomalous years over the study period by calculating mean climate anomalies, ex-²⁴⁰ pressed as the deviation of climate variables from their 30-year historical average, over each of ²⁴¹ the two-year census intervals. To this aim, we calculated the 30-year baseline (historical mean, ²⁴² $\mu_{m,\text{baseline}}$ and standard deviation ($\sigma_{m,\text{baseline}}$) for each month of the year (1-12) for the 1991-2021 ²⁴³ period. We then calculated standardised anomalies for each month (1-12) *m* at time *t* as the deviation of the mean monthly climate variable $\mu_{m,t}$ from the long-term monthly baseline $\mu_{m,\text{baseline}}$ ²⁴⁵ (Bauman et al. 2022a, Rifai et al. 2018). To investigate the effect of mean yearly anomalies on ₂₄₆ tree growth at two-year census interval (CA_t) , we averaged the standardised monthly anomalies over the $M = 24$ months prior to census t (July_t-July_{t-2}) as (Aubry-Kientz et al. 2015, Bauman ²⁴⁸ et al. 2022a, Rifai et al. 2018):

$$
CA_t = \frac{1}{M} \sum_{m=1}^{M} \frac{\mu_{m,t} - \mu_{m,\text{baseline}}}{\sigma_{m,\text{baseline}}}
$$
 (eqn 3)

²⁴⁹ Using this approach, two-year periods with positive anomalies can directly be interpreted as periods ²⁵⁰ in which trees experienced higher solar radiation (SRad), heat (Tmax), atmospheric (VPD) or soil ²⁵¹ water (CWD) drought stress than usual.

²⁵² **Functional trait data**

²⁵³ We determined six leaf traits pertaining to water relations (Table 1): stomatal density (SD), ²⁵⁴ oxygen (δ^{18} O) and carbon (δ^{13} C) stable isotope composition, water potential at turgor loss point $(1.50)(\pi_{\text{tlp}})$, minimum conductance (g_{min}) and saturated water content (LSWC). We sampled 645 trees ²⁵⁶ belonging to 102 of the most abundant species (from 76 genera and 34 families) that together

Census Years

Figure **2.** Mean standardised climate anomalies at Paracou for the two-year census intervals over the study period for a) solar radiation, b) maximum temperature, c) vapour pressure deficit and d) climatic water deficit. Red and blue lines and circles represent positive and negative anomalies respectively.

 represent 70% of all unique individual stems and 72% of all growth measurements at Paracou. We randomly-selected 645 individuals located within the permanent 15 forest plots at Paracou, using the following criteria: (i) botanical identification at the species level, (ii) DBH within species- specific 10th and 90th percentile (to avoid sampling unusually small and large tree individuals), (iii) topographic position corresponding with the species topographic preference (i.e. individuals located at a certain topographic position (bottomland, slope, ridge) were only sampled if more than 10% of distinct stems of that species occur on the given topographic position). As the 102 target species include understory, mid-story and canopy species, we sampled only leaves from branches that were exposed to medium light conditions (Dawkins index: 3-4, Alder and Sunnott 1992), thus excluding branches in full sun or shade. For each of the 645 individuals, we collected one fully expanded and healthy leaf from a healthy branch in the dry season in 2020-2021 for each leaf trait measurement. After harvest with a sling shot in the field, leaves were immediately put into ziplock bags with a moist paper towel and kept in the dark in a cooler to minimize transpiration during transport to the lab. No fieldwork permit was required.

To measure leaf SD, we used nail polish or silicon-based dental paste (Speedex Universal Silicon

 Activator and Speedex Light Body, Coltène/Whaledent AG, Switzerland) to take imprints from the abaxial leaf surface at three positions of the widest part of the leaf blade, while avoiding major veins (Voleníková and Tichá 2001). Imprints were mounted on microscope slides, and pictures were taken at a magnification of 300x using a light microscope (Keyence VHX 7000, Osaka, Japan). We took a minimum of 3 pictures, 1 per imprint when possible, for each sampled leaf. We used Fiji ImageJ (Schindelin et al. 2012) to count all complete stomata per image area. Imprint quality and resulting picture quality varied strongly between species and individuals, so we obtained stomata counts for 470 individuals and 87 species.

 Leaf osmotic potential was measured using a vapor pressure osmometer (VAPRO 5520, Wescor, $_{281}$ Logan, UT, USA). Leaf water potential at turgor loss point (π_{tlp}) was linearly estimated from osmotic potential at full hydration, following the standard protocol of Bartlett et al. (2012b,a), which has been validated for species in French Guiana (Maréchaux et al. 2016).

 Leaf g_{min} was estimated from leaf drying curves. We monitored leaf weight loss over time, following the protocol of Sack et al. (2003). Time intervals for weight measurements were set to 30 min. Leaves of tropical species loose water very rapidly, so that consecutive measurements over 3 to 4 h were sufficient to assess minimum conductance.

 We measured leaf saturated and dry weights with a digital scale at a 0*.*0001 g precision (Mettler Toledo, Switzerland). We obtained saturated weights following the 'partial' rehydration protocol of Vendramini et al. (2002), which provides comparable measurements to the 'complete' rehydration protocol by Garnier et al. (2001) (Vaieretti et al. 2007). We wrapped fresh leaves into paper towels that were previously soaked with de-ionised water, and kept them for 24 hours in the dark and at ²⁹³ low temperature (4 °C) for rehydration. Dry weights were measured after drying leaves for at least ²⁹⁴ 72 hours at 60 °C (Sapes and Sala 2021). We calculated LSWC from saturated and dry weights (Barrs and Weatherley 1962).

296 Bulk leaf δ^{18} O and δ^{13} C were calculated from a composite sample of dried and coarsely chopped 297 leaves, including those used for π_{tlp} , LSWC and g_{min} measurements. Samples were ground to fine powder using a ball mill (Retsch MM200, Hanau, Germany) for five minutes at maximum

speed (100 rotations per minute) and further dried in a drying oven for 24 h at 85 °C. Aliquots of $0.2 - 0.4$ mg per sample were weighed into silver capsules and δ^{18} O signatures were measured with ³⁰¹ a high temperature pyrolysis unit (TCEA) coupled to a Delta V isotope ratio mass spectrometer instrument (Thermo Scientific, Vienna, Austria). Aliquots of 1 *−* 2 mg per sample were weighed $_{303}$ into tin capsules and δ^{13} C signatures were subsequently determined using an elemental analyser (Flash EA), linked by a Conflo IV Universal Interface to a Delta V isotope ratio mass spectrometer.

 We combined our measurements for the six leaf traits with data from previous field campaigns conducted at Paracou (Levionnois et al. 2021, Maréchaux et al. 2015, 2019, Ziegler et al. 2019). We further compiled trait data for the 102 focal tree species for six leaf and wood traits that relate to light capture and carbon use from previous work conducted in French Guiana (Baraloto et al. 2010, Fortunel et al. 2012, Vleminckx et al. 2021). We included leaf area (LA), specific leaf area (SLA), leaf thickness (L_{thick}), leaf toughness (L_{though}), bark thickness (Bark_{thick}) and wood specific gravity (WSG, hereafter referred to as wood density) (see: Table 1). For all subsequent analyses including trait values, we considered species means.

Models

 We used Bayesian hierarchical models to evaluate the separate and interactive effects of climate anomalies and neighbourhood crowding on individual absolute growth rates (AGR). We modeled the natural logarithm of absolute growth rates (log(AGR)) because of the high heteroscedasticity within the growth data set and to reduce the influence of outliers (Bauman et al. 2022a, Fortunel et al. 2018, Hérault et al. 2011). To deal with negative and zero growth before taking the logarithm, 319 we added to each observed growth value the constant $1.1 |AGR_{min}|$, where $|AGR_{min}|$ is the absolute minimum value of observed growth of the dataset (Bauman et al. 2022a).

 The hierarchies of the models consist of modelling individual-level growth rates to simultaneously fit a community-level regression and a species-level regression. The community level modelled AGR responses to covariates via hyperparameters (i.e. statistical distributions from which species-level intercepts and slope coefficients arose), while the species level captured species-specific AGR sensi- tivities to covariates (i.e. slope parameters) and species-level intercepts (hereafter intrinsic AGR) - i.e. species-level deviations from the community average parameters. While the community-level regression (eqn 4a) was kept the same across all models, the species-level regressions (capturing species-specific differences in intrinsic growth rates and growth responses to model covariates) only depend on species identity in M1-models (eqn 4c), while species parameters are a function of 330 species traits in M2-models (eqn 5). The models further include a covariance matrix (eqn 4e) to estimate correlations ρ among species-level parameters. To capture part of the unexplained growth 332 variation related to individuals and plots we allowed intercepts to vary by plots γ_p and individuals ϵ_i (eqn 4f) (Bauman et al. 2022a, Fortunel et al. 2018).

³³⁴ We modeled individual $\log(AGR_t)$ as a linear function of (i) tree size at the beginning of the census 335 interval (DBH_{t^{-2}), (ii) monthly climate anomalies averaged over the census interval (CA_t), (iii)} neighbourhood crowding at the beginning of the census interval (NCI*^t−*²) and (iv) the interaction between climate anomalies and neighbourhood crowding (CA*^t ×* NCI*^t−*²) (Bauman et al. 2022a, Fortunel et al. 2018). To allow direct comparison between covariates and ease assignment of plausible weakly-informative prior to the parameters, log(AGR*t*) and all covariates, except for climate anomalies (Bauman et al. 2022a), were standardised to mean zero and unit standard deviation (Bauman et al. 2022a, Fortunel et al. 2018, McElreath 2020, Uriarte et al. 2016a). As we 342 assumed tree growth to have a non-linear, multiplicative relationship with DBH and NCI (Canham et al. 2004, Fortunel et al. 2016, Uriarte et al. 2010), we log transformed DBH and NCI prior to standardisation to achieve an additive and linear relationship (Kunstler et al. 2016, Uriarte et al. 2016a, Fortunel et al. 2018). As species strongly varied in their mean tree size, we standardised DBH to mean zero and unit standard deviation within species, to prevent confounding species ³⁴⁷ differences in growth response to tree size with interspecific variation in mean DBH (Fortunel et al. 2018, Uriarte et al. 2016a).

M1: "no trait models"

 To manage model complexity in assessing tree growth response to the separate and interactive effects of climate anomalies and neighbourhood crowding and examine the influence of forest disturbance history, we fitted models separately for each of the four climate variables (SRad, Tmax, VPD and CWD) and for (i) all 15 plots combined, (ii) the six undisturbed plots and (iii) the nine disturbed plots, resulting in a total of 12 M1-models. To ensure sufficient statistical power, M1-models were run on observed growth for all species with at least 100 growth measurements. M1-models were therefore based on 48,203, 20,688 and 27,515 individuals, belonging to 310, 306 and 307 species, respectively. To allow for direct comparison of parameter estimates between control and disturbed forest, we standardised and centered growth and all model covariates before 359 splitting the data set (Ren et al. 2021). For each individual *i* of species *s* between censuses $t-2$ and *t*, we modeled tree growth as:

$$
\log(\text{AGR}_{i,s,t,p}) \sim \mathcal{N}(\mu_{i,s,t,p}, \sigma) \tag{eqn 4a}
$$

$$
\mu_{i,s,t,p} = \alpha_s + \beta_{1s} \times \log(\text{DBH}_{i,t-2}) + \beta_{2s} \times \text{CA}_t + \beta_{3s} \times \log(\text{NCI}_{i,t-2})
$$

+
$$
\beta_{4s} \times \text{CA}_t \times \log(\text{NCI}_{i,t-2}) + \gamma_p + \epsilon_i
$$
 (eqn 4b)

$$
\begin{pmatrix}\n\alpha_s \\
\beta_{1s} \\
\beta_{2s} \\
\beta_{3s} \\
\beta_{4s}\n\end{pmatrix}\n\sim MVNormal\n\begin{pmatrix}\n\alpha_0 \\
\beta_{1,0} \\
\beta_{2,0} \\
\beta_{3,0} \\
\beta_{4,0}\n\end{pmatrix}, S
$$
\n
$$
\begin{pmatrix}\n\alpha_0 \\
\beta_{1,0} \\
\beta_{3,0} \\
\beta_{4,0}\n\end{pmatrix}
$$
\n
$$
\begin{pmatrix}\n\sigma_{\alpha_s} & 0 & 0 & 0 & 0 \\
0 & 0 & 0 & 0 \\
0 & 0 & 0 & 0 & 0\n\end{pmatrix}
$$

363

362

$$
S = \begin{pmatrix} \sigma_{\alpha_s} & 0 & 0 & 0 & 0 \\ 0 & \sigma_{\beta_{1s}} & 0 & 0 & 0 \\ 0 & 0 & \sigma_{\beta_{2s}} & 0 & 0 \\ 0 & 0 & 0 & \sigma_{\beta_{3s}} & 0 \\ 0 & 0 & 0 & 0 & \sigma_{\beta_{4s}} \end{pmatrix} \times R \times \begin{pmatrix} \sigma_{\alpha_s} & 0 & 0 & 0 & 0 \\ 0 & \sigma_{\beta_{1s}} & 0 & 0 & 0 \\ 0 & 0 & \sigma_{\beta_{2s}} & 0 & 0 \\ 0 & 0 & 0 & \sigma_{\beta_{3s}} & 0 \\ 0 & 0 & 0 & 0 & \sigma_{\beta_{4s}} \end{pmatrix}
$$
(eqn 4d)

364

$$
R = \begin{pmatrix} 1 & \rho_{\alpha_s, \beta_{1s}} & \rho_{\alpha_s, \beta_{2s}} & \rho_{\alpha_s, \beta_{3s}} & \rho_{\alpha_s, \beta_{4s}} \\ \rho_{\alpha_s, \beta_{1s}} & 1 & \rho_{\beta_{1s}, \beta_{2s}} & \rho_{\beta_{1s}, \beta_{3s}} & \rho_{\beta_{1s}, \beta_{4s}} \\ \rho_{\alpha_s, \beta_{2s}} & \rho_{\beta_{1s}, \beta_{2s}} & 1 & \rho_{\beta_{2s}, \beta_{3s}} & \rho_{\beta_{2s}, \beta_{4s}} \\ \rho_{\alpha_s, \beta_{3s}} & \rho_{\beta_{1s}, \beta_{3s}} & \rho_{\beta_{2s}, \beta_{3s}} & 1 & \rho_{\beta_{3s}, \beta_{4s}} \\ \rho_{\alpha_s, \beta_{4s}} & \rho_{\beta_{1s}, \beta_{4s}} & \rho_{\beta_{2s}, \beta_{4s}} & \rho_{\beta_{3s}, \beta_{4s}} & 1 \end{pmatrix}
$$
 (eqn 4e)

³⁶⁵ using the following weakly informative priors:

$$
\gamma_p \sim \mathcal{N}(0, \sigma_\gamma)
$$

\n
$$
\epsilon_i \sim \mathcal{N}(0, \sigma_\epsilon)
$$

\n
$$
\alpha_0, \beta_{1-4,0} \sim \mathcal{N}(0, 0.5)
$$

\n
$$
\sigma_{\alpha_s}, \sigma_{\beta_{1-4s}}, \sigma_\gamma, \sigma_\epsilon, \sigma \sim \exp(1)
$$

\n
$$
R \sim LKJcorr(2)
$$

366 where α_s represents the species-level intrinsic growth and β_{1s} , β_{2s} , β_{3s} and β_{4s} represent species-level ³⁶⁷ growth responses to tree size (DBH), climate anomalies (CA), neighbourhood crowding (NCI) and 368 the interactive effect of climate anomalies and neighbourhood crowding (eqn 4a). Negative values ³⁶⁹ of *α^s* indicate species whose growth is lower and positive values indicate species whose growth is higher than the average AGR across all species. Negative values of *β*¹*−*3*^s* indicate decreases in growth with increasing model covariate. The interpretation of *β*4*^s* depends on the respective 372 directions of β_{2s} and β_{3s} . If the sign of β_{4s} is the same as the signs of β_{2s} and β_{3s} , the interaction between neighbourhood crowding and climate can accentuate their separate effects. Conversely, if β_{374} the sign of β_{4s} is opposite from both signs of β_{2s} and β_{3s} , the interaction between climate anomalies and neighbourhood crowding can attenuate their separate effects. The strength and implications of the interactions usually require a visual exploration to provide a full interpretation.

 Species intrinsic growth α_s and growth response to covariates β_{1-4s} for the *s* species were modeled 378 as in eqn 4c, where α_0 represents the community-level intrinsic growth rate and $\beta_{1-4,0}$ the overall effect of covariates on growth across all species. To assess correlations (*ρ*) among species-level in-380 trinsic growth α_s and growth response to model covariates β_{1-4s} , we fitted a matrix of correlation coefficients among all pairs of species-level parameters. In addition, modeling all species-level pa- rameters as a multivariate normal distribution (eqn 4c) allows to share information across species, thus improving the fit for poorly represented species, while preventing overfitting (Bauman et al. 2022a, McElreath 2020). The variance-covariance matrix *S* was constructed as shown in eqn 4d, 385 where R is the correlation matrix of species-level parameters (eqn 4e). Varying intercepts for the *p* plots γ_p and the *i* individuals ϵ_i were modeled as given in eqn 4f with a mean centered on 0 (i.e. mean intercept across all plots and individuals respectively). We specified weakly informative priors which are given in eqn 4f.

M2: "trait models"

390 To assess the effect of traits on species intrinsic growth α_s , and growth response to climate anoma- β_{2s} , neighbourhood crowding β_{3s} and their joint effects β_{4s} , we fitted M2-models for a subset of 81-102 tree species (representing over 40,000 trees), for which we had available trait data (Table 1). To manage model complexity, M2-models were fitted separately for each of the four climate variables (SRad, Tmax, VPD and CWD) and for each of the 12 functional traits, resulting in a total of 48 M2-models. While the hierarchical structure, variance-covariance matrix and equation 396 of the community level hierarchy of the M2-models is identical to M1-models (eqn $4a$ -eqn $4b$; eqn $397 \text{ }\text{4d-equ }4f$, all species-level parameters, except species-level growth response to tree size β_{1s} , were modeled in species-level regressions to capture species-specific differences, while simultaneously evaluating the effect of species-mean functional traits within these responses (eqn 5) (Fortunel et al. 2018, Uriarte et al. 2016a, Bauman et al. 2022a). More specifically, species-level parameters were modelled as:

$$
\begin{pmatrix}\n\alpha_s \\
\beta_{1s} \\
\beta_{2s} \\
\beta_{3s} \\
\beta_{4s}\n\end{pmatrix}\n\sim MVN\left[\n\begin{pmatrix}\n\alpha_0 + \alpha_1 \times Traits \\
\beta_{1,0} \\
\beta_{2,0} + \beta_{2,1} \times Traits \\
\beta_{3,0} + \beta_{3,1} \times Traits \\
\beta_{4,0} + \beta_{4,1} \times Traits\n\end{pmatrix}, S\n\right]
$$
\n(eqn 5)

402 where α_0 represents the community-level intrinsic growth (i.e. grand intercept) and $\beta_{2-4,0}$ the ⁴⁰³ community-level response to model covariates (i.e grand slopes) whose posterior distributions rep- α_4 resent the effect of covariates on AGR across all species. α_1 and $\beta_{2-4,1}$ represent the departure from ⁴⁰⁵ the community level AGR and community level response to model covariates with one standard ⁴⁰⁶ deviation change in the value of *T rait^s* of species *s*. Traits were standardised to mean zero and 407 unit standard deviation across species, thus implying that parameter β_{2-4} equals $\beta_{2-4,0}$ for the ⁴⁰⁸ mean trait value of the data set. Given their skewed distributions, we log transformed values of 409 stomatal density (SD), leaf saturated water content (LSWC), minimum conductance (g_{min}) , leaf 410 area (LA), leaf thickness (L_{thick}) , leaf toughness (L_{though}) and bark thickness $(Bark_{thick})$, prior to 411 standardisation, to prevent the tail of the distribution from dominating the model fit (Fortunel ⁴¹² et al. 2018). Species AGR responses to model covariates can thus be accentuated (i.e. the sign of ⁴¹³ the trait effect $\beta_{2-4,1}$ is the same as that of the corresponding community level response $\beta_{2-4,0}$) 414 or attenuated (i.e. the sign of the trait effect $\beta_{2-4,1}$ is the opposite of that of the corresponding 415 community level response $\beta_{2-4,0}$) with increasing values of *Trait_s* (Bauman et al. 2022a, Fortunel ⁴¹⁶ et al. 2018, Uriarte et al. 2016a).

Model fitting

 All models were fitted in R environment (R Core Team 2021, RStudio Team 2020), on the Meso@LR HPC cluster using the package brms (Bürkner 2017). Bayesian updating of param- eters was performed via the No-U-Turn Sampler (NUTS) in Stan (Carpenter et al. 2017) using CmdStanR (Stan Development Team 2022), which allows intra-chain parallelisation in the back- end. We used four chains and 4000 iterations (2000 warm up) per chain. Chains of all models mixed well and generally converged within 2000 iterations (Rhat between 1 and 1.05). Model parame- ter posteriors were summarised through their median and 90%-highest posterior density interval (HPDI) using the packages tidyverse (Wickham et al. 2019) and tidybayes (Kay 2022). Model covariates were considered to have a clear effect when their slope coefficients 90%-HPDIs did not encompass zero. To assess the model goodness of fit, we calculated conditional and marginal R2 ⁴²⁸ using the bayes $R2$ () function of the brms package (Bürkner 2017), which calculates a Bayesian $_{429}$ version of R2 for regression models (Gelman et al. 2019). M1- and M2-models had high explanatory power, with a mean conditional R2 of 0.62 and 0.64, respectively. For more detailed information on conditional and marginal R2 for each model fit see Table S1.

Results

Tree growth sensitivity to climate anomalies and neighbourhood crowd-ing

 Across the 15 plots, anomalies in all four climate variables influenced tree growth at Paracou (Fig. 3). Positive anomalies in solar radiation (SRad) enhanced tree growth, whereas positive anomalies in maximum temperature (Tmax), vapour pressure deficit (VPD) and climatic water deficit (CWD) reduced tree growth. Neighbourhood crowding reduced tree growth, and this effect was stronger than that of any of the climate anomalies. Moreover, neighbourhood crowding showed positive interactive effects with climate anomalies linked to heat and drought stress (Tmax, VPD and CWD) on tree growth, but negative interactive effects with anomalies in SRad on tree growth. More crowded trees suffered less from negative effects of Tmax, VPD and CWD anomalies, while ⁴⁴³ also benefiting less from the positive effects of SRad anomalies.

Figure **3.** Community-level effects of climate anomaly $(\beta_{2,0})$, neighbourhood crowding (NCI, $\beta_{3,0}$), and their interaction (Climate anomaly×NCI, *β*4*,*0) on tree growth (M1-models). Standardised coefficients from a) SRad, b) Tmax, c) VPD and d) CWD models are shown for all 15 plots combined, the 6 undisturbed plots and the 9 disturbed plots respectively. Circles show posterior medians of standardised coefficients (community level responses, see: *β*2*−*4*,*⁰ in eqn 4c), and lines indicate 90% highest posterior density intervals [HPDI]. Colored circles indicate clear negative (red) and positive (blue) effects (i.e. slope coefficient 90%-HPDIs not encompassing zero), while white circles indicate no effects. Positive $\beta_{2-3,0}$ values indicate faster growth, while negative $\beta_{2-3,0}$ values indicate slower growth with increasing model covariate (details in Table S2). The interpretation of *β*4*,*⁰ depends on the direction of the climate effect $\beta_{2,0}$. If the sign of $\beta_{4,0}$ is the same as the sign of $\beta_{2,0}$, neighbourhood crowding accentuates the effect of climate, inversely if the sign of β_{4s} is in the opposite of the sign of $\beta_{2,0}$, neighbourhood crowding attenuates the effect of climate. Community-level intercepts (α_0) and tree size effects (DBH, $\beta_{1,0}$) are shown in Fig. S1 A.

444 Tree growth sensitivity to climate anomalies (β_{2s}) and neighbourhood crowding (β_{3s}) varied among the focal 310 species: 39% showed clear responses to SRad anomalies (of which 98% showed positive and only 2% negative responses). Moreover, 45% responded to Tmax anomalies (of which 99% responded negatively and only 1% positively) and 21% responded to VPD anomalies (of which 86% responded negatively and 14% positively). Finally, 21% responded to anomalies in CWD (of which 94% responded negatively and 6% positively). On average across the four climate models, 88% of all species showed marked responses to NCI, all of them suppressed by higher crowding intensities.

 Overall, individual-level tree growth response to the interaction between climate anomalies and neighbourhood crowding was weakly related to species identity. Yet, species showed a consistent directional response (i.e. no species showed an opposite effect) to the interactive effects of neigh- bourhood crowding and anomalies (*β*4*s*) in SRad, Tmax and CWD, but less in VPD. Specifically, neighbourhood crowding clearly attenuated positive effects (negative *β*4*s*) of SRad anomalies in 457 18%, and negative effects (positive β_{4s}) of Tmax and CWD anomalies in 27% and 8% of all species. Neighbourhood crowding modulated negative effects of VPD anomalies in 7% of all species, either 459 attenuating (81% of positive β_{4s}) or reinforcing (19% of negative β_{4s}) these effects. For details see Table S3.

 ϵ_{461} Species-level intrinsic AGR (M1, α_s) were negatively associated with AGR response to anomalies μ_{462} (M1, β_{2s}), as shown by the median $\tilde{\rho}$, lower 5% quantile $q_5(\rho)$ and upper 5% quantile $q_{95}(\rho)$ of the 463 highest posterior density, in SRad ($\tilde{\rho} = -0.20$, $q_5(\rho) = -0.32$, $q_{95}(\rho) = -0.09$), Tmax ($\tilde{\rho} = -0.69$, $q_5(\rho) = -0.76$, $q_{95}(\rho) = -0.62$, VPD ($\tilde{\rho} = -0.59$, $q_5(\rho) = -0.66$, $q_{95}(\rho) = -0.52$) and CWD ϕ_6 ($\tilde{\rho}$ = −0.55, $q_5(\rho)$ = −0.65, q_{95} = −0.46) as well as neighbourhood crowding ($\tilde{\rho}$ = −0.43, $q_5(\rho)$ = *−*0*.*51, *q*95(*ρ*) = *−*0*.*33, mean across climate models). Tree growth sensitivity to neighbourhood β_{467} crowding (M1, β_{3s}) was negatively associated with the response to SRad anomalies ($\tilde{\rho} = -0.17$, ⁴⁶⁸ $q_5(\rho) = -0.30, q_{95}(\rho) = -0.04$) and positively with the response to anomalies in Tmax ($\tilde{\rho} = 0.26$, $q_5(\rho) = 0.13, q_{95}(\rho) = 0.37$ and CWD ($\tilde{\rho} = 0.28, q_5(\rho) = 0.15, q_{95}(\rho) = 0.41$) For details see Table S4: All plots).

Variation in tree growth response to climate and neighbours with forest disturbance history

 The effects of climate anomalies (M1, *β*2*,*0) and neighbourhood crowding (M1, *β*3*,*0) were consistent between the six undisturbed and nine disturbed plots, but were usually stronger in disturbed plots (except for VPD anomalies). In particular, tree growth sensitivity to neighbourhood crowding was twice as strong in disturbed plots than undisturbed plots. In addition, the interactive effects between climate anomalies and neighbourhood crowding were usually greater in disturbed plots, further attenuating their separate effects in these forests. More crowded trees in disturbed plots were more buffered against climate anomalies (except for the VPD anomalies:NCI interaction). However, while negative effects of CWD anomalies on growth were mitigated by neighbourhood crowding in disturbed plots, they were reinforced in undisturbed plots (Fig. 3).

 Although species intrinsic growth rates varied slightly between undisturbed and disturbed plots 483 (Fig. S2), correlations between species intrinsic growth rates (M1, intercept α_s) and their growth response to climate anomalies (M1, *β*2*s*) and neighbourhood crowding (M1, *β*3*s*) were overall con-485 sistent between undisturbed and disturbed plots (Fig. a-b). However, while species that suffered stronger growth declines due to neighbourhood crowding (M1, *β*3*s*) were also more sensitive to pos- itive CWD anomalies (M1, *β*2*s*) in disturbed plots, we found no clear relationship in undisturbed plots (Fig. 4 c). For details on correlations of species-level parameters see: Table S4.

Figure 4. Correlations between a) species intrinsic growth rates α_s and species-level growth response to CWD anomalies (β_{2s}) , b) species intrinsic growth rates α_s and species-level growth response to neighbourhood crowding (NCI, β_{3s}) and c) species-level growth response to NCI (β_{3s}) and CWD anomalies (β_{2s}), estimated by the CWD M1-models for undisturbed (left) and disturbed (right) plots, respectively (eqn 4c). Circles represent species coefficients (i.e. species median posteriors). Vertical and horizontal lines represent coefficients 90%-highest posterior density intervals (HPDI). Blue and red regression lines indicate positive and negative correlations, respectively. *ρ* values were obtained from the variance-covariance matrix (see: eqn 4e). Median, lower and upper 90%-HPDI are given in the upper right-hand or left-hand corner of the figures.

⁴⁸⁹ **Trait effects on tree growth response to climate anomalies and neigh-**⁴⁹⁰ **bourhood crowding**

⁴⁹¹ Parameter estimates of community-level tree growth regressions were consistent between the M1- ⁴⁹² models (310 focal species) and M2-models (subset of species with available trait information, Table ⁴⁹³ S2 and Table S5). Fast-growing tree species (i.e. with high intrinsic growth rates, *αs*) had higher 494 SD, but lower $\delta^{18}O$, LA, L_{thick} , L_{though} and WSG (Fig. S1 B and Table S5). Species traits influenced ⁴⁹⁵ tree growth response to climate anomalies, neighbourhood crowding and their interaction, either 496 reinforcing (values of $\beta_{2-4,1}$ moved $\beta_{2-4,8}$ further away from zero with increasing trait value) or 497 attenuating them (values of $\beta_{2-4,1}$ moved $\beta_{2-4,8}$ closer to zero with increasing trait value) (Fig. 5).

 μ_{98} Species with lower LSWC and SLA, but higher δ^{13} C, L_{thick} and Bark_{thick} grew faster with increasing ⁴⁹⁹ SRad anomalies (Fig. 5 a). Species with lower SD, but higher $\delta^{18}O$, L_{though} and WSG were less ⁵⁰⁰ sensitive to anomalies in Tmax and VPD (Fig. 5 b-c). Additionally, species with higher δ^{13} C, $_{501}$ g_{min}, LA and L_{thick} showed lower sensitivities to VPD anomalies (Fig. 5 c). Species with lower π_{tlp} but higher WSG were less sensitive to CWD anomalies. (Fig. 5 d). Species with lower SD, ⁵⁰³ but higher δ¹⁸O, L_{though} and WSG were less sensitive to neighbourhood crowding (Fig. 5 a-d). $_{504}$ Finally, species with lower δ^{13} C but higher SLA benefited more from SRad anomalies when found 505 in crowded environments (Fig. 5 a). Species with higher WSG profited less from the mitigating ⁵⁰⁶ effects of neighbours against Tmax anomalies (Fig. 5 b), while species with higher SD and SLA, $_{507}$ but lower LA, L_{thick}, L_{though} were less buffered against VPD anomalies by their neighbours (Fig. ⁵⁰⁸ 5 c). Species traits did not mediate the interactive effects of CWD anomalies and neighbourhood $_{509}$ crowding (Fig. 5 d).

Figure 5. M2-models: trait effects on tree growth response to climate anomalies $\beta_{2,1}$, neighbourhood crowding $\beta_{3,1}$ and the climate anomaly-crowding interaction $\beta_{4,1}$ (see: eqn 5). Standardised coefficients from a) SRad, b) Tmax, c) VPD and d) CWD models are shown for each of the 12 trait models: stomatal density (SD), leaf oxygen ($\delta^{18}O$) and carbon ($\delta^{13}C$) isotope composition, water potential at turgor loss point (π_{tlp}) , leaf saturated water content (LSWC), minimum conductance (g_{min}) , leaf area (LA), specific leaf area (SLA), leaf thickness (L_{thick}), leaf toughness (L_{though}), bark thickness (Barkthick) and wood specific gravity (WSG). Circles show posterior medians of standardised coefficients, and lines indicate 90% HPDIs. Filled circles indicate negative and positive effects (i.e. slope coefficient 90%-HPDI not encompassing zero) and empty circles indicate no clear effects. Positive $\beta_{2-4,1}$ values indicate faster growth with increasing trait values, while negative $\beta_{2-4,1}$ values indicate slower growth with increasing trait values (details in Table S5).

⁵¹⁰ **Discussion**

 $_{511}$ In this study, we examined the separate and interactive effects of climate anomalies and neighbours ⁵¹² on tree growth, combining 30 years of climate and tree growth data for over 300 species from 15 permanent forest plots, with contrasting disturbance histories. For the most abundant species (ca. 100), we further evaluated the effect of 12 traits pertaining to water relations, light capture and carbon use (Table 1), in mediating tree growth response to climate and neighbours.

Climate anomalies and neighbourhood crowding reduce tropical tree growth

 High temperatures (Tmax) were associated with tree growth decline in half of the 300 species at Paracou (*β*2*s*; Table S3), indicating that in years without strong positive anomalies, some species at the site may already operate near their temperature optimum. Indeed, higher Tmax (and minimum night time temperatures) can increase respiration but decrease stomatal conductance and photosynthesis, thus reducing carbon assimilation (Anderegg et al. 2015, Doughty and Goulden 2008) and gross primary production (Aguilos et al. 2018, Crous et al. 2022), which can lead to inter-annual declines in tropical woody productivity (Sullivan et al. 2020, Way and Oren 2010) and carbon sinks (i.e. net ecosystem exchange) (Anderegg et al. 2015). As global temperatures are predicted to increase in the tropics with ongoing climate change (Shukla et al. 2022), this thermal sensitivity of tropical species may lead to important reductions in tree growth in the future, which may alter long-term forest dynamics (Aubry-Kientz et al. 2019, Sullivan et al. 2020). Although Tmax directly influences VPD (Table S6 and Fig. S2), VPD was a less important driver of tree growth at Paracou. While positive VPD anomalies likely reduce tree growth by reducing stomatal conductance and thereby carbon assimilation (Grossiord et al. 2020, Sanginés de Cárcer et al. 2018), it does not affect respiration. Conversely, positive anomalies in CWD strongly reduced tree growth at Paracou (Wagner et al. 2012, Aubry-Kientz et al. 2015), in contrast to previous work in the Australian wet tropics indicating that VPD may limit tree growth before soil water becomes scarce (Bauman et al. 2022a, Choat et al. 2012, Rifai et al. 2018). Our results suggest that trees at Paracou may be more sensitive to soil water stress than to atmospheric water stress. Alternatively, atmospheric water stress may not be severe enough at Paracou to lead to important growth reductions. In addition, trees allocated more to diameter growth with higher anomalies in solar radiation (SRad), which contrasts with recent work showing negative effects of light anomalies on tropical tree growth in Australia's wet tropics (Bauman et al. 2022a). Light availability is an important driver of tree growth and primary production at Paracou (Aguilos et al. 2018, Wagner et al. 2012, 2014a), a typical light-limited tropical forest (Guan et al. 2015, Wagner et al. 2016). Furthermore, years with positive anomalies in SRad, did not necessarily translate into years with increased heat stress and atmospheric drought stress (Table S6 and Fig. S2), therefore reducing confounding and opposing effects of light availability and heat or atmospheric drought stress on tree growth. We explored the effects of different climate drivers separately, but there could be interactive effects between climate factors because they are causally linked. Future studies could further improve our understanding of tropical forest sensitivity to ₅₄₉ climate by embedding the causal associations among climate variables into the model structure, and by explicitly modelling their interactive effects (Allen et al. 2010, 2015, Hammond et al. 2022, Zuidema et al. 2022). Neighbourhood crowding was the strongest driver of tree growth at Paracou, confirming the importance of neighbourhood interactions for tropical forest dynamics (Dawkins 1959, Canham et al. 2004, Fortunel et al. 2018, Uriarte et al. 2016a, Zambrano et al. 2017).

 Neighbourhood crowding overall reduced tree growth across the 300 focal species, in line with previous work across forest systems (Kunstler et al. 2016). Importantly, we provide first evidence that climate anomalies can interact with neighbourhood crowding to shape tropical tree growth, thereby attenuating their respective separate effects. In particular, the negative (Tmax, VPD, CWD) and positive (SRad) effects of climate anomalies on tree growth were attenuated in more crowded neighbourhoods. Denser neighbourhoods may result into higher levels of canopy closure, decreasing light availability (Gaudio et al. 2017, Ghuman and Lal 1987), but simultaneously pro- viding thermal insulation (Fetcher et al. 1985, Gaudio et al. 2017, Ghuman and Lal 1987), thereby sustaining a more favourable micro-climate during extreme macro-climatic events (De Frenne et al. 2013, 2019). Contrary to our expectation, more crowded trees were also less sensitive to CWD anomalies, indicating that buffering effects might out-weigh increased competition for soil water during drought. During periods of lower soil water content, more crowded trees may have benefited from complementary neighbourhood water uptake, hydraulic redistribution or improved water re-tention through higher root densities (Grossiord et al. 2014c,b, Grossiord 2020, Hafner et al. 2021). Previous studies have also shown mitigating effects of neighbourhood diversity or species richness on tree growth response to drought across forest biomes (Fichtner et al. 2020, Gillerot et al. 2021, Grossiord et al. 2014a, O'Brien et al. 2017, Pardos et al. 2021), in contrast with findings showing exacerbating effects of stand or neighbourhood densities on tree demographic response to climatic stresses in temperate (Bottero et al. 2017) and Mediterranean (Astigarraga et al. 2020, Gómez- Aparicio et al. 2011) forests. Our results suggest that accounting for the interactive effects of climate and neighbours offers a promising way towards an improved understanding of tropical forest resistance and resilience to climate change.

Disturbance increases tropical forest sensitivity to climate anomalies and neighbourhood crowding

 As expected, disturbed plots were more sensitive (i.e. showed greater tree growth reductions) to separate and joint effects of climate anomalies and neighbourhood crowding than undisturbed plots at Paracou (Fig. 3 a-d), highlighting the lasting impact of human activities such as log- ging and thinning on tropical forest dynamics (Hérault et al. 2010, Hérault and Piponiot 2018, Hiltner et al. 2021, Fargeon et al. 2016, Piponiot et al. 2016). Differences between disturbed and undisturbed plots in effect sizes of climate anomalies and neighbourhood crowding may originate from differences in tree community composition. Indeed, disturbed plots at Paracou harbor more fast-growing species (Mirabel et al. 2020, 2021), which tend to be more sensitive to climate anoma- lies, as shown in other tropical forests (Bauman et al. 2022a, Esquivel-Muelbert et al. 2020). The higher abundance of fast growing species and individuals on disturbed plots was also reflected $\frac{1}{588}$ in the higher community-level intrinsic growth rate (α_0) of disturbed plots (Fig. S1 and Table S2). While higher sensitivity of disturbed communities to environmental drivers is consistent with expectations from the 'fast-slow' spectrum (Reich 2014), we show that this trend extends to in- $\frac{591}{2}$ creased susceptibility to biotic stressors such as neighbourhood crowding (Fig. 4 a-b, Table S4). In addition, greater canopy openness in disturbed plots may lead to more severe climate effects through tighter coupling of micro- and macro-climatic conditions (De Frenne et al. 2013, 2019, Ghuman and Lal 1987, Von Arx et al. 2012).

 We further demonstrate that interactive effects of climate and neighbours (i.e. *β*4*,*0) can shift from negative to positive depending on forest disturbance history at Paracou. In particular, denser neighbourhoods reinforced soil water stress in undisturbed plots, while they acted as a buffer against soil water stress in disturbed plots. This suggests that in disturbed plots, positive biotic interactions (e.g. complementary resource use between neighbours and direct or indirect facilita- tive effects of neighbouring trees) can mitigate competitive effects for soil water during drought. Disturbed plots typically have more and larger canopy gaps than undisturbed plots, and soil drying (and thus water availability) is controlled by evaporation from the soil surface in large forest gaps, but by root water extraction in the dense understory and small forest gaps (Marthews et al. 2008). In line with the stress gradient hypothesis, our results suggest that positive biotic interactions may become more important under more stressful environmental conditions at Paracou (Bertness and Callaway 1994, Brooker et al. 2007, Callaway 1995, Callaway and Walker 1997, Isbell et al. 2015).

 Neutral crowding indices have recently been suggested to quantify silvicultural treatments (Yue et al. 2022), but in this study they were insufficient to capture all dimensions in disturbance history. Logging effects on forest communities involve not only alterations in forest structure, but also in forest composition. In addition, previous work in temperate and Mediterranean forests suggested that forest management via thinning may help to alleviate the negative impact of climate change on forest dynamics (Astigarraga et al. 2020, Gómez-Aparicio et al. 2011), but our study provides a more nuanced perspective in tropical forests where the interactive effects of neighbourhoods and climate may change from negative to positive depending on the specific climatic variable. Given that more than half of the world's tropical forests are designated production forests (Blaser et al. 2011), our results warrant closer examination of how logging and land use practices affect neighbourhood interactions and the resistance and resilience of tropical forests to future climate change (Hiltner et al. 2021)

Traits mediate tree growth response to climate anomalies and neigh-bourhood crowding

 In line with expectations from the 'fast-slow' economic spectrum (Reich 2014), high species in- trinsic growth rates were related to acquisitive strategies for carbon and water (Fig. S1 B). Fast ϵ_{23} growing species had high SD and low $\delta^{18}O$, highlighting the key role of stomatal conductance and transpiration rates for efficient carbon assimilation (Drake et al. 2013, Hepworth et al. 2015, Jarvis 1981, Poorter and Bongers 2006, Wong et al. 1979). Fast-growing species also exhibited leaf and $\frac{626}{100}$ wood traits pertaining to acquisitive strategies such as lower L_{thick} , L_{though} , and WSG, supporting previous findings that a species investment into tissue longevity comes at higher construction costs (Hérault et al. 2011, Philipson et al. 2014, Poorter et al. 2008, 2010, Reich 2014, Wright et al. $629 \quad 2010$).

 ϵ_{630} Focusing on ca. 100 species that make up $>70\%$ of total stems and basal area at Paracou, our study showed that the effects of climate on tree growth were mediated by different traits depending on the climate variable (Fig. 5 b-c). Leaf water-related traits overall mediated the negative effects of climate anomalies related to heat and drought stress. More specifically, species that prioritise ω_{4} water-conservation over carbon gain (e.g. low SD and high $\delta^{18}O$) were less sensitive to heat and atmospheric drought stress. These functional strategies reduced species' potential for fast growth, but may allow species to maintain photosynthesis and *CO*² assimilation during periods of elevated temperature and atmospheric evaporative demands (Caine et al. 2019, Hepworth et al. 2015, Reich 2014, Wang and Wen 2022a).

 The mitigating effects of traits linked to carbon-water relations were less clear for tree growth response to soil water stress (CWD), potentially because individual topographic position may lead to a decoupling between growth sensitivity to soil water stress and water conservation capacity 642 (Esteban et al. 2021, Hammond et al. 2022). Still, the growth of species with more negative π_{th} 643 was less impacted by soil drought stress. Lower π_{tlp} (Table 1) captures species' drought-related mortality risks (Powell et al. 2017) and distribution along aridity gradients (Bartlett et al. 2012b), but previous work from different biomes found mixed evidence for species drought-related growth reductions (McGregor et al. 2021, Smith‐Martin et al. 2023, Song et al. 2022).

 $\frac{647}{100}$ While we predicted species with higher g_{min} would suffer stronger drought-related growth declines (Blackman et al. 2016, Duursma et al. 2019), we found they were less vulnerable to atmospheric drought stress. As elevated VPD leads to stomatal closure (Grossiord et al. 2020), residual water loss could maintain evaporative cooling (given non limiting soil water conditions), which potentially prevents irreversible leaf tissue damage through overheating (Drake et al. 2018). The maintainance of evaporative cooling could thus allow species to quickly resume carbon assimilation after atmo- spheric drought stress, while simultaneously reducing post-drought respiration rates linked to tissue repair (Drake et al. 2018, Powell et al. 2017).

 Leaf and wood morphological traits also captured tree growth response to climate anomalies at Paracou. Species with conservative tissues (i.e. lower SLA, higher L_{thick} and $Bark_{thick}$) grew faster at higher solar radiation (Fig. 5 a). During atmospheric drought and heat stress, species with thicker and tougher leaves may experience less tissue damage through desiccation, thus reducing post-drought costs for tissue repair (Poorter et al. 2010, Nardini 2022), while species with larger leaves have higher evaporative cooling through transpiration (Gates 2003, Schuepp 1993, Wright et al. 2017), leading to less growth reductions. Moreover, species with high WSG were less vulner- able to heat stress, atmospheric and soil drought stress at Paracou, confirming previous findings (Greenwood et al. 2017, Wagner et al. 2014b).

 Slow-growing, resource-conservative species tend to be more tolerant to competition at Paracou. $\frac{665}{100}$ In particular, species conservative for carbon (with higher L_{though} and WSG) were less sensitive to neighbourhood crowding, as found in previous studies across forest biomes (Fortunel et al. 2016, Kunstler et al. 2016). In addition, species conservative for water (with lower SD, stomatal conduc- tance and transpiration rates) were also less sensitive to neighbourhood crowding, highlighting the role of water limitation in these tropical wet forests. This is a similar suite of traits that mediated ϵ_{670} growth response to heat stress, suggesting that species suffering most from Tmax anomalies are also experiencing stronger neighbourhood effects (Table S4).

However, species traits poorly captured the interactive effects of climate anomalies and neighbour-

 ϵ_{673} hood crowding on tree growth at Paracou (Fig. 5 a-d). Trait differences between neighbouring trees rather than the traits of the focal tree alone may be playing a key role in driving the inter- active effects of climate anomalies and neighbourhood crowding on tree growth, in particular as trait differences can reflect niche partitioning and competitive hierarchies between neighbouring individuals (Fortunel et al. 2016, Kunstler et al. 2016). To clarify this, future work could explore the relative contribution of focal tree traits and trait differences in mediating tree growth response to the interaction between climate extremes and neighbourhood crowding.

Conclusion

 Tropical trees are particularly sensitive to climate stress and interactions with neighbours, sug- gesting that ongoing climate change and associated changes in forest composition and structure may result in a long-term slowdown of carbon sequestration in tropical forests. Our study shows that climate anomalies and neighbourhood crowding can interact positively in shaping tropical tree growth, suggesting that trees growing in more crowded neighbourhood may be less sensitive to climate extremes. Logged forests suffered considerably stronger growth reductions in response to water stress, suggesting lower resistance of managed tropical forests to climate change. Our findings reveal that functional traits, in particular related to water relations, can capture differ- ences in species sensitivities to atmospheric and soil water stress and neighbourhood crowding (though not their interactive effects), that can be implemented in models to improve predictions of the future of these highly diverse ecosystems.

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Author Contributions

 Daniela Nemetschek, Claire Fortunel and Géraldine Derroire designed the study. Daniela Nemetschek, Marion Boisseaux, Johanna Auer, Vincyane Badouard, Christopher Baraloto, Quentin Le Blaye,

 Damien Bonal, Sabrina Coste, Elia Dardevet, Claire Fortunel, Patrick Heuret, Sébastien Levion- nois, Isabelle Maréchaux, Clément Stahl, Jason Vleminckx, Wolfgang Wanek and Camille Ziegler collected or contributed trait data. Daniela Nemetschek formatted and vetted the plot census, climate and functional trait data, with help from Claire Fortunel, Géraldine Derroire, Eric Marcon and Marion Boisseaux. Daniela Nemetschek, Claire Fortunel and Géraldine Derroire designed the tree growth models, with input from Eric Marcon, David Bauman, Sean M. McMahon and Mélaine Aubry-Kientz. Daniela Nemetschek performed the analyses, with help from Claire Fortunel, Géral- dine Derroire and Eric Marcon. Daniela Nemetschek and Claire Fortunel led the interpretation of the results and the writing of the first draft, with frequent input from Géraldine Derroire. All authors contributed to revisions and gave final approval for publication.

Data Availability Statement

Plot census data was extracted from the Paracou Station database, for which access is available at

https://dataverse.cirad.fr/dataverse/paracou, with corresponding DOIs: 10.18167/DVN1/NSCWF0;

10.18167/DVN1/Q8V2YI; 10.18167/DVN1/LIVCEK; 10.18167/DVN1/HWTD4U;

 10.18167/DVN1/HIGNWQ. Species mean values for the six leaf water-related traits are available at 10.5281/zenodo.10396927

Conflict of interest

The authors declare there are no competing interests.

References

- Aakala, T., Berninger, F., and Starr, M. The roles of competition and climate in tree growth variation in northern boreal old-growth forests. *Journal of Vegetation Science*, 29(6):1040–1051, 2018. doi: 10.1111/jvs.12687.
- Abatzoglou, J. T., Dobrowski, S. Z., Parks, S. A., and Hegewisch, K. C. TerraClimate, a high- resolution global dataset of monthly climate and climatic water balance from 1958–2015. *Sci-entific Data*, 5(1):170191, 2018. doi: 10.1038/sdata.2017.191.
- Aguilos, M., Hérault, B., Burban, B., Wagner, F., and Bonal, D. What drives long-term variations in carbon flux and balance in a tropical rainforest in French Guiana? *Agricultural and Forest Meteorology*, 253-254:114–123, 2018. doi: 10.1016/j.agrformet.2018.02.009.
- Alder, D. and Sunnott, T. *Permanent sample plot techniques for mixed tropical forest*. Oxford Forestry Institute, University of Oxford, 1992.
- Allen, C. D., Macalady, A. K., Chenchouni, H., Bachelet, D., McDowell, N., Vennetier, M., Kitzberger, T., Rigling, A., Breshears, D. D., Hogg, E. T., Gonzalez, P., Fensham, R., Zhang, Z., Castro, J., Demidova, N., Lim, J.-H., Allard, G., Running, S. W., Semerci, A., and Cobb, N. A global overview of drought and heat-induced tree mortality reveals emerging cli- mate change risks for forests. *Forest Ecology and Management*, 259(4):660–684, 2010. doi: 10.1016/j.foreco.2009.09.001.
- Allen, C. D., Breshears, D. D., and McDowell, N. G. On underestimation of global vulnerability to tree mortality and forest die-off from hotter drought in the Anthropocene. *Ecosphere*, 6(8): 129, 2015. doi: 10.1890/ES15-00203.1.
- Anderegg, W. R. L., Ballantyne, A. P., Smith, W. K., Majkut, J., Rabin, S., Beaulieu, C., Birdsey, R., Dunne, J. P., Houghton, R. A., Myneni, R. B., Pan, Y., Sarmiento, J. L., Serota, N., Shevliakova, E., Tans, P., and Pacala, S. W. Tropical nighttime warming as a dominant driver of variability in the terrestrial carbon sink. *Proceedings of the National Academy of Sciences*, 112(51):15591–15596, 2015. doi: 10.1073/pnas.1521479112.
- Anderegg, W. R. L., Konings, A. G., Trugman, A. T., Yu, K., Bowling, D. R., Gabbitas, R., Karp, D. S., Pacala, S., Sperry, J. S., Sulman, B. N., and Zenes, N. Hydraulic diversity of forests regulates ecosystem resilience during drought. *Nature*, 561(7724):538–541, 2018. doi: 10.1038/s41586-018-0539-7.
- Astigarraga, J., Andivia, E., Zavala, M. A., Gazol, A., Cruz-Alonso, V., Vicente-Serrano, S. M., and Ruiz-Benito, P. Evidence of non-stationary relationships between climate and forest re- sponses: Increased sensitivity to climate change in Iberian forests. *Global Change Biology*, 26 (9):5063–5076, 2020. doi: 10.1111/GCB.15198.
- Aubry-Kientz, M., Rossi, V., Wagner, F., and Hérault, B. Identifying climatic drivers of tropical forest dynamics. *Biogeosciences*, 12(19):5583–5596, 2015. doi: 10.5194/bg-12-5583-2015.
- Aubry-Kientz, M., Rossi, V., Cornu, G., Wagner, F., and Hérault, B. Temperature rising would slow down tropical forest dynamic in the Guiana Shield. *Scientific Reports*, 9(1):10235, 2019. doi: 10.1038/s41598-019-46597-8.
- Baraloto, C., Goldberg, D. E., and Bonal, D. Performance trade-offs among tropical tree seedlings in contrasting microhabitats. *Ecology*, 86(9):2461–2472, 2005. doi: 10.1890/04-1956.
- Baraloto, C., Timothy Paine, C. E., Poorter, L., Beauchene, J., Bonal, D., Domenach, A.-M.,
- Hérault, B., Patiño, S., Roggy, J.-C., and Chave, J. Decoupled leaf and stem economics in rain
- forest trees. *Ecology Letters*, 13(11):1338–1347, 2010. doi: 10.1111/j.1461-0248.2010.01517.x.
- Baraloto, C., Hérault, B., Paine, C. E. T., Massot, H., Blanc, L., Bonal, D., Molino, J.-F., Nicol- ini, E. A., and Sabatier, D. Contrasting taxonomic and functional responses of a tropical tree community to selective logging. *Journal of Applied Ecology*, 49(4):861–870, 2012. doi: $781 \quad 10.1111/j.1365-2664.2012.02164.x.$
- Barros, F. d. V., Bittencourt, P. R. L., Brum, M., Restrepo‐Coupe, N., Pereira, L., Teodoro,
- G. S., Saleska, S. R., Borma, L. S., Christoffersen, B. O., Penha, D., Alves, L. F., Lima, A.
- J. N., Carneiro, V. M. C., Gentine, P., Lee, J., Aragão, L. E. O. C., Ivanov, V., Leal, L. S. M.,
- Araujo, A. C., and Oliveira, R. S. Hydraulic traits explain differential responses of Amazonian
- forests to the 2015 El Niño‐induced drought. *New Phytologist*, 223(3):1253–1266, 2019. doi: $787 \quad 10.1111/\text{nph}.15909.$
- Barrs, H. and Weatherley, P. A Re-Examination of the Relative Turgidity Technique for Estimating Water Deficits in Leaves. *Australian Journal of Biological Sciences*, 15(3):413–428, 1962. doi: 10.1071/bi9620413.
- Bartlett, M. K., Zhang, Y., Yang, J., Kreidler, N., Sun, S.-W., Lin, L., Hu, Y.-H., Cao, K.-F., and Sack, L. Drought tolerance as a driver of tropical forest assembly: resolving spatial signatures for multiple processes. *Ecology*, 97(2):503–514, 2016. doi: 10.1890/15-0468.1.
- Bartlett, M. K., Scoffoni, C., Ardy, R., Zhang, Y., Sun, S., Cao, K., and Sack, L. Rapid determi- nation of comparative drought tolerance traits: using an osmometer to predict turgor loss point. *Methods in Ecology and Evolution*, 3(5):880–888, 2012a. doi: 10.1111/j.2041-210X.2012.00230.x.
- Bartlett, M. K., Scoffoni, C., and Sack, L. The determinants of leaf turgor loss point and prediction of drought tolerance of species and biomes: a global meta-analysis. *Ecology Letters*, 15(5):393– 405, 2012b. doi: 10.1111/j.1461-0248.2012.01751.x.
- Bauman, D., Fortunel, C., Cernusak, L. A., Bentley, L. P., McMahon, S. M., Rifai, S. W., Aguirre‐Gutiérrez, J., Oliveras, I., Bradford, M., Laurance, S. G. W., Delhaye, G., Hutchin- son, M. F., Dempsey, R., McNellis, B. E., Santos‐Andrade, P. E., Ninantay‐Rivera, H. R., Chambi Paucar, J. R., Phillips, O. L., and Malhi, Y. Tropical tree growth sensitivity to climate is driven by species intrinsic growth rate and leaf traits. *Global Change Biology*, 28(4):1414–1432, 2022a. doi: 10.1111/gcb.15982.
- Bauman, D., Fortunel, C., Delhaye, G., Malhi, Y., Cernusak, L. A., Bentley, L. P., Rifai, S. W.,
- Aguirre-Gutiérrez, J., Menor, I. O., Phillips, O. L., McNellis, B. E., Bradford, M., Laurance,
- S. G. W., Hutchinson, M. F., Dempsey, R., Santos-Andrade, P. E., Ninantay-Rivera, H. R.,
- Chambi Paucar, J. R., and McMahon, S. M. Tropical tree mortality has increased with rising
- atmospheric water stress. *Nature*, 608(7923):528–533, 2022b. doi: 10.1038/s41586-022-04737-7.
- Bertness, M. D. and Callaway, R. Positive interactions in communities. *Trends in Ecology & Evolution*, 9(5):191–193, 1994. doi: 10.1016/0169-5347(94)90088-4.
- 813 Bertolino, L. T., Caine, R. S., and Gray, J. E. Impact of Stomatal Density and Morphology on Water-Use Efficiency in a Changing World. *Frontiers in Plant Science*, 10:6, 2019. doi: 815 10.3389/fpls.2019.00225.
- Blackman, C. J., Aspinwall, M. J., Resco de Dios, V., Smith, R. A., and Tissue, D. T. Leaf photosynthetic, economics and hydraulic traits are decoupled among genotypes of a widespread species of eucalypt grown under ambient and elevated CO2. *Functional Ecology*, 30(9):1491– 1500, 2016. doi: $10.1111/1365-2435.12661$.
- Blackman, C. J., Creek, D., Maier, C., Aspinwall, M. J., Drake, J. E., Pfautsch, S., O'Grady, A., Delzon, S., Medlyn, B. E., Tissue, D. T., and Choat, B. Drought response strategies and hydraulic traits contribute to mechanistic understanding of plant dry-down to hydraulic failure. *Tree Physiology*, 39(6):910–924, 2019. doi: 10.1093/treephys/tpz016.
- Blaser, J., Sarre, A., Poore, D., and Johnson, S. Status of Tropical Forest Management 2011. ⁸²⁵ Technical report, International Tropical Timber Organization, Yokohama, Japan, 2011.
- Bottero, A., D'Amato, A. W., Palik, B. J., Bradford, J. B., Fraver, S., Battaglia, M. A., and Asherin, L. A. Density-dependent vulnerability of forest ecosystems to drought. *Journal of Applied Ecology*, 54(6):1605–1614, 2017. doi: 10.1111/1365-2664.12847.
- Brodribb, T. J., Powers, J., Cochard, H., and Choat, B. Hanging by a thread? Forests and drought. *Science*, 368(6488):261–266, 2020. doi: 10.1126/science.aat7631.
- Brooker, R. W., Maestre, F. T., Callaway, R. M., Lortie, C. L., Cavieres, L. A., Kunstler, G., Liancourt, P., Tielbörger, K., Travis, J. M., Anthelme, F., Armas, C., Coll, L., Corcket, E., Delzon, S., Forey, E., Kikvidze, Z., Olofsson, J., Pugnaire, F. I., Quiroz, C. L., Saccone, P., Schiffers, K., Seifan, M., Touzard, B., and Michalet, R. Facilitation in plant communities: the past, the present, and the future. *Journal of Ecology*, 96(1):18–34, 2007. doi: 10.1111/j.1365-
- 836 2745.2007.01295.x.

 Bürkner, P. C. brms: An R Package for Bayesian Multilevel Models Using Stan. *Journal of Statistical Software*, 80(1):1–28, 2017. doi: 10.18637/JSS.V080.I01.

 Caine, R. S., Yin, X., Sloan, J., Harrison, E. L., Mohammed, U., Fulton, T., Biswal, A. K., Dionora, J., Chater, C. C., Coe, R. A., Bandyopadhyay, A., Murchie, E. H., Swarup, R., Quick, ⁸⁴¹ W. P., and Gray, J. E. Rice with reduced stomatal density conserves water and has improved drought tolerance under future climate conditions. *New Phytologist*, 221(1):371–384, 2019. doi: 10.1111/nph.15344.

- Callaway, R. M. Positive interactions among plants. *The Botanical Review*, 61(4):306–349, 1995. doi: 10.1007/BF02912621.
- Callaway, R. M. and Walker, L. R. Competition and facilitation: A synthetic approach to in- teractions in plant communities. *Ecology*, 78(7):1958–1965, 1997. doi: doi.org/10.1890/0012- 848 9658(1997)078[1958:CAFASA]2.0.CO;2.
- Canham, C. D., LePage, P. T., and Coates, K. D. A neighborhood analysis of canopy tree competition: effects of shading versus crowding. *Canadian Journal of Forest Research*, 34(4): 778–787, 2004. doi: 10.1139/x03-232.
- Cannon, C. H., Peart, D. R., Leighton, M., and Kartawinata, K. The structure of lowland rainforest after selective logging in West Kalimantan, Indonesia. *Forest Ecology and Management*, 67(1-3): $854 \qquad 49-68, 1994. \text{ doi: } 10.1016/0378-1127(94)90007-8.$
- Carpenter, B., Gelman, A., Hoffman, M. D., Lee, D., Goodrich, B., Betancourt, M., Brubaker, M. A., Guo, J., Li, P., and Riddell, A. Stan: A Probabilistic Programming Language. *Journal*
- *of Statistical Software*, 76(1):1–32, 2017. doi: 10.18637/JSS.V076.I01.
- Carreño-Rocabado, G., Peña-Claros, M., Bongers, F., Alarcón, A., Licona, J.-C., and Poorter, L. Effects of disturbance intensity on species and functional diversity in a tropical forest. *Journal*
- *of Ecology*, 100(6):1453–1463, 2012. doi: 10.1111/j.1365-2745.2012.02015.x.
- Cernusak, L. A., Ubierna, N., Winter, K., Holtum, J. A. M., Marshall, J. D., and Farquhar, G. D.
- Environmental and physiological determinants of carbon isotope discrimination in terrestrial plants. *New Phytologist*, 200(4):950–965, 2013. doi: 10.1111/nph.12423.
- Chave, J., Coomes, D., Jansen, S., Lewis, S. L., Swenson, N. G., and Zanne, A. E. Towards a worldwide wood economics spectrum. *Ecology Letters*, 12:351–366, 2009. doi: 10.1111/j.1461- 866 0248.2009.01285.x.
- Choat, B., Jansen, S., Brodribb, T. J., Cochard, H., Delzon, S., Bhaskar, R., Bucci, S. J., Feild, T. S., Gleason, S. M., Hacke, U. G., Jacobsen, A. L., Lens, F., Maherali, H., Martínez-Vilalta,
- J., Mayr, S., Mencuccini, M., Mitchell, P. J., Nardini, A., Pittermann, J., Pratt, R. B., Sperry,
- J. S., Westoby, M., Wright, I. J., and Zanne, A. E. Global convergence in the vulnerability of forests to drought. *Nature*, 491(7426):752–755, 2012. doi: 10.1038/nature11688.
- Choat, B., Brodribb, T. J., Brodersen, C. R., Duursma, R. A., López, R., and Medlyn, B. E. Triggers of tree mortality under drought. *Nature*, 558(7711):531–539, 2018. doi: 10.1038/s41586- $874 \qquad 018-0240-x.$
- Clark, J. A. and Covey, K. R. Tree species richness and the logging of natural forests: A meta-analysis. *Forest Ecology and Management*, 276:146–153, 2012. doi: 10.1016/j.foreco.2012.04.001.
- 877 Crous, K. Y., Uddling, J., and De Kauwe, M. G. Temperature responses of photosynthesis and respiration in evergreen trees from boreal to tropical latitudes. *New Phytologist*, 234(2):353–374, 2022. doi: $10.1111/pph.17951$.
- Cunningham, S. C. Photosynthetic responses to vapour pressure deficit in temperate and tropical evergreen rainforest trees of Australia. *Oecologia*, 142(4):521–528, 2005. doi: 10.1007/s00442- $882 \qquad 004-1766-1.$
- Dawkins, H. The volume increment of natural tropical high-forest and limitations on its improve-ments. *Empire Forestry Review*, 38:175–180, 1959.
- De Frenne, P., Rodríguez-Sánchez, F., Coomes, D. A., Baeten, L., Verstraeten, G., Vellend, M.,
- Bernhardt-Römermann, M., Brown, C. D., Brunet, J., Cornelis, J., Decocq, G. M., Dierschke,
- H., Eriksson, O., Gilliam, F. S., Hédl, R., Heinken, T., Hermy, M., Hommel, P., Jenkins, M. A.,
- Kelly, D. L., Kirby, K. J., Mitchell, F. J. G., Naaf, T., Newman, M., Peterken, G., Petřík, P.,
- Schultz, J., Sonnier, G., Van Calster, H., Waller, D. M., Walther, G.-R., White, P. S., Woods,
- K. D., Wulf, M., Graae, B. J., and Verheyen, K. Microclimate moderates plant responses to
- macroclimate warming. *Proceedings of the National Academy of Sciences*, 110(46):18561–18565,
- 2013. doi: 10.1073/pnas.1311190110.
- De Frenne, P., Zellweger, F., Rodríguez-Sánchez, F., Scheffers, B. R., Hylander, K., Luoto, M.,
- Vellend, M., Verheyen, K., and Lenoir, J. Global buffering of temperatures under forest canopies. *Nature Ecology & Evolution*, 3(5):744–749, 2019. doi: 10.1038/s41559-019-0842-1.
- Derroire, G., Hérault, B., Rossi, V., Blanc, L., Gourlet-Fleury, S., and Schmitt, L. *Paracou Biodiversity Plots*. CIRAD Dataverse, 2022a. doi: 10.18167/DVN1/NSCWF0.
- Derroire, G., Hérault, B., Rossi, V., Blanc, L., Gourlet-Fleury, S., and Schmitt, L. *Paracou Dis-turbance Experiment - Control Plots*. CIRAD Dataverse, 2022b. doi: 10.18167/DVN1/Q8V2YI.
- Derroire, G., Hérault, B., Rossi, V., Blanc, L., Gourlet-Fleury, S., and Schmitt, L. *Para- cou Disturbance Experiment - Level1 Treatment Plots*. CIRAD Dataverse, 2022c. doi: 902 10.18167/DVN1/LIVCEK.
- Derroire, G., Hérault, B., Rossi, V., Blanc, L., Gourlet-Fleury, S., and Schmitt, L. *Para- cou Disturbance Experiment - Level2 Treatment Plots*. CIRAD Dataverse, 2022d. doi: 905 10.18167/DVN1/HWTD4U.
- Derroire, G., Hérault, B., Rossi, V., Blanc, L., Gourlet-Fleury, S., and Schmitt, L. *Para- cou Disturbance Experiment - Level3 Treatment Plots*. CIRAD Dataverse, 2022e. doi: 908 10.18167/DVN1/HIGNWQ.
- Doughty, C. E. and Goulden, M. L. Are tropical forests near a high temperature threshold? *Journal of Geophysical Research: Biogeosciences*, 113(G00B07), 2008. doi: 10.1029/2007JG000632.
- Drake, J. E., Tjoelker, M. G., Vårhammar, A., Medlyn, B. E., Reich, P. B., Leigh, A., Pfautsch, S.,
- Blackman, C. J., López, R., Aspinwall, M. J., Crous, K. Y., Duursma, R. A., Kumarathunge, D.,
- De Kauwe, M. G., Jiang, M., Nicotra, A. B., Tissue, D. T., Choat, B., Atkin, O. K., and Barton, C. V. M. Trees tolerate an extreme heatwave via sustained transpirational cooling and increased leaf thermal tolerance. *Global Change Biology*, 24(6):2390–2402, 2018. doi: 10.1111/gcb.14037.
- Drake, P. L., Froend, R. H., and Franks, P. J. Smaller, faster stomata: scaling of stomatal size, rate of response, and stomatal conductance. *Journal of Experimental Botany*, 64(2):495–505, 918 2013. doi: 10.1093/jxb/ers347.
- Duursma, R. A., Blackman, C. J., Lopéz, R., Martin‐StPaul, N. K., Cochard, H., and Medlyn, B. E. On the minimum leaf conductance: its role in models of plant water use, and ecological and environmental controls. *New Phytologist*, 221(2):693–705, 2019. doi: 10.1111/nph.15395.
- Esquivel-Muelbert, A., Phillips, O. L., Brienen, R. J. W., Fauset, S., Sullivan, M. J. P., Baker,
- T. R., Chao, K.-J., Feldpausch, T. R., Gloor, E., Higuchi, N., Houwing-Duistermaat, J., Lloyd,
- J., Liu, H., Malhi, Y., Marimon, B., Marimon Junior, B. H., Monteagudo-Mendoza, A., Poorter,
- L., Silveira, M., Torre, E. V., Dávila, E. A., del Aguila Pasquel, J., Almeida, E., Loayza, P. A.,
- Andrade, A., Aragão, L. E. O. C., Araujo-Murakami, A., Arets, E., Arroyo, L., Aymard C.,
- G. A., Baisie, M., Baraloto, C., Camargo, P. B., Barroso, J., Blanc, L., Bonal, D., Bongers,
- F., Boot, R., Brown, F., Burban, B., Camargo, J. L., Castro, W., Moscoso, V. C., Chave, J.,
- Comiskey, J., Valverde, F. C., da Costa, A. L., Cardozo, N. D., Di Fiore, A., Dourdain, A.,
- Erwin, T., Llampazo, G. F., Vieira, I. C. G., Herrera, R., Honorio Coronado, E., Huamantupa-
- Chuquimaco, I., Jimenez-Rojas, E., Killeen, T., Laurance, S., Laurance, W., Levesley, A., Lewis,
- S. L., Ladvocat, K. L. L. M., Lopez-Gonzalez, G., Lovejoy, T., Meir, P., Mendoza, C., Morandi,
- P., Neill, D., Nogueira Lima, A. J., Vargas, P. N., de Oliveira, E. A., Camacho, N. P., Pardo,
- G., Peacock, J., Peña-Claros, M., Peñuela-Mora, M. C., Pickavance, G., Pipoly, J., Pitman, N.,
- Prieto, A., Pugh, T. A. M., Quesada, C., Ramirez-Angulo, H., de Almeida Reis, S. M., Rejou-
- Machain, M., Correa, Z. R., Bayona, L. R., Rudas, A., Salomão, R., Serrano, J., Espejo, J. S.,
- Silva, N., Singh, J., Stahl, C., Stropp, J., Swamy, V., Talbot, J., ter Steege, H., Terborgh, J.,
- Thomas, R., Toledo, M., Torres-Lezama, A., Gamarra, L. V., van der Heijden, G., van der Meer,
- P., van der Hout, P., Martinez, R. V., Vieira, S. A., Cayo, J. V., Vos, V., Zagt, R., Zuidema, P.,
- and Galbraith, D. Tree mode of death and mortality risk factors across Amazon forests. *Nature Communications*, 11(1):5515, 2020. doi: 10.1038/s41467-020-18996-3.
- Esteban, E. J. L., Castilho, C. V., Melgaço, K. L., and Costa, F. R. C. The other side of droughts: wet extremes and topography as buffers of negative drought effects in an Amazonian forest. *New Phytologist*, 229(4):1995–2006, 2021. doi: 10.1111/nph.17005.
- Fargeon, H., Aubry-Kientz, M., Brunaux, O., Descroix, L., Gaspard, R., Guitet, S., Rossi, V., and Hérault, B. Vulnerability of Commercial Tree Species to Water Stress in Logged Forests of the Guiana Shield. *Forests*, 7(12):105, 2016. doi: 10.3390/f7050105.
- Farquhar, G. D., Ehleringer, J. R., and Hubick, K. T. Carbon Isotope Discrimination and Pho- tosynthesis. *Annual Review of Plant Physiology and Plant Molecular Biology*, 40(1):503–537, 950 1989. doi: 10.1146/ANNUREV.PP.40.060189.002443.
- Farquhar, G. D., Cernusak, L. A., and Barnes, B. Heavy Water Fractionation during Transpiration. *Plant Physiology*, 143(1):11–18, 2007. doi: 10.1104/pp.106.093278.
- Fetcher, N., Oberbauer, S. F., and Strain, B. R. Vegetation effects on microclimate in lowland tropical forest in Costa Rica. *International Journal of Biometeorology*, 29:145–155, 1985. doi: 955 10.1007/BF02189035.
- Fichtner, A., Schnabel, F., Bruelheide, H., Kunz, M., Mausolf, K., Schuldt, A., Härdtle, W., and von Oheimb, G. Neighbourhood diversity mitigates drought impacts on tree growth. *Journal of Ecology*, 108(3):865–875, 2020. doi: 10.1111/1365-2745.13353.
- Finegan, B. and Camacho, M. Stand dynamics in a logged and silviculturally treated Costa Rican rain forest, 1988–1996. *Forest Ecology and Management*, 121(3):177–189, 1999. doi: $961 \qquad 10.1016/S0378-1127(98)00550-7.$
- Finegan, B., Camacho, M., and Zamora, N. Diameter increment patterns among 106 tree species in a logged and silviculturally treated Costa Rican rain forest. *Forest Ecology and Management*, $964 \qquad 121(3):159-176, 1999.$ doi: $10.1016/S0378-1127(98)00551-9.$
- Fortunel, C., Fine, P. V. A., and Baraloto, C. Leaf, stem and root tissue strategies across 758 Neotropical tree species. *Functional Ecology*, 26(5):1153–1161, 2012. doi: 10.1111/j.1365- 967 2435.2012.02020.x.
- Fortunel, C., Valencia, R., Wright, S. J., Garwood, N. C., and Kraft, N. J. Functional trait differences influence neighbourhood interactions in a hyperdiverse Amazonian forest. *Ecology letters*, 19:1062–1070, 2016. doi: 10.1111/ele.12642.
- Fortunel, C., Lasky, J. R., Uriarte, M., Valencia, R., Wright, S. J., Garwood, N. C., and Kraft, N. Topography and neighborhood crowding can interact to shape species growth and distribution in a diverse Amazonian forest. *Ecology*, 99(10):2272–2283, 2018. doi: 10.1002/ecy.2441.
- Garnier, E., Shipley, B., Roumet, C., and Laurent, G. A standardized protocol for the determi- nation of specific leaf area and leaf dry matter content. *Functional Ecology*, 15(5):688–695, 10 976 2001. doi: $10.1046/J.0269-8463.2001.00563.x$.
- Gates, D. M. Transpiration and Leaf Temperature. *Annual Review of Plant Physiology*, 19:211–238, 978 2003. doi: 10.1146/ANNUREV.PP.19.060168.001235.
- Gatti, R. C., Reich, P. B., Gamarra, J. G., Crowther, T., Hui, C., Morera, A., Bastin, J. F., De-Miguel, S., Nabuurs, G. J., Svenning, J. C., Serra-Diaz, J. M., Merow, C., Enquist, B., Kamenetsky, M., Lee, J., Zhu, J., Fang, J., Jacobs, D. F., Pijanowski, B., Banerjee, A., Gi- aquinto, R. A., Alberti, G., Zambrano, A. M. A., Alvarez-Davila, E., Araujo-Murakami, A., Avitabile, V., Aymard, G. A., Balazy, R., Baraloto, C., Barroso, J. G., Bastian, M. L., Birn- baum, P., Bitariho, R., Bogaert, J., Bongers, F., Bouriaud, O., Brancalion, P. H., Brearley, F. Q., Broadbent, E. N., Bussotti, F., da Silva, W. C., César, R. G., Češljar, G., Moscoso, V. C., Chen, H. Y., Cienciala, E., Clark, C. J., Coomes, D. A., Dayanandan, S., Decuyper, M., Dee, L. E., Del Aguila Pasquel, J., Derroire, G., Djuikouo, M. N. K., van Do, T., Dolezal, J., Đorđevic, I., Engel, J., Fayle, T. M., Feldpausch, T. R., Fridman, J. K., Harris, D. J., Hemp,
- A., Hengeveld, G., Herault, B., Herold, M., Ibanez, T., Jagodzinski, A. M., Jaroszewicz, B.,
- Jeffery, K. J., Johannsen, V. K., Jucker, T., Kangur, A., Karminov, V. N., Kartawinata, K.,
- Kennard, D. K., Kepfer-Rojas, S., Keppel, G., Khan, M. L., Khare, P. K., Kileen, T. J., Kim,
- H. S., Korjus, H., Kumar, A., Kumar, A., Laarmann, D., Labrière, N., Lang, M., Lewis, S. L., Lukina, N., Maitner, B. S., Malhi, Y., Marshall, A. R., Martynenko, O. V., Monteagudo Men- doza, A. L., Ontikov, P. V., Ortiz-Malavasi, E., Pallqui Camacho, N. C., Paquette, A., Park, M., Parthasarathy, N., Peri, P. L., Petronelli, P., Pfautsch, S., Phillips, O. L., Picard, N., Piotto, D., Poorter, L., Poulsen, J. R., Pretzsch, H., Ramírez-Angulo, H., Correa, Z. R., Rodeghiero, M., Del Pilar Rojas Gonzáles, R., Rolim, S. G., Rovero, F., Rutishauser, E., Saikia, P., Salas- Eljatib, C., Schepaschenko, D., Scherer-Lorenzen, M., Šebeň, V., Silveira, M., Slik, F., Sonké, B., Souza, A. F., Stereńczak, K. J., Svoboda, M., Taedoumg, H., Tchebakova, N., Terborgh, J., Tikhonova, E., Torres-Lezama, A., van der Plas, F., Vásquez, R., Viana, H., Vibrans, A. C., Vilanova, E., Vos, V. A., Wang, H. F., Westerlund, B., White, L. J., Wiser, S. K., Zawiła- Niedźwiecki, T., Zemagho, L., Zhu, Z. X., Zo-Bi, I. C., and Liang, J. The number of tree species on Earth. *Proceedings of the National Academy of Sciences of the United States of America*, 119 (6):e2115329119, 2022. doi: 10.1073/pnas.2115329119.
- Gaudio, N., Gendre, X., Saudreau, M., Seigner, V., and Balandier, P. Impact of tree canopy on thermal and radiative microclimates in a mixed temperate forest: A new statistical method to analyse hourly temporal dynamics. *Agricultural and Forest Meteorology*, 237-238:71–79, 2017. doi: 10.1016/J.AGRFORMET.2017.02.010.
- Gelman, A., Goodrich, B., Gabry, J., and Vehtari, A. R-squared for Bayesian Regression Models. *The American Statistician*, 73(3):307–309, 2019. doi: 10.1080/00031305.2018.1549100.
- Ghuman, B. S. and Lal, R. Effects of partial clearing on microclimate in a humid tropical forest. *Agricultural and Forest Meteorology*, 40(1):17–29, 1987. doi: 10.1016/0168-1923(87)90051-7.
- Gillerot, L., Forrester, D. I., Bottero, A., Rigling, A., and Lévesque, M. Tree Neighbourhood Diversity Has Negligible Effects on Drought Resilience of European Beech , Silver Fir and Norway Spruce. *Ecosystems*, 24:20–36, 2021. doi: 10.1007/s10021-020-00501-y.
- Gleason, S. M., Blackman, C. J., Cook, A. M., Laws, C. A., and Westoby, M. Whole-plant capacitance , embolism resistance and slow transpiration rates all contribute to longer desiccation
- times in woody angiosperms from arid and wet habitats. *Tree Physiology*, 34:275–284, 2014. doi: 1019 10.1093/treephys/tpu001.
- Gómez-Aparicio, L., García-Valdés, R., Ruíz-Benito, P., and Zavala, M. A. Disentangling the rel- ative importance of climate, size and competition on tree growth in Iberian forests: implications for forest management under global change. *Global Change Biology*, 17(7):2400–2414, 2011. doi: 10^{23} 10.1111/J.1365-2486.2011.02421.X.
- Gourlet-Fleury, S., Guehl, J.-M., and Laroussinie, O. E. *Ecology and management of a neotropical rainforest: lessons drawn from Paracou, a long-term experimental research site in French Guiana*. Elsevier, Paris, 2004.
- Gourlet-Fleury, S. and Houllier, F. Modelling diameter increment in a lowland evergreen rain forest in French Guiana. *Forest Ecology and Management*, 131(1-3):269–289, 6 2000. doi: $1029 \qquad 10.1016/S0378-1127(99)00212-1.$
- Greenwood, S., Ruiz-Benito, P., Martínez-Vilalta, J., Lloret, F., Kitzberger, T., Allen, C. D., Fensham, R., Laughlin, D. C., Kattge, J., Bönisch, G., Kraft, N. J., and Jump, A. S. Tree mortality across biomes is promoted by drought intensity, lower wood density and higher specific leaf area. *Ecology Letters*, 20(4):539–553, 2017. doi: 10.1111/ele.12748.
- Grossiord, C. Having the right neighbors: how tree species diversity modulates drought impacts on forests. *New Phytologist*, 228(1):42–49, 2020. doi: 10.1111/nph.15667.
- Grossiord, C., Gessler, A., Granier, A., Berger, S., Bréchet, C., Hentschel, R., Hommel, R., Scherer- Lorenzen, M., and Bonal, D. Impact of interspecific interactions on the soil water uptake depth in a young temperate mixed species plantation. *Journal of Hydrology*, 519(Part D):3511–3519, 2014a. doi: 10.1016/j.jhydrol.2014.11.011.
- Grossiord, C., Gessler, A., Granier, A., Pollastrini, M., Bussotti, F., and Bonal, D. Inter- specific competition influences the response of oak transpiration to increasing drought stress in a mixed Mediterranean forest. *Forest Ecology and Management*, 318:54–61, 2014b. doi: 1043 10.1016/j.foreco.2014.01.004.
- Grossiord, C., Granier, A., Ratcliffe, S., Bouriaud, O., Bruelheide, H., Chećko, E., Forrester, D. I., Dawud, S. M., Finér, L., Pollastrini, M., Scherer-Lorenzen, M., Valladares, F., Bonal, D., and
- Gessler, A. Tree diversity does not always improve resistance of forest ecosystems to drought.
- *Proceedings of the National Academy of Sciences of the United States of America*, 111(41):
- 14812–14815, 2014c. doi: 10.1073/pnas.1411970111.
- Grossiord, C., Buckley, T. N., Cernusak, L. A., Novick, K. A., Poulter, B., Siegwolf, R. T., Sperry,
- J. S., and McDowell, N. G. Plant responses to rising vapor pressure deficit. *New Phytologist*, 226:1550–1566, 2020. doi: 10.1111/nph.16485.
- Guan, K., Pan, M., Li, H., Wolf, A., Wu, J., Medvigy, D., Caylor, K. K., Sheffield, J., Wood,
- E. F., Malhi, Y., Liang, M., Kimball, J. S., Saleska, S. R., Berry, J., Joiner, J., and Lyapustin,
- A. I. Photosynthetic seasonality of global tropical forests constrained by hydroclimate. *Nature Geoscience 2014 8:4*, 8(4):284–289, 2015. doi: 10.1038/ngeo2382.
- Guerrieri, R., Belmecheri, S., Ollinger, S. V., Asbjornsen, H., Jennings, K., Xiao, J., Stocker, B. D., Martin, M., Hollinger, D. Y., Bracho-Garrillo, R., Clark, K., Dore, S., Kolb, T., Munger, J. W., Novick, K., and Richardson, A. D. Disentangling the role of photosynthesis and stom- atal conductance on rising forest water-use efficiency. *Proceedings of the National Academy of Sciences*, 116(34):16909–16914, 2019. doi: 10.1073/pnas.1905912116.
- Haberstroh, S. and Werner, C. The role of species interactions for forest resilience to drought. *Plant Biology*, 24(7):1098–1107, 2022. doi: 10.1111/plb.13415.
- Hafner, B. D., Hesse, B. D., and Grams, T. E. Friendly neighbours: Hydraulic redistribution accounts for one quarter of water used by neighbouring drought stressed tree saplings. *Plant, Cell & Environment*, 44(4):1243–1256, 2021. doi: 10.1111/PCE.13852.
- Hammond, W. M., Williams, A. P., Abatzoglou, J. T., Adams, H. D., Klein, T., López, R., Sáenz- Romero, C., Hartmann, H., Breshears, D. D., and Allen, C. D. Global field observations of tree die-off reveal hotter-drought fingerprint for Earth's forests. *Nature Communications*, 13(1): 1761, 2022. doi: 10.1038/s41467-022-29289-2.
- Hepworth, C., Doheny-Adams, T., Hunt, L., Cameron, D. D., and Gray, J. E. Manipulating stomatal density enhances drought tolerance without deleterious effect on nutrient uptake. *New Phytologist*, 208(2):336–341, 2015. doi: 10.1111/nph.13598.
- Hérault, B. and Piponiot, C. Key drivers of ecosystem recovery after disturbance in a neotropical forest. *Forest Ecosystems*, 5(2):1–15, 2018. doi: 10.1186/s40663-017-0126-7.
- Hérault, B., Ouallet, J., Blanc, L., Wagner, F., and Baraloto, C. Growth responses of neotropical trees to logging gaps. *Journal of Applied Ecology*, 47(4):821–831, 2010. doi: 10.1111/j.1365- 1077 2664.2010.01826.x.
- Hérault, B., Bachelot, B., Poorter, L., Rossi, V., Bongers, F., Chave, J., Paine, C. E. T., Wagner,
- F., and Baraloto, C. Functional traits shape ontogenetic growth trajectories of rain forest tree

species. *Journal of Ecology*, 99(6):1431–1440, 2011. doi: 10.1111/j.1365-2745.2011.01883.x.

- Hiltner, U., Huth, A., Hérault, B., Holtmann, A., Bräuning, A., and Fischer, R. Climate change alters the ability of neotropical forests to provide timber and sequester carbon. *Forest Ecology and Management*, 492:119166, 2021. doi: 10.1016/J.FORECO.2021.119166.
- Isbell, F., Craven, D., Connolly, J., Loreau, M., Schmid, B., Beierkuhnlein, C., Bezemer, T. M.,
- Bonin, C., Bruelheide, H., de Luca, E., Ebeling, A., Griffin, J. N., Guo, Q., Hautier, Y., Hector,
- A., Jentsch, A., Kreyling, J., Lanta, V., Manning, P., Meyer, S. T., Mori, A. S., Naeem, S.,
- Niklaus, P. A. N., Polley, H. W., Reich, P. B. R., Roscher, C., Seabloom, E. W., Smith, M. D. S.,
- Thakur, M. P., Tilman, D., Tracy, B. F., van der Putten, W. H., van Ruijven, J., Weigelt, A.,
- Weisser, W. W., Wilsey, B., and Eisenhauer, N. Biodiversity increases the resistance of ecosystem
- productivity to climate extremes. *Nature*, 526:574–577, 2015. doi: 10.1038/nature15374.
- Jarvis, P. Stomatal conductance, gaseous exchange and transpiration. In Grace, J., Ford, E. D., and Jarvis, P. G., editors, *Plants and their Atmospheric Environment*, page 419. Blackwell Sci.Publ., Oxford, 1981.
- Kay, M. tidybayes: Tidy Data and Geoms for Bayesian Models, 2022.
- Kitajima, K. and Poorter, L. Tissue‐level leaf toughness, but not lamina thickness, predicts sapling leaf lifespan and shade tolerance of tropical tree species. *New Phytologist*, 186(3):708–721, 2010. 1097 doi: 10.1111/j.1469-8137.2010.03212.x.
- Kunstler, G., Falster, D., Coomes, D. A., Hui, F., Kooyman, R. M., Laughlin, D. C., Poorter, L.,
- Vanderwel, M., Vieilledent, G., Wright, S. J., Aiba, M., Baraloto, C., Caspersen, J., Cornelissen,
- J. H. C., Gourlet-Fleury, S., Hanewinkel, M., Herault, B., Kattge, J., Kurokawa, H., Onoda, Y.,
- Peñuelas, J., Poorter, H., Uriarte, M. M., Richardson, S., Ruiz-Benito, P., Sun, I. F., Ståhl, G.,
- Swenson, N. G., Thompson, J., Westerlund, B., Wirth, C., Zavala, M. A., Zeng, H., Zimmerman,
- J. K., Zimmermann, N. E., and Westoby, M. Plant functional traits have globally consistent effects on competition. *Nature*, 529(7585):204–207, 2016. doi: 10.1038/nature16476.
- Kuusipalo, J., Jafarsidik, Y., Ådjers, G., and Tuomela, K. Population dynamics of tree seedlings in a mixed dipterocarp rainforest before and after logging and crown liberation. *Forest Ecology and Management*, 81(1-3):85–94, 1996. doi: 10.1016/0378-1127(95)03654-7.
- Lasky, J. R., Uriarte, M., Boukili, V. K., and Chazdon, R. L. Trait-mediated assembly pro- cesses predict successional changes in community diversity of tropical forests. *Proceedings of the National Academy of Sciences*, 111(15):5616–5621, 2014. doi: 10.1073/pnas.1319342111.
- Levionnois, S., Ziegler, C., Heuret, P., Jansen, S., Stahl, C., Calvet, E., Goret, J. Y., Bonal, D., and Coste, S. Is vulnerability segmentation at the leaf-stem transition a drought resistance mechanism? A theoretical test with a trait-based model for Neotropical canopy tree species. *Annals of Forest Science*, 78(4):1–16, 2021. doi: 10.1007/S13595-021-01094-9/FIGURES/4.
- Loram-Lourenço, L., Farnese, F. S., Alves, R. D. F. B., Dario, B. M. M., Martins, A. C., Aun, M. A., Batista, P. F., Silva, F. G., Cochard, H., Franco, A. C., and Menezes-Silva, P. E. Vari- ations in bark structural properties affect both water loss and carbon economics in neotropical savanna trees in the Cerrado region of Brazil. *Journal of Ecology*, 110(8):1826–1843, 2022. doi: 1119 10.1111/1365-2745.13908.
- Luo, Y., Ho, C.-L., Helliker, B. R., and Katifori, E. Leaf Water Storage and Robustness to
- Intermittent Drought: A Spatially Explicit Capacitive Model for Leaf Hydraulics. *Frontiers in Plant Science*, 12:2269, 2021. doi: 10.3389/fpls.2021.725995.
- Machado, R., Loram-Lourenço, L., Farnese, F. S., Alves, R. D. F. B., de Sousa, L. F., Silva, F. G., Filho, S. C. V., Torres-Ruiz, J. M., Cochard, H., and Menezes-Silva, P. E. Where do leaf water leaks come from? Trade-offs underlying the variability in minimum conductance across tropical savanna species with contrasting growth strategies. *New Phytologist*, 229(3):1415–1430, 2021. doi: 10.1111/nph.16941.
- Malhi, Y., Roberts, J. T., Betts, R. A., Killeen, T. J., Li, W., and Nobre, C. A. Climate Change, Deforestation, and the Fate of the Amazon. *Science*, 319(5860):169–172, 2008. doi: 10.1126/sci-1130 ence.1146961.
- Maréchaux, I., Bartlett, M. K., Sack, L., Baraloto, C., Engel, J., Joetzjer, E., and Chave, J. Drought tolerance as predicted by leaf water potential at turgor loss point varies strongly across species within an Amazonian forest. *Functional Ecology*, 29(10):1268–1277, 2015. doi: 1134 10.1111/1365-2435.12452.
- Maréchaux, I., Bartlett, M. K., Gaucher, P., Sack, L., and Chave, J. Causes of variation in leaf- level drought tolerance within an Amazonian forest. *Journal of Plant Hydraulics*, 3:e004, 2016. doi: 10.20870/jph.2016.e004.
- Maréchaux, I., Bonal, D., Bartlett, M. K., Burban, B., Coste, S., Courtois, E. A., Dulormne, M., Goret, J. Y., Mira, E., Mirabel, A., Sack, L., Stahl, C., and Chave, J. Dry-season decline in tree sapflux is correlated with leaf turgor loss point in a tropical rainforest. *Functional Ecology*, 32 (10):2285–2297, 2018. doi: 10.1111/1365-2435.13188.
- Maréchaux, I., Saint-André, L., Bartlett, M. K., Sack, L., and Chave, J. Leaf drought tolerance cannot be inferred from classic leaf traits in a tropical rainforest. *Journal of Ecology*, 108(3): 1030–1045, 2019. doi: 10.1111/1365-2745.13321.
- Marthews, T. R., Burslem, D. F., Paton, S. R., Yangüez, F., and Mullins, C. E. Soil drying in
- a tropical forest: Three distinct environments controlled by gap size. *Ecological Modelling*, 216 (3-4):369–384, 2008. doi: 10.1016/J.ECOLMODEL.2008.05.011.
- McElreath, R. *Statistical rethinking: A Bayesian course with examples in R and Stan*. CRC Press, 2020.
- McGregor, I. R., Helcoski, R., Kunert, N., Tepley, A. J., Gonzalez-Akre, E. B., Herrmann, V., Zailaa, J., Stovall, A. E. L., Bourg, N. A., McShea, W. J., Pederson, N., Sack, L., and Anderson-Teixeira, K. J. Tree height and leaf drought tolerance traits shape growth responses across droughts in a temperate broadleaf forest. *New Phytologist*, 231(2):601–616, 2021. doi: 1154 10.1111/NPH.16996.
- Mirabel, A., Hérault, B., and Marcon, E. Diverging taxonomic and functional trajectories following disturbance in a Neotropical forest. *Science of the Total Environment*, 720:137397, 2020. doi: 1157 10.1016/j.scitotenv.2020.137397.
- Mirabel, A., Marcon, E., and Hérault, B. 30 Years of postdisturbance recruitment in a Neotropical forest. *Ecology and Evolution*, 11(21):14448–14458, 2021. doi: 10.1002/ECE3.7634.
- Moreno-Gutiérrez, C., Dawson, T. E., Nicolás, E., and Querejeta, J. I. Isotopes reveal contrast- ing water use strategies among coexisting plant species in a Mediterranean ecosystem. *New Phytologist*, 196(2):489–496, 2012. doi: 10.1111/J.1469-8137.2012.04276.X.
- Nardini, A. Hard and tough: the coordination between leaf mechanical resistance and drought tolerance. *Flora*, 288:152023, 2022. doi: 10.1016/J.FLORA.2022.152023.
- Needham, J., Merow, C., Chang-Yang, C. H., Caswell, H., and McMahon, S. M. Inferring forest
- fate from demographic data: from vital rates to population dynamic models. *Proceedings of the Royal Society B: Biological Sciences*, 285(1874), 2018. doi: 10.1098/RSPB.2017.2050.
- Nunes, C. A., Berenguer, E., França, F., Ferreira, J., Lees, A. C., Louzada, J., Sayer, E. J.,
- Solar, R., Smith, C. C., Aragão, L. E. O. C., Braga, D. d. L., de Camargo, P. B., Cerri, C.
- E. P., de Oliveira, R. C., Durigan, M., Moura, N., Oliveira, V. H. F., Ribas, C., Vaz-de Mello,
- F., Vieira, I., Zanetti, R., and Barlow, J. Linking land-use and land-cover transitions to their ecological impact in the Amazon. *Proceedings of the National Academy of Sciences*, 119(27): e2202310119, 2022. doi: 10.1073/pnas.2202310119.
- O'Brien, M. J., Reynolds, G., Ong, R., and Hector, A. Resistance of tropical seedlings to drought is mediated by neighbourhood diversity. *Nature Ecology and Evolution*, 1(11):1643–1648, 2017. 1176 doi: 10.1038/s41559-017-0326-0.
- Osazuwa-Peters, O. L., Chapman, C. A., and Zanne, A. E. Selective logging: does the imprint remain on tree structure and composition after 45 years? *Conservation Physiology*, 3(1):cov012, 2015. doi: 10.1093/conphys/cov012.
- Osnas, J. L. D., Lichstein, J. W., Reich, P. B., and Pacala, S. W. Global Leaf Trait Relation- ships: Mass, Area, and the Leaf Economics Spectrum. *Science*, 340(6133):741–744, 2013. doi: 1182 10.1126/science.1231574.
- Pan, Y., Birdsey, R. A., Fang, J., Houghton, R., Kauppi, P. E., Kurz, W. A., Phillips, O. L., Shvidenko, A., Lewis, S. L., Canadell, J. G., Ciais, P., Jackson, R. B., Pacala, S. W., McGuire, A. D., Piao, S., Rautiainen, A., Sitch, S., and Hayes, D. A large and persistent carbon sink in the world's forests. *Science*, 333(6045):988–993, 2011. doi: 10.1126/SCI-1187 ENCE.1201609/SUPPL_FILE/PAPV2.PDF.
- Pardos, M., del Río, M., Pretzsch, H., Jactel, H., Bielak, K., Bravo, F., Brazaitis, G., Defossez, E., Engel, M., Godvod, K., Jacobs, K., Jansone, L., Jansons, A., Morin, X., Nothdurft, A., Oreti, L., Ponette, Q., Pach, M., Riofrío, J., Ruíz-Peinado, R., Tomao, A., Uhl, E., and Calama, R. The greater resilience of mixed forests to drought mainly depends on their composition: Analysis along a climate gradient across Europe. *Forest Ecology and Management*, 481:118687, 2021. doi: 1193 10.1016/j.foreco.2020.118687.
- Peña-Claros, M., Peters, E. M., Justiniano, M. J., Bongers, F., Blate, G. M., Fredericksen, T. S., and Putz, F. E. Regeneration of commercial tree species following silvicultural treatments
- in a moist tropical forest. *Forest Ecology and Management*, 255(3-4):1283–1293, 2008. doi: 1197 10.1016/J.FORECO.2007.10.033.
- Philipson, C. D., Dent, D. H., O'Brien, M. J., Chamagne, J., Dzulkifli, D., Nilus, R., Philips, S., Reynolds, G., Saner, P., and Hector, A. A trait-based trade-off between growth and mortality: Evidence from 15 tropical tree species using size-specific relative growth rates. *Ecology and Evolution*, 4(18):3675–3688, 2014. doi: 10.1002/ece3.1186.
- Pillay, R., Venter, M., Aragon-Osejo, J., González-del Pliego, P., Hansen, A. J., Watson, J. E., and Venter, O. Tropical forests are home to over half of the world's vertebrate species. *Frontiers in Ecology and the Environment*, 20(1):10–15, 2022. doi: 10.1002/FEE.2420.
- Piponiot, C., Sist, P., Mazzei, L., Peña-Claros, M., Putz, F. E., Rutishauser, E., Shenkin, A., Ascarrunz, N., de Azevedo, C. P., Baraloto, C., França, M., Guedes, M., Coronado, E. N., D'Oliveira, M. V., Ruschel, A. R., da Silva, K. E., Sotta, E. D., de Souza, C. R., Vidal, E., West, T. A., and Hérault, B. Carbon recovery dynamics following disturbance by selective logging in amazonian forests. *eLife*, 5:e21394, 2016. doi: 10.7554/eLife.21394.
- Poorter, L., Wright, S. J., Paz, H., Ackerly, D. D., Condit, R., Ibarra-Manríquez, G., Harms, K. E.,
- Licona, J. C., Martínez-Ramos, M., Mazer, S. J., Muller-Landau, H. C., Peña-Claros, M., Webb,
- C. O., and Wright, I. J. Are functional traits good predictors of demographic rates? Evidence from five neotropical forests. *Ecology*, 89(7):1908–1920, 2008. doi: 10.1890/07-0207.1.
- Poorter, L. and Bongers, F. Leaf traits are good predictors of plant performance across 53 rain forest species. *Ecology*, 87(7):1733–1743, 2006. doi: 10.1890/0012- 1216 9658(2006)87[1733:LTAGPO]2.0.CO;2.
- Poorter, L., McDonald, I., Alarcón, A., Fichtler, E., Licona, J.-C., Peña-Claros, M., Sterck, F., Villegas, Z., and Sass-Klaassen, U. The importance of wood traits and hydraulic conductance for the performance and life history strategies of 42 rainforest tree species. *New Phytologist*, 185 (2):481–492, 2010. doi: 10.1111/j.1469-8137.2009.03092.x.
- Poorter, L., Mcneil, A., Hurtado, V. H., Prins, H. H., and Putz, F. E. Bark traits and life-history strategies of tropical dry- and moist forest trees. *Functional Ecology*, 28(1):232–242, 2014. doi: 1223 10.1111/1365-2435.12158.
- Powell, T. L., Wheeler, J. K., de Oliveira, A. A., da Costa, A. C. L., Saleska, S. R., Meir, P., and Moorcroft, P. R. Differences in xylem and leaf hydraulic traits explain differences in drought tolerance among mature Amazon rainforest trees. *Global Change Biology*, 23(10):4280–4293, 1227 2017. doi: 10.1111/gcb.13731.
- Powers, J. S., Vargas G., G., Brodribb, T. J., Schwartz, N. B., Pérez-Aviles, D., Smith-Martin, C. M., Becknell, J. M., Aureli, F., Blanco, R., Calderón-Morales, E., Calvo-Alvarado, J. C., Calvo-Obando, A. J., Chavarría, M. M., Carvajal-Vanegas, D., Jiménez-Rodríguez, C. D., Murillo Chacon, E., Schaffner, C. M., Werden, L. K., Xu, X., and Medvigy, D. A catas- trophic tropical drought kills hydraulically vulnerable tree species. *Global Change Biology*, 26 (5):3122–3133, 2020. doi: 10.1111/GCB.15037.
- Prieto, I., Querejeta, J. I., Segrestin, J., Volaire, F., and Roumet, C. Leaf carbon and oxygen isotopes are coordinated with the leaf economics spectrum in Mediterranean rangeland species. *Functional Ecology*, 32(3):612–625, 2018. doi: 10.1111/1365-2435.13025.
- R Core Team. *R: A language and environment for statistical computing.* R Foundation for Statistical Computing, Vienna, Austria. URL https://www.R-project.org/., Vienna, Austria, 2021.
- Reich, P. B. The world-wide ' fast slow ' plant economics spectrum: A traits manifesto. *Journal of Ecology*, 102(2):275–301, 2014. doi: 10.1111/1365-2745.12211.
- Ren, J., Fang, S., Lin, G., Lin, F., Yuan, Z., Ye, J., Wang, X., Hao, Z., and Fortunel, C. Tree growth response to soil nutrients and neighborhood crowding varies between mycorrhizal types in an old-growth temperate forest. *Oecologia*, 197(2):523–535, 2021. doi: 10.1007/S00442-021- 05034-2.

- Rifai, S. W., Girardin, C. A., Berenguer, E., Del Aguila-Pasquel, J., Dahlsjö, C. A., Doughty, C. E., Jeffery, K. J., Moore, S., Oliveras, I., Riutta, T., Rowland, L. M., Murakami, A. A., Addo-Danso, S. D., Brando, P., Burton, C., Ondo, F. E., Duah-Gyamfi, A., Amézquita, F. F., Freitag, R., Pacha, F. H., Huasco, W. H., Ibrahim, F., Mbou, A. T., Mihindou, V. M., Peixoto, K. S., Rocha, W., Rossi, L. C., Seixas, M., Silva-Espejo, J. E., Abernethy, K. A., Adu-Bredu, S., Barlow, J., Da Costa, A. C., Marimon, B. S., Marimon-Junior, B. H., Meir, P., Metcalfe,
- D. B., Phillips, O. L., White, L. J., and Malhi, Y. ENSO Drives interannual variation of forest
- woody growth across the tropics. *Philosophical Transactions of the Royal Society B: Biological*
- *Sciences*, 373(1760):20170410, 2018. doi: 10.1098/rstb.2017.0410.
- Rosell, J. A., Gleason, S., Méndez-Alonzo, R., Chang, Y., and Westoby, M. Bark functional ecol- ogy: evidence for tradeoffs, functional coordination, and environment producing bark diversity. *New Phytologist*, 201(2):486–497, 2014. doi: 10.1111/NPH.12541.
- Rowland, L., Oliveira, R. S., Bittencourt, P. R., Giles, A. L., Coughlin, I., Costa, P. d. B., Domingues, T., Ferreira, L. V., Vasconcelos, S. S., Junior, J. A., Oliveira, A. A., da Costa, A. C., Meir, P., and Mencuccini, M. Plant traits controlling growth change in response to a drier climate. *New Phytologist*, 229(3):1363–1374, 2021. doi: 10.1111/NPH.16972.
- RStudio Team. RStudio: Integrated Development Environment for R, 2020.
- Sack, L., Cowan, P. D., Jaikumar, N., and Holbrook, N. M. The 'hydrology' of leaves: co-ordination of structure and function in temperate woody species. *Plant, Cell & Environment*, 26(8):1343– 1356, 2003. doi: 10.1046/J.0016-8025.2003.01058.X.
- Sanginés de Cárcer, P., Vitasse, Y., Peñuelas, J., Jassey, V. E., Buttler, A., and Signarbieux, C. Vapor–pressure deficit and extreme climatic variables limit tree growth. *Global Change Biology*, 24(3):1108–1122, 2018. doi: 10.1111/GCB.13973.
- Sapes, G. and Sala, A. Relative water content consistently predicts drought mortality risk in seedling populations with different morphology, physiology and times to death. *Plant, Cell & Environment*, 44(10):3322–3335, 2021. doi: 10.1111/PCE.14149.
- Scheidegger, Y., Saurer, M., Bahn, M., and Siegwolf, R. Linking stable oxygen and carbon isotopes with stomatal conductance and photosynthetic capacity: A conceptual model. *Oecologia*, 125 (3):350–357, 2000. doi: 10.1007/S004420000466/METRICS.
- Schindelin, J., Arganda-Carreras, I., Frise, E., Kaynig, V., Longair, M., Pietzsch, T., Preibisch,
- S., Rueden, C., Saalfeld, S., Schmid, B., Tinevez, J. Y., White, D. J., Hartenstein, V., Eliceiri,
- K., Tomancak, P., and Cardona, A. Fiji: an open-source platform for biological-image analysis.
- *Nature Methods*, 9(7):676–682, 2012. doi: 10.1038/nmeth.2019.
- Schuepp, P. H. Leaf boundary layers. *New Phytologist*, 125(3):477–507, 1993. doi: 10.1111/J.1469- 1280 8137.1993.TB03898.X.
- Serra-Maluquer, X., Gazol, A., Anderegg, W. R., Martínez-Vilalta, J., Mencuccini, M., and Ca- marero, J. J. Wood density and hydraulic traits influence species' growth response to drought across biomes. *Global Change Biology*, 28(12):3871–3882, 2022. doi: 10.1111/GCB.16123.
- Shukla, P. R., Skea, J., Slade, R., Al Khourdajie, A., Van Diemen, R., McCollum, D., Pathak, M., Some, S., Vyas, P., Fradera, R., Belkacemi, M., Hasija, A., Lisboa, G., Luz, S., Malley, J., Luz, S., and Malley, J., editors. *IPCC: Climate Change 2022: Mitigation of Climate Change Working Group III to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge University Press, Cambridge, UK and New York, NY, USA, 2022. doi: 1289 10.5194/egusphere-2022-471.
- Silva, J. N., de Carvalho, J. O., Lopes, J. d. C., de Almeida, B. F., Costa, D. H., de Oliveira, L. C., Vanclay, J. K., and Skovsgaard, J. P. Growth and yield of a tropical rain forest in the Brazilian Amazon 13 years after logging. *Forest Ecology and Management*, 71(3):267–274, 1995. doi: 10.1016/0378-1127(94)06106-S.
- Smith‐Martin, C. M., Muscarella, R., Ankori‐Karlinsky, R., Delzon, S., Farrar, S. L., Salva‐Sauri, M., Thompson, J., Zimmerman, J. K., and Uriarte, M. Hydraulic traits are not robust predictors of tree species stem growth during a severe drought in a wet tropical forest. *Functional Ecology*, 37(2):447–460, 2023. doi: 10.1111/1365-2435.14235.

 Song, Y., Sterck, F., Sass-Klaassen, U., Li, C., and Poorter, L. Growth resilience of conifer species decreases with early, long-lasting and intense droughts but cannot be explained by hydraulic traits. *Journal of Ecology*, 110(9):2088–2104, 2022. doi: 10.1111/1365-2745.13931.

Stan Development Team. CmdStanR: the R interface to CmdStan, 2022.

- Sullivan, M. J., Lewis, S. L., Affum-Baffoe, K., Castilho, C., Costa, F., Sanchez, A. C., Ewango,
- C. E., Hubau, W., Marimon, B. B. H., Monteagudo-Mendoza, A., Qie, L., Sonké, B., Martinez,
- R. V., Baker, T. R., Brienen, R. J., Feldpausch, T. R., Galbraith, D., Gloor, M., Malhi, Y.,
- 1305 Aiba, S. I., Alexiades, M. N., Almeida, E. C., De Oliveira, E. A., Dávila, E. ., Loayza, P. A.,
- Andrade, A., Vieira, S. A., Aragão, L. E., Araujo-Murakami, A., Arets, E. J., Arroyo, L.,
- Ashton, P., Gerardo Aymard, C., Baccaro, F. B., Banin, L. F., Baraloto, C., Camargo, P. B.,
- Barlow, J., Barroso, J., Bastin, J. F., Batterman, S. A., Beeckman, H., Begne, S. K., Bennett,
- A. C., Berenguer, E., Berry, N., Blanc, L., Boeckx, P., Bogaert, J., Bonal, D., Bongers, F.,
- Bradford, M., Brearley, F. Q., Brncic, T., Brown, F., Burban, B., Camargo, J. L., Castro, W.,
- Céron, C., Ribeiro, S. C., Moscoso, V. C., Chave, J., Chezeaux, E., Clark, C. J., De Souza, F. C.,
- Collins, M., Comiskey, J. A., Valverde, F. C., Medina, M. C., Da Costa, L., Dančsák, M., Dargie,
- G. C., Davies, S., Cardozo, N. D., De Haulleville, T., De Medeiros, M. B., Del Aguila Pasquel, J.,
- Derroire, G., Di Fiore, A., Doucet, J. L., Dourdain, A., Droissart, V., Duque, L. F., Ekoungoulou,
- R., Elias, F., Erwin, T., Esquivel-Muelbert, A., Fauset, S., Ferreira, J., Llampazo, G. F., Foli,
- E., Ford, A., Gilpin, M., Hall, J. S., Hamer, K. C., Hamilton, A. C., Harris, D. J., Hart, T. B.,
- Hédl, R., Herault, B., Herrera, R., Higuchi, N., Hladik, A., Coronado, E. H., Huamantupa-
- Chuquimaco, I., Huasco, W. H., Jeffery, K. J., Jimenez-Rojas, E., Kalamandeen, M., Djuikouo,
- M. N. K., Kearsley, E., Umetsu, R. K., Kho, L. K., Killeen, T., Kitayama, K., Klitgaard, B.,
- Koch, A., Labrière, N., Laurance, W., Laurance, S., Leal, M. E., Levesley, A., Lima, A. J.,
- Lisingo, J., Lopes, A. P., Lopez-Gonzalez, G., Lovejoy, T., Lovett, J. C., Lowe, R., Magnusson,
- 1322 W. E., Malumbres-Olarte, J., Manzatto, G., Marimon, B. B. H., Marshall, A. R., Marthews,
- T., De Almeida Reis, S. M., Maycock, C., Melgaço, K., Mendoza, C., Metali, F., Mihindou,
- V., Milliken, W., Mitchard, E. T., Morandi, P. S., Mossman, H. L., Nagy, L., Nascimento,
- H., Neill, D., Nilus, R., Vargas, P. N., Palacios, W., Camacho, N. P., Peacock, J., Pendry, C.,
- Mora, M. C. P., Pickavance, G. C., Pipoly, J., Pitman, N., Playfair, M., Poorter, L., Poulsen, J. R., Poulsen, A. D., Preziosi, R., Prieto, A., Primack, R. B., Ramírez-Angulo, H., Reitsma, J., Réjou-Méchain, M., Correa, Z. R., De Sousa, T. R., Bayona, L. R., Roopsind, A., Rudas, A., Rutishauser, E., Salim, K. A., Salomão, R. P., Schietti, J., Sheil, D., Silva, R. C., Espejo, J. S., Valeria, C. S., Silveira, M., Simo-Droissart, M., Simon, M. F., Singh, J., Shareva, Y. C. S., Stahl, C., Stropp, J., Sukri, R., Sunderland, T., Svátek, M., Swaine, M. D., Swamy, V., Taedoumg, H., Talbot, J., Taplin, J., Taylor, D., Ter Steege, H., Terborgh, J., Thomas, R., Thomas, S. C., Torres-Lezama, A., Umunay, P., Gamarra, L. V., Van Der Heijden, G., Van Der Hout, P., Van Der Meer, P., Van Nieuwstadt, M., Verbeeck, H., Vernimmen, R., Vicentini, A., Vieira, I. C. G., Torre, E. V., Vleminckx, J., Vos, V., Wang, O., White, L. J., Willcock, S., Woods, J. T., Wortel, V., Young, K., Zagt, R., Zemagho, L., Zuidema, P. A., Zwerts, J. A., and Phillips, O. L. Long-term thermal sensitivity of earth's tropical forests. *Science*, 368(6493): 869–874, 5 2020. doi: 10.1126/science.aaw7578.
- Uriarte, M., Canham, C. D., Thompson, J., and Zimmerman, J. K. A neighborhood analysis of tree growth and survival in a hurricane-driven tropical forest. *Ecological Monographs*, 74(4): 591–614, 2004. doi: 10.1890/03-4031.
- Uriarte, M., Swenson, N. G., Chazdon, R. L., Comita, L. S., John Kress, W., Erickson, D., Forero- Montaña, J., Zimmerman, J. K., and Thompson, J. Trait similarity, shared ancestry and the structure of neighbourhood interactions in a subtropical wet forest: implications for community assembly. *Ecology Letters*, 13:1503–1514, 2010. doi: 10.1111/j.1461-0248.2010.01541.x.
- Uriarte, M., Lasky, J. R., Boukili, V. K., and Chazdon, R. L. A trait-mediated, neighbourhood approach to quantify climate impacts on successional dynamics of tropical rainforests. *Functional Ecology*, 30(2):157–167, 2016a. doi: 10.1111/1365-2435.12576.
- Uriarte, M., Schwartz, N., Powers, J. S., Marín-Spiotta, E., Liao, W., and Werden, L. K. Impacts of climate variability on tree demography in second growth tropical forests: the importance of regional context for predicting successional trajectories. *Biotropica*, 48(6):780–797, 2016b. doi: 1352 10.1111/btp.12380.
- Vaieretti, M. V., Díaz, S., Vile, D., and Garnier, E. Two measurement methods of leaf dry matter content produce similar results in a broad range of species. *Annals of Botany*, 99(5):955–958, 2007. doi: 10.1093/aob/mcm022.
- Vancutsem, C., Achard, F., Pekel, J.-F., Vieilledent, G., Carboni, S., Simonetti, D., Gallego, J., Aragão, L. E. O. C., and Nasi, R. Long-term (1990–2019) monitoring of forest cover changes in the humid tropics. *Science Advances*, 7(10), 2021. doi: 10.1126/sciadv.abe1603.
- Vendramini, F., Díaz, S., Gurvich, D. E., Wilson, P. J., Thompson, K., and Hodgson, J. G. Leaf traits as indicators of resource-use strategy in floras with succulent species. *New Phytologist*, 154(1):147–157, 2002. doi: 10.1046/J.1469-8137.2002.00357.X.
- Vile, D., Garnier, ff., Shipley, B., Laurent, G., Navas, M. L., Roumet, C., Lavorel, S., Díaz, S.,
- Hodgson, J. G., Lloret, F., Midgley, G. F., Poorter, H., Rutherford, M. C., Wilson, P. J., and
- 1364 Wright, I. J. Specific Leaf Area and Dry Matter Content Estimate Thickness in Laminar Leaves. *Annals of Botany*, 96(6):1129–1136, 11 2005. doi: 10.1093/aob/mci264.
- Vleminckx, J., Fortunel, C., Valverde-Barrantes, O., Timothy Paine, C. E., Engel, J., Petronelli, P., Dourdain, A. K., Guevara, J., Béroujon, S., and Baraloto, C. Resolving whole-plant economics from leaf, stem and root traits of 1467 Amazonian tree species. *Oikos*, 130(7):1193–1208, 2021. doi: 10.1111/OIK.08284.
- Voleníková, M. and Tichá, I. Insertion Profiles in Stomatal Density and Sizes in Nicotiana Tabacum L. Plantlets. *Biologia Plantarum*, 44:161–165, 2001. doi: 10.1023/A:1017982619635.
- Von Arx, G., Dobbertin, M., and Rebetez, M. Spatio-temporal effects of forest canopy on under- story microclimate in a long-term experiment in Switzerland. *Agricultural and Forest Meteorol-ogy*, 166-167:144–155, 2012. doi: 10.1016/j.agrformet.2012.07.018.
- Wagner, F., Rossi, V., Stahl, C., Bonal, D., and Hérault, B. Water Availability Is the Main Climate Driver of Neotropical Tree Growth. *PLoS ONE*, 7(4):e34074, 2012. doi: 10.1371/jour-1377 nal.pone.0034074.
- Wagner, F., Rossi, V., Aubry-Kientz, M., Bonal, D., Dalitz, H., Gliniars, R., Stahl, C., Trabucco, A., and Hérault, B. Pan-Tropical Analysis of Climate Effects on Seasonal Tree Growth. *PLoS ONE*, 9(3):e92337, 2014a. doi: 10.1371/journal.pone.0092337.
- Wagner, F., Rossi, V., Baraloto, C., Bonal, D., Stahl, C., and Hérault, B. Are commonly measured functional traits involved in tropical tree responses to climate? *International Journal of Ecology*, 2014:389409, 2014b. doi: 10.1155/2014/389409.
- Wagner, F. H., Hérault, B., Bonal, D., Stahl, C., Anderson, L. O., Baker, T. R., Sebastian Becker, G., Beeckman, H., Boanerges Souza, D., Cesar Botosso, P., Bowman, D. M., Bräuning, A., Brede, B., Irving Brown, F., Julio Camarero, J., Camargo, P. B., Cardoso, F. C., Carvalho, F. A., Castro, W., Koloski Chagas, R., Chave, J., Chidumayo, E. N., Clark, D. A., Regina Capel- lotto Costa, F., Couralet, C., Henrique Da Silva Mauricio, P., Dalitz, H., Resende De Castro, V., Milani, J. E. D. F., Consuelo De Oliveira, E., De Souza Arruda, L., Devineau, J. L., Drew, D. M., Dünisch, O., Durigan, G., Elifuraha, E., Fedele, M., Ferreira Fedele, L., Figueiredo Filho, A., Finger, C. A. G., César Franco, A., Jnior, L. F., Galvão, F., Gebrekirstos, A., Gliniars, R., Maurício Lima De Alencastro Graça, P., Griffiths, A. D., Grogan, J., Guan, K., Homeier, J., Raquel Kanieski, M., Khoon Kho, L., Koenig, J., Valerio Kohler, S., Krepkowski, J., Lemos- Filho, J. P., Lieberman, D., Eugene Lieberman, M., Sergio Lisi, C., Longhi Santos, T., Ayala, J. L. L., Eijji Maeda, E., Malhi, Y., Maria, V. R., Marques, M. C., Marques, R., Maza Chamba, H., Mbwambo, L., Liana Lisboa Melgaço, K., Angela Mendivelso, H., Murphy, B. P., O'Brien, J. J., F. Oberbauer, S., Okada, N., Plissier, R., Prior, L. D., Alejandro Roig, F., Ross, M., Ro- drigo Rossatto, D., Rossi, V., Rowland, L., Rutishauser, E., Santana, H., Schulze, M., Selhorst, D., Rodrigues Silva, W., Silveira, M., Spannl, S., Swaine, M. D., Toledo, J. J., Miranda Toledo, M., Toledo, M., Toma, T., Tomazello Filho, M., Ignacio Valdez Hernández, J., Verbesselt, J., Aparecida Vieira, S., Vincent, G., Volkmer De Castilho, C., Volland, F., Worbes, M., Lea Bolzan Zanon, M., and Aragão, L. E. Climate seasonality limits leaf carbon assimilation and wood productivity in tropical forests. *Biogeosciences*, 13(8):2537–2562, 2016. doi: 10.5194/BG-

1404 13-2537-2016.

- Wang, J. and Wen, X. Divergence and conservative of stomatal conductance in coexisting species in response to climatic stress in Tibetan Plateau. *Ecological Indicators*, 138:108843, 2022a. doi: 1407 10.1016/J.ECOLIND.2022.108843.
- Wang, J. and Wen, X. Limiting resource and leaf functional traits jointly determine distribution patterns of leaf intrinsic water use efficiency along aridity gradients. *Frontiers in Plant Science*, 13:2559, 2022b. doi: 10.3389/fpls.2022.909603.
- Way, D. A. and Oren, R. Differential responses to changes in growth temperature between trees from different functional groups and biomes : a review and synthesis of data. *Tree Physiology*, 30:669–688, 2010. doi: 10.1093/treephys/tpq015.
- Whitmore, T. C. Canopy Gaps and the Two Major Groups of Forest Trees. *Ecology*, 70(3):536–538, 1989. doi: 10.2307/1940195.
- Wickham, H., Averick, M., Bryan, J., Chang, W., McGowan, L. D., François, R., Grolemund, G., Hayes, A., Henry, L., Hester, J., Kuhn, M., Pedersen, T. L., Miller, E., Bache, S. M., Müller, K., Ooms, J., Robinson, D., Seidel, D. P., Spinu, V., Takahashi, K., Vaughan, D., Wilke, C., Woo, K., and Yutani, H. Welcome to the tidyverse. *Journal of Open Source Software*, 4(43): 1686, 2019. doi: 10.21105/joss.01686.
- Wong, S. C., Cowan, I. R., and Farquhar, G. D. Stomatal conductance correlates with photosyn-thetic capacity. *Nature*, 282(5737):424–426, 1979. doi: 10.1038/282424a0.
- Wright, I. J., Reich, P. B., Westoby, M., Ackerly, D. D., Baruch, Z., Bongers, F., Cavender-Bares,
- J., Chapin, T., Cornellssen, J. H., Diemer, M., Flexas, J., Garnier, E., Groom, P. K., Gulias, J.,
- Hikosaka, K., Lamont, B. B., Lee, T., Lee, W., Lusk, C., Midgley, J. J., Navas, M. L., Niinemets,
- ff., Oleksyn, J., Osada, H., Poorter, H., Pool, P., Prior, L., Pyankov, V. I., Roumet, C., Thomas,
- S. C., Tjoelker, M. G., Veneklaas, E. J., and Villar, R. The worldwide leaf economics spectrum.
- *Nature*, 428:821–827, 2004. doi: 10.1038/nature02403.
- Wright, I. J., Dong, N., Maire, V., Prentice, I. C., Westoby, M., Díaz, S., Gallagher, R. V., Jacobs,
- B. F., Kooyman, R., Law, E. A., Leishman, M. R., Niinemets, ff., Reich, P. B., Sack, L., Villar,
- R., Wang, H., and Wilf, P. Global climatic drivers of leaf size. *Science*, 357(6354):917–921, 2017. doi: 10.1126/science.aal4760.
- Wright, S. J., Kitajima, K., Kraft, N. J. B., Reich, P. B., Wright, I. J., Bunker, D. E., Condit, R., Dalling, J. W., Davies, S. J., DíAz, S., Engelbrecht, B. M. J., Harms, K. E., Hubbell, S. P., Marks, C. O., Ruiz-Jaen, M. C., Salvador, C. M., and Zanne, A. E. Functional traits and the growth-mortality trade-off in tropical trees. *Ecology*, 91(12):3664–74, 2010. doi: 10.1890/09- 2335.1.
- Yue, Q., Geng, Y., von Gadow, K., Fan, C., Zhang, C., and Zhao, X. Effects of neighborhood inter- action on tree growth in a temperate forest following selection harvesting. *Ecological Indicators*, 136:108663, 2022. doi: 10.1016/J.ECOLIND.2022.108663.
- Zambrano, J., Marchand, P., and Swenson, N. G. Local neighbourhood and regional climatic contexts interact to explain tree performance. *Proceedings of the Royal Society B: Biological Sciences*, 284:20170523, 2017. doi: 10.1098/rspb.2017.0523.
- Zambrano, J., Beckman, N. G., Marchand, P., Thompson, J., Uriarte, M., Zimmerman, J. K., Umaña, M. N., and Swenson, N. G. The scale dependency of trait-based tree neighborhood models. *Journal of Vegetation Science*, 31:581–593, 2020. doi: 10.1111/jvs.12880.
- Ziegler, C., Coste, S., Stahl, C., Delzon, S., Levionnois, S., Cazal, J., Cochard, H., Esquivel- Muelbert, A., Goret, J. Y., Heuret, P., Jaouen, G., Santiago, L. S., and Bonal, D. Large hydraulic safety margins protect Neotropical canopy rainforest tree species against hydraulic failure during drought. *Annals of Forest Science*, 76:115, 2019. doi: 10.1007/s13595-019-0905-0.
- Zuidema, P. A., Babst, F., Groenendijk, P., Trouet, V., Abiyu, A., Acuña-Soto, R., Adenesky-
- Filho, E., Alfaro-Sánchez, R., Aragão, J. R. V., Assis-Pereira, G., Bai, X., Barbosa, A. C.,
- Battipaglia, G., Beeckman, H., Botosso, P. C., Bradley, T., Bräuning, A., Brienen, R., Buckley,
- B. M., Camarero, J. J., Carvalho, A., Ceccantini, G., Centeno-Erguera, L. R., Cerano-Paredes,
- J., Chávez-Durán, ff. A., Cintra, B. B. L., Cleaveland, M. K., Couralet, C., D'Arrigo, R., del
- Valle, J. I., Dünisch, O., Enquist, B. J., Esemann-Quadros, K., Eshetu, Z., Fan, Z. X., Ferrero,
- M. E., Fichtler, E., Fontana, C., Francisco, K. S., Gebrekirstos, A., Gloor, E., Granato-Souza,
- D., Haneca, K., Harley, G. L., Heinrich, I., Helle, G., Inga, J. G., Islam, M., Jiang, Y. m.,
- Kaib, M., Khamisi, Z. H., Koprowski, M., Kruijt, B., Layme, E., Leemans, R., Leffler, A. J.,
- Lisi, C. S., Loader, N. J., Locosselli, G. M., Lopez, L., López-Hernández, M. I., Lousada, J.
- L. P. C., Mendivelso, H. A., Mokria, M., Montóia, V. R., Moors, E., Nabais, C., Ngoma, J.,
- Nogueira Júnior, F. d. C., Oliveira, J. M., Olmedo, G. M., Pagotto, M. A., Panthi, S., Pérez-
- De-Lis, G., Pucha-Cofrep, D., Pumijumnong, N., Rahman, M., Ramirez, J. A., Requena-Rojas,
- E. J., Ribeiro, A. d. S., Robertson, I., Roig, F. A., Rubio-Camacho, E. A., Sass-Klaassen, U.,
- Schöngart, J., Sheppard, P. R., Slotta, F., Speer, J. H., Therrell, M. D., Toirambe, B., Tomazello-
- Filho, M., Torbenson, M. C., Touchan, R., Venegas-González, A., Villalba, R., Villanueva-Diaz,
- J., Vinya, R., Vlam, M., Wils, T., and Zhou, Z. K. Tropical tree growth driven by dry-season
- climate variability. *Nature Geoscience*, 15:269–276, 2022. doi: 10.1038/S41561-022-00911-8.

List of Supporting Information

See PDF file SupportingInformation_1.pdf

Figure S1. M1-model standardised regression coefficients of community level intercepts, effect of

DBH and M2-model trait effects on intercepts

 Figure S2. M1-CWD model standardised regression coefficients of species-level intercepts in undisturbed and disturbed plots.

- **Figure S3.** Bi-plot of principal component analysis (PCA) on climate variables.
- **Supplementary Methods S1.** Corrections of tree inventory data.
- **Supplementary Methods S2.** R code for M1- and M2-models.

- See Excel file SupportingInformation_2.xlsx
- **Table S1.** Conditional and margignal R2 estimates for M1- and M2-models
- **Table S2.** Standardised regression coefficients of community level parameters and group-level sigmas for M1-models.
- **Table S3.** Summarised species-level responses to model covariates for M1-models.
- **Table S4.** Estimates of correlations (*ρ*) between species-level parameters for M1-models.
- **Table S5.** Standardised regression coefficients of community level parameters and group-level sigmas for M2-models.
- **Table S6.** Pearson correlation coefficients among mean anomalies of the 4 climate variables.