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4 **Assessing agri-environmental strategies to reduce pesticide concentrations in surface**
5 **drinking water sources, Coastal Charente River basin, SW France**

6

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19

20 **Abstract**

21 Most surface water bodies are deteriorating, resulting in health risks and
22 environmental damage. This study aimed to assess the performance of the modeled agri-
23 environmental transition scenarios for restoring freshwater quality within the protection
24 perimeter of strategic surface water drinking water catchments. Realistic agri-environmental
25 scenarios with combined agronomic orientations were tested using a well-calibrated
26 ecohydrological model. Their performance was quantified based on pesticide concentration
27 reductions in freshwater bodies, focusing on temporal and spatial variability, using non-
28 parametric tests. Our results show that it is possible to differentiate cropping mitigation
29 scenarios according to their ability to limit pesticide transfers, taking into account the most
30 influential biogeochemical factors in the catchment, their agronomic orientation, and their
31 progressive implementation. Our study is also applicable to other contexts.

32 *Keywords: agriculture; drinking water; pesticide; scenarios; watershed; water*
33 *source protection*

34 **1. Introduction**

35 In Europe, despite decades of directives and policies aimed at protecting water
36 resources (EC, 2000), and notably the quality of water intended for human consumption (EC,
37 1998, recast with EC, 2020), and the sustainable use of pesticides (EC, 2009), approximately
38 60% of surface water bodies exhibit poor ecological and chemical conditions (István, 2020).
39 This deterioration not only impacts ecosystems long-term after their ban (Veselá, et al., 2020)
40 but also poses risks to drinking water abstraction points, thereby endangering human health
41 (Budzinski and Couderchet, 2018; Baldi et al., 2022).

42 Effective water quality restoration requires comprehensive action programs
43 addressing diffuse pollution, incorporating holistic farm-level interventions and broader
44 watershed management strategies (Bieroza et al., 2021), and collectively involving territorial

45 stakeholders and water users and managers (Amblard, 2019). However, because financial
46 resources are limited, water managers target the most contaminated areas, identified by water
47 development and management plans as "Priority Areas" (PAs), with agri-environmental and
48 climate measures. Despite these efforts, persistent diffuse pollution continues to affect
49 drinking water sources (Amblard, 2019). In France, around 500 catchments affected by
50 diffuse agricultural pollution have been classified as 'Grenelle' strategic catchments as part of
51 the Grenelle de l'Environnement initiative launched in 2007 to prioritize actions to protect
52 drinking water. This number has since increased to an estimated 3000 catchments (Amblard,
53 2019).

54 Decision-support tools are needed to characterize the spatiotemporal dynamics of
55 agricultural diffuse pollution and improve water quality governance at the catchment scale.
56 Identifying critical source areas (CSAs) (Giri et al., 2016) and understanding the temporal
57 variability in pollution levels are essential for the efficient implementation of diffuse
58 pollution mitigation strategies.

59 Ecohydrological modeling simulates the agro-ecosystems and agricultural practices
60 associated with soils and plants, soils and climate heterogeneity, and the non-linearity of
61 biogeochemical watershed processes responsible for pollutant fate and transport, as well as
62 hydrosystem response to land use and climate change (Kohne et al., 2009; Porporato et al.,
63 2015; Veselá et al., 2020). Thus, it helps improve long-term watershed management
64 (Strehmel et al., 2016). However, for enhanced decision support and considering the
65 complexity of scenarios and biophysical variations in time and space, further analysis is
66 required to discriminate the respective contributions of anthropogenic and biophysical
67 factors, in time and space and evaluate the statistical efficiencies of the modeled scenarios.

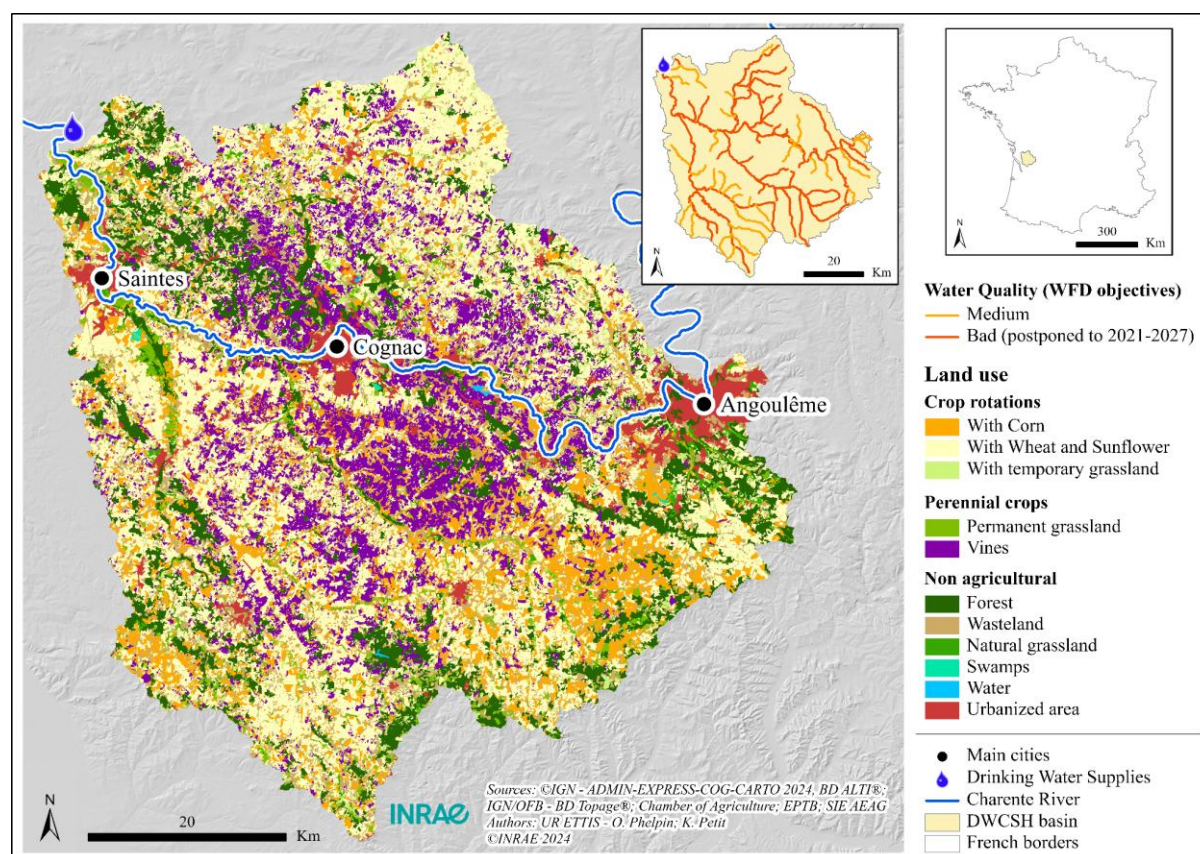
68 This study focusses on a 3,800 km² wide priority surface water catchment area,
69 endangered by pesticide diffuse pollution, and aims to shed light on the anthropogenic and

70 biophysical factors that significantly contribute to the reduction of plant protection product
71 (PPP) pollution in freshwater using statistical methods on the ecohydrological model output
72 of the tested mitigation scenarios. We aim to differentiate these scenarios based on their
73 ability to limit pesticide transfers by considering the most influential biogeochemical factors
74 in the catchment, their agronomic orientation, and their progressive implementation in time
75 and space. By applying indirect methods through a well-calibrated ecohydrological model
76 (SWAT) and using statistical analysis of its outputs, this study evaluates mitigation strategies
77 for improving water quality. The findings provide insights applicable to other contexts and
78 contribute to enhancing the overall approach to water quality management in agricultural
79 landscapes.

80 **2. Materials and Methods**

81 **2.1. Study area**

82 The Drinking Water Coulonge Saint-Hippolyte (DWSCH) basin, situated in
83 southwest France within the Charente basin ($0^{\circ}45'W$ $45^{\circ}58'N$ / $0^{\circ}16'E$ $45^{\circ}14'N$), serves as a
84 critical area or protection area for surface drinking water abstraction sources, spanning 3,800
85 km² (Fig. 1). The Charente River extends 381 km from east to west and is the primary
86 drainage system in the basin, ultimately discharging into the Atlantic Ocean. With an average
87 flow rate of $95 \text{ m}^3 \text{ s}^{-1}$, the river discharge varies from a few $\text{m}^3 \text{ s}^{-1}$ to $815 \text{ m}^3 \text{ s}^{-1}$ at the
88 furthest downstream gauging station, located approximately 40 km upstream from the Ocean
89 estuary, where tidal influences become significant. The main outlet of the DWSCH basin is
90 located about 30 km from the coast, however, due to the meandering course of the river, it
91 can be up to 40 km upstream from the estuary. While tidal effects are more pronounced
92 closer to the estuary, they exert less influence within the DWSCH basin. The alluvial beds of
93 the river exhibit low density, and dams, canals, and reservoirs throughout the basin
94 significantly influence the natural hydrological regime (Fig. 1).



95
96 Figure 1: Map of the study area.

97 The DWCSH basin's water supply area is considered strategic for several reasons: it
 98 produces 170 millions $\text{m}^3 \text{y}^{-1}$ to serve 290,000 residents, with this number increasing to over
 99 500,000 during the summer months; it is highly vulnerable to pesticides, particularly
 100 herbicides, excessive concentrations of which are regularly observed despite action plans;
 101 and protection of this threatened resource is required by the Water Framework Directive
 102 (WFD) (EC, 2000, EC, 2008). To comply these requirements, local water managers aim to
 103 reduce pesticide peak concentrations and maintain levels below $0.05 \mu\text{g/L}$ for each pesticide,
 104 which is set as half of the maximum acceptable concentration (MAC) outlined in the EU
 105 regulations. To comply with the Water Framework Directive (WFD) and other EU
 106 regulations, local water managers aim to reduce pesticide peak concentrations and maintain
 107 levels below $0.05 \mu\text{g/L}$ for each pesticide, which corresponds to half of the maximum
 108 acceptable concentration (MAC) outlined in the EU regulations.

109 The climate is generally temperate, with an average rainfall of 800 mm yr⁻¹ over the
110 last 30 years and an average temperature of 13°C. However, variability exists: the climate is
111 mild and dry and sunny in the center (Cognac area), oceanic on the coast, and wetter and less
112 mild in the eastern part of the plateau. The topography is mostly flat, with a minimum
113 elevation of -1 m to 0 m a.s.l. in the west and a maximum elevation of 500 m a.s.l. in the
114 East. Agriculture is the predominant activity in the area, significantly impacting pesticide
115 concentrations. The most frequent exceedances were over 40% for metaldehyde and
116 mancozeb, over 25% for glyphosate and 35% for its metabolite (Source: SIE AEAG).

117 **2.2. Reference scenario**

118 The "reference scenario" serves as the baseline against which all mitigation scenarios
119 are evaluated. It is important to note that this scenario is not the current state of the catchment
120 but rather a baseline established in 2017 with the expertise and consensus of various
121 stakeholders, including agricultural experts, environmental scientists, and local water
122 managers (Vernier, 2017). The reference scenario reflects the agricultural practices, land use,
123 and management strategies that were prevalent at that time. In this scenario, the dominant
124 land use consists mainly of intensive conventional agricultural practices. Land use patterns
125 vary with soil characteristics and agricultural goals. Unlike many mesoscale crop modeling
126 studies that cover only 2-6 crop types, our study simulates 41 distinct cropping systems (Fig.
127 1).

128 These systems are associated with specific agricultural practices related to eight soil
129 features (Table 2) and encompass 17 irrigated and unirrigated crops. The crops grown in the
130 region include: i) Traditional crops: sunflower, soft winter wheat, durum wheat, oilseed rape,
131 grain corn, maize corn, winter barley, spring barley, and temporary grassland. ii) Innovative
132 and organic options: alfalfa, field pea, temporary pasture, soybean, clover, and triticale.

133 **Table 1** *Agricultural Soil Types Table*

Common name, cover age (%), FAO Types	Description	Agronomic Properties & Use	Vulnerabilities
Groies, 36%. Calcosols, rendosols.	Calcareous clay-loam soils ; differentiated into shallow (GrSup) that are stony, low water-holding capacity; and silty clay soils over marl or limestone, non-hydromorphic, irregular stone content, with local clayey hydromorphic zones (GrMar).	Fertile, with GrSup requiring irrigation; and suitable for polyculture; requires irrigation for corn due to drought; GrMar suitable for cereals, easy to cultivate due to its well-drained structure.	Moderate runoff in summer for GrSup; low runoff throughout the year for GrMar.
Campanian soils, 33%. Calcisols, calcosols, rendosols..	Clayey-calcareous soils, gray to black in color, overlying chalk, marl, or soft limestone, differentiated into shallow (ChampS) with high calcium carbonate content from the surface; Deeper soils (ChampP) have greater water capacity, benefiting from capillary rise.	High potential for ChampS used in dry farming, and vineyards; ChampP are very suitable for corn and mixed crops.	High runoff potential, especially in winter, due to low surface infiltration, especially in winter.
Doucins, 16% Planosols, neoluvisols, brunisols, calcisols, calcosols.	Mixed clayey, loamy, sandy soils, acidic, with low water-holding capacity, subdivided into hydromorphic sandy-loamy (Dsab) with low fertility and sandy-clayey (Dmar), often waterlogged.	Soil pH management and structure improvement required. Dsab moderately suitable for cereals, requiring irrigation, Dmar poor to cultivate, requiring minimum to no tillage, in dry conditions.	Vulnerable to surface runoff, desiccation and leaching (Dsab); strong runoff and waterlogging (Dmar).
Pays Bas, 7%. Calcosols, rendosols.	Very heavy clay soils (up to 60% clay content) with poor drainage located in the depressions in northern Cognac (Tpbas).	Limited agricultural use; best suited to specific crop rotations with adapted management practices.	High risk of waterlogging and surface runoff, especially in winter; slow infiltration..
Valleys and terraces 7%. Fluvisols, histosols, calcosols.	Alluvial soils ranging from sandy to loamy textures with variable hydromorphology, in valleys (VaMt).	Suitable for pasture; some cereal farming in well-drained areas.	Moderate runoff risk of winter runoff due to fluctuating water levels; low risk of water saturation and seasonal flooding.

134 The soils are predominantly sedimentary, varied in composition and generally thin,

135 with calcareous and marly characteristics shaped by the local geology. The main soil types

136 include calcosols, calcisols and rendosols, which largely determine the agronomic potential
137 of the area (Table 1).

138 Representative herbicides such as glyphosate, aclonifen, MCPA, and tebuconazole,
139 along with fungicides like mancozeb and the molluscicide metaldehyde, were studied. These
140 pesticides were selected based on their frequent detection in freshwater environments and
141 their extensive use in areas dedicated to field crop and vine cultivation. The selection criteria
142 were based on current agricultural practices' prevalence, environmental impact, and
143 relevance (Vernier, 2017).

144 The widespread use of these pesticides in agriculture is a significant concern due to
145 their potential genotoxic effects on both the environment and human health. Glyphosate, the
146 most widely used herbicide, causes DNA damage and oxidative stress (Lebailly et al., 1998).
147 Mancozeb has been associated with genotoxicity for over 40 years (Perocco et al., 1989),
148 while tebuconazole has demonstrated similar effects (Andrioli et al., 2023). Aclonifen is
149 known to cause developmental abnormalities (Lee et al., 2021; Park et al., 2022), and the
150 molluscicide metaldehyde poses significant risks to aquatic life (Macar et al., 2023). The
151 increased use of the herbicide MCPA is also of concern.

152 Exposure to a mixture of plant protection products (PPPs) increases the risk of
153 genotoxicity (Dhananjayan et al., 2019). Although the EU banned the sale of mancozeb and
154 metaldehyde by the EU in 2020 and 2021, respectively, these substances are still used as anti-
155 mildew and anti-slug agents and continue to be detected in streamflows within the DWSCH
156 basin.

157 The persistent presence of these pesticides in excess underscores the pressing need for
158 effective management strategies to mitigate their impact on drinking water sources.

159 **2.3. The Charente SWAT model**

160 The ecohydrological Soil and Water Assessment Tool (SWAT) was utilized to
161 analyze the effects of agri-environmental mitigation strategies aimed at reducing pesticide
162 contamination in surface waters and assessing long-term impacts on ecosystems in mesoscale
163 and large-scale watersheds (Arnold et al., 2007). SWAT is recognized as the most suitable
164 tool for simulating pesticide transport and fate in agricultural watersheds due to its
165 comprehensive capabilities and ability to handle complex land-use change (LUC) scenarios
166 (Wang et al., 2019; Kohne et al., 2009) through interfaces such as GenLU2 and Best
167 Management Practices (BMPs) like Vegetative Filter Strips (VFSs) (White and Arnold,
168 2009).

169 The GenLU2-SWAT model was extensively calibrated and validated using observed
170 streamflow and pesticide concentration data from the DWSCH basin, previously collected
171 and referenced (Pryet et al., 2016), (Vernier et al. 2017). These data, sourced from
172 institutional datasets (SIE AEAG and Météo-France), ensured accurate simulation of
173 hydrological processes and contaminant transport. The calibration process included both
174 streamflow measurements and pesticide concentration data, allowing the model to reliably
175 simulate pollutant transfers within the hydrosystem at four available gauging stations and
176 over 15 water quality stations. Although no new field measurements of pesticide
177 concentrations were collected for this study, the previously validated SWAT model was used
178 to simulate contaminant transfers and hydrosystem responses to land use changes (LUCs) and
179 agronomic interventions, accros the 106 modeled sub-basins. The statistical analyses
180 employed in this study (Kruskal-Wallis, Dunn's post hoc, Mann-Kendall, and PCA) are
181 designed to explore spatio-temporal trends and relationships among biophysical factors based

182 on model outputs. This indirect modeling approach provides insight into how these
 183 interventions affect water quality and pollutant transfer to surface waters.

184 2.4. The tested mitigation scenarios

185 The mitigation scenarios were developed to explore the impact of various
 186 agricultural practices and land-use changes (LUCs) on environmental quality, considering
 187 stakeholder interests (Table 2).

188 **Table 2** *Description and objectives of the tested mitigation scenarios*

Scenario	Description & LUC (ha; % Total Area; % UAA)	Objectives
sRef	Reference scenario (2558; 70; 100)	Assess model performance, detect sensitive parameters.
Depp	Conversion of mixed/temporary grasslands to crops (33; 0.9; 1.3)	Model decreased livestock activity.
Sdci1	Full-basin innovative crop systems (Sdci) (2558; 70; 100)	Assess effects of large-scale Sdci practices.
Sdci2	Sdci in Protected Areas (PAs) (1353; 37; 53)	Test implementation within Pas.
Sdci3	Low-input field crops (1949; 53; 76)	Isolate impacts of changes in field crops.
Sdci4	Low-input vineyards (566; 15; 22)	Isolate impacts of changes in vineyards.
Sdci5	Intermediate Sdci in PAs (503; 14; 20)	Test intermediate Sdci in PAs.
Sdci6	20% Sdci, 15% Organic (503; 14; 20)	Test combined practices.
Org1	Double organic area (23; 0.6; 0.9)	Test growth in existing and marginal areas.
Org2	5-fold organic area increase (93; 2.5; 3.6)	Evaluate short-term organic expansion.
Org3	10-fold organic area increase (208; 5.7; 8.1)	Evaluate medium-term organic expansion.
Org4	20% organic vineyards in select areas (203; 5.5; 7.9)	Test organic vineyards in specific areas.
Org5	Organic on low-potential soils (367; 10; 14)	Explore organic farming on less fertile soils.
Org6	Organic on low-to-medium potential soils	Demonstrate potential on diverse soils.

	(904; 25; 35)	
Herb1, 2, 3	Grassed areas with VFSs near watercourses: - Herb1: sRef practices (378; 10.3; 15). - Herb2: Sdci1 practices (2558; 70; 100). - Herb3: Org6 practices (1064; 29; 42).	Test crop-to-pasture conversion with VFSs (5-30 m).

189 Scenarios are grouped according to their agronomic drivers as follows:

190 i) Organic Farming (Org) explore different scales and targets of organic farming expansion,
191 with prioritization in soils of lower agronomic potential (valleys, wetlands, and terraced
192 areas). Organic practices involve reduced fertilization, cessation of synthetic pesticide use,
193 extended crop rotations, and residue management.

194 ii) Innovative Crop Systems (Sdci) integrate transitioning practices to organic farming for
195 vineyards and field crops, at different scales. These systems limit pollutant use and
196 transfer, lengthen crop rotations, promote intercropping, and provide alternative molecules
197 for plant protection.

198 iii) Grassed Areas (Herb): integrate temporary grasslands into crop rotations, in
199 combination with practices from the baseline (Herb1), innovative crop systems (Herb2)
200 and organic farming (Herb3). Each of this Herb scenario has its VFSs scenarios, with filter
201 strips from 5 to 30 m wide along the 551 fields at high pollutant transfer risk to
202 watercourses.

203 Modifications in farming practices lead to variations in field crop distribution. For example,
204 wheat cultivation increased from 22% to 24% of the UAA between the reference scenario
205 (sRef) and the innovative low-input scenario (Sdci1), reflecting current trends. Sunflower
206 cultivation decreases from 19% to 13% and corn cultivation decreases from 15% to 12% in
207 all scenarios.

208 Across all scenarios, vineyard acreage remains constant at 22% of the Utilized Agricultural
209 Area (UAA). However, innovative scenarios show longer crop rotations and increased
210 surfaces of cereals, spring field peas, alfalfa, and temporary grasslands, with reduced areas
211 for sunflower, rapeseed, and corn. These changes occur while maintaining overall vineyard
212 acreage stability.

213 Crop rotations vary in length from 1 year (e.g., maize monoculture or temporary
214 pasture) to 7 years (e.g., Sdci scenario with three years of maize, followed by barley, then
215 three years of alfalfa, or two years of silage corn, sunflower, winter wheat, and three years of
216 alfalfa). These repeated cycles ensured that the simulations realistically captured the impacts
217 of agronomic practices over time, providing reliable predictions of potential environmental
218 benefits.

219 The mitigation scenarios were simulated to assess potential improvements in
220 freshwater quality over an extended period. This long-term approach accounts for ecosystem
221 inertia and response to agricultural changes. By systematically modeling crop rotations, soil
222 practices, and agricultural techniques across reference and mitigation scenarios, we aimed to
223 provide reliable predictions of environmental benefits.

224 **2.5.Mitigation scenario performance evaluation methods**

225 Simulated daily hydro-meteorological and water-quality outputs exhibited a strong
226 right skew and several concentration peaks, as observed. The well-calibrated Soil and Water
227 Assessment Tool (SWAT) model was used to route the PPPs from fields to water reach,
228 reporting at each of the 106 outlets.

229 Statistical analyses were performed on each of the simulated PPP across the 106
230 outlets over the last 4,383 days. Various scripts using R v4.3.1 package executed statistical
231 computations and visualization (R Core Team, 2022).

232 To evaluate the effectiveness of each modeled mitigation scenario compared to the
233 reference scenario (sRef), the non-parametric Kruskal-Wallis (KW) test (Kruskal and Wallis,
234 1952) and post hoc Dunn's test with the Benjamini-Hochberg correction was used. The KW
235 test was chosen for its robustness over the ANOVA or t-test in the presence of non-normality
236 (Gibbons and Fielden, 1993). The null hypothesis (H0) states that each mitigation scenario is
237 neither greater nor less than sRef at the 95% confidence level. For visualization, p-values
238 from the KW pairs were prepared and presented as boxplots of the simulated PPP
239 concentrations. Scenario performance was considered significant when $p < 0.05$, as follows:
240 *: $p \leq 0.05$; **: $p \leq 0.01$; ***: $p \leq 0.001$; ****: $p \leq 0.0001$.

241 Principal Component Analysis (PCA) was employed to identify the contributions of
242 biogeochemical factors and analyze the spatial variability in individual sub-basins. PCA was
243 chosen because it is suitable for water quality studies, even in the presence of nonlinear
244 processes, as it can optimally transform data (Gamble and Babbar-Sebens, 2012). Biplots
245 were used as part of PCA to graphically represent the results by combining scores and
246 loadings into a single plot. This approach allows simultaneous analysis of categorical (e.g.,
247 sub-basin identities) and quantitative (e.g., pollutant concentrations) multivariate data
248 (Agudelo-Jaramillo et al., 2016). Biplots facilitate the identification of relationships and
249 contributions among different variables, making them a powerful tool for interpreting the
250 complex data generated in this study.

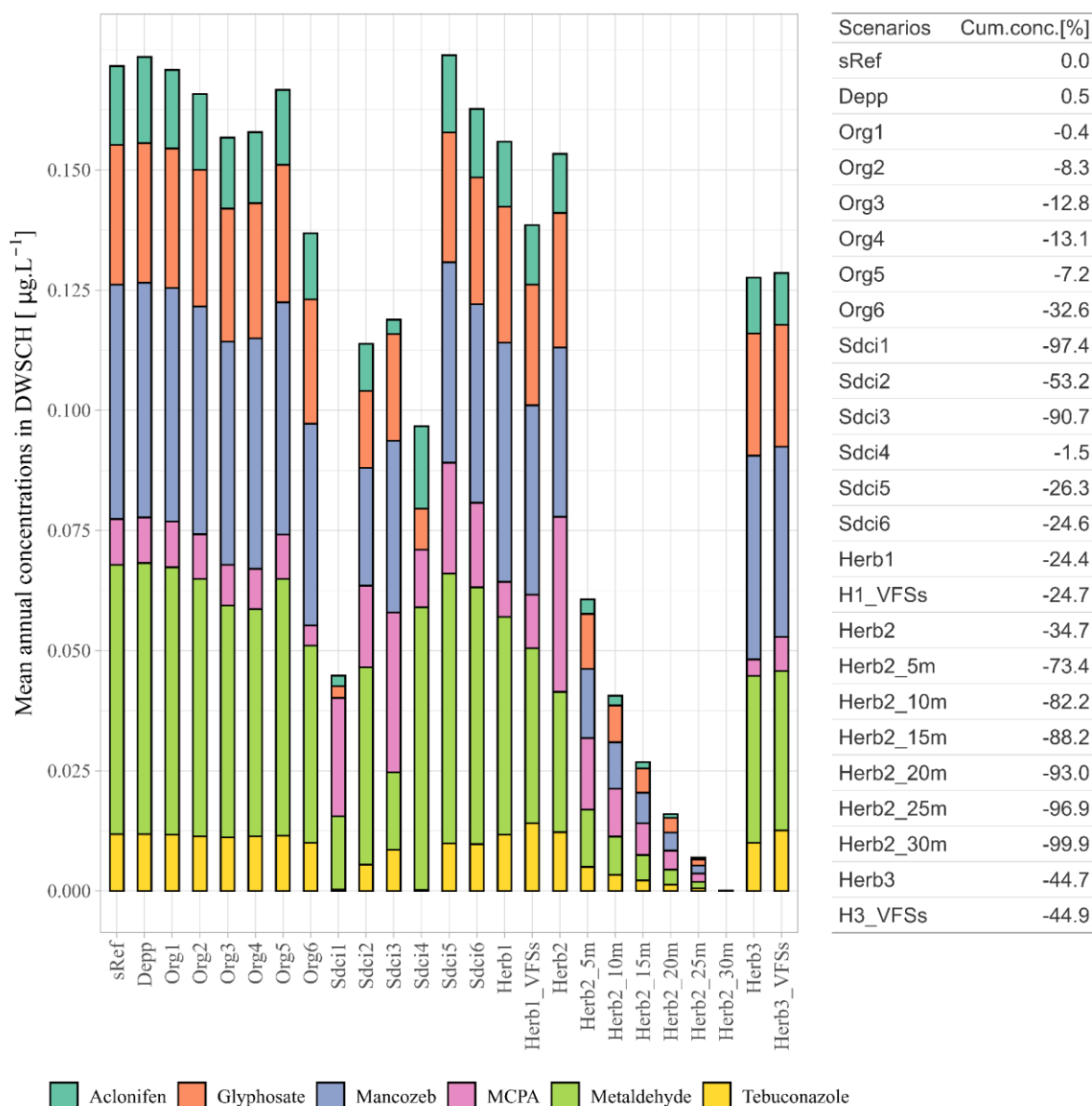
251 Temporal variability within sub-basins was investigated using the non-parametric
252 Mann-Kendall (MK) test (Gilbert, 1987), chosen for its robustness (lower sensitivity to gross

253 errors) and efficiency (smaller asymptotic variance) with adjustments for ties (Kendall and
254 Gibbons, 1990). The MK test evaluated trends in daily water quality outputs, with the null
255 hypothesis (H0) suggesting no a priori notion of an increasing or decreasing trend in
256 mitigation scenarios compared to sRef at an alpha level of 0.05. A p-value of less than 0.05
257 led to the rejection of the null hypothesis, indicating a significant monotonic trend (H1).

258 **3. Results and discussion**

259 In the DWSCH watershed, trends in mitigation scenario performance indicate varying
260 average modeled PPP concentrations over the past 12 years of the daily simulation. Scenarios
261 like Sdci1 and Herb3 (Table 1), which were extensively implemented along with their
262 associated VFS variants, demonstrated substantial reductions, achieving between 97% and
263 98% of the cumulative average PPP concentration drop in water flows over the simulation
264 period. Innovative scenarios targeting PAs, such as Sdci3, also show favorable results, with a
265 notable 90% reduction in cumulative PPP concentration.

266 Organic farming (Org) scenarios exhibit promising results, especially with significant
267 LUCs. For instance, Org6, which accounts for 35% of LUCs devoted to organic farming,
268 achieves a 33% reduction in the concentration. However, scenarios with conversion areas of
269 less than 4% of the UAA, like Org1 and Org2, showed less significant relative reductions
270 (Fig. 2).



271
272

Figure 2: Simulated average PPP concentrations throughout the DWSCCH basin.

273 **3.1. Statistical Significance of Scenarios (KW Test)**

274 To assess the effectiveness of the mitigation scenarios in safeguarding drinking water
 275 catchments, our analysis focused on the downstream sub-basin, i.e. the sub 6, which mainly
 276 consisted of wheat-based rotations with sunflower, corn, and rapeseed-based rotations.

277 **3.1.1. Organic (Org) and Innovative (Sdci) scenarios**

278 Boxplots reveal significant reductions in PPP concentrations across the organic and
279 innovative system scenarios, particularly for Sdc1 and Org6, where the peak concentrations
280 remain below half the MAC. Innovative scenarios successfully maintained concentrations
281 below the MAC threshold, except for MCPA.

282 The KW pairwise test provides further evidence of the effectiveness of these last
283 massively implemented scenarios as significantly effective, but also some other targeted
284 scenarios based on the progressive shift from intensive farming (Fig.3). Combining various
285 agronomic orientations within a single scenario is effective for water quality restoration. For
286 example, scenario Sdc6, which integrates innovative and organic farming practices, exhibits
287 significant efficacy in reducing aconifen and MCPA concentrations, moderate efficacy in
288 reducing mancozeb and metaldehyde concentrations, and limited efficacy in reducing
289 glyphosate concentrations. In addition, the scenarios characterized by a single, widely
290 implemented agronomic orientation, such as Sdc1 and Sdc3, demonstrated significant
291 statistical efficiencies in reducing aconifen and metaldehyde concentrations. Notably,
292 alterations in practices within these innovative scenarios led to partial substitution of
293 aconifen by MCPA and a decrease in metaldehyde applications. However, scenarios
294 targeting innovative systems on PAs (Sdc2) and vines (Sdc4) also significantly reduce the
295 concentration of mancozeb in watercourses. Targeting LUCs efficiently on relatively small
296 surfaces using these combined scenarios is interesting for water management with limited
297 financial resources.

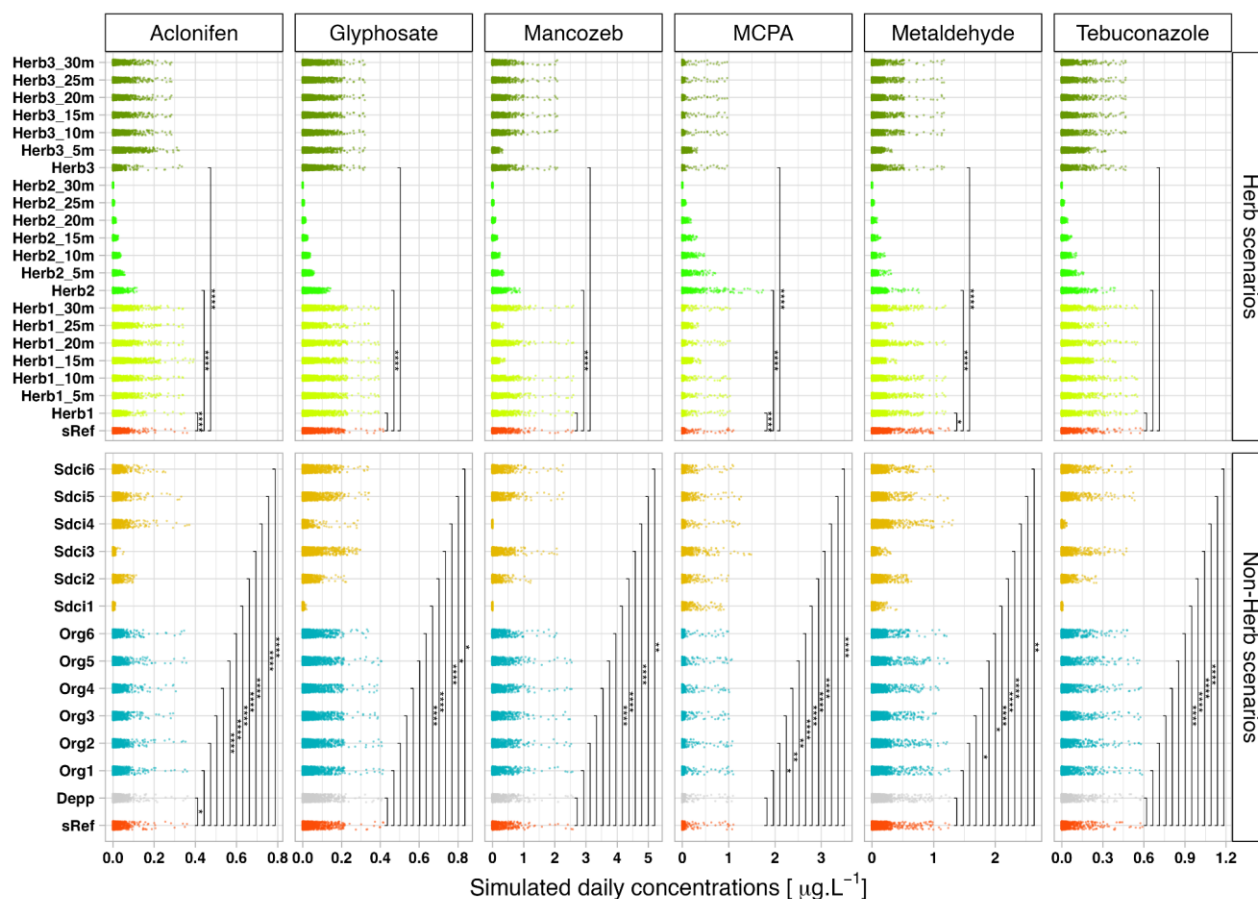
298 Bio-geophysical factors help explain the sub-basin discrepancies between targeted
299 actions and observed efficiency. For example, the Org1-Org4 scenarios receive runoff from
300 the upstream sub-basins where these scenarios are implemented. Similarly, the Depp scenario
301 results in increased aconifen concentrations due to neighboring sub-basins experiencing

302 grassland abandonment for corn cultivation, exacerbated by the presence of silty soils
303 (DHyLi), which may favor the particulate transport of this adsorbable contaminant (Fig. 3).

304 **3.1.2. Herb Scenarios and their VFS variants**

305 The Herb2 and Herb3 scenarios generally lead to greater reductions in pollutant
306 concentrations than Herb1, especially in targeted areas. Although VFS scenarios are
307 primarily effective for nitrogen reduction (Abimbola et al., 2021), they also demonstrated the
308 efficiency of PPP reduction (Tang et al., 2012). Predicting the VFS efficiency of pesticide
309 trapping is a complex task due to lumped and in-stream empirical processes and parameter
310 uncertainties (Munoz-Carpena et al., 2010). However, a general gradient between VFS width
311 and efficiency was observed, with the combined Herb2 and Herb3 scenarios being the most
312 effective at reducing various pollutant concentrations.

313 Combined Herb2-VFS scenarios are most effective at reducing various pollutant
314 concentrations, with Herb2 alone effectively reducing glyphosate and mancozeb
315 concentrations. Combined Herb2 with the largest VFS scenario maintained pollutant
316 concentrations below the MAC, whereas combined Herb3 scenarios maintained the MAC
317 below half and were highly effective at reducing metaldehyde concentrations.



318
319 Figure 3: Distributions and KW efficiencies at basin outlet.

320 **3.2. Investigation of relative contributions of anthropogenic and environmental**
321 **variables**

322 Anthropogenic practices and biogeochemical factors play pivotal roles in PPP
323 transfers, with soil type and climate significantly influencing the temporal distribution of
324 diffuse pollution in watercourses. Although precipitation (PCP) worsens this phenomenon, it
325 is noteworthy that PCP alone cannot determine the extent of pollution (Boithias et al., 2014).

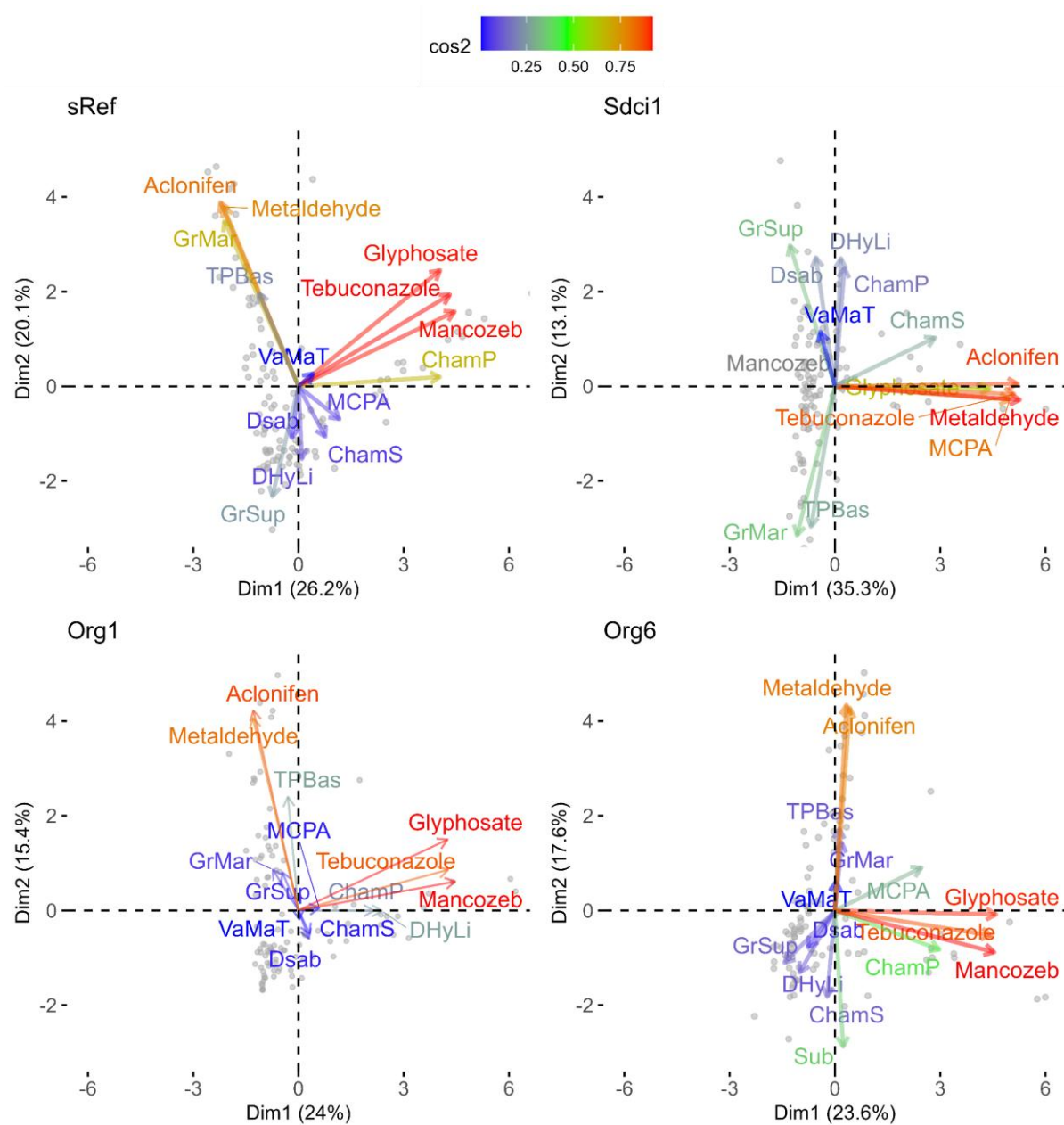
326 Shallow soils, such as GrSup and ChamS, along with silty loams like DHyLi and
327 sandy loams, such as Dsab, exhibit strong correlations with PCP and are primary contributors
328 to PPP transfers.

329 Within the four illustrated scenarios (Fig. 4), aclonifen and metaldehyde demonstrated
330 significant correlations with certain soil types, particularly GrMar, TPBas, and VaMaT.
331 Conversely, glyphosate, tebuconazole, and mancozeb exhibited larger variances within sub-

332 basins, with tebuconazole showing an average transfer capacity across the sampled PPPs.
333 Tebuconazole is notable for its ability to transfer in both dissolved and particulate phases,
334 whereas mancozeb primarily transfers in the adsorbed phase. Agricultural practices further
335 intensify PPP transfers, with intensive conventional agriculture, especially in vineyards and
336 wheat-dominated rotations, significantly contributing to metaldehyde, aconifen, and MCPA
337 concentrations in sub-basins. The hydrological behavior of certain sub-basins, particularly
338 those experiencing significant groundwater-river exchanges, also plays a role in intensifying
339 pollution transfers. Regarding agricultural practices, sub-basins with high contributions of
340 tebuconazole and mancozeb are located in PAs, except for certain sub-basins characterized
341 by intensive conventional agriculture with vine and wheat-dominated rotations.
342 Tebuconazole is applied to vines and dry and irrigated durum wheat at specific rates, whereas
343 mancozeb is applied at different rates to vines.

344 Conversely, sub-basins contributing the most to metaldehyde, aconifen and, to a
345 lesser extent, MCPA are correlated with the minority TPBa heavy soil type (Fig. 4).
346 Metaldehyde is applied to rapeseed (113 g ha^{-1}), maize (110 g ha^{-1}) and all other straw cereals
347 and sunflower, depending on the soil type and plant, at a rate of between 9 and 38 g ha^{-1} .
348 However, sub-basins 27 and 17 are only PAs. One hypothesis for the factors aggravating
349 transfers to sub-basins 5 and 37 is the hydrological behavior of these sub-basins, which are
350 part of the northern median of the DWSCH basin and are characterized by significant

351 groundwater-river hydrological exchanges.



352

353 Figure 4: PCA Biplots of factors at the basin outlet.

354 The analysis of the sRef scenario highlights the significant contributions of various
 355 PPPs to spatial variance, except for mancozeb. Sub-basins with high contributions of
 356 tebuconazole and mancozeb are strongly correlated with the ChamP agricultural soil type
 357 (Fig 4). Tebuconazole has an average transfer capacity among the PPPs, being able to

358 transfer in both the dissolved and particulate phases, whereas mancozeb primarily transfers in
359 the adsorbed phase.

360 Principal component analysis of the scenarios for conversion to organic farming
361 revealed reductions in the contributions of aconifen and metaldehyde transfers with targeted
362 actions, even in small areas. Taking sub-basin 27 as an example, the conversion of 148 ha
363 and 212 ha significantly reduced the aconifen and metaldehyde transfers. This indicates the
364 effectiveness of organic farming in mitigating diffuse pollution. Additionally, the variance in
365 individual sub-basin contribution changes as the conversion of land to organic farming
366 intensifies. Conversion to organic farming is a potential mitigation strategy, with targeted
367 actions significantly reducing PPP contributions, particularly aconifen and metaldehyde. The
368 effectiveness of organic farming scenarios underscores the potential of small-scale
369 interventions to mitigate diffuse pollution markers. Spatializing changes in practices, such as
370 intercropping and alternative techniques, to reduce active substance quantities demonstrate
371 promise in mitigating PPP transfers and reducing effluent volumes.

372 The results highlight the complex interplay among anthropogenic, biogeochemical,
373 and agricultural practices in influencing PPP transfers within the basin studied. Furthermore,
374 the effectiveness of targeted interventions underscores the importance of tailored strategies
375 for mitigating diffuse pollution and protecting water resources.

376 The biplot on the left shows that Sdc1 significantly reduces PPP contributions, except
377 MCPA, which is used as a substitute herbicide. Notably, the most contributive sub-basins
378 (CSAs) align primarily with specialization in winegrowing downstream (particularly Cognac)
379 and the concentration of field crops upstream. These areas face significant phytosanitary
380 pressure, with numerous vineyard treatments generating PPP effluents that require
381 management. More than a dozen treatments of vineyards with phytosanitary substances
382 generate PPP effluents that need to be spread or treated.

383 Across all Sdci scenarios, there was a consistent reduction in transfers in the central
384 and downstream parts of the main section and its northern tributaries. Although certain
385 molecules showed a relatively low secondary effect of PPP substitution in terms of
386 cumulative concentrations, the innovative scenarios, particularly Sdci1 and Sdci3, which
387 target all field crops, proved effective. These scenarios are also highly effective in certain
388 sub-basins, including priority zones, for specific molecules, while demonstrating a moderate
389 effect on other molecules. This highlights the need for spatializing changes in practices to
390 address diverse pollution sources effectively.

391 Sdci scenarios implement measures such as intercropping to limit PPP transfer by
392 runoff, thus safeguarding water resources. Additional techniques like reducing active
393 substance quantities through methods like the grassing of vines and mechanical weeding,
394 along with adherence to good practices in PPP application, contribute to reducing effluent
395 volumes. However, understanding the distribution of surface and underground inputs,
396 particularly in karstic soils, remains a challenge. This underscores the need for ongoing
397 research and management strategies to address complex pollutant pathways.

398 **3.3. Temporal variability analysis using the Mann-Kendall test (MK)**

399 The spatial analysis of the Mann-Kendall (MK) test results, focusing on tebuconazole and
400 scenarios of conversion to organic farming, revealed notable monotonic trends ranging from
401 low to high significance (Fig. 5):

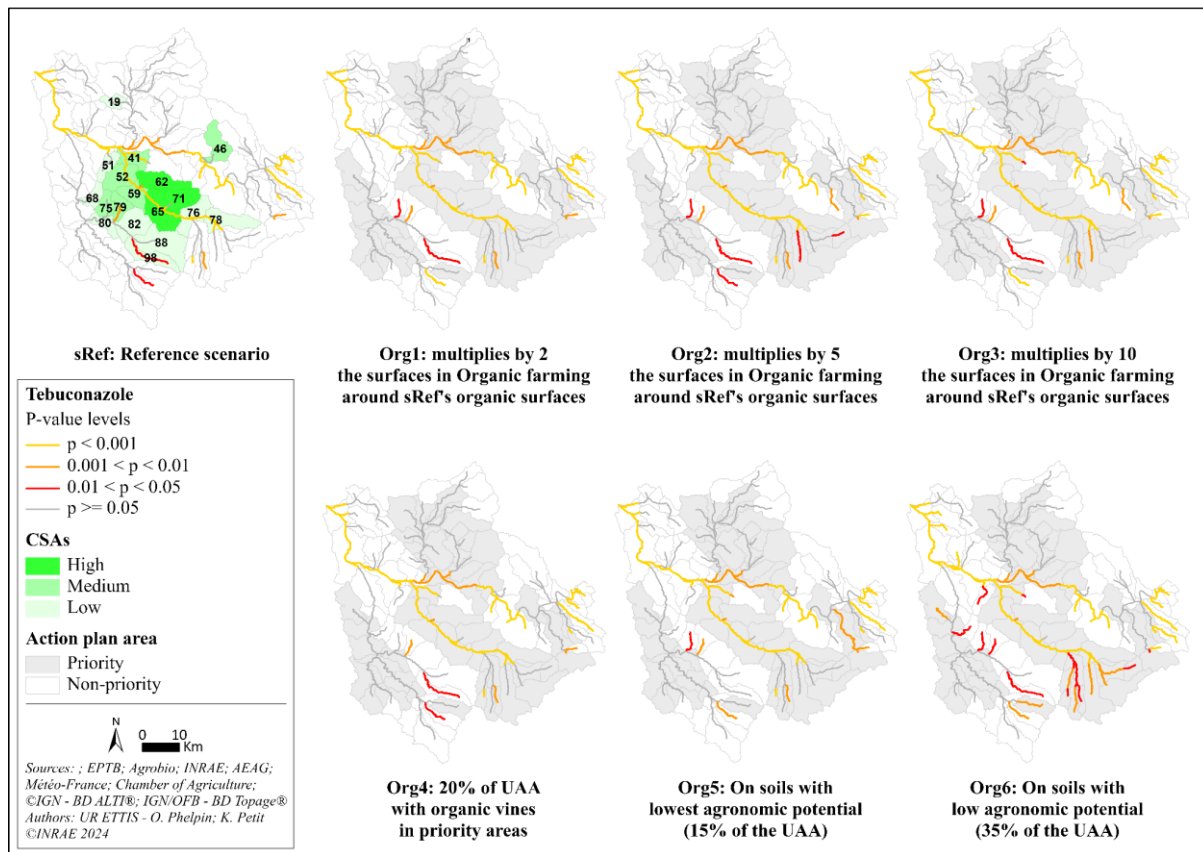
402 i. Across the main river continuum, excluding the headwaters, a substantial number of
403 monotonic trends was observed across all scenarios, including the reference scenario (sRef).
404 The increase in areas converted from conventional to organic farming is correlated with the
405 increase in the number of river sections that exhibit significant monotonic trends, which is
406 particularly evident along the main southern tributary, notably in the Org6 scenario.

407 ii. High levels of diffuse tebuconazole pollution were identified in the southern and
408 southwestern regions, specifically sub-basins 62, 65, and 71. In the same impluvium situated
409 in priority zones, some CSAs fall within priority areas designated by water governance,
410 whereas others, such as CSAs 75, 79, 51, 68, 82, 88, and 98, do not. Nonetheless, significant
411 monotonic trends were detected outside priority areas; the potential benefits of implementing
412 biological scenarios to mitigate tebuconazole contamination. Geological and hydrological
413 factors may contribute to the observed disparities between actions targeted at PAs and actual
414 responses over time, particularly given shared groundwater bodies and dynamic surface
415 runoff dynamics.

416 iii. In the northern region, sub-basin 19, which is characterized as a low contributor, exhibits
417 significant monotonic trends associated with parts heavily cultivated with irrigated durum
418 wheat. This study highlights tebuconazole as a marker of specific agronomic practices,
419 particularly durum wheat cultivation.

420 iv. To the east, agricultural sub-basin 46, a moderate contributor, shows potential benefits
421 from inclusion in priority action zones, especially for its cultivation of irrigated durum wheat
422 on silty soils (DH_YLi), which may facilitate contaminant transfers.

423 The distribution and mean concentrations of tebuconazole in the agri-environmental
424 scenarios were notably smaller and lower in amplitude than those in the reference scenario.
425 These monotonic downward trends indicate significant reductions in pollutant concentrations
426 over the 12-year study period, indicating sustainable and substantial improvements in water
427 quality attributed to LUC for this particular PPP in select river sections.



428

429 Figure 5: MK trend significance (p-values) for tebuconazole concentrations under organic
 430 scenarios vs. baseline (sRef).

431 4. Conclusions

432 Our SWAT-GenLU2 spatialized agrohydrological modeling yielded valuable insights
 433 into the effectiveness of various cropping scenarios in limiting PPP transfers. By considering
 434 combined agronomic orientations, such as organic farming, innovative systems, grass strips,
 435 and meadows, along with their progressive implementation, we have elucidated the nuanced
 436 relationship between agricultural practices and water quality outcomes. The statistical
 437 methods applied to model outputs have provided complementary perspectives for assessing
 438 the efficacy of agro-environmental mitigation scenarios, identifying the contributions of
 439 anthropogenic and soil factors, and discerning monotonic trends over time.

440 Our findings highlight that scenarios widely deployed in the UAA tend to be more
441 effective over time. However, certain sub-basins exhibit stronger responses to changes in
442 practices and agrosystems influenced by specific climatic, biophysical, and geomorphological
443 characteristics. Targeting changes within specific agrosystems, such as field crops and
444 vineyards, can significantly improve surface water quality, especially when considering the
445 unique characteristics of each sub-watershed.

446 As agroecology gains traction in the study area, it becomes imperative to address the
447 heterogeneous nature of organic farming practices. The increasing adoption of organic
448 farming highlights the need to incorporate PPPs authorized for organic farming into modeling
449 efforts. Identifying highly vulnerable areas to PPP transfer, such as Critical Source Areas
450 (CSAs), can help define PAs for targeted mitigation efforts, particularly in the context of
451 climate change.

452 Although our modeling approach provides valuable insights, it is essential to
453 acknowledge its limitations. The complexity of water quality processes and the simplified
454 representation in the model necessitate caution in its application, which requires validation by
455 field experts, as was the case in this study. Improved PPP monitoring, along with model
456 calibration and validation efforts, can enhance the understanding of diffuse pollution and
457 reduce modeling uncertainties.

458 Moving forward, integrating socio-economic drivers into decision support methods
459 could further enhance our ability to assess and manage water quality effectively. By
460 considering uncertainties and leveraging multi-criteria decision-support methods, integrated
461 water quality modeling, and scenario planning, stakeholders can develop more robust
462 strategies for sustainable water resource management.

463 Despite these challenges, our spatialization modeling approach offers valuable
464 insights into the environmental implications of agro-environmental changes and can be
465 extrapolated to other contexts. Continued research efforts and interdisciplinary collaborations
466 are essential for addressing the complex challenges associated with water quality
467 management in agricultural landscapes.

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