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## Harmonizing Campylobacter risk assessments across European countries – can the pooled process hygiene criteria data be used in the Danish risk assessment model?

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1 **Harmonizing *Campylobacter* risk assessments across European countries – can**  
2 **the pooled process hygiene criteria data be used in the Danish risk assessment**  
3 **model?**

4

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16 **Highlights**

- 17 • French *Campylobacter* data 2020-21 on testing broilers neck skin (NS) pools was used
- 18 • Human risk of disease was estimated by a Danish model developed for leg skins (LS)
- 19 • The 2021 risk relative to 2020 (RR) was 1.22 meaning an increase of risk of  $\approx 22\%$
- 20 • A harmonized risk assessment could be carried out using pooled NS data
- 21 • Limitations for harmonizing risk assessment across countries were addressed

## 22 **Abstract**

23 This study investigated the possibility of harmonizing quantitative microbiological risk assessment  
24 (QMRA) for *Campylobacter spp.* across European (EU) countries.

25 French *Campylobacter* data (2020-2021) from neck skin (NS) pools, sampled at slaughterhouses  
26 under the European surveillance component for Food Business Operators (FBOs), were adapted  
27 to inform a QMRA model that, among others, has been used within the Danish Action Plan against  
28 *Campylobacter*, on the basis of single leg skins (LS) data.

29 Datasets included culture results (in colony forming unit per gram, CFU/g) from 1,284 broiler  
30 flocks slaughtered at 13 slaughterhouses representing broiler production in western France. Five  
31 pools (of 2-4 NS samples each) per flock were tested. One pool per tested flock was randomly  
32 chosen for the analysis. After conducting descriptive statistics (on flock prevalence and meat  
33 contaminations across months and years), three contamination transformation factors (CTFs) were  
34 estimated to translate NS pools contaminations into single LS contamination, based on data from  
35 French and Danish studies. A reference simulation scenario (ScenRef) was set with  $CTF = 3.2$  (i.e.  
36 NS pool concentration divided by 3.2); while other 13 scenarios represented an alternative scenario  
37 analysis to investigate the impact of: the CTF value (ScenMin with  $CTF = 2$  and ScenMax with  
38  $CTF = 10$ ), censored test results (ScenUncens) and random choice of pool per flock  
39 (ScenSampling-1 to 10), on the risk estimates. The average monthly/annual risk of human disease  
40 per poultry meal and the monthly/annual relative risk (RR) of 2021 compared to 2020, were  
41 estimated.

42 In ScenRef, the annual RR was 1.22, suggesting an increase of risk of  $\approx 22\%$  in 2021 compared to  
43 2020. The impact of CTFs, censored data and randomized pool sampling per flock, on the annual  
44 and (most) monthly RRs, appeared limited.

45 This study gives an overview of the strengths and limitations to be considered for adapting the  
46 French FBO data into the Danish model and to harmonize risk assessments across EU countries,  
47 accordingly. To reduce uncertainty in risk estimates, it could be considered increasing  
48 representativeness of NS tested flock populations and/or using LS rather than NS samples; because  
49 LS samples are more representative of actually retailed meat contaminations. If NS pools are

50 maintained, the relationships between concentrations on NS pools and those on consumed meat  
51 requires further investigation.

52 **Key words:** *Campylobacter; Neck skin pools; Surveillance data; Harmonization; Risk*  
53 *assessment; Contamination transformation factor*

54 **1. Introduction**

55 *Campylobacter spp.* is the most commonly reported cause of foodborne gastrointestinal infection  
56 in humans in Europe (EU). In 2022, 137,107 human cases were reported, corresponding to a  
57 notification rate of 43.1 per 100,000 inhabitants (EFSA and ECDC, 2023). In France, the  
58 notification has been rising steadily since 2017 and was at its highest in 2022 (67.0 cases per  
59 100,000 inhabitants) (EFSA and ECDC, 2019, 2023). As the number of confirmed cases of  
60 campylobacteriosis constitutes only a portion of the cases that actually occurred (due to  
61 underreporting and underdiagnoses) (Havelaar et al., 2013; Monteiro Pires et al., 2020), the  
62 average annual number of symptomatic cases in France was estimated at  $\approx 493,000$ , of which  $\approx$   
63 392,000 are thought to be foodborne over the period 2008-2013. For those estimates,  
64 underreporting and underdiagnoses factors were applied on national reference laboratory data (Van  
65 Cauteren et al., 2017). Moreover, EFSA estimated that broiler meat may account for 20–30% of  
66 campylobacteriosis human cases in the EU, while 50–80% of these may be attributed to the chicken  
67 reservoir as a whole (EFSA BIOHAZ Panel, 2010). Campylobacteriosis is mainly characterised  
68 by gastrointestinal symptoms, but it can also cause a serious post infectious neuro-paralytic  
69 complication: Guillain-Barré syndrome (Allos, 1997).

70 The public health importance of *Campylobacter* contamination and survival have been  
71 investigated throughout the poultry sectors (from farm to fork) using Quantitative Microbial Risk  
72 Assessment (QMRA) models, which include a model of the consumer phase (Chapman et al.,  
73 2016). These models can be used to inform evidence-based risk mitigation actions along the food  
74 chain. In Denmark, a QMRA model developed by Nauta *et al.* (Nauta et al., 2008; Nauta et al.,  
75 2012) has been used since 2014, to assess the risk of campylobacteriosis due to *Campylobacter*  
76 contaminated fresh broiler meat. In the Danish Action Plan against *Campylobacter*, the results of  
77 culture testing single leg skin (LS) samples, are fed into this model, to assess the mean annual risk  
78 per serving (i.e. per meal) and its changes in relation to a reference year (2013) (Foddai et al.,  
79 2022a, 2023; Foddai et al., 2022b).

80 In 2017, a process hygiene criterion (PHC) for *Campylobacter* in broiler meat, to be complied with  
81 by Food Business Operators (FBOs), was introduced at EU level. The criterion is defined by  
82 Regulation (EU) 2017/1495 (Anonymous, 2017a) and sets the microbiological limit in colony

83 forming units per gram (CFU/g) for pooled neck skin (NS) samples taken from broiler flocks at  
84 slaughterhouses. Five pooled NS samples are analysed per sampling session per week.

85 The PHC aims to inform risk mitigation actions at the single slaughterhouses rather than at national  
86 level. Still, FBOs need to inform the competent authorities about non-conformities upon request  
87 (Anonymous, 2017b), and also to report the annual number of pooled NS samples with a  
88 concentration larger than 1,000 CFU/g (Anonymous, 2019).

89 If the NS surveillance data obtained from individual FBOs under Regulation (EU) 2017/1495,  
90 could be used as input in a QMRA model, there would be an important international added value,  
91 e.g. for harmonizing national quantitative risk assessments across EU countries. In fact, during  
92 recent years, there has been an increased demand for standardizing procedures of inter-sectorial  
93 data integrations, data analysis and risk assessments within the EU (Colangeli et al., 2023; One  
94 Health EJP, 2018, 2023).

95 The objective of this study was to give a proof of concept (with strengths and limitations to  
96 consider) for potential harmonization of quantitative risk assessment for *Campylobacter spp.*  
97 across EU countries, by using as example the Danish QMRA model (Nauta et al., 2008; Nauta et  
98 al., 2012), informed by the French NS surveillance data.

99

## 100 **2. Materials and methods**

101 The QMRA model used to assess the risk per serving and its relative changes across surveillance  
102 periods (months and years) requires concentrations that are representative for single carcasses, not  
103 from pooled broiler samples. Therefore, in the risk assessments done for the Danish Action Plan  
104 (Foddai et al., 2022a, 2023), single LS samples are used. Those are more representative for the  
105 carcass contaminations to which the consumer is actually exposed (usually, NS are not sold for  
106 consumption, although sampling does not represent any meat loss to the FBO). As a consequence,  
107 the *Campylobacter* concentration quantified on the French NS pooled samples had to be  
108 transformed into concentrations for single LS samples, before being fed into the model. Data  
109 handling and analyses were carried out using the package “*base*” in R version 4.3.1 and RStudio  
110 software version 2023.6.2.561 (Posit team, 2023; R Core Team, 2023), whereas the risk  
111 assessment was carried out using the Danish QMRA model (Nauta et al., 2008; Nauta et al., 2012).



112 **2.1 The original data on pooled NS samples**

113 Data 2020-2021 from thirteen slaughterhouses were used for the study. These are located in  
114 western France and were estimated to represent approximately 48% of the French tonnage of  
115 broilers slaughtered in 2021 (Agreste, 2022). The metadata included: the date of slaughter (i.e.  
116 sampling), the identifier of the broiler flock, the analysed food matrix (i.e. NS pools) and the  
117 *Campylobacter* concentration according to the culture standard EN ISO 10272-2 (ISO 10272-2,  
118 2017). A flock was defined as the group of broilers reared in the same house unit of the same farm  
119 (conventional or organic), and slaughtered on the same day.

120 The data was collected in Excel files (version Excel 2021) in two stages. Firstly, testing results  
121 were collected directly from five slaughterhouses' quality department teams. Secondly, for the  
122 other eight slaughterhouses, the surveillance results were extracted from the Donavol platform of  
123 the French federation of poultry industries (in French: Fédération des Industries Avicoles (FIA))  
124 (FIA, 2020). This federation is responsible for collecting and transmitting to the Ministry of  
125 Agriculture the self-checking data required by the regulations in poultry slaughterhouses  
126 (Anonymous, 2017a, 2017b, 2019).

127 According to Regulation (EU) 2017/1495 (Anonymous, 2017a), five pools of NS samples (a pool  
128 is made of up to four carcasses from the same flock of origin) must be analysed at each  
129 slaughterhouse, each week of the year. In practice, the five NS pools come from the same flock.  
130 There may be three main types of pooled NS samples, according to regulations and analyses to be  
131 performed (Anonymous, 2017a, 2021): a) Pools of two samples when used solely for  
132 *Campylobacter* testing; b) pools of three samples when used for *Salmonella* and *Campylobacter*  
133 testing in the same laboratory; and c) pools of four samples when used for *Salmonella* and  
134 *Campylobacter* testing in different laboratories. In the data used for this study, each pool was  
135 obtained from 2-4 NS single samples (i.e. all options: a, b, and c were applicable).

136 In total, the data consisted of 6,574 NS pools collected from 1,284 flocks, over the two years: from  
137 620 flocks in 2020 and from 664 flocks in 2021. For 14 of the flocks, one to four NS pools out of  
138 five were not available due to logistical or sample analysis problems. For other 36 flocks, 10 NS  
139 pools were tested.

140

141 **2.2 Standardization of concentration ranges across slaughterhouses**

142 For the eight slaughterhouses which did not provide the data directly (Section 2.1), the  
143 *Campylobacter* counts <100 and >10,000 CFU/g were censored in the Donavol database, because  
144 the FBOs are not required entering CFU/g below or beyond these values (FIA, 2020). In a context  
145 of adaptation of individual FBO data into a national level risk assessment, for France, the Donavol  
146 database would represent the main national data source and format. Therefore, all the datasets from  
147 the thirteen FBOs were standardized similarly, to represent at national level, the overall data  
148 analysis (Section 2.3) and the simulation reference scenario (Section 2.4-2.5). Accordingly, the  
149 NS pools recorded as “<100” CFU/g were set as = 0 (i.e. considered negative) and the pools  
150 recorded as “>10,000” CFU/g were set equal to 10,000 CFU/g. Concentrations in between these  
151 two limits were kept as originally recorded and classified positive<sup>1</sup>.

152

153 **2.3 Descriptive statistics and statistical data analysis**

154 From the standardized data, for each tested flock, one of the five NS pools was randomly selected  
155 using the function “*sample*” in the package “*base*” in R (R Core Team, 2023), to simulate the  
156 random selection of the leg sampled per flock at the abattoir in the Danish Action Plan. Thus, the  
157 overall number of pools selected for the descriptive statistics was 620 for 2020 and 664 for 2021.

158 The procedure and the interface tool from Foddai et al. 2024 (under review), were used to visualize  
159 and interpret outputs of data analysis, together with their related risk estimates (see Section 2.5).  
160 The descriptive statistics included: the number of tested and positive flocks, with their  
161 corresponding apparent prevalence (AP), i.e., the prevalence not corrected for the diagnostic test’s  
162 error (Rogan and Gladen, 1978; Moreno-Torres et al., 2016; Foddai et al., 2022b). Moreover, the  
163 distributions of concentrations across the positive NS pools were log<sub>10</sub> transformed and compared  
164 between surveillance months and years, through box plots produced in R.

165 A multiple linear regression model was also set for statistical analysis (significance level at P-  
166 value < 0.05) to compare log<sub>10</sub> *Campylobacter* concentrations across the two surveillance years  
167 and the thirteen slaughterhouses, and to assess whether there was any interaction between abattoir

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<sup>1</sup> N.B. The impact of this data standardization procedure is addressed through the alternative scenario analysis, in Section 2.6.

168 and year. For that purpose, the “*lm*” function of the R statistical package “*stats*” was used to set a  
169 linear regression model (Wilkinson and Rogers, 1973; Chambers and Hastie, 1992) as:

$$170 \log_{10} \text{concentration} = \beta_0 + \beta_1 \text{year} + \beta_2 \text{slaughterhouse} + \beta_3 \text{year} * \text{slaughterhouse} + \varepsilon_i.$$

171 The term “*year \* slaughterhouse*” represented the interaction between these two variables. The  
172 results of the model were observed using the function “*summary*” from the package R “*base*”.  
173 However, the significance of each overall variable was assessed by analysis of variance (ANOVA)  
174 using the function “*anova*” from the package “*stats*”. Moreover, the suggested model according  
175 to the AIC criterion was also checked using the “*drop1*” function from the package “*stats*”  
176 (Chambers and Hastie, 1992). Both variables were categorical. For the variable “*year*” the 2020  
177 was set as reference category, while abattoir No.4 (Ab\_04) was set as reference for the variable  
178 “*slaughterhouse*”. These were the categories with the lowest log<sub>10</sub> concentration across the two  
179 years.

180

#### 181 ***2.4 Transforming concentration from neck skin pools into (simulated) single leg skin*** 182 ***contaminations***

183 In France and Denmark, two publications examined the different *Campylobacter* concentrations  
184 between paired pooled NS samples vs. single LS samples (Duqué et al., 2018; Ellis-Iversen et al.,  
185 2020). In both cases, the concentrations of *Campylobacter spp.* found on the single LS was lower  
186 than these found on the pooled NS. The data from both studies were used to estimate the  
187 contamination transformation factors (CTFs), which enabled the approximation of single LS  
188 concentration from the NS pools concentration.

189

#### 190 ***2.5 The quantitative risk assessment model and the reference risk assessment scenario***

191 The risk assessment model (Nauta et al., 2012), as applied in Denmark (Foddai et al., 2022a, 2023),  
192 relates the apparent prevalence (AP) of carcass positive flocks and their *Campylobacter*  
193 concentrations detected on single LS samples collected at slaughterhouses, with the risk of human  
194 campylobacteriosis per broiler meat serving (i.e. per meal). Accordingly, the model describes:

- 195 i) The difference between concentrations on single LS samples and broiler meat (after  
196 skin removal) (Nauta et al., 2012),
- 197 ii) The effect of preparing a meal with broiler meat, on the bacterial transfer and survival,  
198 so that the ingested dose can be estimated, as described in an observational study (Nauta  
199 et al., 2008; Nauta and Christensen, 2011), and
- 200 iii) A widely used dose-response model for *Campylobacter* (Teunis and Havelaar, 2000).

201 The original version of the risk assessment model is a stochastic model that captures the variability  
202 between servings (Nauta et al., 2012). Here we use the association between the input and the output  
203 of this model, which provides a deterministic numerical relationship between the LS  
204 concentrations and the mean probability of illness per serving (Figure 1), as found after  $91 \times 50,000$   
205 iterations in a Monte Carlo simulation of the stochastic version (Nauta et al., 2012). This  
206 deterministic relationship is also used in the Danish Action Plan and has been previously applied  
207 in different studies, as a generic risk assessment model for *Campylobacter* based on broiler skin  
208 data (EFSA BIOHAZ Panel, 2011; Nauta and Christensen, 2011; Nauta et al., 2012; EFSA  
209 BIOHAZ Panel et al., 2020). We refer to Nauta et al. (2012) for a more detailed description.

210 For our purpose and within the Danish Action Plan, the model provides estimates of the average  
211 monthly risk per serving based on monthly collected skin samples, as well as the mean annual risk,  
212 which is equal to the mean of the 12 monthly risk estimates of the same year. Next, the monthly  
213 and annual relative changes in risk (i.e. the relative risk (RR) estimates) can be calculated, by  
214 comparing two years. In this study, 2020 was used as year of reference (or baseline) and the  
215 monthly/annual RRs were estimated for 2021; by dividing the monthly/annual risk estimates from  
216 2021 by these of 2020.

217 An  $RR < 1$  or  $> 1$  suggested a decrease or an increase (respectively) of the monthly/annual risk of  
218 2021, compared to that of 2020, whereas  $RR = 1$  suggested no change in risk across the months  
219 and years.

220 For the simulation reference scenario (ScenRef), the CTF was derived from Ellis-Iversen et al.,  
221 (2020), where the mean concentrations were estimated = 886 CFU/g for the pools of three NS  
222 samples and = 276 CFU/g for the LS single samples, so  $CTF = 886/276 = 3.2$ . Accordingly, in the  
223 ScenRef, the estimates for the single LS concentrations were obtained by dividing the  
224 concentrations observed in the (randomly) selected NS pools by 3.2. Accordingly, by using the

225 CTF values  $> 1$  it was reflected the fact that: a) the NSs are usually more contaminated than LSs,  
226 due to the carcass position (upside down) in the slaughter line, which makes the former more likely  
227 to become contaminated (e.g. due to viscera ruptures); and b) that the mean concentration in a  
228 pooled sample is expected to be larger than that in a single sample, when the distribution of  
229 concentrations is lognormal. Next, within the model, the LS simulated concentrations were  $\log_{10}$   
230 transformed.

231

## 232 **2.6 Alternative scenario analysis**

233 An alternative scenario analysis was conducted, to examine the sensitivity of RR estimates (the  
234 main output of interest) due to three key factors: a) CTFs (i.e. on the simulated single LS  
235 contaminations); b) censored data; c) randomized pool sampling (i.e. the random selection of one  
236 NS pool from five per flock). Therefore, thirteen alternative risk assessment scenarios were  
237 compared to the ScenRef (Section 2.5):

238 I) two scenarios were carried out using minimum CTF = 2 (ScenMin) and maximum CTF = 10  
239 (ScenMax), based on the data from the French and the Danish studies (Duqué et al., 2018; Ellis-  
240 Iversen et al., 2020);

241 II) one scenario (ScenUncens) replaced the originally NS censored results recorded as “<100”  
242 CFU/g and “>10,000” CFU/g in the standardized data (which was used also from the ScenRef),  
243 with values sampled from two distinct PERT distributions, which were set in R using the function  
244 “*rpert*” in the package “*mc2d*” (Pouillot and Delignette-Muller, 2010). These distributions were  
245 defined using the data from the five slaughterhouses quality departments, which had broader  
246 CFU/g ranges available (as described in Section 2.1). This scenario allowed assessing the impact  
247 of the censored data (“<100” and “>10,000”) and of the data standardization procedure described  
248 in Section 2.2, on the risk estimates. Therefore, here a broader range of concentration values were  
249 included, compared to all other scenarios: specifically, also contaminations below 100 CFU/g and  
250 above 10,000 CFU/g. Hence, for all 13 slaughterhouses, contaminations between 100 and 10,000  
251 CFU/g were used. Next, for the five slaughterhouses having values below 100 and above 10,000,  
252 all the original contaminations (from 10 to 150,000) were used (i.e. these given by the companies  
253 and before the data standardization made in Section 2.2), and censored data “<10”, “>15,000” or  
254 “>150,000” were replaced by 0, 15,000 and 150,000 respectively. For the remaining eight

255 slaughterhouses for which data was extracted from the Donavol database, the censored values  
256 reported as “<100” and “>10,000” were replaced by simulated values according to information  
257 from the other five companies and the mentioned PERT distributions, to decide which value to use  
258 into the censored data cells. Accordingly, a PERT distribution was based on contamination data  
259 below 100 CFU/g (i.e. PertLow: min = 0; median = 10; max = 99) and another on the data  
260 exceeding 10,000 CFU/g (i.e. PertHigh: min = 10,001; median = 25,000; max = 150,000).  
261 Thereafter, all concentrations (simulated and not) were divided by CTF = 3.2 (as in the ScenRef)  
262 and flocks with simulated LS results below 10 CFU/g were considered as negative and set at “0”  
263 into the model, as is currently (mid 2024) the case within the Danish Action Plan.

264 III) ten scenarios, each based on a different random sampling of NS pool per flock, were created  
265 but still using CTF = 3.2 (ScenSampling-1 to ScenSampling-10).

266 Accordingly, scenarios ScenMin, ScenMax and ScenSampling (1 to 10) investigated the impact of  
267 between and within-flock contaminations variability on the risk estimates, while ScenUncens  
268 investigated the impact of censored data.

269

### 270 **3. Results**

271 Sections 3.1 and 3.2 present the results from the descriptive statistics and from the data analysis  
272 (i.e. the AP estimates and the concentration distributions), while Section 3.3 shows the results of  
273 the risk assessment (i.e. the risk and the RR estimates) under the reference scenario. Finally,  
274 Section 3.4 presents the outputs of the alternative scenario analysis.

275

#### 276 ***3.1 Apparent prevalence (AP) of positive flocks***

277 Across the two investigated years, according to the culture testing results from the randomly  
278 selected NS pools of the standardized data (Section 2.2) used in the ScenRef (Section 2.4-2.5), the  
279 overall prevalence of positive flocks (i.e., of sampled pools  $\geq 100$  CFU/g) was 69.5% (892/1,284).  
280 The annual prevalence was 68.7% (426/620) in 2020 and 70.2% (466/664) in 2021.

281 Across the two years, the overall monthly median number of tested flocks was 52 (min = 45 in  
282 February and max = 59 in December) and 55 (min = 49 in October; max = 62 in March),  
283 respectively (Figure, 2A).

284 The overall monthly median number of positive flocks was = 36 in 2020 (min = 22 in February;  
285 max = 45 in June) and 40 in 2021 (min = 25 in January; max = 52 in August) (Figure, 2B).

286 The overall monthly median AP was 69.4% (min = 48.9% in February; max = 84.9% in June) and  
287 72.5% (min = 47.2% in January; max = 89.7% in August), during 2020 and 2021, respectively. In  
288 both years, the monthly AP showed seasonality and increased during summer months, compared  
289 to the rest of the year (Figure, 2C).

290

### 291 ***3.2 Concentration distributions on standardized NS pools according to surveillance period***

292 Considering the positive flocks shown in Figure 2B, across the two investigated years, the overall  
293 mean *Campylobacter*'s concentration was 3.03 log<sub>10</sub> CFU/g (2.5<sup>th</sup> percentile = 2.00; 1<sup>st</sup> quartile =  
294 2.46; 3<sup>rd</sup> quartile = 3.67; 97.5<sup>th</sup> percentile = 4.00). During 2020 and 2021, the overall mean  
295 concentrations were 2.96 log<sub>10</sub> CFU/g (2.00; 2.42; 3.54; 4.00) and 3.10 log<sub>10</sub> CFU/g (2.00; 2.52;  
296 3.75; 4.00); respectively.

297 In the multiple linear regression model, the interaction between the two variables “*year*” and  
298 “*slaughterhouse*” appeared significant only for the two combinations “2021 \* *Ab\_01*” and “2021  
299 \* *Ab\_08*” (p < 0.05); whereas the overall interaction did not appear significant by ANOVA (p =  
300 0.1799) and was suggested for removal from the model based on the AIC criterion. Without  
301 considering this interaction, the difference in mean concentrations between the two years appeared  
302 significant by ANOVA (p < 0.001), as did the difference between abattoirs (p < 0.001).

303 A visual comparison of the concentration distributions, across monthly surveillance periods, is  
304 shown in the boxplots from Figure 3. It can be noted that no particular seasonal patterns were  
305 observed in any of the two investigated years. The months with bigger difference in mean log<sub>10</sub>  
306 concentrations across years, were June (higher for 2021) and December (higher for 2020).  
307 Furthermore, the mean standard deviation of concentration between positive NS pools of the same  
308 flock was 0.31 log<sub>10</sub> CFU/g (min = 0.00; 2.5<sup>th</sup> percentile = 0.00; 1<sup>st</sup> quartile = 0.15; 3<sup>rd</sup> quartile =  
309 0.44; 97.5<sup>th</sup> percentile = 0.76; max = 1.10), indicating the variability across NS positive pools  
310 contamination levels within flocks.

311

### 312 ***3.3 Risk and relative risk (RR) estimates according to reference scenario***

313 Under the reference simulation scenario (ScenRef with CTF = 3.2 and censored data <100 and  
314 >10,000 CFU/g), during 2020, the monthly risk (i.e., the average probability of campylobacteriosis  
315 per broiler meat meal) ranged from 0.35% in January to 0.98% in May (Figure 4A, middle green  
316 line). The overall mean annual risk (across the 12 months) was 0.67%. In 2021, the monthly risk  
317 ranged from 0.44% in January and 1.16% in June (Figure 4A, middle yellow bar), while the overall  
318 annual average risk was 0.82%.

319 During 2021, the RRs were < 1 in four months (March, May, November and December) and >1 in  
320 the remaining eight months of the year (Figure 4B, middle yellow bars). Between those (in  
321 hierarchical order) June, October and April showed the highest RRs. During the year, the monthly  
322 RR ranged from 0.75 in December (i.e. showing a reduction of 25% in risk compared to December  
323 2020), and 1.96 in June (i.e. almost doubled risk).

324 The overall annual RR obtained by the ratio of the two annual risk estimates was  $0.82\% / 0.67\%$   
325  $= 1.22$ ; thus indicating for 2021, an overall increase of risk per serving of  $\approx 22\%$  compared to 2020  
326 (Figure 4C, middle yellow bar).

327

### 328 ***3.4 Outputs of alternative scenario analysis***

#### 329 ***3.4.1 Impact of concentration transformation factors (CTFs)***

330 The alternative scenarios ScenMin (CTF = 2) and ScenMax (CTF = 10) showed the impact of the  
331 assumptions made for the CTF on the risk estimates (i.e. when transforming the *Campylobacter*  
332 *spp.* contaminations observed in the NS pools, into the simulated single LS contaminations)  
333 (Figure 4A). For example, in May 2020, which was the month with the highest risk during that  
334 year, the risk changed from 0.98% in the ScenRef to 1.23% in the ScenMin and 0.52% in the  
335 ScenMax (Figure 4A, green lines). For the three scenarios, the average annual risk (across the 12  
336 months) was 0.67%, 0.86% and 0.35%; respectively. Instead, for May 2021, estimates changed  
337 from 0.76% (ScenRef) to 0.97% (ScenMin) and 0.40% (ScenMax) (Figure 4A, yellow bars). In  
338 that year, the average annual risk was 0.82%, 1.04% and 0.42%; respectively.

339 The effect of the assumed CTFs, on each monthly RR of 2021, is shown in Figure 4B, by  
340 comparing the numbers reported over the three yellow bars of the same month. The biggest



341 variation of RRs occurred in June, when changing from the ScenRef (RR = 1.96) to the ScenMax  
342 (2.18).

343 Regarding the annual RR (the main output of interest), it changed from 1.22 in the ScenRef to 1.21  
344 and 1.26 in the ScenMin and ScenMax scenarios (Figure 4C).

345

### 346 *3.4.2 Impact of censored data*

347 The scenario ScenUncens demonstrated the impact of replacing the censored concentrations (i.e.  
348 concentrations recorded as “<100” CFU/g and “>10,000” CFU/g), with values sampled from the  
349 PERT distributions (which were set as described in Section 2.6).

350 This is the only scenario that, next to the concentration distributions, affected also the prevalence  
351 of simulated LS positive flocks, because concentrations between 10 and 100 CFU/g were also  
352 considered positive. For instance, in ScenRef, the flocks with censored values “<100” were  
353 translated into LS contaminations = 0 (i.e. <10 CFU/g and LS negative). In contrast, in the  
354 ScenUncens, the same censored flock could get a value >0 (e.g. 90 CFU/g from the PERT  
355 distribution), which could give a LS simulated concentration >10 CFU/g (e.g.  $90/3.2 = 28$  CFU/g,  
356 i.e. LS positive).

357 Thus, the number of positive flocks per month was higher in the ScenUncens for 2020 (median =  
358 41) and 2021 (median = 43) compared to the ScenRef (median = 36 in 2020, 40 in 2021).  
359 Consequently, the monthly and the annual LS-AP were also consistently higher in the ScenUncens  
360 – varying between 62.2% and 90.6% in 2020 and between 50.9% and 93.1% in 2021 – than in the  
361 ScenRef (where it ranged from 48.9% to 84.9% and from 47.2% to 89.7%, respectively).

362 The simulated LS mean concentrations also increased for both years, when passing from the  
363 ScenRef (2.74 log<sub>10</sub> CFU/g for 2020 and 2.86 log<sub>10</sub> CFU/g for 2021) to the ScenUncens (3.12 log<sub>10</sub>  
364 CFU/g in 2020; 3.22 log<sub>10</sub> CFU/g in 2021).

365 Accordingly, the monthly risk was always higher in the ScenUncens than in the ScenRef in both  
366 2020 (compare green lines in Figure 5A) and 2021 (compare yellow bars in Figure 5A).

367 The monthly RRs were higher in the ScenUncens than in ScenRef for February, March, April,  
368 June, July, and October, but lower for the remaining months. The largest increase in monthly RRs

369 was observed for June (RR = 1.96 in the ScenRef and 2.18 in the ScenUncens), while the largest  
370 decrease was seen for September (1.33 and 1.12) (Figure 5B). The annual RR changed from 1.22  
371 to 1.20 when comparing the ScenRef vs. the ScenUncens.

372

### 373 *3.4.3 Impact of randomized sampling of one NS pool per flock.*

374 Ten scenarios (ScenSampling-1 to 10) were generated based on repeated random selections of one  
375 NS pool per flock out of the five tested, and simulated into single LS contaminations with CTF =  
376 3.2. Thus, these ten alternative scenarios illustrated the variability in RR, which was due to the  
377 selection of a single result out of five available per flock, to simulate LS contamination. The largest  
378 monthly RR difference between ScenRef and the ten ScenSampling scenarios were observed for  
379 March (largest increase from ScenRef RR = 0.94 to ScenSampling-3 RR = 1.40) and for June  
380 (largest decrease from ScenRef RR = 1.96 to ScenSampling-7 RR = 1.52) (Figure 6).

381 Finally, the annual median RR for 2021 was 1.23, varying from 1.18 to 1.28 across the ten  
382 scenarios, and representing a maximum variation of +/- 5% compared to the ScenRef (1.22).

383

## 384 **4. Discussion**

385 We studied the use of the NS surveillance data collected at slaughterhouses under Regulation (EU)  
386 2017/1495, for an harmonized quantitative risk assessment approach for *Campylobacter spp.*  
387 across different EU countries, i.e. by using French surveillance data in a Danish procedure for risk  
388 assessment (Nauta et al., 2008; Nauta et al., 2012; Foddai et al., 2022a, 2023; Foddai et al., 2022b;  
389 Foddai et al., 2024 under review).

390 The main advantage of having harmonized data analysis and related risk assessments at national  
391 level, would be that the risk estimates become more comparable between countries, alongside their  
392 related flock prevalence and meat concentrations. Such a harmonization can be possible when  
393 countries use the same national farm/flock/slaughterhouse populations' coverage, type of samples  
394 and diagnostic test. A good candidate for such a harmonization would be Regulation (EU)  
395 2017/1495. The purpose of the sampling described on the EU regulation is however different,  
396 which leads to some complications when applying the collected surveillance data (i.e. NS pooled  
397 samples) for the mentioned risk assessment.

398 Therefore, in the Sections below, are discussed the main results of the study (Section 4.1) with the  
399 related sources of uncertainty (Section 4.2) and limitations to be considered for harmonizing the  
400 estimation of flock prevalence, carcass contaminations and risk of human disease, at national level.  
401 Finally, suggestions for potential solutions to address the mentioned limitations are provided in  
402 Section 4.3.

403

#### 404 ***4.1 Interpretation of outputs according to reference simulation scenario***

405 Currently, for France, the reference simulation scenario (ScenRef), would be the most realistic, if  
406 the Donavol data format was kept unchanged and is used for all flocks tested in the country (i.e. if  
407 culture results <100 and >10,000 were kept as censored). The impact of such data format, was  
408 investigated in the alternative scenario analysis (see next section).

409 According to the annual RR of the ScenRef, the average annual risk per serving increased notably  
410 (by almost one fourth) for 2021 compared to 2020 (Figure 4C). This relative increase of annual  
411 risk, partly aligns with the increased number of confirmed cases of campylobacteriosis in France  
412 in 2021 (8875) compared to 2020 (7920) (EFSA and ECDC, 2023).

413 Nevertheless, the increase of risk was not homogeneous during the 12 months of 2021 (Figures  
414 4A-B), neither did it show a particular seasonality. Actually, in one third of the surveillance months  
415 (i.e. in 4 out of 12), the RR was <1. Therefore, the data and the model used in this study, allowed  
416 identifying the surveillance periods during which the most important changes (and especially) the  
417 highest relative increases of risk, occurred during the year of interest (2021).

418 For instance, looking at April, June, and October it can be noted that those months showed the  
419 highest RRs (Figure 4B). Accordingly, a more detailed investigation can be carried out  
420 retrospectively, to understand if the mentioned relative increases of risk, were caused by an  
421 increased prevalence (namely the AP) and/or by increased *Campylobacter* concentrations (namely  
422 the monthly CFU/g distributions), in line with (Foddai et al., 2024, under review). In fact, the risk  
423 estimates used to calculate the RRs, are affected by the combination of both the AP and the  
424 *Campylobacter* concentration distributions fed into the QMRA model (Nauta et al., 2008; Nauta  
425 et al., 2012; Foddai et al., 2022a, 2023, 2024 under review). From previous risk assessments for  
426 *Campylobacter*, it is known that the variance in risk is most often affected by the variance in

427 prevalence rather than by the variance in concentrations (FAO, 2001; Christensen et al., 2001;  
428 Nauta and Havelaar, 2008; Nauta et al., 2009; Foddai et al., 2023). Still, relatively high meat  
429 concentrations can also cause high risk values during some surveillance periods (even when the  
430 AP is relatively low).

431 In this study, the overall annual AP was higher during 2021 than in 2020, and during both years,  
432 the monthly AP showed a seasonal pattern (Figure 2C), similarly to other countries like Denmark  
433 (Kuhn et al., 2020; Foddai et al., 2022b), where the flock prevalence is known to increase during  
434 summer months. From the descriptive statistics, it can be seen that the monthly flock sample sizes  
435 were similar between the months and years (Figure, 2A), and thus, the observed seasonal changes  
436 of AP, represented actual epidemiological patterns in the monthly occurrence of *Campylobacter*  
437 positive flocks (Figure 2B) within the population of studied French flocks.

438 At the same time, for the meat concentration distributions, no particular seasonality was observed  
439 within the two years (Figure 3). When looking at the three months mentioned above (those with  
440 the worst RRs), the boxplots from Figure 3, showed higher median  $\log_{10}$  concentration in 2021  
441 than in 2020. Accordingly, it could be concluded that, for April and June 2021, the  $RR > 1$  was  
442 caused mainly by the worsening of concentration distributions, rather by relevant increases in AP.  
443 Whereas for October, both parameters increased and contributed to the worsened risk per serving  
444 of 2021 compared to 2020. Consequently, for all the three months, it could be also speculated that  
445 the relative increase of risk was mainly caused by flocks which arrived at the slaughterhouse  
446 already contaminated (from the farm). In fact, usually, flocks contaminated from the pre-harvest  
447 are those contributing the most to the AP detected at slaughterhouses and have significantly higher  
448 concentrations than flocks cross-contaminated on carcasses during slaughter. In Denmark for  
449 example, cross-contaminated flocks represented approximately 8-9% of all carcasses positive  
450 flocks and had a median concentration which was relatively low compared to flocks originally  
451 contaminated from the farm (Foddai et al., 2022b).

452 The meat contaminations were also studied using the multiple linear regression modelling based  
453 on year and slaughterhouse set as categorical variables. The interaction between these two  
454 variables appeared significant only for a few combinations (2021 with Ab\_01 and 2021 with  
455 Ab\_08) and the number of observations was similar for both slaughterhouses across both years  
456 (Ab\_01: 33 in 2020, 35 in 2021; Ab\_08: 41 in 2020, 38 in 2021). Hence, based on these data, the

457 mentioned few interactions apparently lack a biological explanation if we only consider the year  
458 and the slaughterhouse in the model. It is therefore not possible excluding that unaccounted factors,  
459 such as flock type (e.g. organic vs. conventional) and/or their origin (i.e. farm's effect), might have  
460 partly affected the results of the statistical analysis. On the other hand, since the overall interaction  
461 did not appear significant; if the two individual factors (year and slaughterhouse) were considered  
462 independently, it seems that both significantly affected the NS contaminations and year 2021  
463 showed a significantly higher mean  $\log_{10}$  concentration than 2020.

464 All this information obtained from the combination of data analysis and risk assessment, when  
465 interpreted systematically and retrospectively, across surveillance periods (e.g. year by year and/or  
466 month by month) allows tracing the flocks and their farms/house units of origin (through the ID  
467 metadata) which contribute the most to the AP, to the concentrations on the meat, and consequently  
468 to the risk of disease in a country (Foddai et al., 2024 under review). Hence, if needed, the risk  
469 assessment can also be carried out at individual slaughterhouse level (using the respective  
470 surveillance data), so that the FBO contributing the most to the overall national risk (and the farms  
471 delivering them the slaughtered flocks) can be traced across surveillance periods and be prioritized  
472 for additional surveillance and/or risk-based control actions along the food chain, as previously  
473 investigated for Denmark (Foddai et al., 2022a, Foddai et al., 2023, Foddai et al., 2024 under  
474 review).

475

#### 476 ***4.2 Information from alternative scenario analysis***

477 The impact of the CTFs appeared limited, especially on the annual RRs (Figure 4C). The biggest  
478 variations in monthly RRs, across simulation scenarios (namely ScenRef vs. ScenMin and  
479 ScenMax); were observed in June when (as mentioned above) the largest difference in the  $\log_{10}$   
480 mean concentration was observed between the two years (Figure 3). Therefore, assumptions on  
481 CTF can have higher impact on the RR, when differences in concentrations are larger across the  
482 considered periods.

483 The scenario ScenUncens, showed that the initial data standardization procedure used in the  
484 ScenRef (i.e. setting contaminations = 0 and = 10,000 CFU/g for cells having original censored  
485 records as: "<100" and ">10,000" CFU/g, respectively) resulted in an underestimation of the risk,

486 for both 2020 and 2021 (Section 3.4.2). These differences between risk estimates of the same year,  
487 by changing from ScenRef to ScenUncens, were due to the differences in the LS simulated flock  
488 prevalence and concentrations. In fact, the considered concentration distributions had a larger  
489 range (from 10 to 150,000 CFU/g) and a larger upper limit ( $>10,000$ ) in the ScenUncens than in  
490 the ScenRef. Consequently, the overall concentrations and the simulated LS-AP were higher in  
491 ScenUncens than in ScenRef, due to the higher frequency of flocks, which were simulated as  
492 carcass positive in LSs (i.e. also with values between 10-100 CFU/g and  $>10,000$ ). Nevertheless,  
493 when combining the effect of the simulated LS-AP and meat concentrations, on the risk of both  
494 years, the impact of the censored data on the annual RR appeared marginal (the RR reduced of 2%  
495 on the ScenUncens compared to the ScenRef, Section 3.4.2), while it was more evident on the  
496 monthly RRs (Figure 5B). For instance, for June, the RR of the ScenUncens was remarkably higher  
497 than that of the ScenRef (Figure 5B: RR 2.18 vs. 1.96), because in that month, the simulation of  
498 concentrations for the censored data, increased more the risk of June 2021 than that of June 2020  
499 (used as baseline). In contrast, for September, the opposite situation occurred (Figure 5B: 1.12 vs.  
500 1.33).

501 The variability of RR estimates associated with the random selection of one pool out of five per  
502 tested flock, was investigated in scenarios ScenSampling-1 to 10. The impact of this pools  
503 selection process on the annual RR appeared limited (RR +/- 5% compared to the ScenRef) (Figure  
504 6), while the effect was more noticeable for monthly RRs due to variations in AP and  
505 *Campylobacter* counts; particularly in March, June, September and October, when there was a  
506 wider range of RRs (Figure 6). This variability across the ScenSampling scenarios, occurred  
507 because fluctuations in *Campylobacter* counts within flocks can happen, as also shown by the  
508 standard deviation distributions observed across standardized NS positive pools of the same flock  
509 (Section 3.2) and as also observed in a previous study based on skin sample data (with more than  
510 one sample taken per flock) (Allen et al., 2007). Such variations affected both the simulated LS  
511 monthly contaminations and flock prevalence, and consequently, the variability of RR estimates.

512 Overall, the alternative scenario analysis shows that the annual RR estimates ranged from 1.21 to  
513 1.26 when comparing different CTFs, from 1.20 to 1.22 when considering the usage of censored  
514 data, and from 1.18 to 1.28 when considering the random selection of one pooled NS sample per  
515 flock. This implies that the major simplifications used (a fixed ratio of single LS and pooled NS

516 concentrations, the simulation of censored data, and the usage of a randomly selected single pooled  
517 NS sample) have little effect on the main output of the model; namely the annual RR. On the other  
518 hand, the current FBO data format and its coverage of national flock populations, can impact the  
519 uncertainty on prevalence, contaminations and monthly RR estimates. Such an impact could be  
520 mitigated by addressing the data limitations, which is suggested before applying/comparing the  
521 shown procedures of data analysis and risk assessment across EU countries and to achieve a  
522 reliable level of international harmonization and comparability of results (see next section).

523

#### 524 ***4.3 Limitations and potential solutions for eventual future international applications***

525 The *Campylobacter* surveillance data used for the study was collected by individual FBOs on a  
526 weekly basis, following EU Regulation 2017/1495. As explained in the introduction, such data is  
527 aimed to inform corrective actions at the single slaughterhouses (i.e. when the PHC is exceeded)  
528 and is not primarily collected to estimate flock prevalence, carcass concentration distributions and  
529 risk of human disease, at national level. For those purposes, flock sample sizes should be set  
530 considering test's performance at the flock level (flock sensitivity and specificity), flock  
531 population's sizes and expected flock prevalence in the country.

532 For instance, in the used data, approximately 650 flocks per year were tested in NS pools collected  
533 from 13 slaughterhouses, which would correspond to approximately 52-53 flocks per FBO or per  
534 month (Figure 2A), out of several hundred slaughtered. To our knowledge, we estimated that the  
535 annual number of flocks slaughtered by the thirteen French FBOs was at least around 17,000  
536 (based on 5 flocks a day for a 5-day slaughtering week over 52-week year). This would mean that  
537 the annual number of tested flocks used in this study represented only  $\approx 650/17,000 = 3.8\%$ . Hence,  
538 by using the sample sizes set at the FBO level, there is a limitation that the NS tested flocks could  
539 be poorly representative of the overall population of flocks processed in a country, when the  
540 monthly/annual flock prevalence needs to be estimated at national level. For that reason, in this  
541 study, we avoided comparing the current risk estimates from Denmark and France, which would  
542 be obtained with the same simulation model; but with different levels of surveillance coverage of  
543 the respective national populations of: farms, flocks and slaughterhouses.

544 Uncertainty on mean RR estimates, could also be due to sample sizes, as incidental high  
545 concentrations, which are expected when the distribution of concentrations is lognormal, have  
546 large effect on the risk estimates (Nauta and Havelaar, 2008). Accordingly, if the data analysis and  
547 the risk assessment proposed in this study, were to be standardized across EU countries, sample  
548 size calculations should be set appositely for the purposes of prevalence and concentration  
549 estimation as well as for risk assessment at national level (i.e. reflecting the aimed confidence and  
550 precision on the estimates). For example, if different tests are used, the impact of their performance  
551 should be considered, by comparing across years and countries, the “true prevalence” estimates  
552 rather than APs (Foddai et al., 2022b). Next, a stochastic version of the QMRA model could be  
553 developed to assess the uncertainty around the estimated mean RRs. Such a model could also allow  
554 estimating the statistical significance of changes in risk across surveillance periods. However, the  
555 deterministic version of the model (used in this study) could be also considered sufficient enough,  
556 for informing decision making within the same country (as in the Danish Action Plan), if the  
557 QMRA is not aimed to produce statistically significant RR estimates, which need to be generalized  
558 across different years (i.e. if the RR is repeatedly assessed on annual bases) and/or if RR estimates  
559 are not aimed for comparison across countries and the sample size of tested flocks /slaughterhouses  
560 remains similar across surveillance periods.

561 Moreover, to apply the same risk assessment procedure across countries, there is a need to report  
562 all the *Campylobacter* counts (CFU/g) from the laboratories into the national databases (i.e.  
563 avoiding censored data), so that similar cut-offs of the concentrations can be used internationally,  
564 to define a flock as positive or negative, and to standardize risk assessment procedures. This would  
565 remove the limitations highlighted when comparing the ScenRef vs. the ScenUncens. Currently,  
566 according to Regulation (EU) 2017/1495, only enumeration is carried out, resulting in some  
567 censored data (< 100 CFU/g and > 10,000 CFU/g), and no detection by enrichment analysis needs  
568 to be performed simultaneously on the same samples. It was therefore impossible to determine  
569 whether *Campylobacter* was absent from a sample, or whether its concentration was below 100  
570 CFU/g.

571 It should also be kept in mind that the NS pools are not fully representative of the carcass  
572 contaminations, because necks are not usually consumed, and most often, a single meal is prepared  
573 from a single carcass. NS pools are usually more contaminated and allow having higher testing



574 sensitivity than the single LSs samples (Ellis-Iversen et al., 2020), but the latter would be more  
575 representative of the actual *Campylobacter* concentrations present on the chilled carcasses. Using  
576 a single CTF to describe the relation between concentrations in pooled NS samples and single LS  
577 samples is a simplification, as the data and theoretical considerations show that the relation is more  
578 complex (Evers and Nauta, 2001; Bahrndorff et al., 2015; Ellis-Iversen et al., 2020). We could not  
579 check the impact of having different number of necks included in the pool (as described in Section  
580 2.1), because the used surveillance data, only represented testing results at pool level. Nevertheless  
581 the use of the CTFs, allowed updating the AP, the CFU/g distributions and the risk estimates  
582 accordingly and simultaneously, when adapting the NS data for a model, which has been  
583 previously used with LS data. However, despite economic considerations, there would be no need  
584 to estimate the CTFs if LS surveillance data could be used (as is currently the case in Denmark),  
585 and thus removing the limitations highlighted when translating NS pools contaminations into  
586 retailed-skin-carcass contaminations.

587 Regarding, the random selection of NS pools, it varied for some flocks due to the number of pools,  
588 which differed from the standard five, for 50 out of 1,284 flocks. This small proportion (<4%) is  
589 unlikely to significantly bias the conclusions of this study.

590 Furthermore, it needs to be mentioned that extrapolations of the risk estimates (e.g. from Fig. 4A)  
591 in comparison with number of reported human cases data, should be made with high caution for  
592 several reasons. For instance, a relevant level of underdiagnoses and underreporting occurs for the  
593 estimated human incidence of disease (Havelaar et al., 2013; Monteiro Pires et al., 2020).  
594 Moreover, in the QMRA model used for the study, there is a margin of error due to the assumptions  
595 and/or the lack of data needed to represent some of the simulated variables. The model includes a  
596 consumer phase section that evaluates bacterial transfer and survival when preparing a meal with  
597 broiler meat, based on a Dutch study (Nauta et al., 2008). This section was found to perform  
598 similarly (but not identically) to other consumer phase models (Nauta and Christensen, 2011); and  
599 is not necessarily representative of all European poultry meat preparations. Additionally, as  
600 explained in EFSA (2020), the majority of *Campylobacter* QMRA studies, including the previous  
601 EFSA opinion (EFSA BIOHAZ Panel, 2011), have applied the “classic” dose-response model  
602 published by Teunis and Havelaar (2000), based on a human challenge study (Black et al., 1988)  
603 in which the *Campylobacter*’s strain A3249 was used. Such a strain is not necessarily a

604 representative strain. New dose-response models have been proposed (Teunis et al., 2018), since  
605 the development of the model utilized in this study, and thus, the uncertainty about these parts of  
606 the risk assessment should be kept into consideration (EFSA BIOHAZ Panel et al., 2020; Nauta et  
607 al., 2022), for eventual future standardizations across EU countries, of risk assessments based on  
608 the Danish QMRA model.

609 Moreover, from a general point of view, it should be reminded that QMRA models tend to  
610 overestimate the risk (Havelaar et al., 2008), compared to results from other kind of  
611 epidemiological studies based on testing data; for example if the QMRA model does not account  
612 for uncertainty on consumer's acquired immunity and/or it does not include other epidemiological  
613 factors affecting the dose-response patterns.

614 Finally, it must be noted that, due to the different sources of uncertainty on the risk estimates, in  
615 this study (as in the Danish Action Plan), the main research focus was addressed to the RR rather  
616 than based on the risk estimate. In the RR, uncertainties due to the consume-phase and dose-  
617 response sections, are largely cancelled out. Hence, from a managerial point of view, it can be  
618 assumed that the estimated RR and (i.e.) most change in risk across surveillance periods  
619 (years/months), will be caused by actual epidemiological changes in prevalence and/or  
620 contamination at the farm and slaughterhouse level; where regulatory control actions can be taken  
621 to reduce risk of human illness. Possible changes in (the dynamics of) transfer and survival of  
622 *Campylobacter* in the post-slaughter food chain stages were disregarded, because systematic  
623 surveillance data on such changes are lacking, and they are not easily enforced by regulations.  
624 Still, such a kind of surveillance information collected between the slaughterhouse and the food-  
625 consumer levels, could improve the risk assessment process aimed to inform inter-sectorial  
626 decision making along the food chain.

627

## 628 **5. Conclusion**

629 This study provides a proof of concept for the potential harmonization of data analysis and related  
630 risk assessment for *Campylobacter spp.*, across European countries using the FBO data. We  
631 showed how pooled NS data collected following Regulation (EU) 2017/1495 in France were  
632 applied in a QMRA model, which is used with single LS data in Denmark, following a procedure

633 that allows direct practical risk management support for decision makers. The impact of the  
634 assumed CTFs, the censored data and the selection of a single NS pool per flock appeared limited  
635 on the annual RR. Advices on potential strengths, limitations and needs for using NS data collected  
636 by individual FBOs, were provided. The study can pave the road for applying standardized data-  
637 driven risk assessment, communication and risk-based control actions against *Campylobacter spp.*  
638 across the European poultry meat chains using NS data.

639

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643

## 644 **Conflict of interest: None**

645

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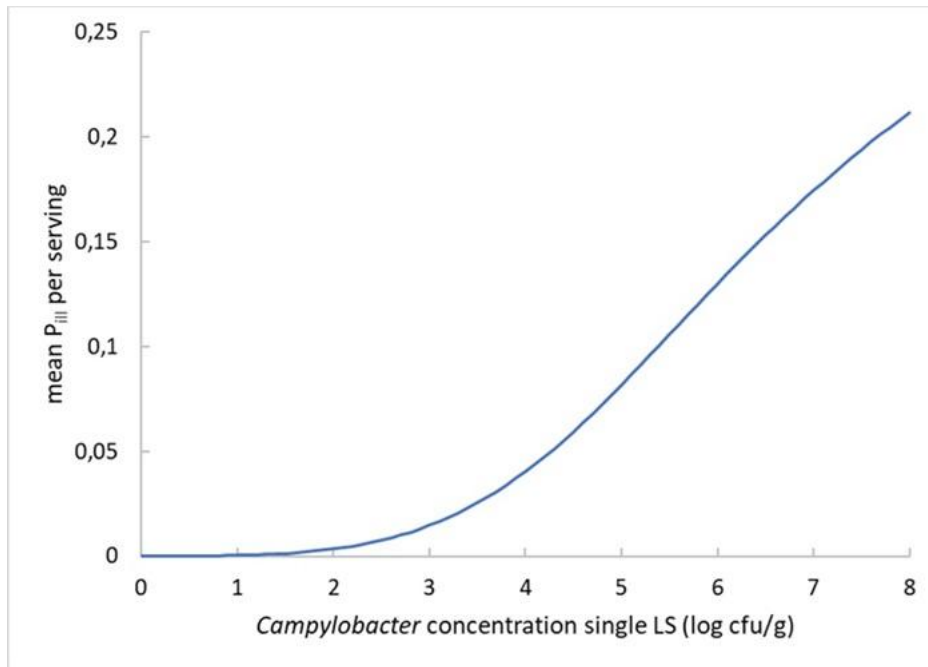
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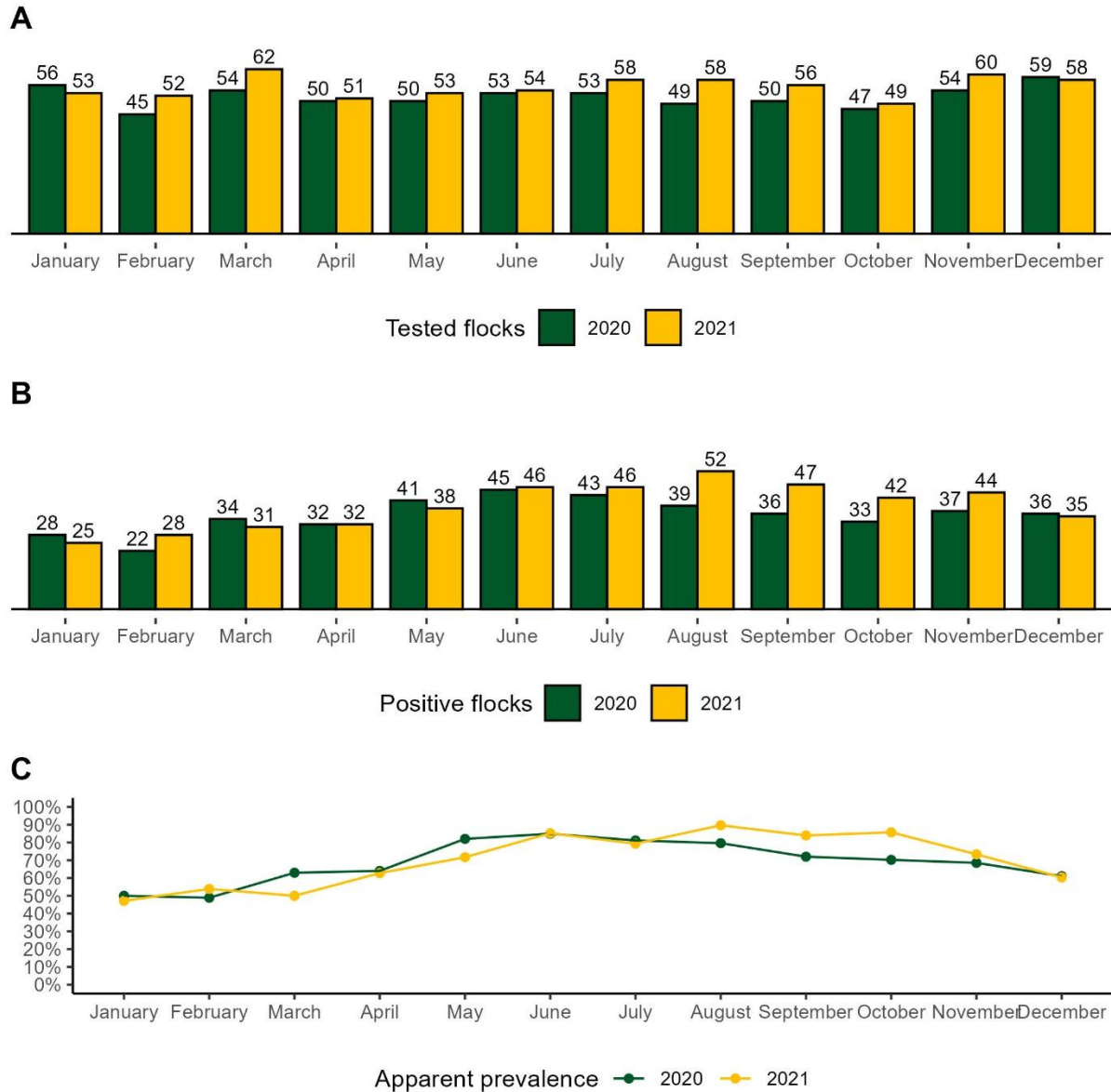
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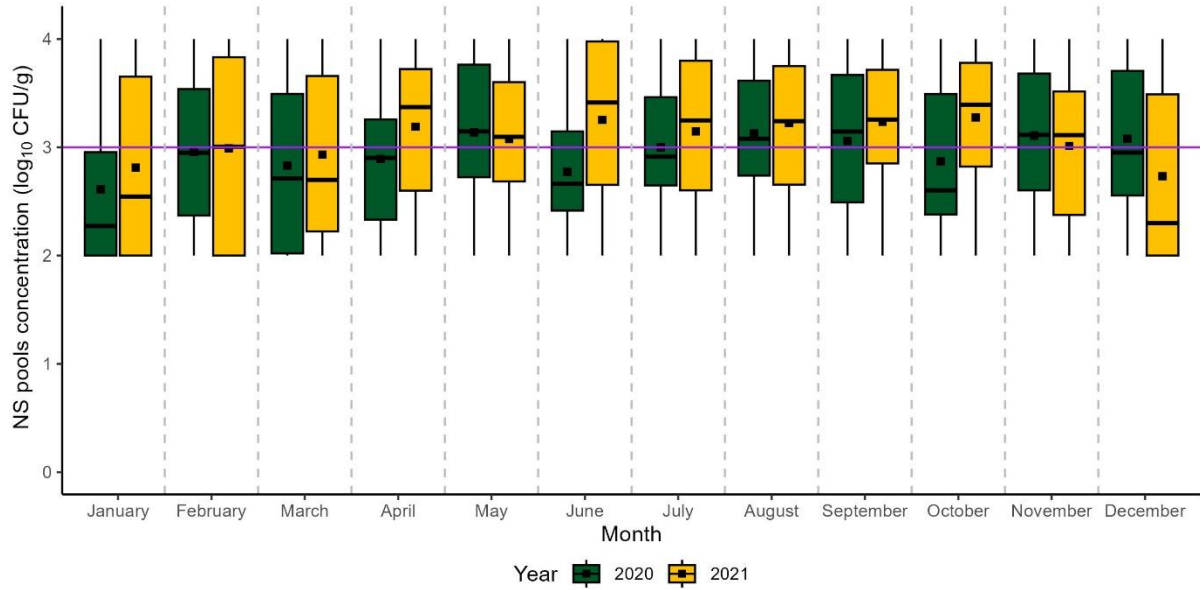
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777 **Figure 1.** Graphical representation of the relationship between *Campylobacter* concentrations found on single leg  
778 skin (LS) samples (x-axis) and the mean probability of illness per serving (y-axis) obtained from the stochastic version  
779 of the QMRA model after  $91 \times 50,000$  iterations in a Monte Carlo simulation (Nauta et al., 2012).



780

781 **Figure 2.** Descriptive statistics on French *Campylobacter* surveillance data 2020-2021, based on testing of broiler  
 782 neck skin (NS) pools, standardized across slaughterhouses as described in Section 2.2 (values <100 set = 0 and values  
 783 >10,000 set = 10,000. **A**) Monthly number of tested flocks per year, **B**) Monthly number of NS positive flocks  
 784 (randomly sampled pool  $\geq 100$  CFU/g) per year, **C**) Monthly apparent prevalence (AP) per year.

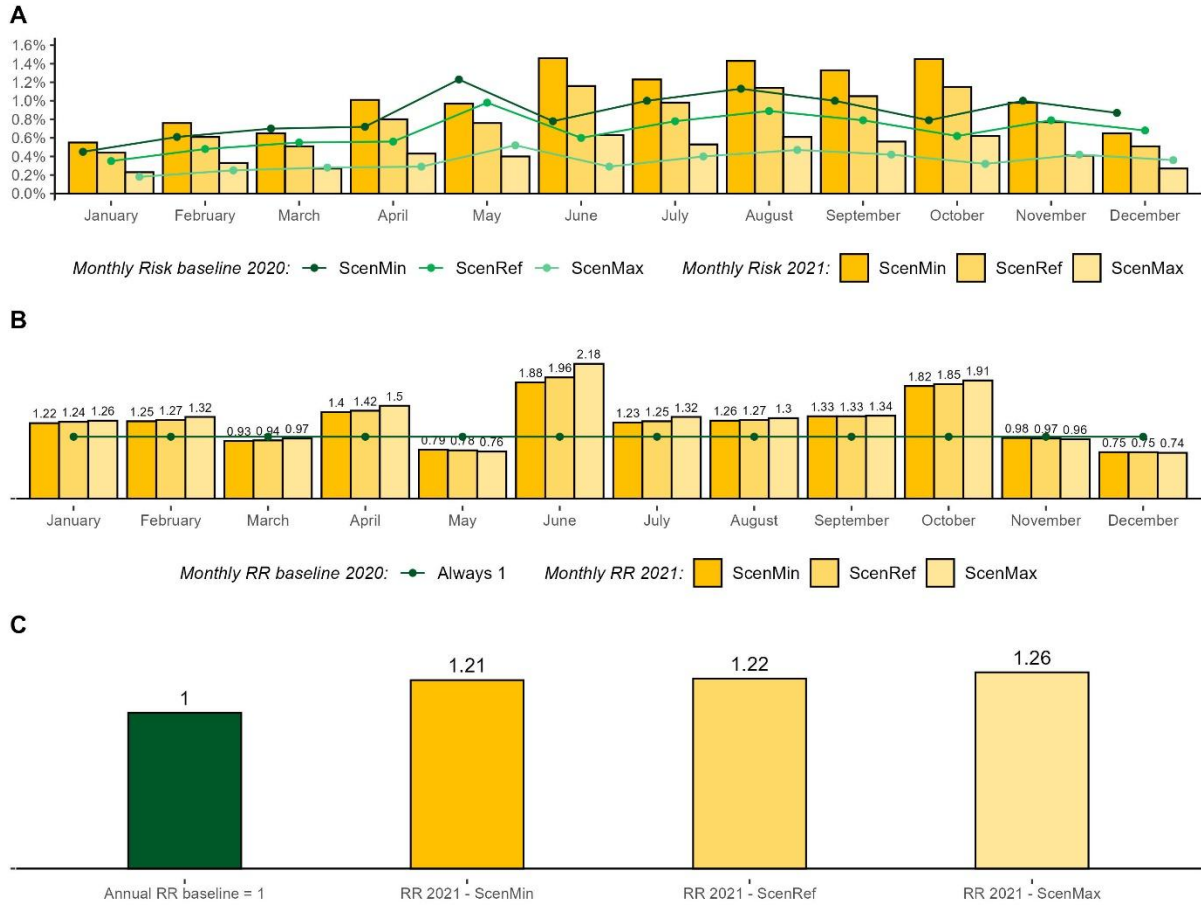


785

786 **Figure 3.** Boxplots comparing the distributions of concentrations across randomly selected and positive (>100  
 787 CFU/g) NS pools (one pool per flock) across months of years 2020 and 2021. NB. Boxplots are constrained between  
 788 the minimum and maximum limits of  $\log_{10} = 2$  and  $\log_{10} = 4$ , as explained in Section 2.2. The violet horizontal line  
 789 shows the limit of 1,000 CFU/g (i.e.  $\log_{10} = 3$ ) set in the European legislation and the black square on each boxplot  
 790 the mean  $\log_{10}$  concentration.

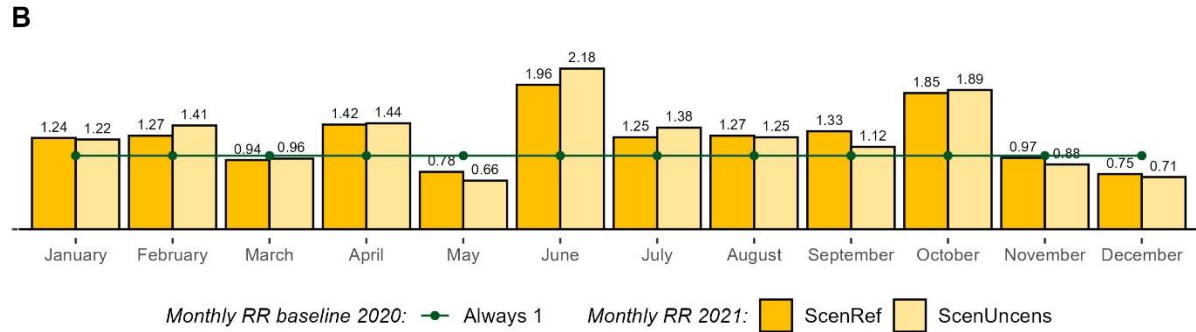
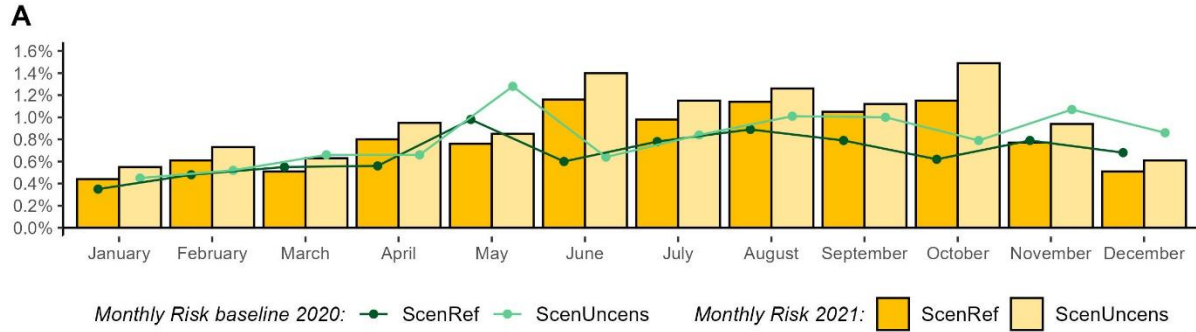
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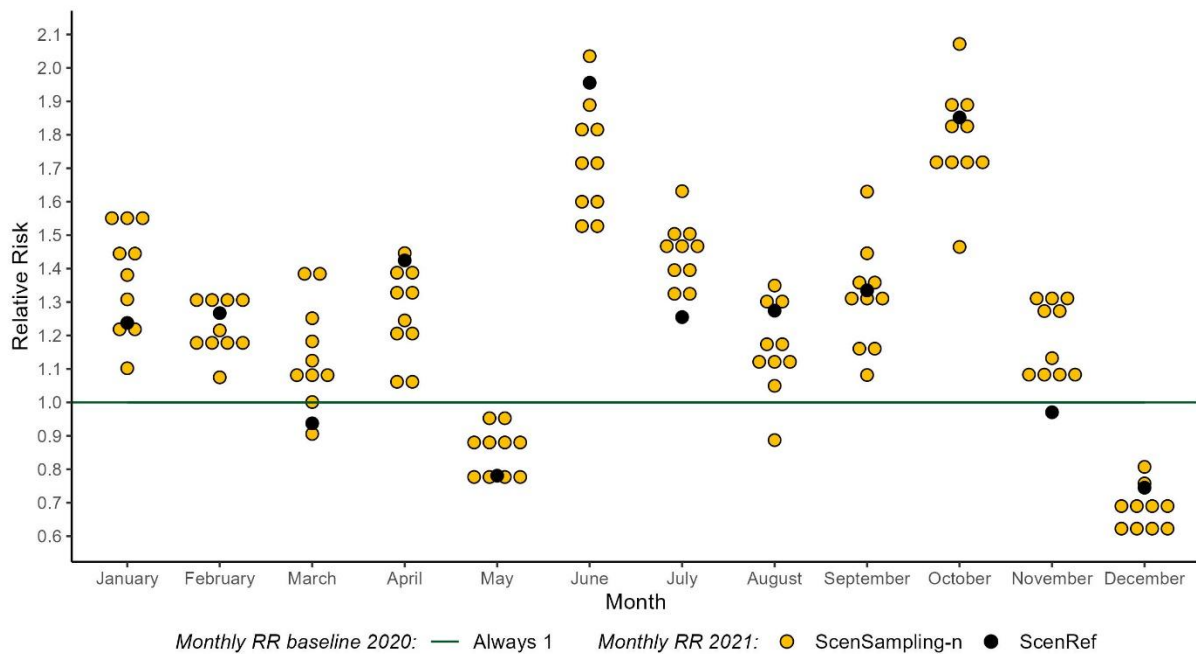
793 **Figure 4.** Outputs of quantitative risk assessment according to year and simulation scenario (ScenMin with  
 794 concentration transformation factor or CTF = 2; ScenRef with CTF = 3.2 and ScenMax with CTF= 10). **A)** Monthly  
 795 absolute risk, **B)** Monthly relative risk (RR), **C)** Overall annual RR. NB. For 2020, the monthly and annual RRs are =  
 796 1, because this was the year used as reference (i.e. baseline).



797

798 **Figure 5.** Outputs of quantitative risk assessment according to simulation scenario and year, to compare the reference  
 799 simulation scenario (ScenRef) and the scenario with censored data simulated (ScenUncens). **A)** Monthly absolute risk,  
 800 **B)** Monthly relative risk (RR). NB For 2020, the monthly RR equals to 1, because this was the year used as reference.

801



802

803 **Figure 6.** Monthly relative risk (RR) according to ten simulation scenarios with CTF = 3.2. For 2020, the monthly  
 804 RR equals to 1, because this was the year used as reference. For 2021, each monthly RR (ScenSampling1 to  
 805 ScenSampling-10) is represented by a yellow dot, while the monthly RR of the reference scenario (ScenRef) is shown  
 806 by a black dot.

807